

1 Constraining aerosol optical models using ground-based, collocated  
2 particle size and mass measurements in variable air mass regimes  
3 during the 7-SEAS/Dongsha Experiment

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## Abstract

During the spring of 2010, NASA Goddard's COMMIT ground-based mobile laboratory was stationed on Dongsha Island off the southwest coast of Taiwan, in preparation for the upcoming 2012 7-SEAS field campaign. The measurement period offered a unique opportunity for conducting detailed investigations of the optical properties of aerosols associated with different air mass regimes including background maritime and those contaminated by anthropogenic air pollution and mineral dust. What appears to be the first time for this region, a shortwave optical closure experiment for both scattering and absorption was attempted over a 12-day period during which aerosols exhibited the most change. Constraints to the optical model included combined SMPS and APS number concentration data for a continuum of fine and coarse-mode particle sizes up to  $PM_{2.5}$ . We also take advantage of an IMPROVE chemical sampler to help constrain aerosol composition and mass partitioning of key elemental species including sea-salt, particulate organic matter, soil, non sea-salt sulphate, nitrate, and elemental carbon. Our results demonstrate that the observed aerosol scattering and absorption for these diverse air masses are reasonably captured by the model, where peak aerosol events and transitions between key aerosols types are evident. Signatures of heavy polluted aerosol composed mostly of ammonium and non sea-salt sulphate mixed with some dust with transitions to background sea-salt conditions are apparent in the absorption data, which is particularly reassuring owing to the large variability in the imaginary component of the refractive indices. Extinctive features at significantly smaller time scales than the one-day sample period of IMPROVE are more difficult to reproduce, as this requires further knowledge concerning the source apportionment of major chemical components in the model. Consistency between the measured and modeled optical parameters serves as an important link for advancing remote sensing and climate research studies in dynamic aerosol-rich environments like Dongsha.

## 1. Introduction

It is well known that aerosols and their optical properties of scattering and absorption play a considerable role in climate forcing and as such have a large impact on public policy (Forster et al., 2007). The radiative effects of aerosols have been a primary focus of numerous research studies over the past decade, particularly as advances in aerosol optical modeling (e.g., Nousiainen et al. 2009) and the rapid growths in field and laboratory measurements (e.g., Reid et al. 2008; Volten et al. 2001) have significantly improved our understanding of the role of aerosols on our environment. The challenges in elucidating these radiative effects lie in the complex spatial and temporal interplay of the aerosol's physicochemical properties with the ambient electromagnetic fields along with changes in the thermodynamic state of the atmosphere. These challenges become even more daunting when aerosols from disparate air mass regimes converge making it exceedingly difficult to separate the extinctive properties and radiative contributions of the component aerosol species.

The fundamental question we seek to answer in this study is how well one can capture the observed optical properties of aerosols in model simulations using simple parameterizations constrained by ground-based measurements. Moreover, how well can these properties be captured in a dynamic atmospheric environment where the air masses and aerosols are constantly changing? To address this question we take advantage of the extended suite of aerosol instrumentation found in NASA Goddard's Chemical, Optical, Microphysical Measurements of In-situ Troposphere (COMMIT) ground-based mobile laboratory (<http://smartlabs.gsfc.nasa.gov/>) which was deployed at Dongsha Island in the northern South China Sea (SCS) during the spring of 2010 in support of the 7-SEAS field experiment and inter-comparison study (Lin et al. 2012 - this issue and the references therein). Further discussions of

instrumentation, site topology, and scientific background for the investigation can be found in Wang et al. (2011) as well as other studies in this special issue. For brevity, the primary instruments and resulting datasets employed in this study are given in Table 1 followed by more detailed descriptions in section 2. Additional information can also be found in various other literature including Li et al. (2012 - this issue) and (2010), Jeong et al. (2008), and at the above web link.

Because of the annual climate of the northern SCS and its tendency to be shaped by changes in the Asian monsoonal winds, Dongsha is ideally suited for this study. Prevailing southwesterly and northeasterly flows during the months of June-September and November-April, respectively can lead to high aerosol loading events stemming from a variety of sources including urban pollution, biomass burning smoke, wind-blown desert dust, and sea-salt over much of the Asian continent (Tsay et al. 2009) and surrounding coastal regions. Consequently the northern SCS is at a confluence of many different aerosols and thus lends a perfect opportunity for not only testing the self-consistency of our ground-based measurements but also attempting optical closure which is crucial for remote sensing and climate studies.

For years, the aerosol community has been actively engaged in the age old, ill-posed problem of bridging together the modeling and observational perspectives of an aerosol's optical properties. The degree of closure is a function of many parameters including how well the measurements are known (i.e. the uncertainty) of the aerosol property being investigated and how well the model is constrained using independent observations. Many past studies have focused exclusively on the light scattering properties of aerosols (e.g., Cai et al. 2011) particularly in regions concerned about air quality and visibility reduction (e.g., Cabada et al. 2004). Others have examined aerosol extinctive properties from specific aerosol events such as

those due to biomass burning activities (e.g., Reid et al. 2005; Mack et al. 2010; Malm et al. 2005) and those derived from urban and industrialized sources (e.g., Highwood et al. 2011; Ma et al. 2011). Lack of key measurements has also required that some studies make assumptions regarding the aerosols' properties, such as for example, chemical composition (e.g., Lee, 2009). To the best of our knowledge, detailed closure studies of scattering and absorption have not yet been conducted at Dongsha in the northern SCS. This recent deployment offers a unique opportunity for using an extensive measurement set to probe the optical properties of multiple aerosol types. This was the primary impetus for the current study given the availability of measurements and diversity of aerosols encountered during the 7SEAS/Dongsha Experiment.

During the deployment (March – May 2010), an extended array of *in-situ* measurements characterized the physicochemical properties of aerosols. This study focuses on a twelve day window from 25 March - 27 April 2010 during which different aerosols were observed using an IMPROVE (A modified Interagency Monitoring of PROtected Visual Environments) chemical sampler. These days were chosen to maximize the different aerosol types analyzed. Continuous sampling revealed a number of key chemical species associated with anthropogenic air pollution, mineral dust, and sea-salt. Atwood et al. (2012 - this issue) also observed similar aerosol components at Dongsha using a collocated 8-stage drum impactor. Back trajectory and subsequent cluster analysis from Wang et al. (2012 - this issue) lend support to the origins of the different air masses and aerosols and their respective transport paths during the experiment. This is later addressed in section 3. To exploit the particle size information of our measurements, SMPS and APS number concentration data are combined for a continuous distribution of both fine and coarse-mode particle sizes up to  $PM_{2.5}$ . Estimated total mass from the particle size measurements is then compared to that from a collocated TEOM for validation purposes.

Furthermore, IMPROVE data along with some assumptions regarding source apportionment, was utilized as a supplementary constraint to help facilitate distributing aerosol mass in the light-scattering model simulations. Lastly, model comparisons with both light scattering (nephelometer) and absorption (PSAP) measurements were assessed to gauge the overall convergence of the shortwave (SW) optical parameters. This is important for two main reasons: (1) to provide a firm anchor point for remote sensing and radiative studies during 7-SEAS and beyond and (2) to later extend the SW optical properties for which most ambient aerosols are measured, into the longer wavelengths to help probe the thermal emissions of the atmosphere (e.g., Hansell et al. 2011).

Despite the advantages of having an extended instrument suite, there are still inherent complexities and uncertainties in characterizing the physicochemical properties of aerosols, particularly for multiple aerosol systems. This includes such unknowns as particle morphology, aerosol mixing state, surface roughness, etc. In addition to the fundamental limitations in measurements and techniques, including temporal and spatial resolution of the instruments, the current study relies on a number of basic assumptions which are listed in Table 2 along with their supporting rationale.

The paper is arranged as follows: Description of instrumentation, data analysis, light scattering model and methodology are presented in section 2; aerosol chemistry and particle size results along with comparisons of the measured and modeled extinctive parameters and their implications are examined in section 3, and finally a summary and future works are discussed in section 4.

## 2. Instrumentation/Data Analysis, Model Overview, and Methodology

### a. *IMPROVE Chemical Sampler*

A modified IMPROVE sampler with an Air Industrial Hygiene Laboratory (AIHL) 2.5  $\mu\text{m}$ -cut cyclone, operated at 22.8 L/min flow rate was used for sampling. As shown in Figure 1, sampling was conducted on three parallel channels equipped with 25 mm Teflon®-membrane (R2PI025; Teflo® PTET membrane with polymethyl propylene support; 2  $\mu\text{m}$  pore size), quartz-fiber (TissuQuartz 2500 QAT-UP), and nylon-membrane (1.0  $\mu\text{m}$  pore size, all filters from Pall Life Sciences, Ann Arbor, MI) filters at a flow rate of 6 L/min per channel (leaving 4.8 L/min on the bypass channel). Details of the IMPROVE chemical analysis performed at the Desert Research Institute (DRI) are given below.

Unexposed filters are subject to pre-treatment, followed by acceptance testing to ensure there is no contamination prior to the measurement process. Teflon-membrane filters are equilibrated in a constant temperature ( $25 \pm 1.5$  °C) and relative humidity ( $35 \pm 5\%$ ) environment for a minimum of 48 hours before weighing. Gravimetric analysis is performed using a microbalance (Mettler MT-5, Hightown, NJ) with a sensitivity of  $\pm 1$   $\mu\text{g}$ . Quartz-fiber filters are pre-fired at 900 °C for four hours to remove organic artifact, sealed, and stored under refrigeration prior to field sampling. Nylon-membrane filters in 47 mm diameter discs are punched to accommodate a 25 mm sampling cassette.

Analysis of 51 elements (Na to U) was conducted by X-ray fluorescence (XRF) spectrometry (PanAlytical Epsilon 5, Almelo, the Netherlands) on Teflon-membrane filters using 13 secondary targets (Watson et al., 1999). Nylon-membrane filters were extracted in 15 ml of distilled-deionized water (DDW) with one hour each sonication and mechanical shaking. Extracts are stored under refrigeration overnight before being submitted for anion analyses of

chloride ( $\text{Cl}^-$ ), nitrate ( $\text{NO}_3^-$ ), and sulfate ( $\text{SO}_4^{2-}$ ) by ion chromatography (IC; Dionex ICS-3000, Sunnyvale, CA; Chow and Watson, 1999); cation analysis of water-soluble ammonium ( $\text{NH}_4^+$ ) by automated colorimetry (AC; Astoria 302A Colorimetry System, Astoria, OR), and water-soluble sodium ( $\text{Na}^+$ ) and potassium ( $\text{K}^+$ ) by atomic absorption spectrometry (AAS; Varian Spectro800, Walnut Creek, CA).

A  $0.5 \text{ cm}^2$  punch of quartz-fiber filter was submitted for organic and elemental carbon (OC and EC, respectively), and thermal carbon fractions (OC1–OC4 at 140, 280, 480, and 580 °C in a 100% helium [He] atmosphere and EC1–EC3 at 580, 780, and 880 °C in a 98% He/2% oxygen [ $\text{O}_2$ ] atmosphere, respectively. Pyrolyzed OC [i.e., OP] is determined based on laser reflectance) following the IMPROVE\_A protocol (Chow et al., 2007; 2011) using the DRI Model 2001 thermal/optical carbon analyzer (Atmoslytic, Calabasas, CA). This standard protocol for OC (sum of OC1+OC2+OC3+OC4+OP) and EC (sum of EC1+EC2+EC3+OP) has been applied in the U.S. long-term non-urban IMPROVE network and the urban Chemical Speciation Network. A minimum of 10% of the samples are submitted for replicate analyses following the quality assurance/quality control (QA/QC) procedure specified in Chow and Watson (2012). The reported concentration in  $\mu\text{g}/\text{m}^3$  is blank subtracted and error propagated following the procedure described in Bevington (1969).

#### ***b. Combined APS-SMPS Aerosol Size Distributions***

Combining data from multiple aerosol-sizing instruments is desirable in order to obtain the largest possible continuum of particle sizing information, however, most instruments measure different approximations of particle size (mobility, aerodynamic, and geometric). The TSI SMPS measures particle mobility size and is used for submicron particles whereas the TSI APS



measures aerodynamic size and measures predominantly super micron particles. These differences must be accounted for when combining the size distributions from these instruments. Khlystov et al. (2004) present an innovative approach to combine the mobility distribution from the SMPS with the aerodynamic distribution from the APS via a statistical approach that requires no a-priori information regarding particle density ( $\rho$ ) or shape.

Converting number concentrations to equivalent surface and volume distributions is straight forward as it only implies assumptions regarding particle shape. For simplicity, all particles in this study are considered spherical although it is recognized that the shapes of aerosol particles vary widely and can include highly irregular non-symmetric particles as in the case of dust. Converting to mass distributions can be a bit more cumbersome as it involves estimates of both particle density and shape factor ( $\chi$ ). Additionally, many of these instruments have engineering limitations when trying to measure denser aspherical particles, which convolute the resulting size distributions even further. For brevity, the largest and most important assumptions are stated below; however the reader is referred to Khlystov et al. (2004) for more detail.

Although the APS and SMPS use fundamentally different approaches, both instruments attempt to characterize the number concentration at discrete size intervals. The aerodynamic diameter measured and reported by the APS is predominantly a function of particle asphericity and density, whereas the mobility diameter measured and reported by the SMPS is a function of asphericity but not density. An in-depth review of the varying terminologies and definitions for conversion between various definitions of diameter can be found in Decarlo et al. (2004). First, we assume the instrument counting efficiency to be unity in the overlapping regions between the two instruments (540-800nm) and second, the SMPS log-weighted number concentration profile for the overlapping region is approximated by a log-linear fit. This smoothes the SMPS profile

for the APS fitting. The log-weighted number concentration of the APS is then shifted along the diameter axis until, via a minimized error approach following equation 1 in Khlystov et al. (2004), it matches with the log-linear fit to the SMPS. The amount the APS is shifted is proportional to the square root of the particle density in the instrument overlap region and the asphericity of the particles. The size correction factor for converting aerodynamic to electro-mobility diameter [Khlystov et al. (2004)] is defined as:

$$\alpha = \sqrt{\frac{\rho_p}{\chi\rho_0}}$$

where  $\rho_p$  is the particle density,  $\rho_0$  is the reference density of  $1 \text{ g cm}^{-3}$ , and  $\chi$  is the shape factor.

Under a spherical assumption, we can then retrieve a rough bulk density approximation from the APS correction (Khlystov et al. 2004 - valid only in the overlap region, but used for the entire size distribution in our study) which is discussed later in section 3. For this study, the APS and SMPS size spectra are taken during 18 second and 3 minute sample periods, respectively. The resulting distributions are then averaged into 2-hour bins yielding  $\sim 360$  and 40 samples, respectively.

### ***c. Light-Scattering Model and Methodology***

Assuming that the aerosols are spherical, we employ the Lorenz-Mie light-scattering code (Mishchenko et al. 1998) to calculate the extinction cross sections used for the bulk extinction parameters from each of the chemical compositions and mixtures at a wavelength of  $\lambda=550\text{nm}$ . We choose  $\lambda=550\text{nm}$  since we can apply the nephelometer (Anderson and Ogren, 1998) and PSAP (Bond et al. 1999) corrections for later comparison of the measurements and model data.

All Mie code computations are performed over the range of particle sizes measured by the combined APS-SMPS instruments for each chemical component. Following (Petty, 2004), the bulk volume extinction coefficients  $\beta_e$  of scattering ( $\beta_s$ ) and absorption ( $\beta_a$ ) are given by:

$$\beta_e = \sum_i \sum_j N_{i,j} \sigma_{e,i,j}$$

where  $\sigma_e$  is the Mie calculated extinction cross section and N is the number concentration from the particle size instruments. Here we assume a homogenous mixture of pure aerosol types (i), i.e., no coatings or inclusions are considered for simplicity, over a range of particle diameters (j).

A summary of the methodology employed in this study is given in Figure 2. We first start with the data from the IMPROVE sampler as our basis for determining the content and distribution of aerosol particulates representing the different air mass regimes encountered during the study. After combining the number concentration profiles which span the particle size range of 10nm to 2.5 $\mu$ m, we partition the IMPROVE data as a function of size. The mass distributions of the combined profiles are then weighted by uniformly adjusting all size bins by the estimated mass fractions from IMPROVE. Next, we take the weighted chemical speciated mass profiles and calculate the implied number distributions using the assumed bulk densities given in Table 3. After running the Mie code, we integrate the size dependent Mie solutions over the number size distributions for each chemical constituent from ~10nm (bottom end of the SMPS) to 2.5 $\mu$ m. Note that fine particles were expected at Dongsha and therefore all *in-situ* extinction instruments had a sharp cyclone size cut of ~2.5 $\mu$ m which was the upper end of our particle sampling size. The integrated results yield rough bulk scattering and absorption coefficients for each of the chemical species, which are linearly combined to give the total optical parameters to be compared with the measurements. Optical consistency in this regard lends credence to the

chemical partitioning and composition assumptions used, which will allow us to further hypothesize about applying this methodology for future inter-comparisons with LW measurements such as those from SMART's AERI interferometer (e.g., Hansell et al. 2008).

### 3. Results and Discussion

Here we examine the resulting IMPROVE chemical data and combined APS-SMPS particle size measurements which are used to constrain the light scattering model. The derived model parameters are then compared with those from measurements in order to gauge the optical consistency.

#### a. *Aerosol Chemistry*

After the chemical analysis, the basic elements were characterized using the mass conversion factors from IMPROVE (Watson, 2002) and Sillanpää (2006) to determine rough allotments of bulk chemical constituent properties. The resulting primary elements are listed in Table 3 along with their appropriate data conversions, corresponding refractive indices at  $\lambda=550\text{nm}$ , and other relevant parameters. These data are used to reconstruct the total mass to compared mass by gravimetry and to provide rough allocations of particulate constituent in order to constrain the optical calculations. The resulting mass fractions of the key chemical species for the period analyzed are shown in Figure 3. The reconstructions account for ~75% of the total weight of the filter depending on the sample. Total mass differences can likely be attributed to several things including (1) unused trace chemical species, (2) measurement uncertainties, and (3) differences in mass conversion factors employed. The latter point is clearly illustrated with sea salt. For example, using Sillanpää (2006), the mass fraction of sea salt can be up to 50%

larger than that obtained following IMPROVE's methodology, which consequently increases the reconstructed mass anywhere between 80-100% of the total PM<sub>2.5</sub> weighed mass. For sea salt, we apply the conversion factors from IMPROVE (Table 3).

For the most part, NSS and sea salt, likely stemming from transport over urban/industrialized regions along the coast, dominate the mass distributions both by quantity and by mass fraction. This finding is corroborated by Wang et al. (2012 - this issue) whose study shows that the majority of air mass transport (~52%) into Dongsha follows two primary paths: one that originates from the inland areas of northern China and Mongolia and then travels towards the coast and a second that originates along the coastal regions of China. Because the coast is known to have higher anthropogenic emissions (Streets et al., 2003), the transported air masses into Dongsha will contain a larger number of industrialized pollutants as reflected in Figure 3. Since Dongsha is an atoll, the sea salt distribution is not surprising and likely represents background aerosol conditions for the region. Other dominant classifications evident from the analysis are ammonium and nitrate which in the form of ammonium nitrate accounts for about 9.7±2.3% of the measured PM<sub>2.5</sub>. Various trace chemical constituents such as As, Cd, Co, Cr, etc, were also identified and used in the reconstruction (labeled as 'other elements' in Figure 3 to maintain consistency with Sillanpää [2006]). Since the combined mass fraction of these trace species is small relative to the other components, these were not considered in the study.

Soil particles (e.g., mineral dust) likely transported from northern China and Mongolia (e.g., Taklimakan and Gobi deserts) were also identified, however, their mass fractions over the 12-day period were generally less than those found for NSS and sea salt. It is noted that a major dust event which affected a large area including Dongsha had occurred just days before the

IMPROVE sampler began recording data (~21 March 2010). Observation and modeling perspectives of this event can be found in Wang et al. (2011) and Bian et al. (2011), respectively.

#### ***b. Merged particle size spectra and estimated density***

Figures 4(a) and 4(b) provide examples of the fitting process and final outcome for 26 March, and the mean fit for 27-29 March, respectively. The aforementioned fitting is done on two-hour averaged particle size data from the SMPS and APS instruments which provide us a time series of estimated size correction factors, and therefore estimated particle densities which were found to mostly vary between  $\sim 1.2$ - $2.2 \text{ g/cm}^3$  (Figure 4c). Note that estimated densities on the lower end of the range reflect higher relative humidity values while those values that are much less than one are due to instrumental effects. To simplify the conversion from number concentration profiles to mass concentration profiles for the modeling portion of our study, an average of the “correction factor” derived density is used which was found to be  $\sim 1.76 \text{ g/cm}^3$ . This is consistent with aerosol that is mostly dominated by NSS (Table 3) however; the range also overlaps with the densities for sea salt and various minerals found in soil aerosol. Although spherical assumptions required for merging the particle size spectra could result in overestimated density values, there is reasonable agreement between the retrieved integrated mass concentrations from the size fitting process and collocated TEOM measurements (Figure 5). Here we use the  $\text{PM}_{2.5}$  TEOM from the Taiwan Environmental Protection Administration (T-EPA, <http://www.epa.gov.tw/en/>) mobile facility for comparison. It is evident that the retrieved mass is somewhat larger than the TEOM values, which could be due to the spherical assumptions used. This offset could also be due to the TEOM losing volatiles (e.g., ammonium nitrate and some organic matter) and/or residual water inside the SMPS/APS. The corresponding linear

correlation coefficient is  $\sim 0.70$  and the resulting fitting parameters are 0.94 (slope) and 7.3 (offset). It is also noted that the total mass computed from the IMPROVE data compared to within about 10% of the TEOM values (not shown).

### ***c. Comparison of optical parameters***

The comparisons between modeled and measured optical parameters are presented in Figure 6, where panels (a-b) and (c-d) show the time series and scatter plots for both absorption and scattering coefficients, respectively. Included on the scatter plots are error bars associated with the measurement uncertainties which are taken to be about 15% and 25% for the nephelometer (Anderson et al. 1996) and PSAP (Bond et al. 1999), respectively. Uncertainties in the model parameters are estimated to be on the order of  $\sim 30\%$  based on previous works (e.g., Ma et al. 2011). The linear correlation coefficients for scattering and absorption were 0.47 and 0.56, respectively. On close inspection of Figures 6(b)-6(d), it is apparent that scattering was over predicted ( $> 3\sigma$  standard deviations) on 26 March likely due to changes in our constant relative mass fractions assumption for the chemical constituents. After removing this point, the linear correlation coefficient for scattering increased to 0.84. The smaller correlation for absorption is not surprising since the imaginary term of the refractive index for these aerosols varies by orders of magnitude (Table 3), which in turn can lead to a much larger variability in the absorption coefficient. Linear fits to the scattering and absorption data (red lines shown in Figure 6) reveal fitting parameters of 0.35 (slope) and 29.61 (intercept) and 0.43 (slope) and 3.0 (intercept), respectively. The large intercepts are related to model uncertainties. If the offsets are forced through zero (i.e., model uncertainties are reduced), the resulting slopes for scattering and absorption are 0.67 and 0.75, respectively.

Given the assumptions (Table 2), the observations were reasonably tracked by the model, although in terms of absolute differences, scattering is continuously underestimated [Figure 6(d)] particularly when the measurements are greater than  $\sim 50 \text{ Mm}^{-1}$ . A plausible reason for this disparity is hygroscopic growth on the particles during periods of higher relative humidity which leads to an increase in scattering. This is related to the study's dry particle assumptions where periods with lower relative humidity are expected to converge more closely to the measured results. Modeled absorption values on the other agreed surprisingly well with the measurements up to  $\sim 10 \text{ Mm}^{-1}$ ; however for higher measured absorption the model data exhibited greater scatter.

Excluding absolute differences, the model peaks in scattering and absorption are nearly consistent with those from measurements. This is particularly apparent during late March and early April, where absorption underwent a series of rapid changes likely related to the differences found in mass fractions of the component aerosols [Figure 3(a)]. The IMPROVE data between 26-31 March for example, show some of the highest recorded levels of pollution (NSS) and soil aerosols during the study period. This is consistent with the analysis from Wang et al. (2012 - this issue) showing high-level transport of air masses from dust source regions in China influenced by urban and industrialized sources along the coast. Higher levels of sulphates mixed with transported and localized sea salt, would help explain the periods when absorption was minimal, for example on 26 March, where transported air masses mostly originated from the coastal regions of China. Following this event, the total aerosol mass decreased by almost 40% so the effects of carbonaceous particles became more prominent leading to a period of higher absorption around 27-28 March. The overestimated absorption as shown is potentially due to the excess soot in the model or it could be due to equally applying the aerosol mass across all particle size bins. Later in March sea salt aerosols become more predominate, coinciding with



the region's background maritime conditions and the absorption begins to fall off again.

It is immediately apparent that for a dynamic aerosol rich environment like Dongsha, one cannot simply prescribe *a priori* an assumed aerosol composition of one type which could cause large errors in the modeled parameters. This is particularly more evident for absorption than it is for scattering since the real part of the refractive index only varies by ~30% over the species analyzed, whereas absorption can vary by several orders of magnitude. Assuming sea salt to be the dominant aerosol for example, will largely underestimate total absorption which consequently will impact studies of radiative transfer. Unless aerosol measurements are made at or near their sources, it is essential to adequately account for any changes in the aerosol's composition. It is further noted that additional knowledge concerning the source apportionment of major chemical components relative to particle size should help improve the model's performance. Instead of uniformly scaling all size bins of the SMPS-APS spectra by an aerosol's mass fractional data, only the size bins appropriate to the species under investigation should be scaled.

#### 4. Conclusions and Future Work

The aim of this study was to not only test the self-consistency of NASA Goddard's COMMIT ground-based measurements but was also to attempt optical closure for the first time in the dynamic aerosol-rich environment at Dongsha Island off the southwest coast of Taiwan during the 7-SEAS/Dongsha experiment (2010). Comprehensive aerosol physicochemical measurements were used for constraining the optical model to compare derived extinctive parameters with those that were measured. Critical to this study was the use of an IMPROVE

sampler to enhance our knowledge of the aerosol's changing chemical composition. Given the assumptions required to partition the chemical and microphysical data, the model is able to reasonably track the observations and identify peak events in both scattering and absorption over a 12-day test period. Transitions in key aerosol types from heavy polluted aerosol composed mostly of ammonium and NSS mixed with some dust to mainly sea salt are evident in the absorption data, which is particularly reassuring owing to the large variability in the imaginary component of the refractive indices. Overall, scattering is shown to be negatively biased because of the dry particle assumptions employed in the study, whereas the modeled absorption agrees reasonably well with the measurements. Uniform scaling of the particle size spectra by an aerosol's mass fractional data may have also contributed to the model's bias.

Instrumental limitations such as not being able to account for the role of water vapor on particle growth rates along with uncertainties in the measurements and the techniques employed, makes it difficult to achieve full optical closure (complete alignment along the 1:1 line) in a region that is characterized by a diversity of air mass regimes and aerosols. However, given the limitations and uncertainties in this study, the amount of closure or optical consistency attained, is encouraging as measured by the ability to adequately reproduce the observed behavior in the extinctive parameters. To improve the convergence of the model data with the measurements will require additional key information including:

- Further knowledge concerning the source apportionment of major chemical components in the optical model at higher temporal resolution for capturing the daily evolution of an aerosol's composition.
- The requirement to only scale the particle size bins appropriate to the chemical species under investigation.
- Selecting appropriate optical constants for the chemical species identified
- Consideration to the aspherical nature of aerosol particles and the use of numerical light scattering codes for handling these more complex geometries.
- Extending measurement capabilities to account for the effects of relative humidity on

particle growth rates and its impact on merging the particle size spectra.

Regarding the final point, the relative humidity issues experienced at Dongsha adds substantially to the error in the APS-SMPS size distribution merger. A combination of hydrophobic and hydrophilic particles skews the bulk density towards that of pure water. However, particle growth due to increased humidity is difficult to parameterize from our existing instrument setup. Current modifications to the COMMIT mobile facility are underway to eliminate the inadvertent role of moisture (by drying out the aerosol stream prior to evaluation) or to quantify the relative growth rates of particles (via tandem DMA's, CCN counters, and wet/dry nephelometry). Furthermore extinction instruments which are not biased by relative humidity or varying corrections due to multiple scattering, such as the Aethelometer and PSAP, are also being integrated into the facility.

This study illustrates the significance of employing a synergy of multiple ground-based measurements for characterizing the physicochemical properties of multiple aerosols in an optical model for closure experiments involving both scattering and absorption. It represents a first step at attempting closure for the dynamic aerosol-rich environment at Dongsha. With further instrument modifications and the application of new techniques, it is anticipated that improved closure can be obtained which will certainly benefit remote sensing and climate studies in this important region.

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**Table 1 Primary Instrument at COMMIT and Data**

Instrument	Data
IMPROVE chemical sampler <sup>(1)</sup>	Chemical composition and mass partitioning of aerosols
APS (TSI 3321) <sup>(2)</sup>	Coarse-mode particle size distributions
SMPS (TSI) <sup>(3)</sup>	Fine-mode particle size distributions
3 Wavelength PSAP (Radiance Research, Inc.) <sup>(4)</sup>	Aerosol absorption coefficients
3 Wavelength Nephelometer (TSI) <sup>(5)</sup>	Aerosol scattering coefficients
TEOM (model 1400ab; R&P Co.) <sup>(6)</sup>	Mass concentrations

- <sup>(1)</sup> Interagency Monitoring of PROtected Visual Environments
- <sup>(2)</sup> Aerodynamical Particle Sizer
- <sup>(3)</sup> Scanning Mobility Particle Sizer
- <sup>(4)</sup> Particle/Soot Absorption Photometer (@0.55µm corrected for sample spot size and flow rate)
- <sup>(5)</sup> Nephelometer @0.55µm with Anderson corrections applied
- <sup>(6)</sup> Tapered Element Oscillating Microbalance;
- Note- for additional comparisons, we used the TEOM from the Taiwan Environmental Protection Administration mobile facility.

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Table 2 Key Assumptions in Study
1. Although aerosol particles are rarely spherical in shape, we assume for simplicity that all particles are spheres and apply the Lorenz-Mie light scattering code (Mishchenko et al. 1995) for calculating the scattering and absorption coefficients. The optical parameters for non-spherical distributions of dust particles, for example, are generally larger than those for spheres (e.g., Kalashnikova et al. 2002 and Hansell et al. 2011) however, this should not impact the ability of the model to track the measurements.
2. Assuming spheres, particle density is inferred from combining SMPS-APS particle size spectra following Khlystov et al. (2004). Although spherical assumptions could result in overestimated density values, SMPS-APS mass measurements show good agreement with those from a collocated TEOM suggesting this is a reasonable assumption.
3. All optical constants used to characterize the elemental species identified by IMPROVE are based on data availability and the representativeness of the data to the measurements. Table 3 identifies the datasets used in the study.
4. For simplicity, all particles are considered to be dry in this experiment, although relative humidity was high at Dongsha. Aerosols with absorbed water exhibit larger $\beta_{ext}$ due to the larger optical cross sections. Lack of hygroscopic growth on particles could underestimate $\beta_{ext}$ . However, the ability to track the measurements should not be impacted. Improved convergence is expected to occur for lower relative humidity cases.
5. Fine particles were expected at Dongsha and therefore all optical instruments had a sharp cyclone size cut of $\sim 2.5\mu m$ . To maintain consistency, a numerical size cut on the combined size distribution of $PM_{2.5}$ was made.
6. All aerosol components are uniformly weighted by the calculated IMPROVE mass fraction (i.e., no assumptions are made to preferentially bin certain species such as dust being greater than $0.5\mu m$ ).
7. IMPROVE data were averaged (obtained) over 24-hour sampling periods. As long as the relative mass fractions of multiple chemical constituents remains constant, this will not impact the findings significantly; however, large differences in mass fraction (e.g., significant changes in air mass during a sample) will invalidate this assumption.

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**Figure 1** Modified Interagency Monitoring of PROtected Visual Environments (IMPROVE) sampler including Air Industrial Hygiene Laboratory (AIHL) 2.5  $\mu\text{m}$ -cut cyclone.

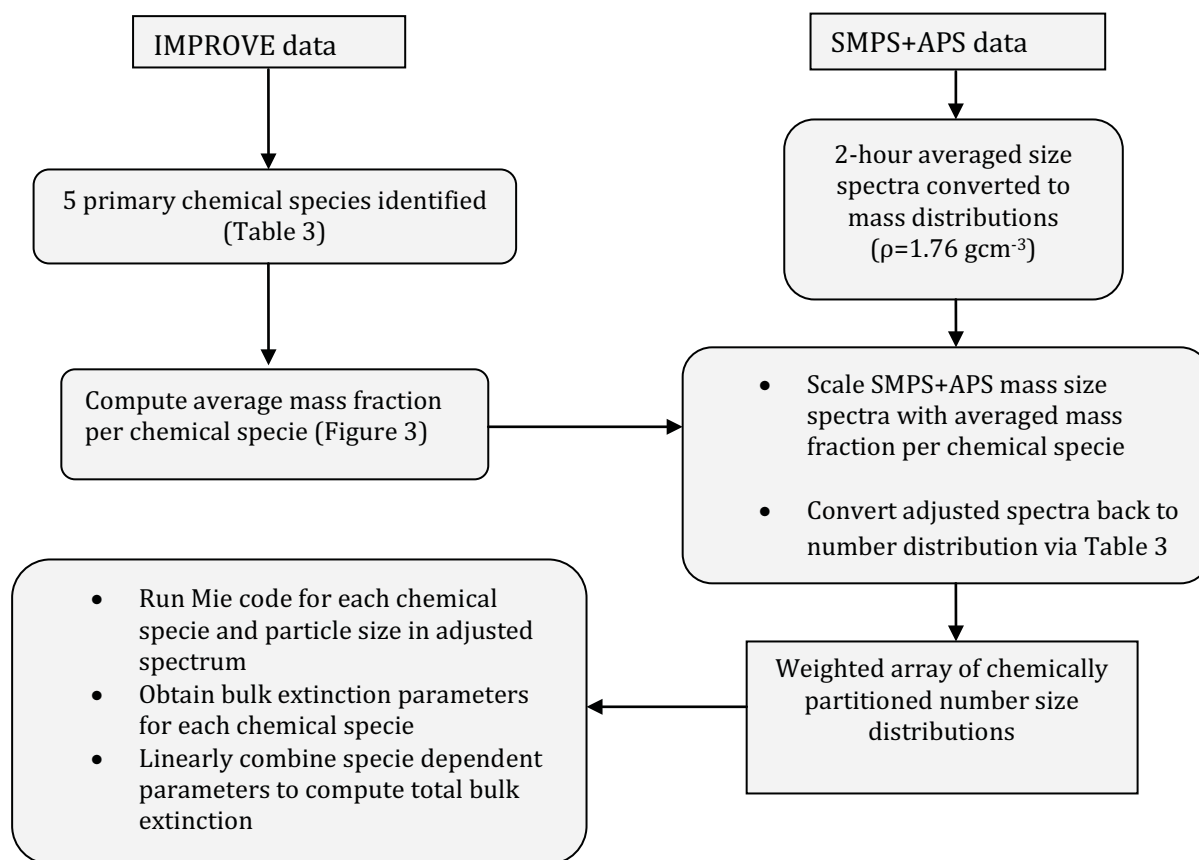


Figure 2 Flow diagram depicting methodology

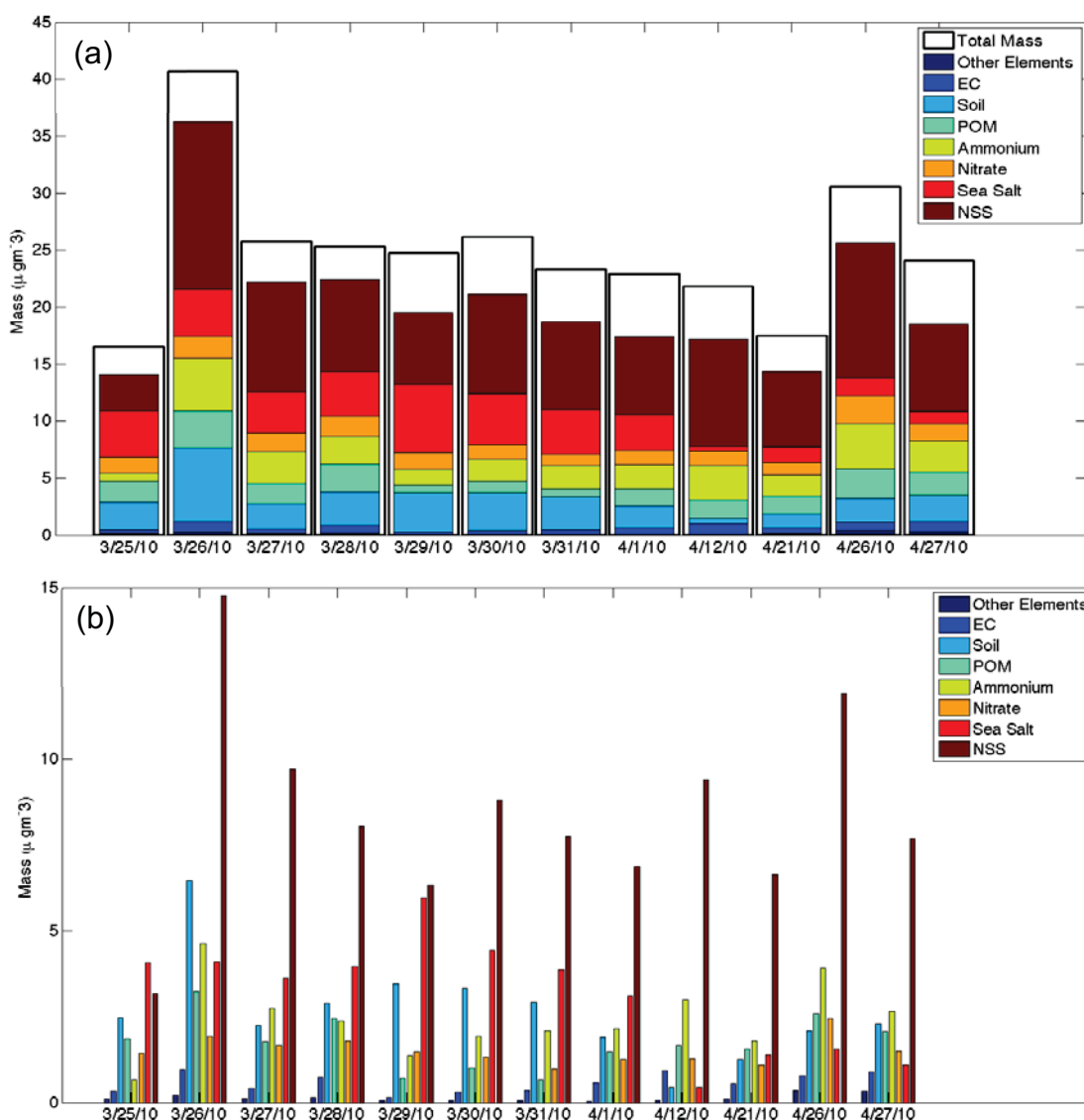


Figure 3 (a) Relative mass of chemical species derived from IMPROVE data for the period analyzed, the black outline is the total mass as measured by the weight of the filter. Our reconstructions account for ~75% of the total weight of the filter depending on the sample. (b) Same as (a) but shown to elucidate individual quantities and trends in the constituents. See text for details.

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768 **Table 3 Chemical Components and Test Parameters**

<b>Chemical Component</b>	<b>Conversion Factor</b>	<b>IOR<sup>(1)</sup> Dataset</b>	<b>IOR (550 nm)</b>	<b>Density (<math>\rho</math>) gcm<sup>-3</sup></b>
<i>Elemental carbon</i>	Total elemental carbon	Soot (Cai, 2011)	1.96+0.66	0.5 <sup>(2)</sup>
<i>Particulate organic matter (POM)</i>	1.6*total organic carbon <sup>(3)</sup>	Organic matter (Schkolnik et al. 2007)	1.4-0i	1.2 <sup>(4,8)</sup>
<i>Soil</i>	2.2Al + 2.49Si + 1.63Ca + 2.42Fe + 1.94Ti <sup>(3)</sup>	Mineral dust (Patterson 1981)	1.56+0.0056	2.5 <sup>(5)</sup>
<i>Sea salt</i>	1.8 Cl- <sup>(3)</sup>	Shettle and Fenn, 1979	1.35+2.9E-09	2.1675 <sup>(6)</sup>
<i>Non-sea salt sulphate (NSS)</i>	Sulphate- (0.246*soluble sodium) <sup>(7)</sup>	Ammonium Sulphate <sup>(8)</sup> (Toon et al. 1976)	1.53+1.0E-07	1.769 <sup>(9)</sup>
<i>Nitrates</i>	Nitrate*1	Ammonium Nitrate <sup>(10)</sup>	1.55+1.0E-09	1.725 <sup>(11)</sup>
<i>Other elements (IMPROVE elemental data)</i>	As, Cd, Co, Cr, Cu, Ni, V, Br, Mm, Pb, Rb, Se, Sr, Zn	N/A <sup>(12)</sup>	N/A <sup>(12)</sup>	N/A <sup>(12)</sup>

769 <sup>(1)</sup> Index of Refraction770 <sup>(2)</sup> Gysel et al. (2011)771 <sup>(3)</sup> From Watson (2002);[Note the 1.6 multiplier for POM is in the range used by IMPROVE (1.4-1.8)]772 <sup>(4)</sup> Turpin B.J. and H.-J. Lim, (2001)773 <sup>(5)</sup> Representative density value based on range presented in Kandler et al. (2007).774 <sup>(6)</sup> Fan and Toon (2010)775 <sup>(7)</sup> Sillanpää (2006);776 <sup>(8)</sup> Ammonium sulphate represents combined masses from NSS and ammonium777 <sup>(9)</sup> Cai (2011)778 <sup>(10)</sup> Software from Andrew Lacis (from <http://gacp.giss.nasa.gov/datasets/>, last visited on 23 January 2012) based on Toon et al. (1976), Gosse et al. (1997), and Tang (1996).779 <sup>(11)</sup> Lloyd et al. (2009)780 <sup>(12)</sup> Not applicable - not included in model analysis

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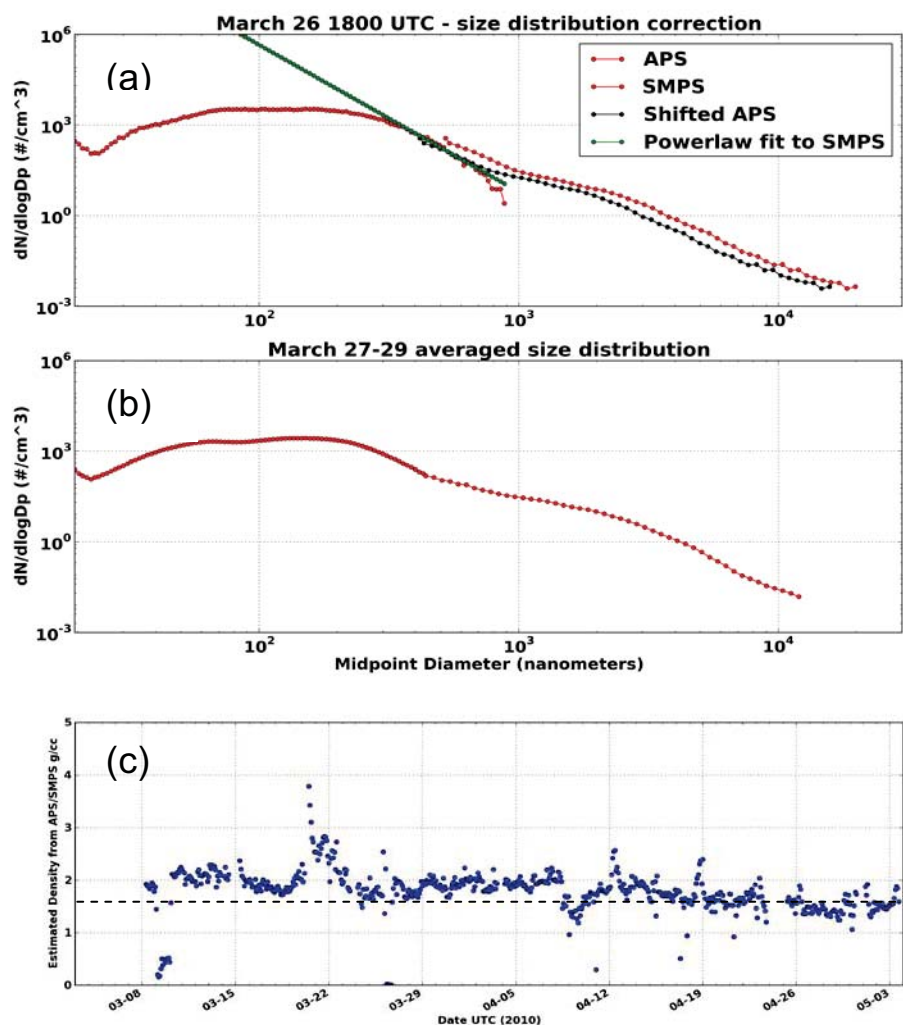


Figure 4 (a) Illustration of fitting process following the methodology of Khlystov et al. 2004 for combined SMPS and APS particle size spectra on 26 March 2010 during the 7-SEAS/Dongsha experiment. (b) Averaged size spectra from combined distribution from measurements during 27-29 March 2010 and (c) estimated particle densities with average value depicted by the dotted line. See text for details.

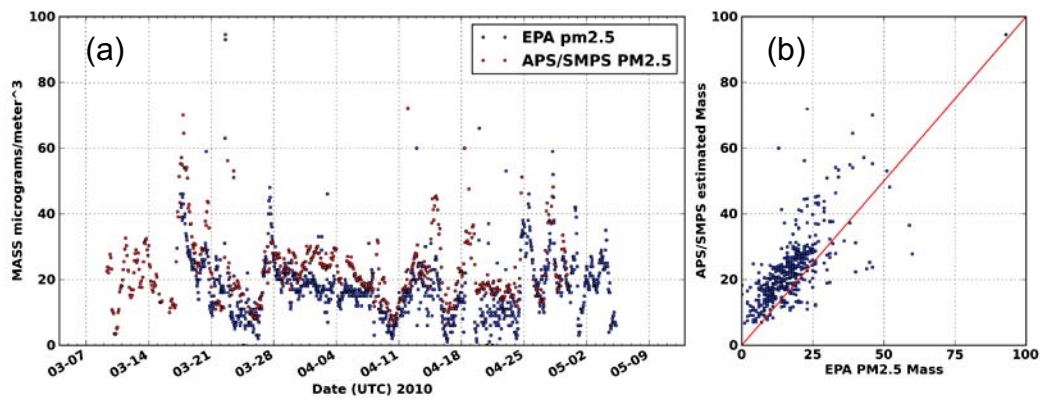
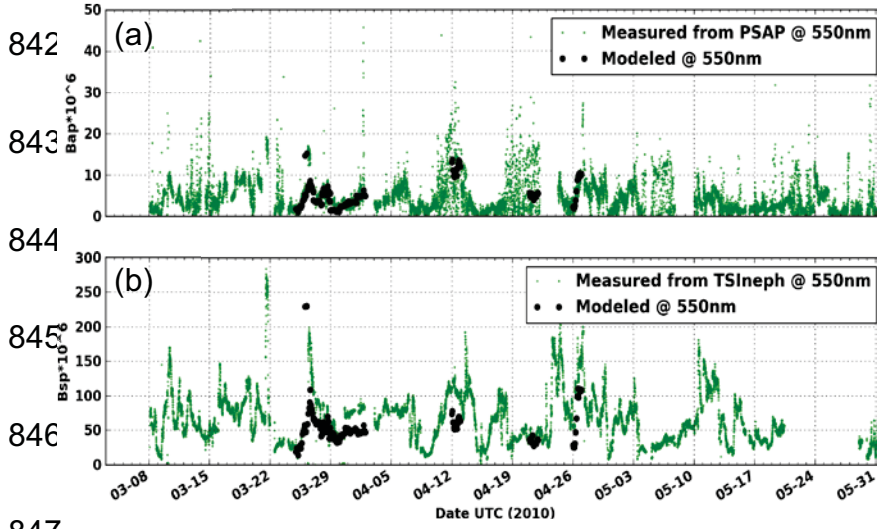


Figure 5 (a) Time series comparison of T-EPA's TEOM PM<sub>2.5</sub> to APS-SMPS PM<sub>2.5</sub> mass using the assumed bulk density for this study (b) scatterplot of the same variables with 1:1 line shown in red.



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849 Figure 6 Time series comparisons between modeled and measured absorption coefficients (a)  
 850 and scattering coefficients (b). Data in green/black are the measured/measured coefficients,  
 851 respectively. Scatter plots of absorption coefficients (c) and scattering coefficients (d). Also  
 852 shown are the corresponding error bars for scattering (15%), absorption (25%), and model  
 853 (30%), along with the 1:1 lines (black) and linear fits (red). All optical coefficients are in units of  
 854  $Mm^{-1}$ . See text for details.

