

1 **Air Pollution Scenario over Pakistan: Characterization and Ranking of Extremely**
2 **Polluted Cities using Long-Term Concentrations of Aerosols and Trace Gases**

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31 **Abstract**

32 Pakistan ranks third in the world in terms of mortality attributable to air pollution, with
33 aerosol mass concentrations ($PM_{2.5}$) consistently well above WHO (World Health Organization)
34 air quality guidelines (AQG). However, regulation is dependent on a sparse network of air quality
35 monitoring stations and insufficient ground data. This study utilizes long-term observations of
36 aerosols and trace gases to characterize and rank the air pollution scenarios and pollution
37 characteristics of 80 selected cities in Pakistan. Datasets used include (1) the Aqua and Terra
38 (AquaTerra) MODIS (Moderate Resolution Imaging Spectroradiometer) Level 2 Collection 6.1
39 merged Dark Target and Deep Blue (DTB) aerosol optical depth (AOD) retrieval products; (2) the
40 CAMS (Copernicus Atmosphere Monitoring Service) reanalysis PM_1 , $PM_{2.5}$, and PM_{10} data; (3) the
41 MERRA-2 (Modern-Era Retrospective analysis for Research and Applications, Version 2)
42 reanalysis $PM_{2.5}$ data, (4) the OMI (Ozone Monitoring Instrument) tropospheric vertical column
43 density (TVCD) of nitrogen dioxide (NO_2), and VCD of sulfur dioxide (SO_2) in the Planetary
44 Boundary Layer (PBL), (5) the VIIRS (Visible Infrared Imaging Radiometer Suite) Nighttime Lights
45 data, (6) MODIS Collection 6 Version 2 global monthly fire location data (MCD14ML), (7)
46 population density, (8) MODIS Level 3 Collection 6 land cover types, (9) AERONET (AErosol
47 RObotic NETwork) Version 3 Level 2.0 data, and (10) ground-based $PM_{2.5}$ concentrations from air
48 quality monitoring stations. Potential Source Contribution Function (PSCF) analyses were
49 performed by integrating with ground-based $PM_{2.5}$ concentrations and the NOAA (National
50 Oceanic and Atmospheric Administration) HYSPLIT (Hybrid Single-Particle Lagrangian Integrated
51 Trajectory) air parcel back trajectories to identify potential pollution source areas which are

52 responsible for extreme air pollution in Pakistan. Results show that the ranking of the top
53 polluted cities depends on the type of pollutant considered and the metric used. For example,
54 Jhang, Multan, and Vehari were characterized as the top three polluted cities in Pakistan when
55 considering AquaTerra DTB AOD products; for PM₁, PM_{2.5}, and PM₁₀ Lahore, Gujranwala, and
56 Okara were the top three; for tropospheric NO₂ VCD Lahore, Rawalpindi, and Islamabad and for
57 PBL SO₂ VCD Lahore, Mirpur, and Gujranwala. The results demonstrate that Pakistan's entire
58 population has been exposed to high PM_{2.5} concentrations for many years, with a mean annual
59 value of 54.7 µg/m³, over all Pakistan from 2003 to 2020. This value exceeds Pakistan's National
60 Environmental Quality Standards (Pak-NEQS, i.e., <15 µg/m³ annual mean) for ambient air
61 defined by the Pakistan Environmental Protection Agency (Pak-EPA) as well as the WHO Interim
62 Target-1 (i.e., mean annual PM_{2.5} <35 µg/m³). The spatial analyses of the concentrations of
63 aerosols and trace gases in terms of population density, nighttime lights, land cover types, and
64 fire location data, and the PSCF analysis indicate that Pakistan's air quality is strongly affected by
65 anthropogenic sources inside of Pakistan, with contributions from surrounding countries.
66 Statistically significant positive (increasing) trends in PM₁, PM_{2.5}, PM₁₀, tropospheric NO₂ VCD,
67 and SO₂ VCD were observed in ~89%, ~67%, ~48%, 91%, and ~88% of the Pakistani cities (80
68 cities), respectively. This comprehensive analysis of aerosol and trace gas levels, their
69 characteristics in spatio-temporal domains, and their trends over Pakistan, is the first of its kind.
70 Results will be helpful to the Ministry of Climate Change (Government of Pakistan), Pak-EPA,
71 SUPARCO (Pakistan Space and Upper Atmosphere Research Commission), policymakers, and the
72 local research community to mitigate air pollution and its effects on human health.

73 **Keywords:** MODIS; AOD; CAMS; MERRA-2; PM₁; PM_{2.5}; PM₁₀; OMI; NO₂; SO₂; PSCF; Pakistan

74 **Highlights:**

- 75 • Lahore, Gujranwala, and Okara are the most polluted city based on PM_{2.5}
- 76 • Jhang, Multan, and Vehari are the most polluted cities based on AOD
- 77 • Aerosols, nighttime lights, population, cropland, and fire show same spatial patterns
- 78 • Pakistan's entire population is exposed to long-term PM_x (x = 1, 2.5, & 10)
- 79 • Pakistan's air quality is mainly affected by local anthropogenic sources

80 **1. Introduction**

81 With the rapid increase in population and overexploitation of natural resources, air pollution
82 is a serious global environmental concern. According to the World Health Organization (WHO
83 2018a), air pollution levels are dangerously high worldwide as 9 out of 10 people breathe polluted
84 air, and each year 7 million deaths are caused by outdoor and indoor aerosol pollutants. Outdoor
85 (ambient) air pollution is due to high concentrations of different species including airborne
86 particulate matter (PM), ozone (O₃), nitrogen dioxide (NO₂), volatile organic compounds (VOC),
87 carbon monoxide (CO), and sulfur dioxide (SO₂), which have adverse health effects (Mannucci
88 and Franchini 2017). Although air pollution is a global problem, the latest WHO air quality
89 database reveals that 97% of affected cities are in low- and middle-income countries with more
90 than 100,000 inhabitants (WHO 2018b). Air pollution is endemic to Pakistan, being listed among
91 low- and middle-income countries as well as being the most urbanized of its South Asian
92 counterparts (77.42 million or 36.37 % of the urban population, with 2.52 % annual growth rate)
93 (UNDP 2019). Purohit et al. (2013) predicted that under current emission control standards, air

94 pollution would decrease life expectancy by more than 100 months by 2030. The Health Effects
95 Institute (2019) reported that since 1990, Pakistan's entire population has been exposed to PM_{2.5}
96 (the integrated dry mass of aerosol particulates with an aerodynamic diameter less than 2.5 µm)
97 annual mean concentrations of 58 µg/m³ in 2017, levels exceeding WHO Interim Target-1 (i.e.,
98 <35 µg/m³). Pakistan ranks third in the world in terms of mortality attributable to air pollution,
99 with an annual loss of 128,000 lives (Government of Pakistan 2019). Recently, on October 30,
100 2019, the Air Quality Index (AQI) was 484 in Lahore (the second-largest city with the highest
101 urbanization rate of 6.12 percent per annum), well above the threshold of 300 for "hazardous"
102 level (Amnesty International 2019). The winter of 2019-2020 witnessed a spate of smog, which
103 compelled authorities in Punjab to close schools for an extended period. The formation of this
104 smog was fueled by the buildup of anthropogenic aerosols having 65% of sources within Pakistan.
105 The principal cause for smog formation is NO_x, which is emitted primarily from Pakistan's 23.6
106 million transport vehicles (58%), followed by industry and power, which accounts for 34% of
107 emissions (Amnesty International 2019; Government of Pakistan 2019; UNDP 2019). According
108 to the Pakistan Air Quality Initiative (PAQI), Lahore, Peshawar, Islamabad, and Karachi are the
109 most polluted cities where air quality does not meet WHO air quality guidelines during autumn
110 and winter (PAQI 2018). Air pollution monitoring throughout Pakistan is challenging due to
111 sparsely distributed air quality monitoring stations, though several remote sensing studies have
112 been conducted.

113 Satellite observations provide spatial distributions of column-integrated concentrations
114 which are related to the near-surface concentrations through meteorological and physico-
115 chemical processes, thus complementing local ground-based observations. Gupta et al. (2013)

116 analyzed MODIS (Moderate Resolution Imaging Spectroradiometer) AOD (Aerosol Optical Depth)
117 retrievals over Lahore and Karachi from 2001 to 2010 and reported higher aerosol loadings near
118 the city center than outside the city. Tariq et al. (2016) analyzed ground-based and satellite-based
119 aerosol optical properties over Lahore during intense haze events in October 2013 and reported
120 crop residue burning and urban-industrial emissions as the main sources of high AOD levels. Bilal
121 et al. (2016) evaluated the performance of the Aqua-MODIS (MYD04) level 2 aerosol products
122 over Lahore and Karachi from 2007 to 2013, and recommended the use of Dark Target (DT) and
123 Deep Blue (DB) algorithms over Karachi and Lahore, respectively, for regional air quality
124 applications, as these cities have different land cover characteristics and aerosol types. Other
125 remote sensing studies have been conducted on atmospheric trace gases, such as ozone (O₃),
126 nitrogen dioxide (NO₂), sulfur dioxide (SO₂), and carbon dioxide CO₂, as well as their trends over
127 time (Khokhar et al. 2016; Khokhar et al. 2015; Tariq and Ali 2015; ul-Haq et al. 2017; ul-Haq et
128 al. 2014; Ul-Haq et al. 2015). Zhang et al. (2020) conducted the first study of the vertical
129 distribution of aerosol optical properties over Pakistan using CALIPSO (Cloud-Aerosol Lidar and
130 Infrared Pathfinder Satellite Observation) data.

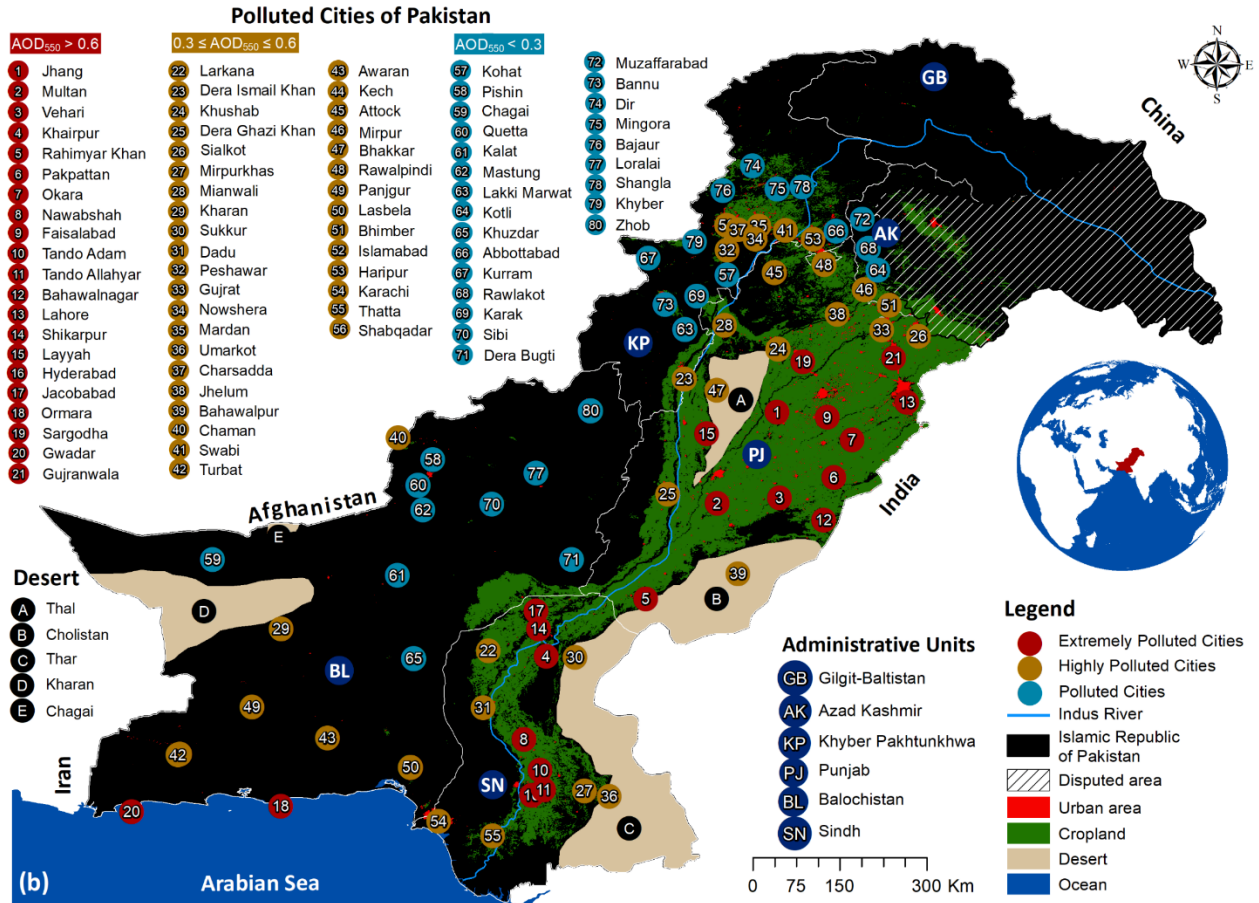
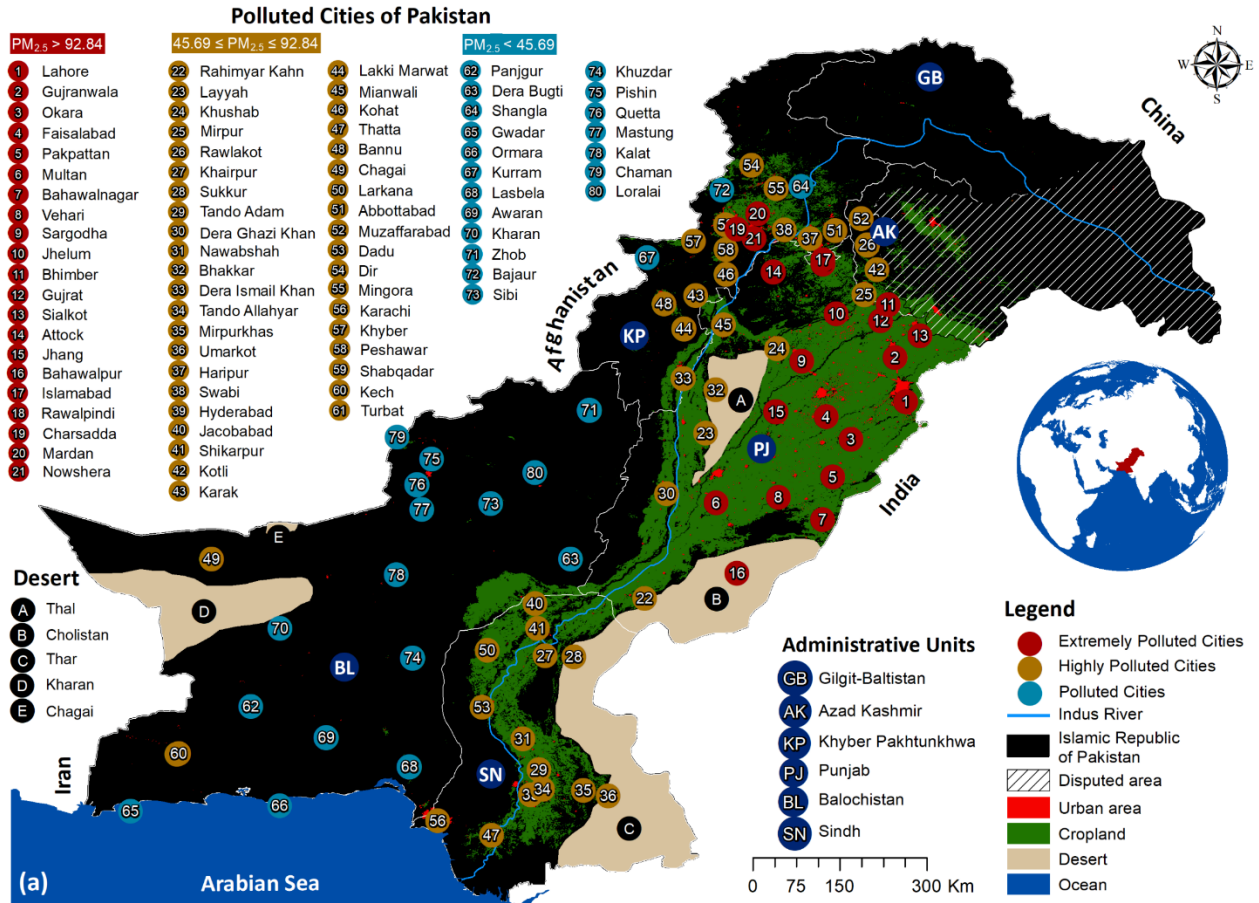
131 Cities are areas of high activity, and every city is a huge source of local anthropogenic aerosols
132 and trace gases from industrial and human activities, which can impact air quality, visibility, and
133 alter the physico-chemical properties of the atmosphere at local, regional, and global scales.
134 Although several studies of AOD and atmospheric trace gases have been conducted over
135 Pakistan, no study has encompassed different particle fractions (PM_x, x = 1, 2.5, and 10) on the
136 national scale, i.e., the dry mass of ultrafine particles with an aerodynamic diameter less than 1
137 μm (PM₁), 2.5 μm (PM_{2.5}) and 10 μm (PM₁₀). PM₁ is part of PM_{2.5}, PM_{2.5} is part of PM₁₀. It is of

138 great importance to identify the cities most affected by different PM_x fractions, as they have
139 different effects on, for instance, health and chemical and physical processes in the atmosphere,
140 and this is the first study to do so. Moreover, very few studies have investigated the long-term
141 trends in pollutant concentrations at city level, which can provide additional insights into the link
142 between concentrations and the changes in emissions. Furthermore, previous studies are not
143 comprehensive enough to answer questions such as: which are the most and least polluted cities
144 of Pakistan, and what are the likely pollution sources? Therefore, this study aims (1) to
145 extensively characterize and rank the extremely polluted cities of Pakistan, considering multiple
146 sources and aerosol mass fractions, for 80 carefully selected cities, representing almost all major
147 urban centers of Pakistan, and (2) to identify the likely pollutant sources by performing PSCF
148 (Potential Source Contribution Function) analysis with the integration of HYSPLIT (Hybrid Single-
149 Particle Lagrangian Integrated Trajectory) back trajectory and ground-based $PM_{2.5}$
150 concentrations. This study is based on long-term combined Aqua and Terra (AquaTerra) MODIS
151 data from 2003 to 2017, OMI (Ozone Monitoring Instrument) data (NO_2 and SO_2) from 2004 to
152 2019, CAMS (Copernicus Atmosphere Monitoring Service) reanalysis PM_1 , $PM_{2.5}$, and PM_{10} data
153 from 2003 to 2019, MERRA-2 (Modern-Era Retrospective analysis for Research and Applications,
154 Version 2) $PM_{2.5}$ data from 2003 to 2020, VIIRS (Visible Infrared Imaging Radiometer Suite)
155 Nighttime Lights from 2012 to 2019, LandScan global population density for 2019, MODIS land
156 cover type for 2019, MODIS global monthly fire location data from 2003 to 2020, ground-based
157 $PM_{2.5}$ concentrations from 2018 to 2020, and AERONET (AErosol RObotic NETwork) AOD
158 measurements from 2006 to 2017. Detailed information on the data used in this study is provided
159 in Section 3.

160 2. Study Area

161 Pakistan, with a population of 212.82 million, is the sixth most populous country in the world.
162 It lies between 23°35' to 37°05' North and 60°50' to 77°50' East, having a diverse geographical
163 landscape bordered by China, the Himalayas, India, Afghanistan, Iran, and the Arabian Sea.
164 Geographically, Pakistan falls into three major regions: the northern highlands, constituting parts
165 of the Hindu Kush, the Karakoram Range, and the Himalayas; the Indus River basin plain in the
166 center and east (65% of the total area i.e. 796,096 km²); and the Balochistan Plateau in the south
167 and west (Government of Pakistan 2019). Administratively, Pakistan has six units: Punjab, Sindh,
168 Khyber Pakhtunkhwa, Balochistan, Azad Kashmir, and Gilgit Baltistan. Punjab is the most
169 populous (112.38 million; 53%) administrative unit of Pakistan, followed by Sindh (49.05 million;
170 23%), Khyber Pakhtunkhwa (36.5 million; 17%), and Balochistan (12.7 million; 6%). Balochistan
171 has the largest area (43.6 %), followed by Punjab (25.8%), Sindh (17.7 %), and Khyber
172 Pakhtunkhwa (12.78%). Sindh is the most urbanized and industrialized administrative unit of
173 Pakistan with 52% urban population. Islamabad (2.1 million; 1%) Capital Territory (ICT), a rather
174 small unit in terms of area (0.1 %), is, in fact, the second most urbanized (50.58%) region of
175 Pakistan, and has an annual urbanization rate of 4.91 %. Currently, 10 cities in Pakistan have a
176 population of over one million, and 7 have higher per-capita incomes than the national average
177 (UNDP 2019). The Pakistan economic survey 2018-19 reports a total cropped area of 22.6 million
178 hectares, and agricultural contributions of 18.5 % to the GDP, compared with 20.3% from the
179 industrial sector (Government of Pakistan 2019).

180 This study covers almost all prominent cities in Pakistan including all administrative units and
181 their capital cities, and the Capital of the country (Figure 1). In summary, the study area analyzes
182 23 cities from the most populated administrative unit, Punjab; Khyber Pakhtunkhwa is also well-
183 represented by 19 urban centers; Balochistan is the least populated but the largest administrative
184 unit, and is represented by 19 cities; 14 other cities exemplify the diversity of Sindh in the South-
185 East, and 5 cities represent the attractive hilly land of Azad Kashmir.



187 **Figure 1:** Geographical and administrative map of Pakistan including a list of cities used in the
188 present study. Cities are characterized using (a) yearly mean CAMS (Copernicus Atmosphere
189 Monitoring Service) reanalysis $PM_{2.5}$ concentrations ($\mu\text{g}/\text{m}^3$) for the years 2003 and 2020, and
190 (b) yearly mean AquaTerra MODIS DTB AOD retrievals at 550 nm from 2003 to 2017. Extremely
191 polluted cities (red color) are defined for $PM_{2.5} > 92.84$ ($AOD > 0.6$) (3rd quartile), highly polluted
192 cities (brown color) for $45.69 \leq PM_{2.5} \leq 92.84$ ($0.3 < AOD < 0.6$) (between 3rd and 1st quartiles),
193 and polluted cities (purple color) for $PM_{2.5} < 45.69$ ($AOD < 0.3$) (1st quartile) using descriptive
194 statistics (Table S1). Cities are not defined as low polluted or clean cities as annual mean $PM_{2.5}$
195 concentrations for all cities exceed Pakistan's National Environmental Quality Standards (Pak-
196 NEQS) for ambient air ($<15 \mu\text{g}/\text{m}^3$ annual mean).

197 **3. Dataset**

198 **3.1 AERONET Data**

199 The AERONET (AErosol RObotic NETwork) (Holben et al. 1998; Holben et al. 2001) is a global
200 network of calibrated Sunphotometers coordinated by NASA (National Aeronautics and Space
201 Administration) which provides regular measurements of spectral AOD at 340 nm, 380 nm, 440
202 nm, 500 nm, 675 nm, 870 nm, 1020 nm, and 1640 nm, and AE at 340–440 nm, 380–500 nm, 440–
203 675 nm, and 500–870 nm at three levels, i.e., Level 1.0 (unscreened), Level 1.5 (cloud-screened),
204 and Level 2.0 (cloud-screened and quality-assured), under cloud-free skies (Smirnov et al. 2000)
205 for every 15 minutes with an uncertainty of 0.01–0.02 (Holben et al. 2001). The present study
206 used Version 3 Level 2.0 AOD at 500 nm (AOD_{500}) and AE at 440–675nm ($AE_{440-675}$) (Giles et al.
207 2019) obtained from the AERONET website (<https://aeronet.gsfc.nasa.gov/>) for the Lahore

208 (31.47987° N, 74.26406° E) and Karachi (24.94574° N, 67.13594° E) sites from 2006 to 2017. The
209 Lahore and Karachi AERONET sites are located in an urban area, and approximately 20 km away
210 from the Arabian Sea coast, respectively.

211 **3.2 AquaTerra MODIS Data**

212 In the present study, Aqua and Terra MODIS C6.1 L2 aerosol products at 10 km spatial
213 resolution are obtained from 2003 to 2017 for Pakistan from the LAADS DAAC
214 (<https://ladsweb.modaps.eosdis.nasa.gov/>). The MODIS aerosol product provides DT AOD
215 retrievals over land and water surfaces (Levy et al. 2013), and DB AOD retrievals only over land
216 (Hsu et al. 2013). The DT and DB AOD retrievals for different collections are extensively validated
217 against Sunphotometer (AERONET) measurements at regional (Bilal et al. 2019b; Bilal et al. 2014;
218 Che et al. 2019; de Leeuw et al. 2018; Fan et al. 2017; Filonchik et al. 2019; Gupta et al. 2013; He
219 et al. 2018; Islam et al. 2019; Livingston et al. 2014; Mhawish et al. 2017; More et al. 2013; Nichol
220 and Bilal 2016; Shen et al. 2018; Shi et al. 2013; Sogacheva et al. 2018; Wang et al. 2017; Wang
221 et al. 2019; Xiao et al. 2016; Xie et al. 2011) and global scales (Bilal et al. 2018a; Bilal et al. 2017;
222 Levy et al. 2013; Levy et al. 2010; Mehta et al. 2016; Remer et al. 2013; Sayer et al. 2013; Sayer
223 et al. 2014; Sayer et al. 2015; Tong et al. 2020). These studies have reported overestimation and
224 underestimation in DT and DB AOD retrievals respectively, due to error in the estimated surface
225 reflectance and aerosol scheme used in the inversion methods, but overall their performance is
226 satisfactory. Previous studies (Bilal et al. 2018a; Bilal and Nichol 2017; Bilal et al. 2017; Bilal et al.
227 2018b; Mei et al. 2019; Sayer et al. 2014) have also reported different spatial coverage of DT and
228 DB AOD retrievals over land due to differences in their approaches, i.e., pixel selection criteria,

229 estimation of surface reflectance, and the cloud mask. Therefore, a new merged Scientific Data
230 Set (SDS: AOD 550 Dark Target Deep Blue Combined) was introduced which contains only the
231 highest quality DT and DB (DTB) AOD retrievals or their average values (Levy et al. 2013). The
232 purpose of this new dataset is to improve spatial coverage over land (Levy et al., 2013; Sayer et
233 al., 2014), i.e., to retrieve AOD in the same image for those regions where either the DT or the
234 DB algorithm does not achieve a successful retrieval (Bilal et al. 2017; Levy et al. 2013). The
235 merged DTB AOD retrievals have been validated at regional and global scales (Ali and Assiri 2019;
236 Bilal et al. 2018a; Bilal and Nichol 2017; Bilal et al. 2017; Sayer et al. 2014; Sogacheva et al. 2018).
237 However, the new customized method-1 (CM1) (Bilal et al. 2017), which is named Simplified
238 Merge Scheme (SMS) in the later publications (Bilal et al. 2018a; Bilal et al. 2018b), provides
239 equally consistent data quality with the combined DTB AOD retrievals available in C6.1, but with
240 significantly improved spatio-temporal coverage.

241 **3.3 CAMS Data**

242 The Copernicus Atmosphere Monitoring Service (CAMS) reanalysis is an atmospheric
243 composition dataset generated by the European Centre for Medium-Range Weather Forecasts
244 (ECMWF). The global CAMS model combines satellite-based observations with chemistry-aerosol
245 modeling using the four-dimensional variational (4D-VAR) data assimilation technique to obtain
246 the mass concentration of aerosols and trace gases. CAMS uses the MACCity inventory at $0.5^\circ \times$
247 0.5° spatial resolution for anthropogenic emissions which covers the period 1960–2010 (Granier
248 et al. 2011). Detailed information about the model and the emission inventory can be found in
249 (Flemming et al. 2017; Flemming et al. 2015). In this study, the ground-based mass concentration

250 of particulate matter, including particles with an aerodynamic diameter of less than 1 μm (PM_{10}),
251 less than 2.5 μm ($\text{PM}_{2.5}$), and less than 10 μm (PM_{10}) was obtained from the CAMS reanalysis data
252 for the years 2003 and 2020. PM_x ($x = 1, 2.5, \& 10$) data were used at two different spatiotemporal
253 resolutions, i.e., (i) CAMS global reanalysis dataset at $0.75^\circ \times 0.75^\circ$ spatial resolution and 3-hourly
254 temporal resolution from 2003 to 2020, and (ii) CAMS near-real time dataset at $0.125^\circ \times 0.125^\circ$
255 spatial resolution and 12-hourly temporal resolution from 2018 to 2020 (Inness et al. 2019). The
256 PM_x data at 0.75° grid size and 3-hourly temporal resolution were used for long-term climatology
257 and for characterizing extremely polluted cities, whereas, the CAMS near-real time data at 0.125°
258 grid size and 12-hourly temporal resolution were used for validation against ground-based $\text{PM}_{2.5}$
259 concentrations obtained from air quality monitoring stations.

260 **3.4 MERRA-2 Reanalysis Data**

261 The MERRA-2 (Modern-Era Retrospective analysis for Research and Applications, Version 2)
262 atmospheric reanalysis is the latest data released by the NASA GMAO (Global Modeling and
263 Assimilation Office) in 2017 (Buchard et al. 2017; Randles et al. 2017). The MERRA-2 aerosol
264 gridded data, i.e., dust, sea salt, sulfate, black carbon, and organic carbon, are simulated with 72
265 vertical layers from the surface to higher than 80 km using the GEOS-5 (GMAO Earth system
266 model version 5) model radiatively coupled to the GOCART (Goddard Chemistry Aerosol
267 Radiation and Transport) model (Chin et al. 2002; Colarco et al. 2010). For anthropogenic
268 emissions, MERRA-2 uses the EDGAR-4.2 emission inventory at $0.1^\circ \times 0.1^\circ$ spatial resolution
269 which covers the period 1970–2008 (Janssens-Maenhout et al. 2013). In this study, the MERRA-
270 2 aerosol gridded data (dust, sea salt, sulfate, black carbon, and organic carbon) at $0.5^\circ \times 0.625^\circ$

271 spatial resolution from 2018 to 2020 were used. More details about MERRA-2 reanalysis data can
272 be found in Randles et al. (2017) and Buchard et al. (2017).

273 **3.5 Ground-based PM_{2.5} Measurements**

274 Ground-based PM_{2.5} measurements were obtained from two different air quality monitoring
275 networks. Firstly, PM_{2.5} data were obtained from 4 air quality stations operated by the US
276 Consulates in Islamabad, Karachi, Lahore, and Peshawar, and secondly, 54 air quality monitoring
277 stations operated by PAQI in Lahore (24 stations), Karachi (15), Islamabad (5) Sialkot (3),
278 Peshawar (2), Rawalpindi (2), Faisalabad (1), Gujranwala (1), and Muridke (1). Due to the lack of
279 a well-developed and standard air quality network of ground-based PM_{2.5} measurements, this
280 study is limited to only these cities for the validation of CAMS and MERRA-2 reanalysis PM_{2.5}
281 gridded data. PM_{2.5} concentrations from the US Consulates are measured by beta gauge
282 attenuation monitors (BAM-1020; Met One Instruments), hereafter referred to as BAM PM_{2.5}
283 concentrations. To increase social awareness in Pakistan, PAQI provides PM_{2.5} data using a
284 nationwide network of low-cost air quality monitors (IQAir AirVisual Pro), hereafter referred to
285 as LCM PM_{2.5} concentrations. In this study, LCM and BAM PM_{2.5} measurements were used for
286 January 2018–December 2019 and January 2019–February 2021, respectively. More details
287 about PAQI (LCM) and US Consulates (BAM) PM_{2.5} data can be found in Shi et al. (2020) and
288 Mhawish et al. (2020), respectively.

289 **3.6 OMI Data**

290 The Ozone Monitoring Instrument (OMI) onboard the Aura satellite was launched in July 2004
291 as a part of the A-Train satellite constellation. OMI is a hyperspectral sensor that measures the
292 radiation reflected from the earth-atmosphere system, in the wavelength range 250–500 nm and
293 provides daily global coverage at a spatial resolution of $13 \times 24 \text{ km}^2$ at nadir. The OMI OMAERUV
294 algorithm utilizes the sensitivity of near-UV spectral regions to aerosol absorption, and it
295 retrieves absorbing aerosol optical depth (AAOD) at 388nm (Torres et al. 2013; Torres et al. 2007).
296 Along with the AAOD, the OMAERUV algorithm also provides an ultraviolet Aerosol Index (UVAI),
297 AOD, and Single Scattering Albedo (SSA). OMI also retrieves the atmospheric trace gases O_3 , NO_2
298 and SO_2 (Carn et al. 2017; Krotkov et al. 2017; Krotkov et al. 2016; Li et al. 2017; Li et al. 2013;
299 Veefkind et al. 2006). In this study, OMAERUV version 3 Level 3 daily cloud-screened (cloud
300 fraction < 30 %) NO_2 tropospheric vertical column density (TVCD) (OMNO2e), and SO_2 VCD in the
301 planetary boundary layer (PBL) (OMSO2e) gridded at $0.25^\circ \times 0.25^\circ$ spatial resolution from 2004
302 to 2019 were used.

303 **3.7 Other Supporting Datasets**

304 Other supporting datasets include (i) annual mean VIIRS nighttime lights data
305 (<https://eogdata.mines.edu/products/vnl/>) from 2012 to 2019 derived from monthly mean data
306 (Elvidge et al. 2021), (ii) MODIS Collection 6 global monthly Fire Location product (MCD14ML)
307 from 2003 to 2020 (<https://firms.modaps.eosdis.nasa.gov/download/>), (iv) MODIS Collection 6
308 Level 3 land cover type product (MCD12Q1) for 2019

309 (<https://ladsweb.modaps.eosdis.nasa.gov/>), and (v) the LandScan population density
310 (<https://landscan.ornl.gov/>) for 2019 (Rose et al. 2020).

311 **4. Research Methodology**

312 To investigate the air pollution scenario over Pakistan and characterize the extremely
313 polluted cities, in this study the following methodology was adopted:

314 1. MODIS AOD retrievals were obtained from the Scientific Data Set (SDS) “Optical Depth
315 Land and Ocean” and “Deep Blue Aerosol Optical Depth 550 Land Best Estimate”. Only
316 the highest quality-assured DT (QA = 3) and DB (QA ≥ 2) retrievals were used, as
317 recommended by previous studies (Bilal et al. 2013; Levy et al. 2013; Mhawish et al. 2019;
318 Sayer et al. 2013). Pakistan has a variety of land cover types, e.g., snow and mountainous
319 land surface in Northern Pakistan, plain and agricultural land surfaces in Central Pakistan,
320 and arid and desert land surfaces in southern Pakistan, where the DT and DB algorithms
321 overestimate and underestimate, respectively. However, the DT algorithm is unable to
322 provide retrievals over the arid and desert land surfaces of Balochistan. Similar results
323 were observed and reported in our previous study over Pakistan (Bilal et al. 2016).
324 Therefore, in the present study, we preferred to generate the combined (merged) DTB
325 AOD₅₅₀ retrievals for both Aqua and Terra MODIS data from 2003 to 2017 using the
326 customized method-1 (CM1) (Bilal et al. 2017), which in later publications is named
327 Simplified Merge Scheme (SMS) (Bilal et al. 2018a; Bilal et al. 2018b), i.e., an average of
328 the DT and DB AOD retrievals or the available one with the highest quality flag (Equation
329 1), to enhance spatio-temporal coverage.

330

$$331 \quad DTB \ AOD_{550} = \left\{ \begin{array}{ll} \text{if only DT AOD exists} & \rightarrow \quad DT \\ \text{if only DB AOD exists} & \rightarrow \quad DB \\ \text{if both DT and DB AOD exist} & \rightarrow \quad (DT + DB)/2 \end{array} \right\} \quad (1)$$

332

333 2. Aqua and Terra MODIS may not provide complete spatial coverage due to cloud cover.

334 On days when Aqua provides AOD retrievals, Terra may not, and vice-versa. Therefore,

335 for more complete spatial coverage between Aqua and Terra as well as to represent an

336 average air pollution scenario between morning and afternoon times with a single

337 dataset, the combined AquaTerra DTB AOD retrievals were generated from the Aqua DTB

338 and Terra DTB AOD retrievals using SMS/CM1, i.e., an average of the Aqua and Terra DTB

339 AOD retrievals or the available one (Equation 2).

$$340 \quad AquaTerra \ AOD = \left\{ \begin{array}{ll} \text{if only Aqua AOD exists} & \rightarrow \quad Aqua \\ \text{if only Terra AOD exists} & \rightarrow \quad Terra \\ \text{if both Aqua and Terra AOD exist} & \rightarrow \quad (Aqua + Terra)/2 \end{array} \right\} \quad (2)$$

341 3. The AquaTerra DTB AOD retrievals are validated against Sunphotometer AOD

342 measurements obtained for Lahore (31.480° N and 74.264° E) and Karachi (24.946° N and

343 67.136° E) AERONET sites. The AERONET Sunphotometer does not provide AOD at 550

344 nm (AOD₅₅₀), AOD₅₅₀ is interpolated using AOD at 500 nm (AOD₅₀₀) and Ångström

345 Exponent at 440-675 nm (AE₄₄₀₋₆₇₅) based on the Ångström Exponent empirical formula

346 (Equation 3) (Eck et al. 1999). Collocated AquaTerra and AERONET AOD retrievals were

347 defined as the average of at least two pixels of DTB within a spatial region of 3 × 3 pixels

348 (at least 2 out of 9 pixels) centered on the AERONET site and the average of at least two
349 AERONET AOD measurements between 10:00 and 14:30 local solar time.

$$350 \quad AOD_{550} = AOD_{500} \left(\frac{550}{500} \right)^{-AE_{440-667}} \quad (3)$$

351 4. Accuracy and errors are reported using the Pearson correlation coefficient (r), the
352 expected error (EE, Equation 4), and relative mean bias (RMB, Equation 5). The slope (β ,
353 Equation 6) and intercept (α , Equation 7) between collocated AquaTerra DTB and
354 AERONET AOD data are calculated using the reduced major axis (RMA) regression which
355 incorporates errors in both independent (AERONET) and dependent (MODIS) variables
356 (Bilal et al. 2019a; Harper 2016). The performance of the Terra and Aqua DT, DB, and DTB
357 AOD retrievals is evaluated based on (i) highest correlation coefficient (r), (ii) highest
358 number of collocated retrievals (N), (iii) the highest percentage of retrievals within the EE,
359 and (iv) lowest RMB. To evaluate the performance of the collocated retrievals, the
360 following criteria are utilized (Bilal et al. 2017): the DT, DB, and DTB retrievals are
361 considered to be of equal quality if the relative difference is within (1) 5% for the
362 correlation coefficient (r), (2) 10% for the collocated retrievals, (3) 10% for the percentage
363 of retrievals is within the EE, and (4) RMB < 25%.

$$364 \quad EE = \pm (0.05 + 0.20 \times AERONET_{AOD}) \quad (4)$$

365 The upper and lower EE envelopes are calculated using Equations 4a and 4b.

$$366 \quad \text{Upper EE envelope} = AERONET_{AOD} + |EE| \quad (4a)$$

367
$$\text{Lower EE envelope} = \text{AERONET}_{AOD} - |EE| \quad (4b)$$

368 The percentage of best retrieved MODIS AOD retrievals within the EE is reported using
 369 Equation 4c.

370
$$\%EE = \text{AERONET}_{AOD} - |EE| \leq \text{MODIS}_{AOD} \leq \text{AERONET}_{AOD} + |EE| \quad (4c)$$

371 Where $|EE|$ is the absolute value of EE.

372
$$\text{RMB} = \frac{(\overline{\text{MODIS}_{AOD}} - \overline{\text{AERONET}_{AOD}})}{\overline{\text{AERONET}_{AOD}}} \times 100 \quad (5)$$

373 Where, $\overline{\text{MODIS}_{AOD}}$ and $\overline{\text{AERONET}_{AOD}}$ are the mean of MODIS and AERONET AOD
 374 retrievals, respectively. $\text{RMB} > 0$ represents overestimation in MODIS AOD compared to
 375 AERONET AOD, $\text{RMB} < 0$ represents underestimation, and $\text{RMB} = 0$ represents no over- and
 376 under-estimations.

377
$$\beta = \frac{\sigma_{\text{MODIS}_{AOD}}}{\sigma_{\text{AERONET}_{AOD}}} \quad (6)$$

378
$$\alpha = \overline{\text{MODIS}_{AOD}} - \left(\frac{\sigma_{\text{MODIS}_{AOD}}}{\sigma_{\text{AERONET}_{AOD}}} \right) \times \overline{\text{AERONET}_{AOD}} \quad (7)$$

379 Where, β , α , $\sigma_{\text{MODIS}_{AOD}}$, and $\sigma_{\text{AERONET}_{AOD}}$ are the slope, intercept, the standard deviation
 380 of MODIS AOD, and standard deviation of AERONET AOD, respectively.

381 5. To show the long-term variation of the mean spatial distributions of AquaTerra AOD over
 382 Pakistan, the AOD retrievals from 2003 to 2017 are used to generate monthly mean

383 spatial AOD maps, and their corresponding pixel counts are calculated for reporting the
384 retrieval performance of both the DT and DB algorithms.

385 6. To assure the quality of the $PM_{2.5}$ data, validation of daily average CAMS and MERRA-2
386 $PM_{2.5}$ data was conducted against in-situ $PM_{2.5}$ measurements obtained from the air
387 quality monitoring stations. The performance was evaluated based on the correlation
388 coefficient (r), RMB (Eq. 5), and slope (Eq. 6). MERRA-2 $PM_{2.5}$ concentrations were
389 calculated based on five aerosol components using Equation 8 (Song et al. 2018), and
390 CAMS $PM_{2.5}$ and PM_{10} concentrations were calculated using Equations 9 and 10 (Rémy et
391 al. 2019).

392

$$393 \quad PM_{2.5} = [Dust_{2.5}] + [SS_{2.5}] + 1.375 \times [SO_4] + [BC] + 1.6 \times [OC] \quad (8)$$

394 Where, $Dust_{2.5}$, $SS_{2.5}$, BC, OC, and SO_4 are the GOCART concentrations of dust, sea salt,
395 black carbon, organic carbon, and sulfate in particles with a diameter smaller than 2.5 μm ,
396 respectively.

$$397 \quad PM_{2.5} = \rho([SS_1]/4.3 + [SS_2]/4.3 + [DD_1] + [DD_2] + 0.7[OM] + [BC] + 0.7[SU] \\ 398 \quad + 0.7[NI_1] + 0.25[NI_2] + 0.7[AM]) \quad (9)$$

399

$$400 \quad PM_{10} = \rho([SS_1]/4.3 + [SS_2]/4.3 + [DD_1] + [DD_2] + 0.4[DD_3] + [OM] + [BC] \\ 401 \quad + [SU] + [NI_1] + [NI_2] + [AM]) \quad (10)$$

402 Where $[SS_{1,2}]$ = sea salt aerosol, $[DD_{1,2,3}]$ = desert dust, $[NI_{1,2}]$ = nitrate, $[OM]$ = organic
403 matter, $[BC]$ = black carbon, $[SU]$ = sulfate, and $[AM]$ = ammonium (concentrations in
404 particles with a diameter smaller than $2.5 \mu\text{m}$ from the CAMS model).

405 7. To characterize extremely polluted cities in Pakistan, the DTB AOD retrieved from
406 AquaTerra, the $PM_{1,}$ $PM_{2.5,}$ and PM_{10} from CAMS data, and the SO_2 VCD and NO_2 TVCD
407 from OMI are used. Polluted months as well as years, for the corresponding polluted
408 cities, are also characterized based on each pollutant.

409 8. To assess recent changes in the concentrations of atmospheric constituents, the non-
410 parametric Mann Kendal test (Kendall and Gibbons 1990; Mann 1945) associated with
411 Theil-Sen's slope (Sen 1968; Theil 1992) was used to estimate and detect trends over the
412 main cities of Pakistan from 2003 to 2020. The non-parametric Mann Kendal test is often
413 used to detect monotonic trends in a time series and is also suitable for non-normally
414 distributed data, or if the data have some missing observations such as environmental
415 data. Further, the bootstrapping technique was used to eliminate serial autocorrelation
416 in the monthly mean aggregated time series data and increase the robustness of the test
417 (Hamed and Ramachandra Rao 1998; Salmi et al. 2002). The significance of the calculated
418 trend was assessed using the two-tailed test method at a 95% confidence interval.

419 9. The NOAA (National Oceanic and Atmospheric Administration) HYSPLIT (Hybrid Single-
420 Particle Lagrangian Integrated Trajectory Model) (Stein et al. 2015), a complete transport,
421 dispersion, and chemical transformation model, is used for back trajectory analysis to
422 determine the origin of air masses (Fleming et al. 2012) and highlight the possible sources
423 of aerosol pollutants affecting the air quality of Pakistan using the PSCF (Potential Source

424 Contribution Function) analysis. In this study, 72 hours HYSPLIT backward trajectories at
425 the height of 500 m above the ground level (AGL) were computed for every 6 hours at
426 seasonal scales from March 2020 to February 2021 using the GDAS (Global Data
427 Assimilation System) meteorological data at $1^\circ \times 1^\circ$ spatial resolution (available at
428 <ftp://arlftp.arlhq.noaa.gov/pub/archives/gdas1>). The PSCF analysis was performed for 4
429 cities selected because of the availability of ground-based $PM_{2.5}$ measurements from the
430 air quality stations operated by the US Consulates, namely, Peshawar, Islamabad, Lahore,
431 and Karachi. The height of 500 m AGL has been reported very useful as it is the
432 approximate height of the mixing layer (Begum et al. 2005). The backward trajectory
433 clustering and investigation of the origins of the particulate matter at the receptor
434 locations were studied using Meteoinfo TrajStat software (Version 2.0, available at
435 <http://meteothink.org/products/trajstat.html>) (Wang et al. 2009) in conjunction with
436 HYSPLIT and Geographic Information System (GIS).

437 The PSCF analysis was performed using 24-hour average ground-based $PM_{2.5}$
438 concentrations over a grid with a resolution of 0.5° , for the days that exceeded the Pak-
439 NEQS 24-hour air quality standards ($35 \mu\text{g}/\text{m}^3$). The PSCF value for a specific grid cell was
440 calculated on the assumption that the trajectory endpoint is located within a cell (i, j) and
441 the trajectory is assumed to collect pollutants emitted from different pocket emission
442 sources within that cell (i, j). The PSCF value can be interpreted as a conditional probability
443 describing the potential contributions of a grid cell to the high $PM_{2.5}$ loadings at the
444 receptor site. The error associated with the trajectory is proportional to the distance from

445 the receptor location (Begum et al. 2005). The PSCF value for the ij^{th} grid cell can be
 446 computed using Equation 11:

$$447 \quad PSCF(i, j) = m_{ij}/n_{ij} \quad (11)$$

448 Where, n_{ij} represents the number of endpoints that fall or pass through the ij^{th} cell
 449 and m_{ij} denotes for the number of endpoints in the ij^{th} cell having a higher pollutant
 450 concentration than the 24-hour Pak-NEQS. The uncertainty arising due to small n_{ij} is
 451 reduced by multiplying an arbitrary weight function $W_{i,j}$, which is multiplied into the
 452 PSCF. In this case, the weight function is given in Equation (12):

$$453 \quad W_{i,j} = \begin{cases} \text{if } n_{ij} > 3\bar{n} \rightarrow 1.00 \\ \text{if } 1.5\bar{n} < n_{ij} \leq 3\bar{n} \rightarrow 0.70 \\ \text{if } \bar{n} < n_{ij} \leq 1.5\bar{n} \rightarrow 0.42 \\ \text{if } n_{ij} \leq \bar{n} \rightarrow 0.15 \end{cases} \quad (12)$$

454 Where \bar{n} denotes the average number of endpoints per cell, which is calculated for each
 455 cell that has at least one endpoint. Therefore, the Weighted PSCF is expressed as Equation
 456 (13):

$$457 \quad WPSCF = W_{i,j} \times PSCF(i, j) \quad (13)$$

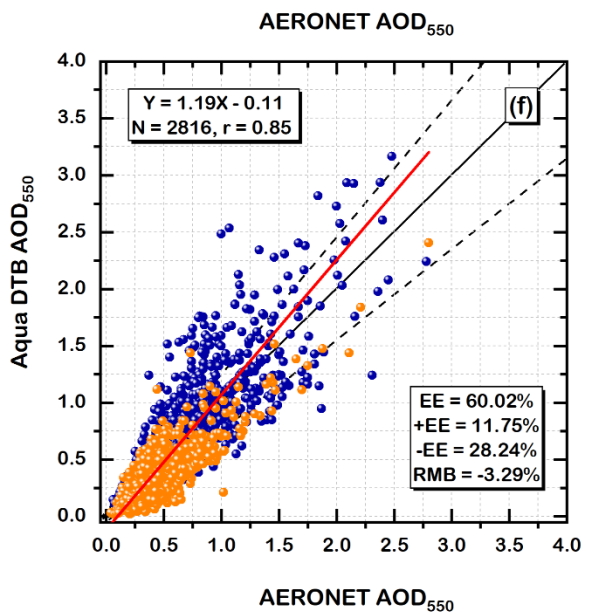
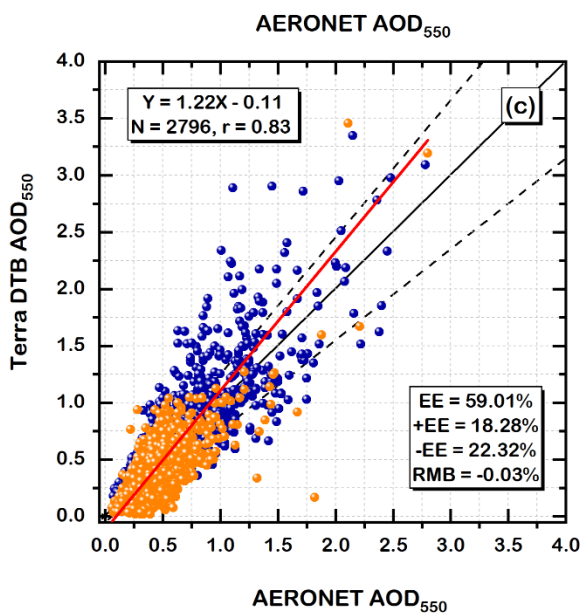
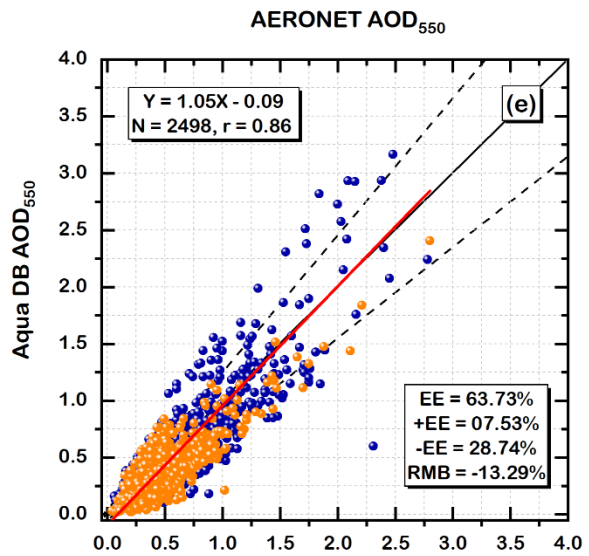
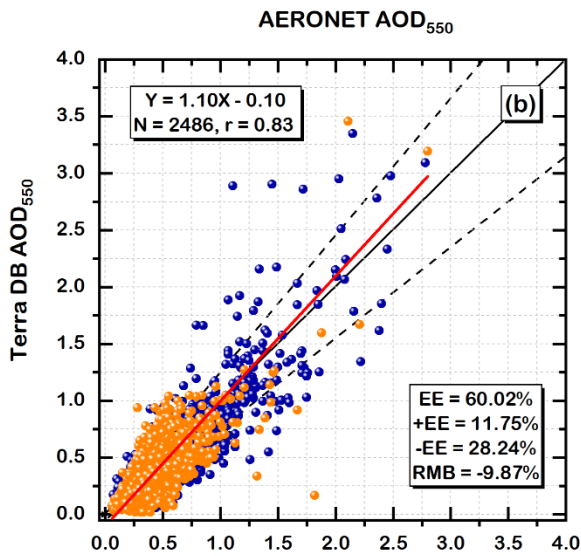
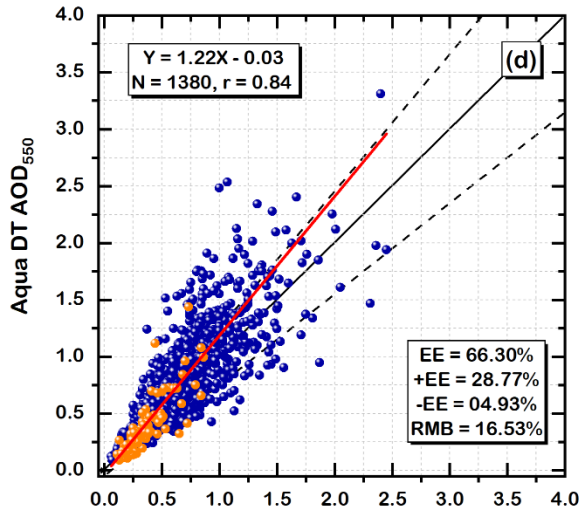
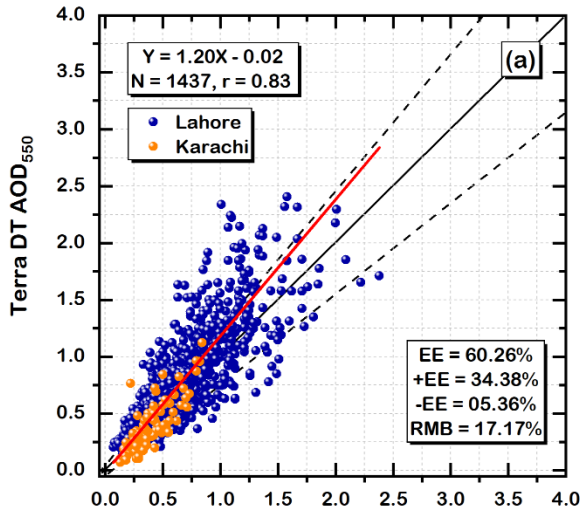
458 **5. Results and Discussion**

459 **5.1 Aqua and Terra MODIS AOD data**

460 **5.1.1 Validation of AOD products against AERONET**

461 The MODIS AOD data used in this paper were evaluated by comparison with the AERONET
 462 AOD data over Lahore and Karachi. The scatterplots in Figure 2 show that Terra DT (Figure 2a),

463 DB (Figure 2b), and DTB (Figure 2c) retrieved AOD are equally correlated ($r = 0.83$) with AERONET-
464 derived AOD, and have the same percentage of retrievals within the EE. However, the number of
465 collocated observations for DTB ($N = 2796$) is significantly higher than for DT ($N = 1437$) and DB
466 ($N = 2486$) i.e., 94.6% and 12.5% more data are available from DTB than from DT and DB,
467 respectively. The AOD retrieved from DT is significantly overestimated (RMB = 17.17%), with
468 34.38% of the data are above the EE (+EE). DB underestimates the AOD (RMB = -9.87%) with
469 28.24% of the data below the EE (-EE). These uncertainties appear to be averaged out in the DTB
470 AOD product, as the overestimations and underestimations are fewer than for DT and DB,
471 individually. Furthermore, the RMB (-0.03%) is significantly improved, being 99.9% and 99.8%
472 lower than for DT and DB, respectively. These results indicate the better performance of the Terra
473 DTB AOD product as compared to DT and DB over Pakistan. Similar to Terra, the performance of
474 the Aqua DTB AOD product (Figure 2f) is much better than for DT (Figure 2d) and DB (Figure 2e)
475 products, with a significantly higher number of collocated AOD values and lower RMB. However,
476 Aqua performs equally as Terra in terms of correlation and the percentage of retrievals within
477 the EE. It is important to mention that a larger number of both DT and DB AOD retrieval products
478 was available for Lahore than for Karachi and also that DB provides a greater number of AOD
479 retrievals over Pakistan than DT. Based on the superior performance of the Aqua and Terra DTB
480 AOD retrievals, the merged AquaTerra DTB AOD product was generated for further analysis (see
481 Figure S1 in the supplementary data for the validation of AquaTerra DTB AOD retrievals).



483 **Figure 2:** Validation of Terra and Aqua DT, DB, and DTB AOD products versus AERONET Version
484 3 Level 2.0 AOD measured in Lahore (for location, see no. 1 in Fig. 1a) and Karachi (for location,
485 see no. 56 in Fig. 1a) from 2006 to 2017. The red line represents the regression line, the solid
486 black line represents the identity line, and the dashed black lines represent the upper and lower
487 EE envelopes. The orange points represent AOD pairs at Karachi, the blue dots at Lahore.

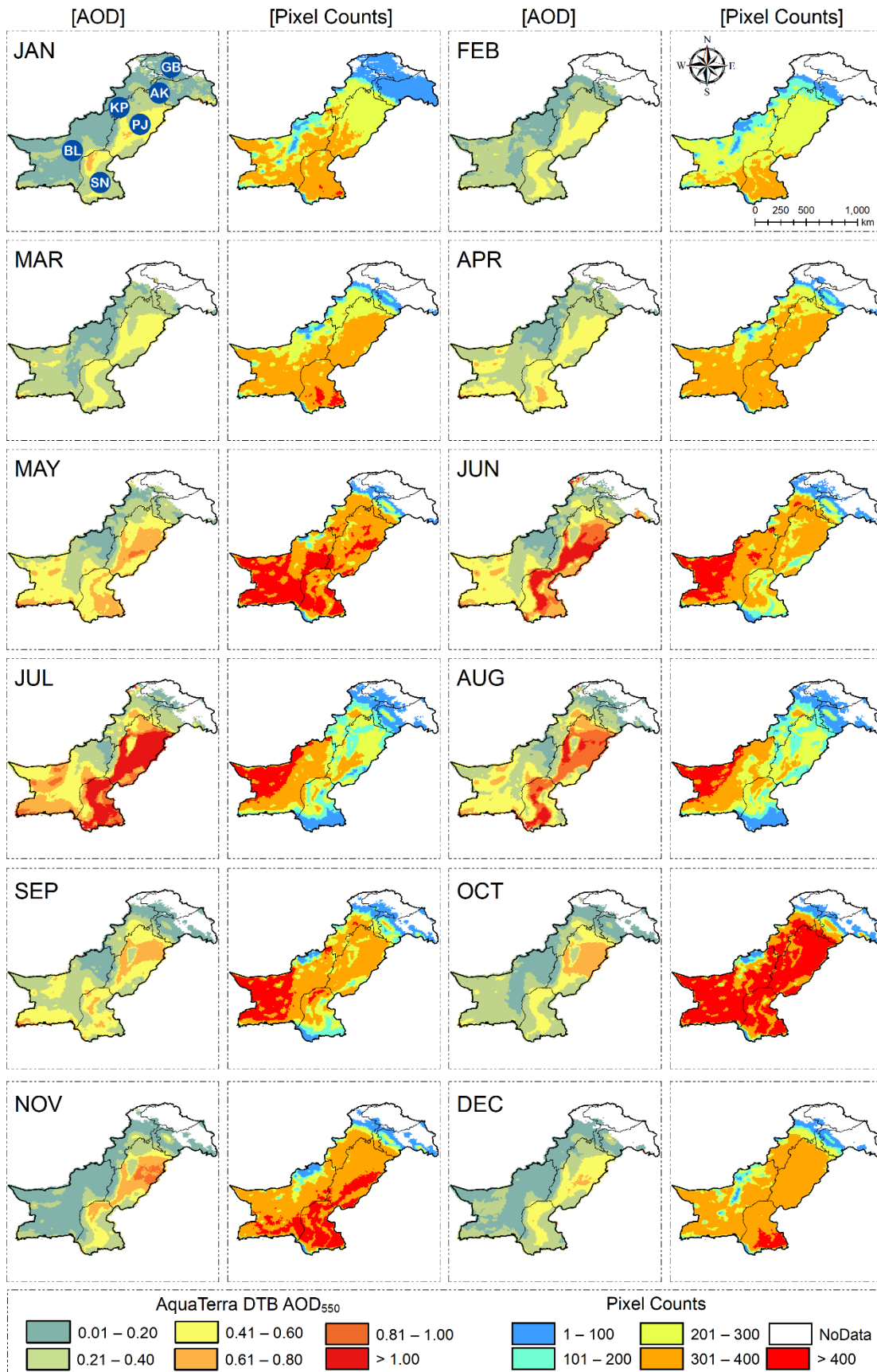
488 **5.1.2 Spatial distribution of AOD retrievals**

489 Figure 3 shows the spatial distributions of the monthly mean AquaTerra DTB AOD over
490 Pakistan together with the corresponding pixel counts (PC) averaged over the years 2003 - 2017.
491 Significant monthly variations in both AOD and PC are observed. AOD retrievals are missing over
492 the Gilgit-Baltistan and Jammu & Kashmir (disputed territory) throughout the year, except for
493 January, as the DT and DB algorithms do not provide AOD retrievals over high mountain regions
494 and snow-covered surfaces. The presence of AOD retrievals during January is because the DB
495 algorithm does not use the MODIS snow mask product directly, and the internal snow/cloud
496 mask does not work well over these regions. Surprisingly, high AOD values > 1.0 are observed
497 during June and July over the Northwestern region of Khyber Pakhtunkhwa, which is a high
498 mountainous region with permanent snow cover. These high AOD values over snow-covered
499 regions could be due to an error in the internal snow/cloud mask of the DB algorithm which has
500 missed these pixels during preprocessing; DT does discard bright pixels during preprocessing.
501 AOD > 1.0 is observed in July followed by June and August over Punjab and Sindh, mainly
502 attributed to hygroscopic growth of the aerosol particles during summer relative humidity is high,
503 similar to other reports using MODIS and MISR aerosol products (Mehta et al. 2016; Mhawish et

504 al. 2021). Most of the major cities of Punjab and Sindh are surrounded by cropland, and the
505 results show that high AOD over Pakistan follows the same spatial pattern as that of the cropland.
506 The AOD over cropland is significantly higher than over non-agricultural (i.e., mainly desert)
507 regions throughout the year, even during late spring and summer when dust storms are
508 considered a major source of aerosols over Punjab and Sindh. Local production of anthropogenic
509 aerosols from urban and industrial emissions and agricultural pre- and post-harvest burning may
510 be responsible for the high pollution levels over the region. Over Balochistan, especially over the
511 desert areas, the AOD is low compared to that in Punjab and Sindh, but still higher than over
512 other administrative units. Over Punjab, the highest AOD values are observed during the post-
513 harvest seasons, i.e., throughout September to November, peaking in November, probably due
514 to biomass (crop residue) burning activities (Jethva et al. 2019; Mhawish et al. 2021). However,
515 if the high AOD levels would only be due to locally produced aerosols, the spatial patterns during
516 each month should be similar, but they are not. Therefore, the transboundary transport of
517 aerosols may contribute to Pakistan's deteriorating air quality. This is confirmed by the well-
518 known smog episodes, occurring every year over Punjab due to both local production of aerosols
519 from crop residue burning and across the border, during which atmospheric visibility is reduced
520 to a few meters in both urban and rural areas. Overall, much higher AOD levels were observed in
521 Pakistan during June, July, and August (summer), followed by September, October, and
522 November (autumn), March, April, and May (spring), and December, January, and February
523 (winter). The higher AOD in the summer is attributed to several reasons, including (i) hygroscopic
524 growth of aerosol particles, due to high relative humidity, which increases the extinction
525 efficiency of the atmospheric aerosols (Dickerson et al. 1997; Li and Wang 2014), (ii) the

526 enhancement of secondary aerosol formation rate due to faster photochemical reactions during
527 higher temperatures (Jacob and Winner 2009; Kulmala et al. 2020), and (iii) the larger
528 contribution of natural aerosols (mainly dust) during the summer monsoon (Mhawish et al.
529 2021).

530 Figure 3 shows a distinct pattern of PC which suggests that the DT and DB algorithms do not
531 perform equally temporally or spatially. For example, between 2003 to 2017, from late spring to
532 early autumn, a large number of AOD retrievals (> 400) per pixel are available over Balochistan
533 and some parts of Punjab, and from late autumn to early spring, a large number of AOD retrievals
534 (> 400) per pixel are available over Sindh and some parts of Punjab. This could be attributed to
535 the seasonality in the surface albedo due to changes in vegetation cover and/or the presence of
536 cloud cover. Only October provides favorable conditions to both the DT and DB algorithms, when
537 more than 400 AOD retrievals are available over Pakistan from both algorithms, except for Gilgit-
538 Baltistan and disputed areas, due to high surface albedo for snow/ice surfaces.



540 **Figure 3:** Monthly mean spatial distributions of AquaTerra DTB AOD₅₅₀ and the total number of
541 corresponding Pixel Counts (PC) over Pakistan, both averaged over the years from 2003 to
542 2017. The six units in Pakistan are indicated in the upper-left figure: GB = Gilgit-Baltistan, AK=
543 Azad Kashmir, KP = Khyber Pakhtunkhwa, PJ = Punjab, BL = Balochistan, and SN = Sindh.

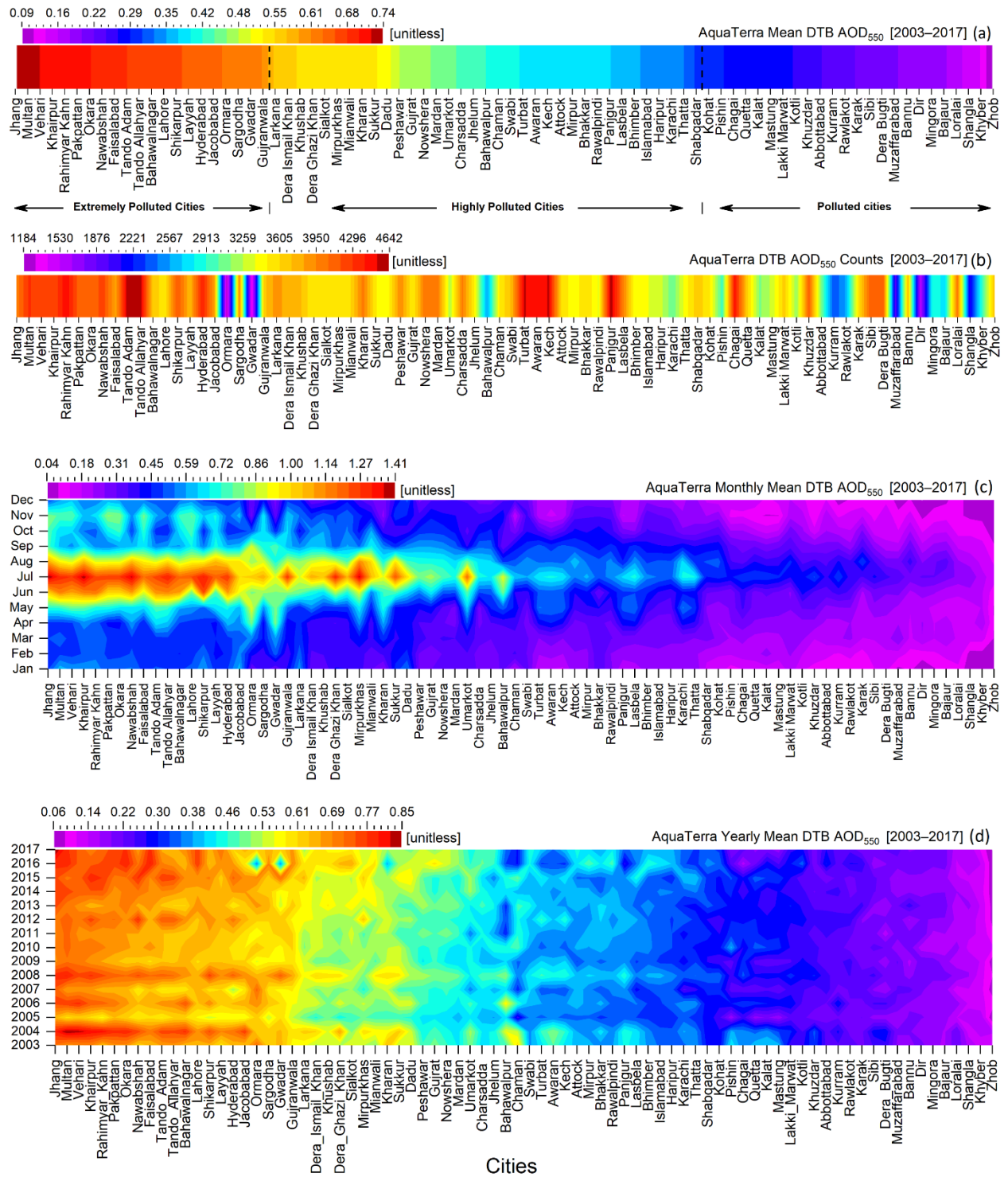
544 **5.1.3 Characterization of extremely polluted cities using MODIS data**

545 Figure 4a shows the mean AOD₅₅₀ retrievals for 80 cities (Figure 1) obtained from the annual
546 mean AquaTerra DTB AOD₅₅₀ images and categorizes the extremely polluted to polluted cities.
547 The thresholds for polluted and extremely polluted cities are defined based on the values of first
548 (Q1) and third (Q3) quartiles respectively, and these quartiles are calculated by analyzing
549 descriptive statistics (Table S1) for the AOD values extracted for 80 cities. Highly polluted cities
550 are defined based on the AOD range between the first and third quartiles. For example, AOD <
551 0.3 (1st quartile) represents polluted cities, $0.3 \leq \text{AOD} \leq 0.6$ (between 1st and 3rd quartiles)
552 represents highly polluted cities and AOD > 0.6 represents extremely polluted cities (3rd quartile).
553 A total of 21 cities fall within the category of extremely polluted cities (Punjab: 12, Sindh: 7, and
554 Balochistan: 2), 35 cities in the category of moderately polluted cities (Punjab: 11, Sindh 7,
555 Balochistan: 7, Khyber Pakhtunkhwa: 8, Azad Kashmir: 2), and 24 cities in the category of low
556 polluted cities (Punjab: 0, Sindh 0, Balochistan: 10, Khyber Pakhtunkhwa: 11, Azad Kashmir: 3).
557 The top 3 polluted cities are Jhang, Multan, and Vehari in Punjab, as Punjab is the most urbanized
558 and populated administrative unit (Figures 1b and 4a), with more vehicles and industries, and
559 also faces severe smog episodes and dust storms, resulting in extremely high AOD levels over the
560 region. Along with anthropogenic aerosols produced locally from cropland, urban and industrial

561 emissions, regional transport of aerosols may be responsible for Punjab's severe air pollution
562 problems which will be investigated using the PSCF analysis based on the HYSPLIT air parcel back
563 trajectory analysis and BAM PM_{2.5} concentrations (see section 5.7).

564 Figure 4b shows the pixel counts (PC) of the daily AOD retrievals for each city from 2003 to
565 2017. Results show a large number of PC for most cities, indicating that the characterization of
566 extremely polluted to polluted cities is based on a large number of PC, which supports the results
567 in Figure 4a and provides confidence in the use of merged AquaTerra DTB AOD products for
568 quantitative research applications over Pakistan. However, it is noted that the lowest number of
569 PC is observed for the coastal (Ormara and Gwadar) and mountainous (Dir) cities, where the
570 inversion scheme of both the DT and DB algorithms needs to be improved.

571 The monthly mean AOD retrievals are plotted to identify the high and low polluted months
572 in Pakistan (Figure 4c). The months of June, July, and August are by far the most polluted, with
573 AOD > 1.20 for extremely polluted cities. A similar pattern of monthly variation in AOD is
574 observed for all other cities, though at lower pollution levels. As mentioned in section 5.1.2, these
575 months may be affected by aerosol pollutants from local sources such as agricultural land, urban
576 and industrial regions, and deserts. Figure 4d, showing inter-annual variations, indicates very high
577 AOD levels for extremely polluted cities throughout the last two decades, with annual mean AOD
578 > 0.60, and with the most polluted years being 2004, 2006, 2008, 2016, and 2017.



580 **Figure 4:** Characterization of extremely polluted to polluted cities in Pakistan using AquaTerra
581 DTB AOD₅₅₀ products from 2003 to 2017. (a) polluted cities based on mean AOD, (b) pixel
582 counts, (c) polluted months based on mean AOD, and (d) polluted years based on mean AOD.

583 **5.2 CAMS and MERRA-2 reanalysis data**

584 **5.2.1 Validation of PM_{2.5} reanalysis data**

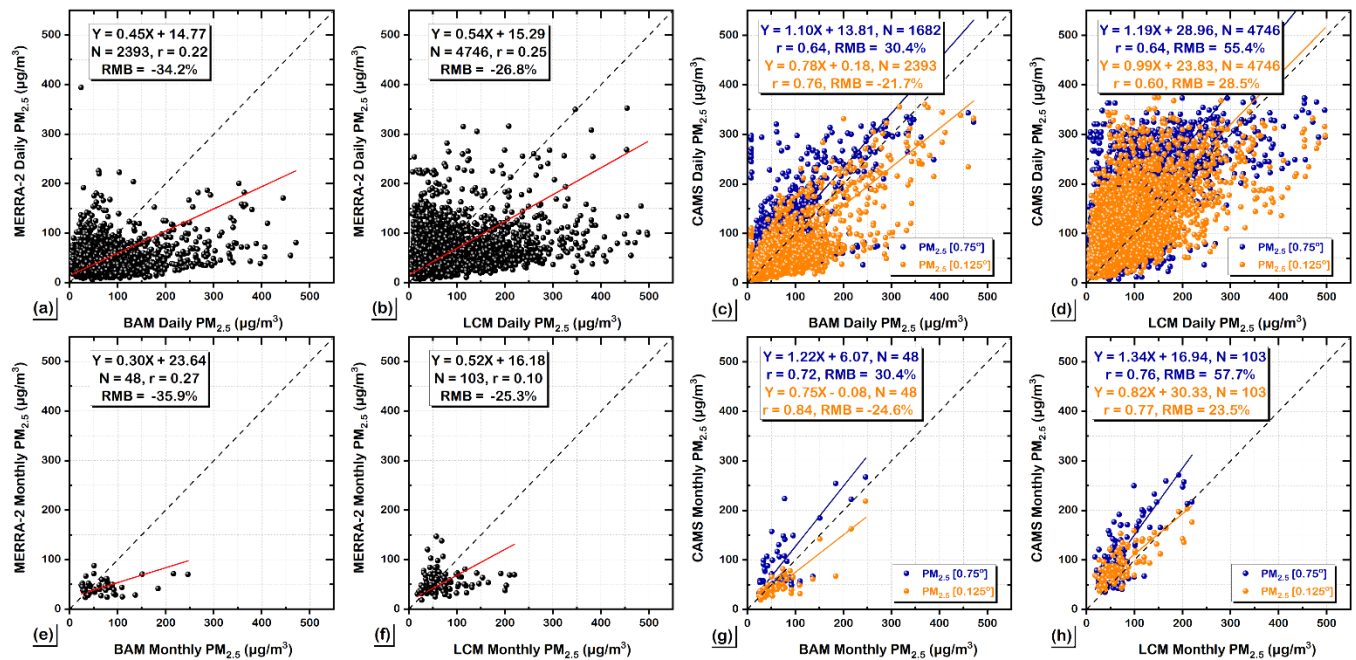
585 Previous studies have evaluated the uncertainties in both CAMS and MERRA-2 PM_{2.5}
586 reanalysis data compared to ground-based PM_{2.5} measurements (Cuevas et al. 2015; He et al.
587 2019; Song et al. 2018; Ukhov et al. 2020). Recently, Ukhov et al. (2020) reported overestimation
588 in CAMS PM_{2.5} over the middle east and west Asia which have been attributed to the deficient
589 size distribution of the emitted dust. Additionally, significant underestimation in MERRA-2 PM_{2.5}
590 was reported over China and India (He et al. 2019; Navinya et al. 2020; Song et al. 2018) which
591 could be due to the lack of nitrate concentrations in the reanalysis data and underestimation of
592 OC emission for urban/suburban areas (Buchard et al. 2016; Provencal et al. 2017).

593 The MERRA-2 and CAMS PM_{2.5} reanalysis data over Pakistan were evaluated by comparison
594 with BAM (beta gauge attenuation monitor) PM_{2.5} concentrations for 2019-2020 provided by the
595 US Consulates and with LCM (low-cost monitor) PM_{2.5} concentrations for 2018-2019 provided by
596 PAQI. The scatterplots in Figure 5 show a significant underestimation of both daily (Figures 5a
597 and 5b) and monthly (Figures 5e and 5f) MERRA-2 PM_{2.5} concentrations compared to both BAM
598 and LCM PM_{2.5} measurements: for the daily data the slopes are 0.45 and 0.54 and the RMB are -
599 34.2% and -26.8%, respectively, and for the monthly data the slopes are 0.30 and 0.52 with -

600 35.9% to 25.3%, respectively. The results also show the weak correlation of MERRA-2 PM_{2.5} data
601 with both BAM and LCM daily ($r = 0.22$ and 0.25 , respectively) and monthly ($r = 0.10$ and 0.27 ,
602 respectively) PM_{2.5} data. The weak correlation suggests that MERRA-2 PM_{2.5} data based on the
603 GOCART aerosol module is unable to accurately reproduce the temporal variations in PM_{2.5}. A
604 significant underestimation of MERRA-2 PM_{2.5} data was also reported over China (He et al. 2019;
605 Song et al. 2018) and India (Navinya et al. 2020), but over Pakistan, the correlation is even
606 weaker. Moreover, the grid size of MERRA2 ($0.5^\circ \times 0.625^\circ$ grid size) could introduce errors due
607 to heterogeneity within the large area that affects the correlation with the in-situ measurements.

608 In comparison with the MERRA-2 data, the correlation coefficients of the CAMS daily (Figures
609 5c and 5d) and monthly (Figures 5g and 5h) PM_{2.5} data versus ground-based in situ PM_{2.5}
610 measurements are substantially higher for both BAM and LCM. However, the data in Figure 5
611 show significant deviations of the CAMS-estimated PM_{2.5} from the ground-based PM_{2.5} values,
612 with over- or under-estimation depending on grid size. For example, CAMS overestimates PM_{2.5}
613 at the 0.75° grid size by 30.4% in comparison with the daily BAM data and by 55.4% in comparison
614 with the daily LCM data. For monthly data, these percentages are 30.4% and 57.4%. In contrast,
615 CAMS underestimates PM_{2.5} at the 0.125° grid size in comparison with BAM data and
616 overestimates in comparison with LCM data. These results suggest that grid size and ground-
617 based PM_{2.5} measurement methods (BAM and LCM) play an important role in the
618 overestimation/underestimation of CAMS PM_{2.5} data. For illustration, in comparison with the
619 BAM PM_{2.5} measurements, CAMS data are overestimated for one grid (0.75°) and
620 underestimated for another grid (0.125°), and CAMS PM_{2.5} data at the same grid size (0.125°) are
621 underestimated when compared with data measured using the BAM method and overestimated

622 when compared with data measured using the LCM method. It is worth mentioning that both
623 MERRA-2 and CAMS simulate 5 types of fine particulate matter components (dust, sea salt,
624 sulfate, organic carbon, and black carbon), but nitrate concentrations are not included. If the lack
625 of nitrate concentrations is the main reason for underestimation in MERRA PM_{2.5} data, as
626 reported by previous studies (Buchard et al. 2016; He et al. 2019; Provencal et al. 2017; Song et
627 al. 2018), then underestimation should be observed in CAMS PM_{2.5} data at 0.75° grid size, but
628 this is not the case. Therefore, the exact reasons for underestimation in both MERRA-2 and CAMS
629 as well as overestimation in CAMS data should be thoroughly investigated in future studies. The
630 results show a higher correlation for CAMS monthly data (Figures 5g and 5h) compared to the
631 daily data (Figures 5c and 5d). Although CAMS monthly data at 0.75° grid size show
632 overestimation, they have a good correlation coefficient ($r = 0.72\text{--}0.76$) with ground-based PM_{2.5}
633 measurements and could be useful for characterizing pollution levels in the cities of Pakistan
634 compared to the MERRA-2. The comparisons in Figure 5 do not provide a strong reason for
635 choosing one data set over the other. We have selected the CAMS data at the 0.75° grid taking
636 into account the deviation in the CAMS data observed in this evaluation, in addition to the large
637 scatter in individual data points which adds uncertainty.



638

639

Figure 5: Validation of MERRA-2 and CAMS PM_{2.5} reanalysis data against BAM (beta gauge

640

attenuation monitor) PM_{2.5} concentrations for 2019-2020 provided by the US Consulates and

641

LCM (low-cost monitor) PM_{2.5} concentrations for 2018-2019 provided by PAQI. Where, (a)

642

MERRA-2 daily PM_{2.5} vs. BAM daily PM_{2.5}, (b) MERRA-2 daily PM_{2.5} vs. LCM daily PM_{2.5}, (c) CAMS

643

daily PM_{2.5} vs. BAM daily PM_{2.5}, (d) CAMS daily PM_{2.5} vs. LCM daily PM_{2.5}, (e) MERRA-2 monthly

644

PM_{2.5} vs. BAM monthly PM_{2.5}, (f) MERRA-2 monthly PM_{2.5} vs. LCM monthly PM_{2.5}, (g) CAMS

645

monthly PM_{2.5} vs. BAM monthly PM_{2.5}, and (h) CAMS monthly PM_{2.5} vs. LCM monthly PM_{2.5}. The

646

dashed line in each figure is the identity line and the blue and orange solid lines are the fit lines

647

with parameters presented in the legends.

648

5.2.2 Characterization of extremely polluted cities using PM₁ and PM_{2.5} concentrations

649

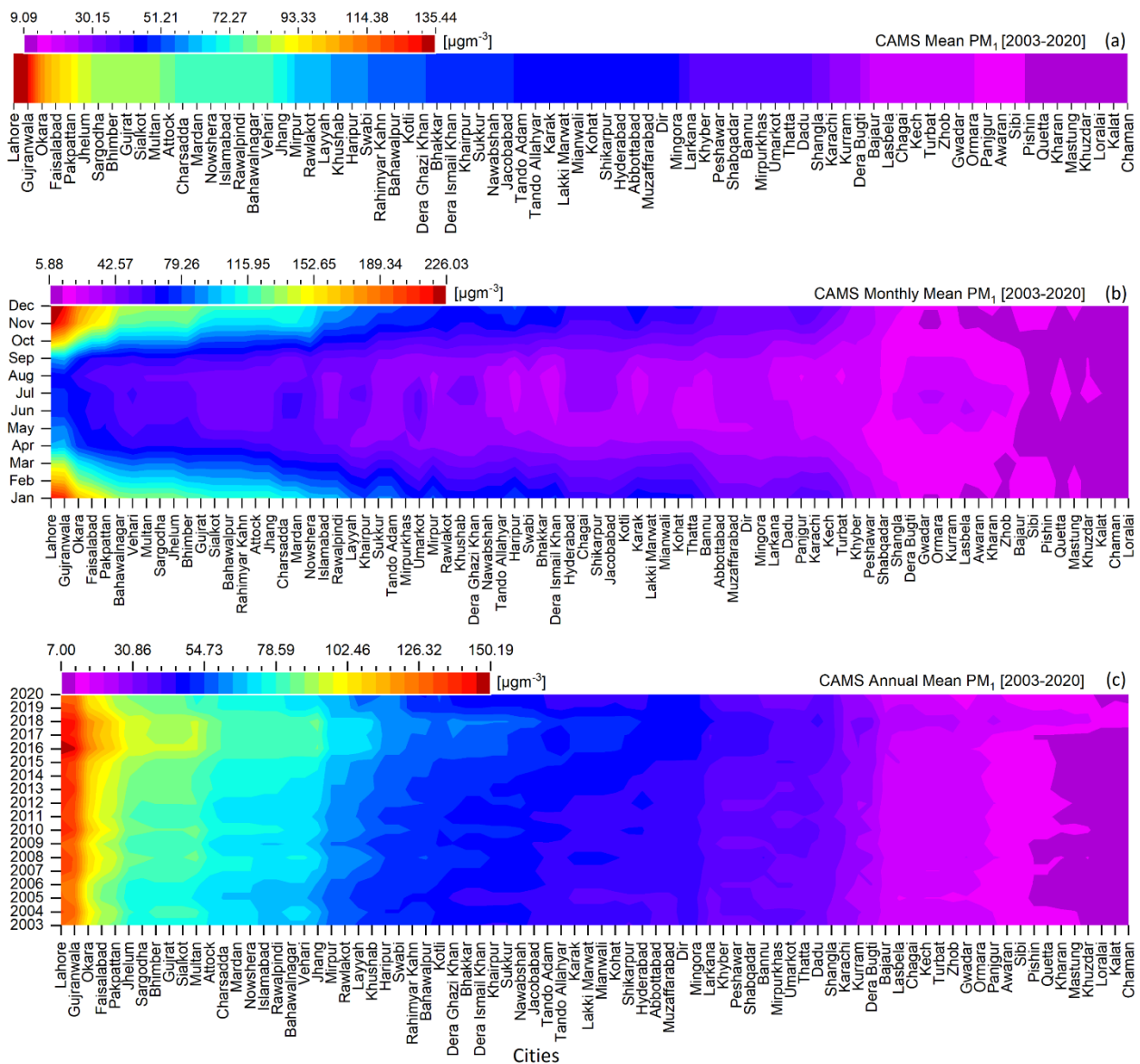
PM₁ and PM_{2.5} are fine particulate matter associated with human health issues. PM₁ is more

650

harmful than PM_{2.5} as it can reach deeper into the lungs and affect the respiratory system (Liu et

651 al. 2013; Meng et al. 2013). A previous study over China reported that most health issues
652 associated with PM_{2.5} were mainly due to greater contributions of PM₁ in PM_{2.5} (Chen et al. 2017).
653 The ranking of extremely polluted to polluted cities in Pakistan according to annual mean CAMS
654 PM₁ concentrations from 2003 to 2020 in Figure 6a indicates that the top 10 extremely polluted
655 cities are Lahore (135.44 µg/m³), Gujranwala (131.99 µg/m³), Okara (107.72 µg/m³), Faisalabad
656 (98.96 µg/m³), Pakpattan (94.06 µg/m³), Jhelum (85.51 µg/m³), Sargodha (84.30 µg/m³), Bhimber
657 (83.99 µg/m³), Gujrat (83.99 µg/m³), and Sialkot (83.99 µg/m³). Similarly, the top 10 extremely
658 polluted cities (Figure 7a) ranked according to PM_{2.5} concentrations are Lahore (170.53 µg/m³),
659 Gujranwala (163.63 µg/m³), Okara (139.43 µg/m³), Faisalabad (129.85 µg/m³), Pakpattan (126.97
660 µg/m³), Multan (113.09 µg/m³), Bahawalnagar (110.81 µg/m³), Vehari (110.81 µg/m³), Sargodha
661 (109.81 µg/m³), and Jhelum (107.68 µg/m³). The WHO air quality guidelines (AQG) are not yet
662 defined for PM₁ as PM₁ is not as widely monitored as PM_{2.5}, therefore the WHO recommended
663 AQG for PM_{2.5} (<10 µg/m³ annual mean) and Pak-NEQS for PM_{2.5} (<15 µg/m³ annual mean) are
664 used for comparison purposes. Not a single city in Pakistan falls within the PM_{2.5} standards
665 defined by Pak-NEQS and WHO, and the values of PM₁ and PM_{2.5} respectively for the top 10 cities
666 are 5.6 (8.4) to 9.0 (13.5) times and 7.2 (10.8) to 11.4 (17.1) times greater than the Pak-NEQS
667 (WHO AQG). For PM₁ and PM_{2.5}, 9 out of 10, and 10 out of 10 cities respectively, are in Punjab.
668 The extremely high pollution level may be due to emissions from local anthropogenic activities,
669 confirming the results of a previous modeling study that suggested local anthropogenic activities
670 as the major cause of high particulate concentrations in Pakistan (Shi et al. 2020). All major cities
671 selected in this study (80 cities) are exposed to PM_{2.5} concentrations during a long period of time
672 (Figures 1a and 7a), which exceed the Pak-NEQS (<15 µg/m³) and 68, 73, and 80, out of 80 cities

673 exceeded the WHO Interim Target-1 ($<35 \mu\text{g}/\text{m}^3$), Target-2 ($<25 \mu\text{g}/\text{m}^3$), and Target-3 (<15
 674 $\mu\text{g}/\text{m}^3$), respectively. These exceedances are set in strong perspective against the much lower
 675 recommended WHO AQG for $\text{PM}_{2.5}$ of $10 \mu\text{g}/\text{m}^3$. These results suggest that the top polluted cities
 676 are extremely hazardous for human health, as an increase of $\text{PM}_{2.5}$ by $10 \mu\text{g}/\text{m}^3$ can increase
 677 mortality, lung cancer, and cardiopulmonary diseases by 8%, 6%, and 4%, respectively, due to
 678 long-term exposure to fine particulates (Pope et al. 2002).



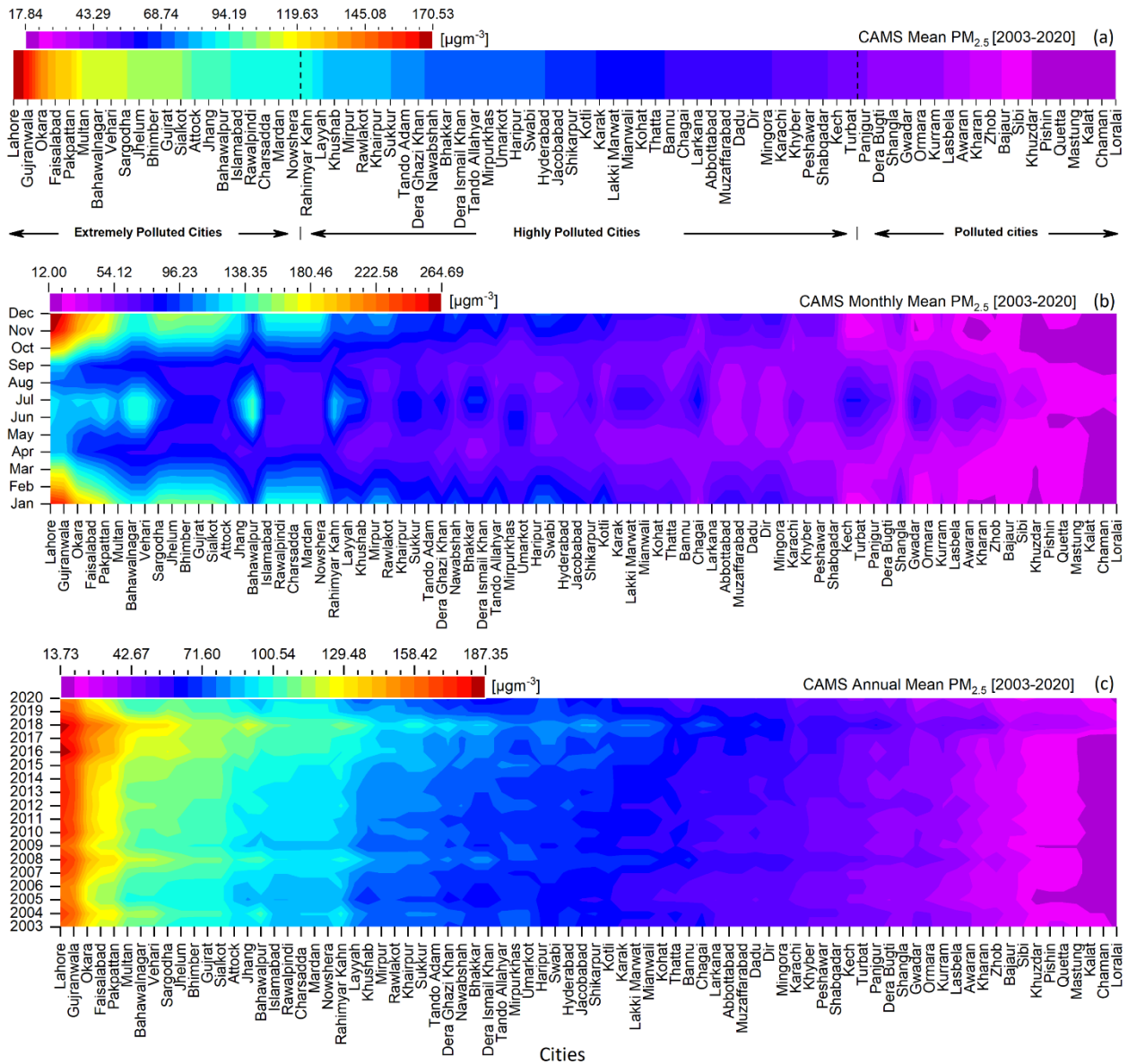
679

680 **Figure 6:** Ranking of extremely polluted to polluted cities in Pakistan according to annual mean
681 CAMS PM₁ concentrations from 2003 to 2020. Where (a) polluted cities based on yearly mean
682 PM₁ averaged over the years 2003-2020, (b) polluted months based on PM₁ averaged over the
683 years 2003-2020, and (c) polluted years based on yearly mean PM₁.

684 Figures 6b and 7b show months with the highest levels of PM₁ and PM_{2.5}, averaged over the
685 years 2003-2020, for the extremely polluted cities. The higher PM₁ and PM_{2.5} concentrations were
686 observed in cold months (October to February) with the maximum concentrations in December
687 and January, while warmer months (March to September) showed lower PM_x concentrations.
688 The high levels of fine particulates in October and November may be attributed to both cross-
689 border transport of aerosol produced from biomass burning activities (from India) as well as
690 locally produced aerosols by anthropogenic activities. As the highest values of fine particulates
691 were observed in December and January which are not the main months of biomass burning
692 activities, these are not likely the main source of the high levels of fine particulates pervasive
693 across these highly polluted cities. At this time of year, less surface heating and less turbulence
694 due to lower intensity of solar irradiation lead to stable and shallow boundary layers.
695 Furthermore, with higher concentrations of light-absorbing aerosols, mainly BC, the atmospheric
696 stability increases due to local heating near the top of the boundary layer, induced by BC, which
697 further lowers the boundary layer height (BLH) (Ding et al. 2016). Stable atmospheric conditions
698 that imply low BLH together with low wind speed, both limiting aerosol transport, lead to the
699 accumulation of aerosols and enhancement of particle concentrations near the surface. As a
700 result, anthropogenic aerosols such as those produced from fossil fuel combustion and other
701 urban and industrial activities may linger for long periods (Mhawish et al. 2020). In October and

702 November, both local and remote (cross-border) biomass (crop residue) burning activities
703 coupled with stable atmospheric conditions have been recognized to cause severe haze and smog
704 episodes, especially over Punjab (Mhawish et al. 2020; Tariq et al. 2015; Tariq et al. 2016). The
705 formation of secondary inorganic aerosol during haze episodes is also responsible for higher
706 $PM_{2.5}$ concentrations as reported from recent studies over China (Nichol et al. 2020; Zhang et al.
707 2018). An increase in $PM_{2.5}$ concentrations was observed in June and July, and PM_1
708 concentrations slightly increased in July. This means that $PM_{2.5}$ exhibited two peaks: the first in
709 winter and the second in summer, whereas a single peak in winter was observed for PM_1 . The
710 second $PM_{2.5}$ peak in summer may be attributed to the fine particulates from dust, as dust storm
711 activities are very common in Pakistan during summer, as well as local anthropogenic activities.
712 The lower peak of $PM_{2.5}$ in the summer, compared to winter, may be due to the unstable
713 atmospheric conditions due to the higher surface heating by solar irradiation, leading to the
714 generation of strong turbulence with rising air and thus strong mixing conditions which promote
715 the vertical dispersion of pollutants.

716 The annual mean concentrations of PM_1 (Figure 6c) and $PM_{2.5}$ (Figure 7c) show strong inter-
717 annual variations with distinct PM_x levels and very poor air quality conditions throughout the last
718 two decades. The annual mean mass concentrations in extremely polluted cities range from 63
719 $\mu\text{g}/\text{m}^3$ to 150.19 $\mu\text{g}/\text{m}^3$ for PM_1 and from 85 $\mu\text{g}/\text{m}^3$ to 187.35 $\mu\text{g}/\text{m}^3$ for $PM_{2.5}$, which are 4.2
720 (6.3)–10 (15) and 5.7 (8.5)–12.5 (18.7) times greater than the Pak-NEQS (WHO AQG),
721 respectively.



722

723

Figure 7: As Figure 6, but for PM_{2.5}.

724

5.2.3 Characterization of extremely polluted cities using PM₁₀ concentrations

725

Figure 8a shows the ranking of polluted cities according to PM₁₀ concentrations. The PM₁₀

726

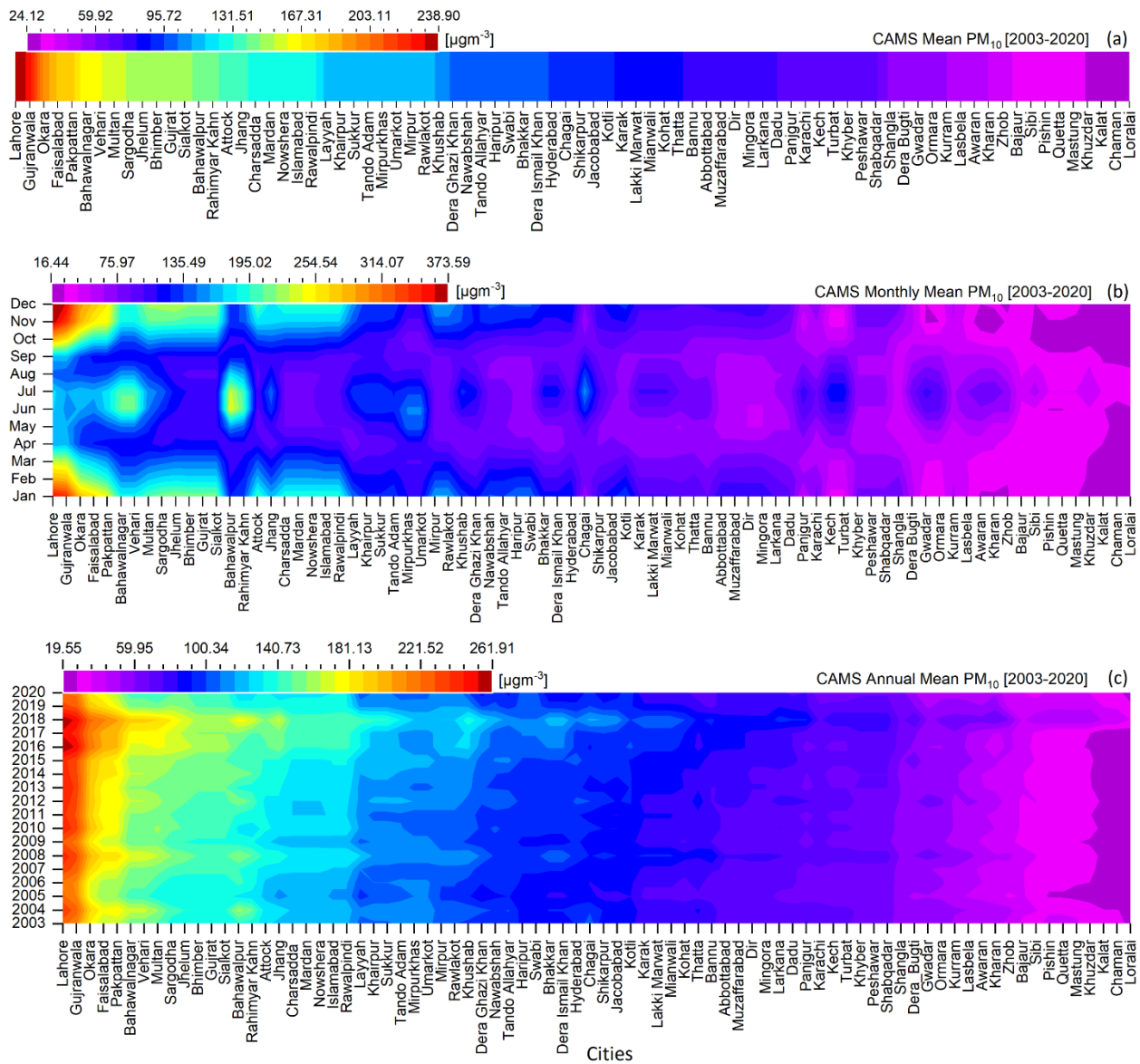
fraction with an aerodynamic diameter larger than PM_{2.5} (PM₁₀-PM_{2.5}), i.e. the mass

727

concentration of coarse particles, mainly originates from natural sources such as desert dust and

728 resuspended soil particles. The top 10 most polluted cities according to the PM₁₀ concentrations
729 are Lahore (238.9 µg/m³), Gujranwala (229.1 µg/m³), Okara (194.5 µg/m³), Faisalabad (180.6
730 µg/m³), Pakpattan (177.9 µg/m³), Bahawalnagar (160.6 µg/m³), Vehari (160.6 µg/m³), Multan
731 (157.5 µg/m³), Sargodha (152.3 µg/m³), and Jhelum (149.7 µg/m³). PM₁₀ concentrations are 1.2
732 to 11.9 times higher than the WHO AQG for PM₁₀ (20 µg/m³ annual mean) for all the cities shown
733 in Figure 8a, suggesting that very poor air quality conditions, hazardous for human life, prevail in
734 all Pakistani cities. Overall, the PM₁₀ temporal trend pattern is very similar to that for PM_{2.5}, i.e.,
735 December is the month with the highest PM₁₀ concentrations, followed by January. In summer,
736 July is the most polluted month followed by June (Figure 8b). Similar to the PM_{2.5} variations, PM₁₀
737 also exhibited peaks in both winter and summer. The higher concentrations during the winter
738 months (i.e. December and January) may be due to increased anthropogenic emission activities
739 along with stable atmospheric conditions (stagnant conditions, and shallower boundary layer).
740 Despite the abundance of coarse particulate matter in spring and summer seasons which are
741 transported from the arid and semiarid regions, the strong convection combined with a deeper
742 boundary layer enhances the dispersion of the near-surface pollutant that decreases the PM₁₀
743 concentrations along with the wet deposition during the rainy summer season. The pre-harvest,
744 harvesting, and post-harvest burning activities along with meteorological conditions such as low
745 wind speed and low boundary layer height may contribute to higher surface PM₁₀ levels
746 especially during October and November as these activities produce both fine (PM₁ and PM_{2.5})
747 and coarse (PM₁₀) particles as reported by (Jain et al. 2020; Singh et al. 2017) over South Asia and
748 by Le Blond et al. (2017) over South American countries.

749 Similar to the annual mean PM_{2.5} variations (Figure 7c), the annual mean PM₁₀ concentrations
 750 also show distinct interannual variations for all cities (Figure 8c), and severe air pollution levels
 751 were observed throughout the last two decades. According to these findings, Pakistani people
 752 are not only exposed to long-term PM_{2.5} but also to PM₁₀ concentrations exceeding the WHO
 753 recommended AQG for PM₁₀ (<20 µg/m³). Overall, these results suggested that Pakistani cities
 754 are a severe threat to human life due to extremely poor air quality conditions.



755

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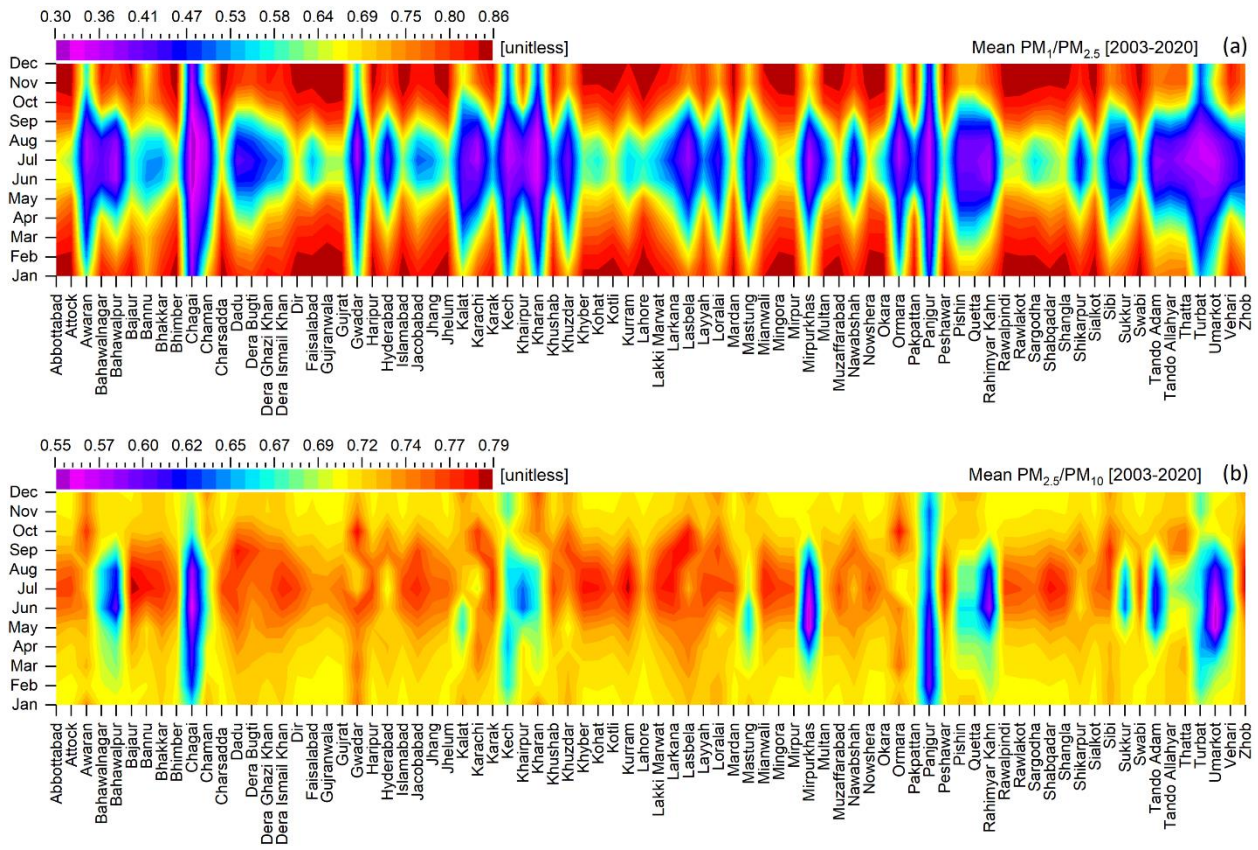
Figure 8: As Figure 6, but for PM_{10} .

757 **5.2.4 $PM_1/PM_{2.5}$ and $PM_{2.5}/PM_{10}$ ratios**

758 The PM_x ratios are very useful for understanding the contributions among particulate size, as
759 revealed by a study in China where PM_1 contributed nearly 80% of $PM_{2.5}$ (Wang et al. 2015),
760 which would have consequences for human health. Over Pakistan, the $PM_1/PM_{2.5}$ (Figure 9a) and
761 $PM_{2.5}/PM_{10}$ (Figure 9b) ratios are lower than those observed over China (Wang et al. 2015),
762 indicating lower contributions of PM_1 to $PM_{2.5}$ and $PM_{2.5}$ to PM_{10} . However, the pattern of ratios
763 is similar to that observed for China, i.e., the $PM_1/PM_{2.5}$ ratios are higher than $PM_{2.5}/PM_{10}$ ratios.
764 Relatively higher $PM_1/PM_{2.5}$ ratios (>75%) are observed from October to March (Figure 9a),
765 indicating a larger fraction of PM_1 in $PM_{2.5}$ due to more anthropogenic activities. The directly
766 emitted PM_1 from the automobile and combustion of fossil fuel, and indirectly by formation from
767 precursor gases, are most likely higher from October to March, leading to the enhanced
768 $PM_1/PM_{2.5}$ ratio. This also suggests that the $PM_{2.5}$ concentrations from October to March are
769 driven by emissions from combustion and secondary aerosols formation (Jain et al. 2020).
770 However, low $PM_1/PM_{2.5}$ ratios are observed from April to September in most of the cities, and
771 low ratios during all months are observed in the cities located in Balochistan, indicating a lower
772 contribution of PM_1 to $PM_{2.5}$, which is mainly dominated by the larger particles especially during
773 summer (June, July, and August) which not contributed to PM_1 .

774 Figure 9b shows large contributions of $PM_{2.5}$ to PM_{10} throughout the year with maximum
775 contributions during summer as indicated by the large $PM_{2.5}/PM_{10}$ ratios. This suggests that the
776 air quality in these cities is mainly (and significantly) influenced by fine particulates, largely from

777 anthropogenic sources. The large $PM_{2.5}/PM_{10}$ ratios in Gwadar and Ormara (Figure 9b), coastal
 778 cities in Balochistan, throughout the year suggest that also in these coastal cities the PM is
 779 dominated by $PM_{2.5}$ particles, which indicates that the PM_{10} is driven by $PM_{2.5}$ which is highly
 780 influenced by anthropogenic sources. Gwadar has the deepest seaport in the world and the ship-
 781 based emissions may be one of the sources of fine anthropogenic particles throughout the year.
 782 However, lower $PM_{2.5}/PM_{10}$ ratios are observed for other cities located in Balochistan, indicating
 783 the greater influence of coarse particulates (mainly desert dust).



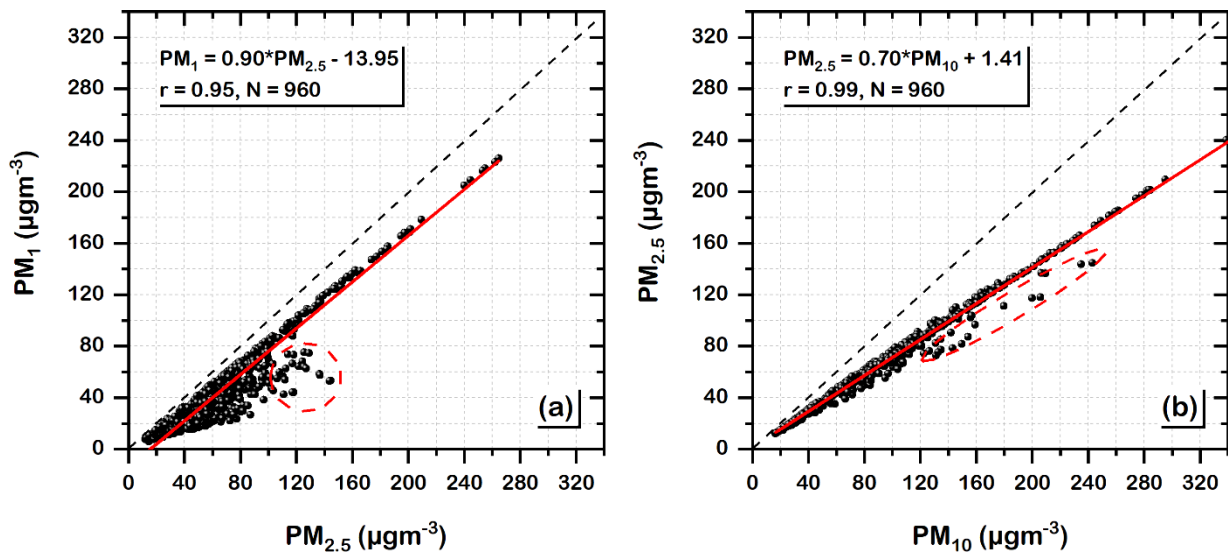
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Figure 9: (a) Monthly mean ratios of $PM_1/PM_{2.5}$ and (b) $PM_{2.5}/PM_{10}$.

786

787 Scatter plots of PM_1 vs. $PM_{2.5}$ (Figure 10a) and $PM_{2.5}$ vs. PM_{10} (Figure 10b) show that the PM_x
788 fractions over Pakistan are well-correlated, with Pearson's correlation coefficients (r) of 0.95 and
789 0.99, and slopes of 0.90 and 0.70, respectively. The strong linear relationship between PM_{10} and
790 $PM_{2.5}$ (higher r values) suggests common sources of fine and coarse particulates compared to
791 PM_1 vs $PM_{2.5}$ relationship. While the higher slope values suggest larger contributions of PM_1 to
792 $PM_{2.5}$ than $PM_{2.5}$ to PM_{10} . Overall, both the contribution of PM_1 to $PM_{2.5}$ and that of $PM_{2.5}$ to PM_{10}
793 are smaller over Pakistan than over China (Wang et al. 2015) as indicated by the PM_x ratios (Figure
794 9) and slope values (Figure 10). This might be due to a higher contribution of anthropogenic
795 emissions to the PM concentrations in China than in Pakistan; however, other processes may also
796 contribute, and unraveling the different contributions requires more detailed research. Figures
797 10a and 10b show some scattered points, within a red circle or ellipse, which represent the data
798 from May to September and these scattered points suggest lower contributions of PM_1 in $PM_{2.5}$
799 and $PM_{2.5}$ in PM_{10} , as also indicated by low PM_x ratios (Figure 9).



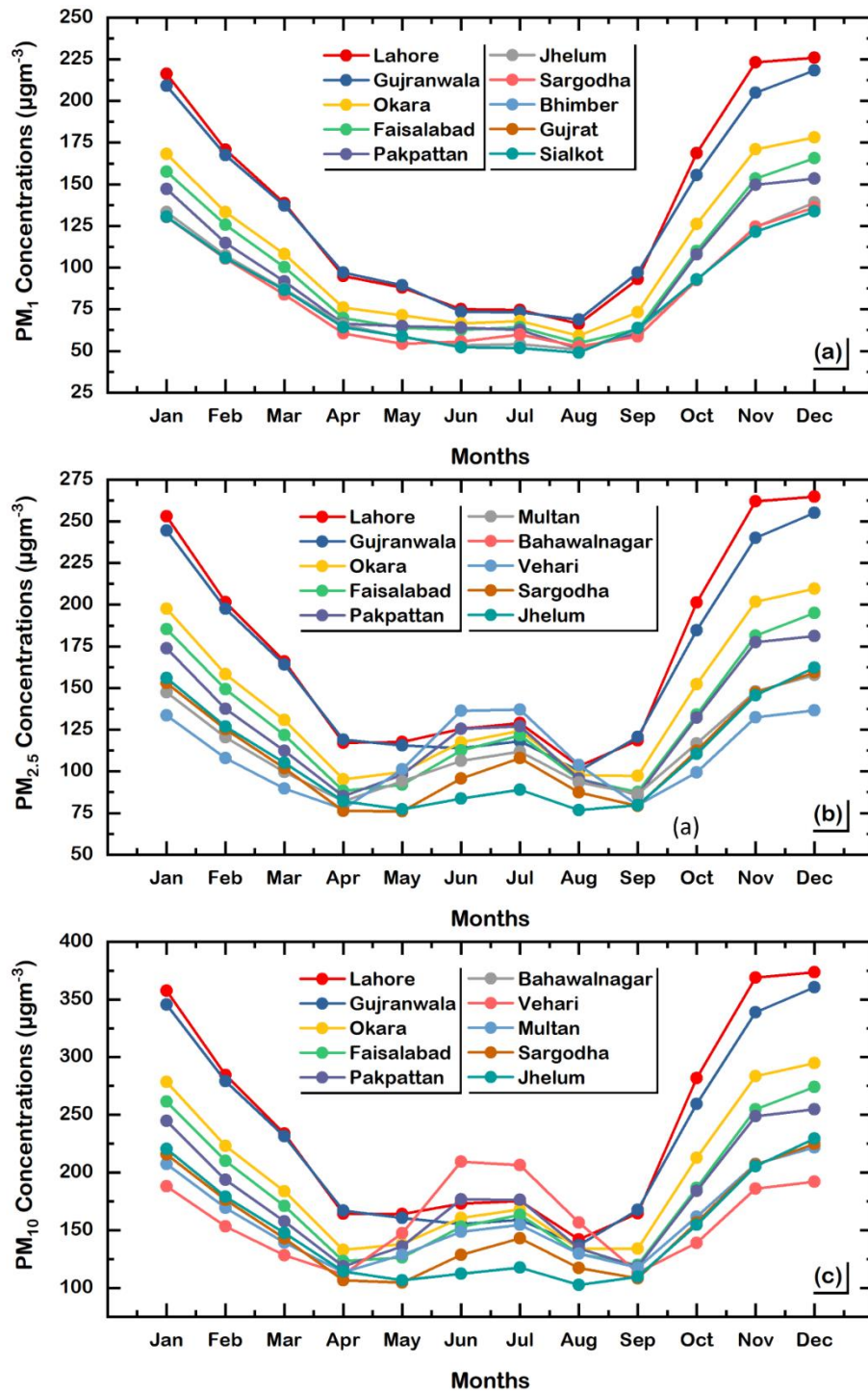
800

801 Figure 10: Scatter plots between (a) PM_1 vs. $PM_{2.5}$ and (b) $PM_{2.5}$ vs. PM_{10} . The red solid line
802 represents the regression line and the black dashed line represents the identity line. The data
803 points in the red circle and ellipse are explained in the text.

804 **5.2.5 Monthly mean temporal trend of PM_1 , $PM_{2.5}$, and PM_{10}**

805 The month-to-month variations of the multi-year (2003–2020) monthly mean PM_1 , $PM_{2.5}$, and
806 PM_{10} concentrations for the top 10 polluted cities are shown in Figure 11. These cities vary
807 according to population growth, the number of automobiles, urbanization, industrialization, city
808 size, land cover types, and climatic conditions, and PM concentrations are expected to behave
809 differently due to these factors. This study follows the hypothesis of our previous study
810 conducted over Hong Kong (Bilal et al. 2019c) i.e., if the PM concentrations have different
811 magnitudes but follow the same temporal pattern at different locations, they are influenced by
812 local as well as regional contributions. Thus for PM_1 concentrations, Figure 11a shows the same
813 pattern for each of the 10 cities, suggesting that both local and regional sources contribute to
814 PM_1 concentrations. For both $PM_{2.5}$ (Figure 11b) and PM_{10} (Figure 11c), similar patterns are only
815 evident from September to April, and dissimilar patterns due to variation in magnitudes are
816 evident from May to August, suggesting more local contributions for the summer months of May
817 to August. This local contribution during summer may be attributed to the frequent dust/sand
818 storms. Similarly, from October to January, the PM_1 , $PM_{2.5}$, and PM_{10} concentrations in Lahore
819 and Gujranwala show similar patterns as in other cities, but with higher concentrations, probably
820 because Lahore and Gujranwala are the largest cities, with consequently more transport, fossil
821 fuel, and industrial emissions, and some local and cross-border biomass burning activities in

822 autumn (Ali et al. 2013; Tariq et al. 2015; Tariq et al. 2016), which reinforce the effects of
823 meteorological impacts, such as shallower boundary layer height and lower wind speed, which
824 result in the accumulation of particulate matter near the surface (Miao et al. 2019; Miao and Liu
825 2019; Miao et al. 2018; Qu et al. 2017; Sun et al. 2019; Wang et al. 2018).



826

827

Figure 11: Multiyear (2003 - 2020) monthly average variations of PM₁, PM_{2.5}, and PM₁₀

828

concentrations in the corresponding top 10 polluted cities (see legend). Cities are plotted with

829

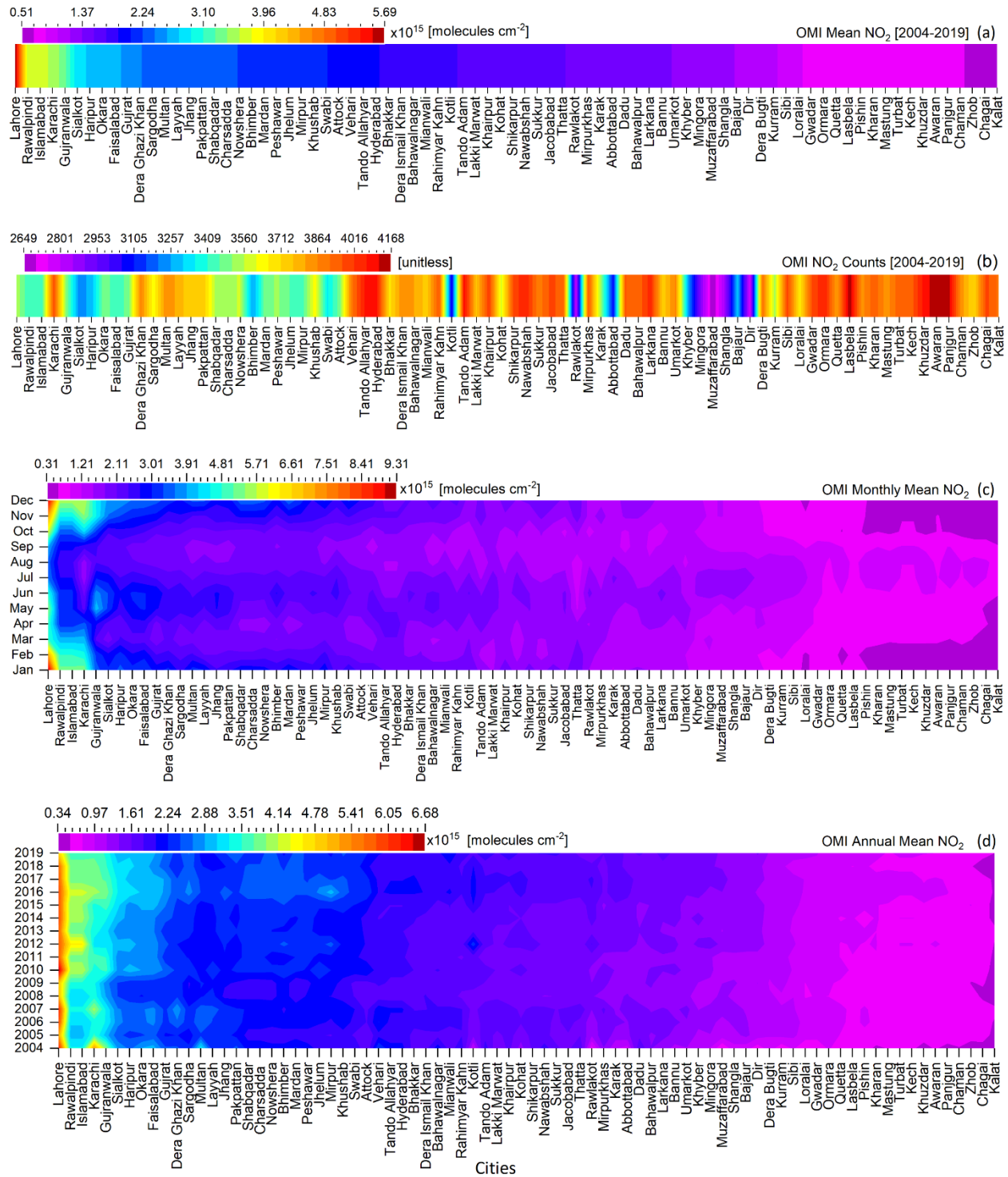
the rank of high to low polluted.

830 5.3 OMI vertical column densities of NO₂ and SO₂

831 5.3.1 Characterization of extremely polluted cities using NO₂ data

832 NO₂ is mainly produced from fossil fuel combustion, industrial emission, automobile
833 emission, biomass burning, natural lightning, and soil microbe emissions (Cheng et al. 2012; Lee
834 et al. 1997; Olivier et al. 1998; Richter and Burrows 2002). NO₂ has an adverse effect on health
835 and contributes to low atmospheric visibility, and poor air quality conditions (Khokhar et al. 2015;
836 ul-Haq et al. 2014). Pakistan's top ten polluted cities according to NO₂, where we use
837 Tropospheric vertical column densities (TVCDs) as a proxy, are those with the highest levels of
838 urbanization, vehicle emissions, and industrialization, suggesting anthropogenic activities to be
839 the major cause. They are Lahore (5.69×10^{15} molecules/cm²), Rawalpindi (3.65×10^{15}
840 molecules/cm²), Islamabad (3.65×10^{15} molecules/cm²), Karachi (3.60×10^{15} molecules/cm²),
841 Gujranwala (3.32×10^{15} molecules/cm²), Sialkot (2.81×10^{15} molecules/cm²), Haripur (2.73×10^{15}
842 molecules/cm²), Okara (2.72×10^{15} molecules/cm²), Faisalabad (2.72×10^{15} molecules/cm²), and
843 Gujrat (2.47×10^{15} molecules/cm²) (Figure 12a). Similar results are reported by Tabinda et al.
844 (2019), Ashraf et al. (2013), and Khanum et al. (2017). In terms of data availability from OMI,
845 Figure 12b indicates the largest number of PC available for Lasbela (4168), Awaran (4154), and
846 Panjgur (4140), all located in Balochistan. On a monthly mean basis, NO₂ (Figure 12c) follows the
847 same patterns as observed for PM₁ and PM_{2.5} concentrations; i.e., higher values in winter,
848 especially for the extremely polluted cities (Lahore, Rawalpindi, Islamabad, and Karachi), which
849 are attributed to emissions of automobiles, industries, and fossil fuel combustion, under stable
850 atmospheric conditions. The NO₂ atmospheric lifetime is higher in winter than in summer due to

851 higher mixing ratio and less sunlight that initiates the breakdown reaction of NO₂; therefore stays
852 longer in the atmosphere in winter than in summer. A different trend observed for cities located
853 in Balochistan, with higher NO₂ in summer, could be due to natural lightning as reported by
854 Khokhar et al. (2015). Figure 12d shows that Lahore, Rawalpindi, Islamabad, and Karachi are
855 polluted in all years from 2004 to 2019, subjecting citizens to long-term exposure associated with
856 respiratory diseases, otitis media, and mortality (Latza et al. 2009).

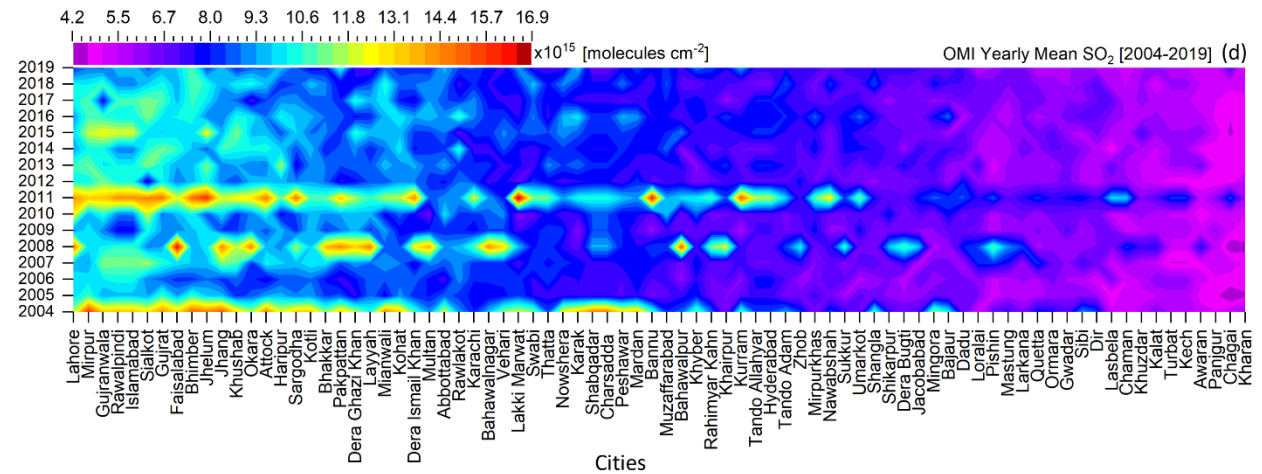
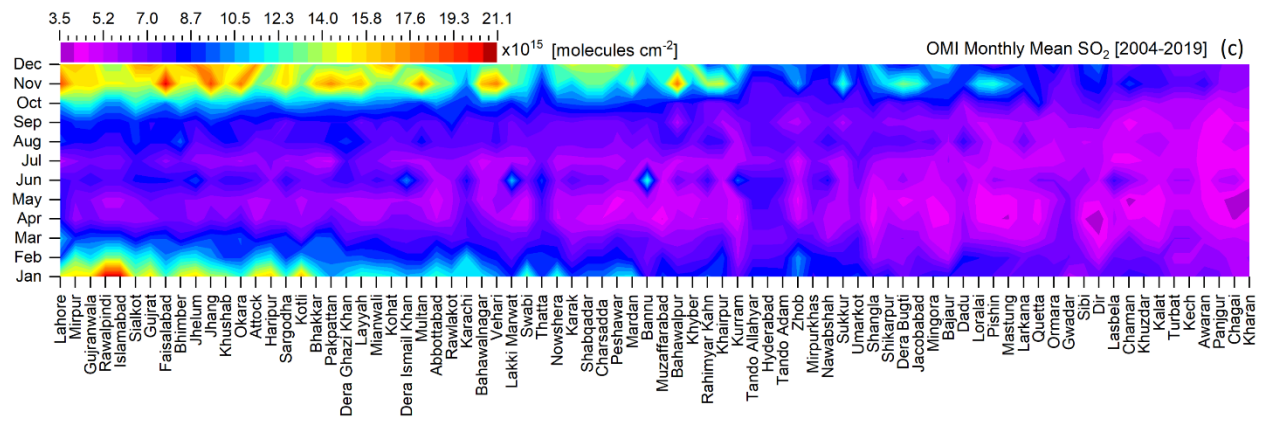
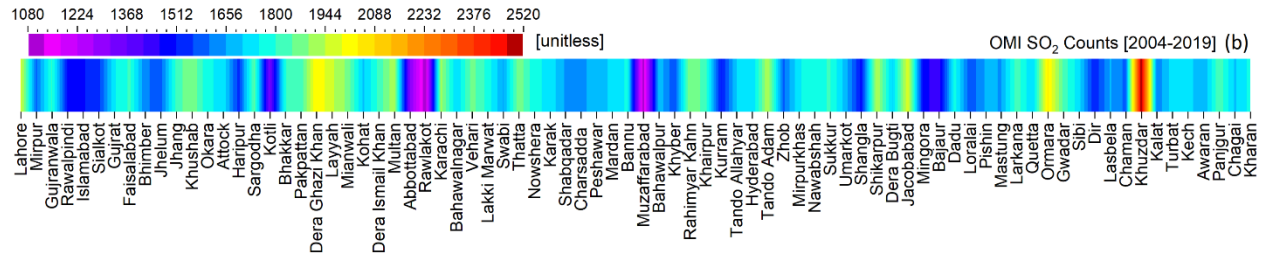
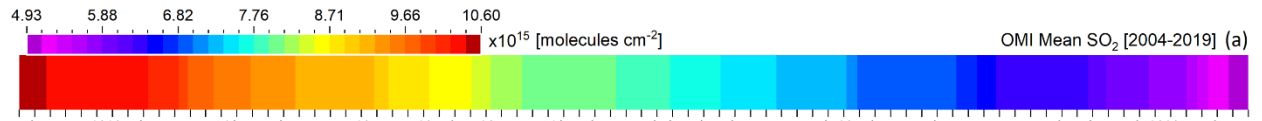


858 **Figure 12:** Ranking of extremely polluted to polluted cities in Pakistan according to OMI NO₂
859 TVCDs (molecules/cm²) from 2004 to 2019. (a) polluted cities based on mean NO₂, (b) pixel
860 counts, (c) polluted months based on mean NO₂, and (d) polluted years based on mean NO₂.

861 **5.3.2 Characterization of extremely polluted cities using SO₂ data**

862 Power plants, oil and gas refineries, and metal smelters are the major sources of
863 anthropogenic SO₂ (Dahiya and Myllyvirta 2019). In Figure 13a, extremely polluted to polluted
864 cities are ranked based on OMI-derived SO₂ vertical column density and the top 10 polluted cities
865 are Lahore (10.6×10¹⁵ molecules/cm²), Mirpur (10.5×10¹⁵ molecules/cm²), Gujranwala (10.3×10¹⁵
866 molecules/cm²), Rawalpindi (10.3×10¹⁵ molecules/cm²), Islamabad (10.3×10¹⁵ molecules/cm²),
867 Sialkot (10.3×10¹⁵ molecules/cm²), Gujrat (10.3×10¹⁵ molecules/cm²), Faisalabad (10.3×10¹⁵
868 molecules/cm²), Bhimber (10.2×10¹⁵ molecules/cm²), and Jhelum (10.2×10¹⁵ molecules/cm²).
869 According to the global SO₂ emission hotspot database (Dahiya and Myllyvirta 2019), five oil
870 power plants near Lahore are the main sources of high SO₂ emissions over Lahore. The lower
871 number (1080–2520) of successful SO₂ retrievals (Figure 13b) as compared to NO₂ retrievals
872 (Figure 12b) is attributed to the high noise level in the OMI-retrieved SO₂ data. Only the relatively
873 strong SO₂ signal over point sources (e.g., power plants, metal smelters) can be detected.
874 (Fioletov et al. 2011; Li et al. 2017; Li et al. 2020). The temporal variation of the monthly mean
875 SO₂ VCDs (Figure 13C) have a pattern similar to that of PM_{2.5} and NO₂ TVCD, with high values in
876 the winter and low in the summer. For the top polluted cities, the high SO₂ observed during
877 November, December, and January may be attributed to the power plants and brick kilns (Dahiya
878 and Myllyvirta 2019; Rahman et al. 2000). Brick kilns are considered as major sources of SO₂

879 resulting in extremely poor air quality. This is clearly observed over Punjab (Adrees et al. 2016;
880 Colbeck et al. 2010; Pervaiz et al. 2021; Ur Rehman et al. 2019). Therefore, every year during late
881 autumn and winter, the government of Pakistan bans these kilns to control pollution levels. The
882 SO₂ accumulates in the BL during the stable atmospheric conditions and shallow BLH at this time
883 of year, in response to the low solar irradiation resulting in little surface heating and turbulence
884 mixing. Unlike NO₂, the contribution of SO₂ to poor air quality in Pakistani cities varies from year
885 to year, as shown in Figure 13d. The SO₂ VCD is higher in 2004, 2008, and 2011 than in other
886 years. The investigation of the year-to-year variability requires a separate study.



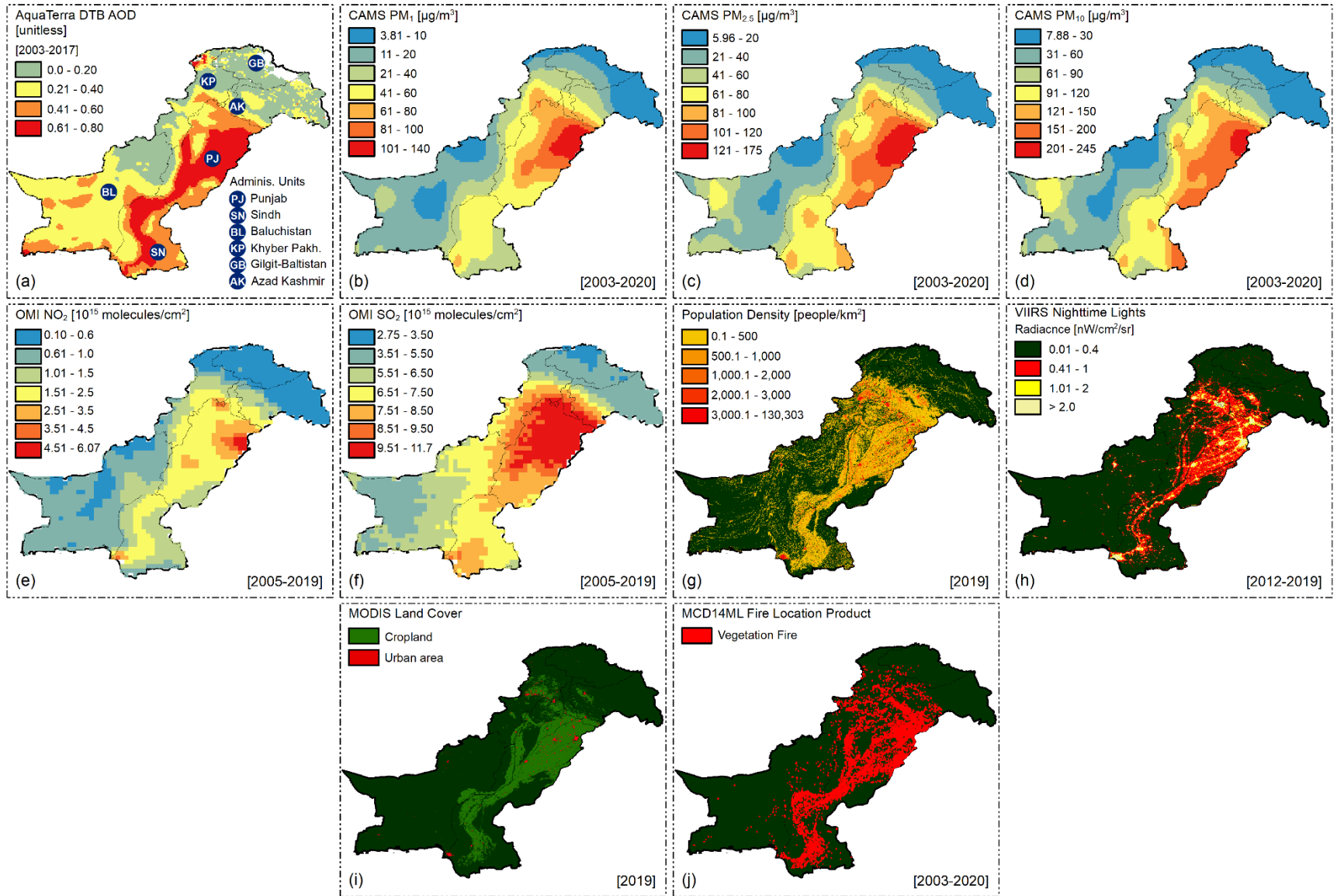
888 **Figure 13:** Ranking of high to low polluted cities in Pakistan according to OMI SO₂ VCDs
889 (molecules/cm²) from 2004 to 2019. (a) polluted cities based on mean SO₂, (b) pixel counts, (c)
890 polluted months based on mean SO₂, and (d) polluted years based on mean SO₂.

891 **5.4 Spatial distributions of aerosols and trace gases**

892 The purpose of this section is to link the spatial distributions of aerosols and trace gases with
893 each other as well as with population density, nighttime lights, land cover types (cropland and
894 urban areas), and presumed vegetation fire activities. Here, the PM_x data are interpolated using
895 cubic convolution (Keys 1981) from 0.75° grid size to 0.125° grid size to better show the smooth
896 spatial distributions over different administrative units. The spatial distributions of the multi-year
897 averaged concentrations of aerosols (AOD, PM_x) and trace gases (VCDs) (Figure 14) show that
898 Punjab is the most polluted region of Pakistan, followed by Sindh. It is significant that other
899 environmental data including population density (Figure 14g), VIIRS nighttime lights (Figure 14h),
900 cropland (Figure 14i), and vegetation fires (Figure 14j) show similar spatial patterns. It is obvious
901 that vegetation fires would have the same spatial pattern as cropland, but not obvious that
902 population density and nighttime lights would have the same pattern. As nighttime lights and
903 vegetation fires represent human activities, having the same spatial patterns suggests that the
904 majority of human settlements including urban, suburban and, industrial regions, are inter-mixed
905 with cropland. Interestingly, these coincident spatial distributions (population, nighttime lights,
906 land cover, and fires) correspond to the higher ranges of pollutants i.e., AOD > 0.4, PM₁ > 20
907 µg/m³, PM_{2.5} > 40 µg/m³, PM₁₀ > 60 µg/m³, NO₂ > 1.0×10¹⁵ molecules/cm², and SO₂ > 6.5×10¹⁵
908 molecules/cm². These results suggested that the primary (directly emitted) and the secondary

909 (gas-to-particles formation) aerosol emissions and trace gases are mainly from local
910 anthropogenic sources such as power plants, oil and gas refineries, vehicular emissions, crop
911 residue burning, and industrial activities including construction, manufacturing of cement,
912 ceramic, and bricks, and metals smelting. These anthropogenic sources are mainly responsible
913 for NO₂, SO₂, and PM_x (Adrees et al. 2016; Shah et al. 2012; Ur Rehman et al. 2019). Among these
914 anthropogenic sources, brick kilns industries are considered a major source. Small-scale
915 traditional brick kilns, located in rural and suburban areas, produce large amounts of gaseous
916 pollutants (NO₂, SO₂, O₃, and CO) and PM_x due to the usage of low-quality fuels including coal,
917 oil, wood, rice straw, rice husk, rubber tires, bagasse, and corncobs (Adrees et al. 2016; Ishaq et
918 al. 2010). Besides this, the combustion of agricultural biomass and crop residue burning are also
919 contributing to deteriorating rural and urban air quality (Irfan et al. 2015; Irfan et al. 2014). Irfan
920 et al. (2015) reported that Punjab produced more aerosol pollutants than Sindh from crop
921 residue burning and among the crop residues, wheat straw is the main contributor of NO_x, SO₂,
922 CO₂, and CO. Pakistan's 23.6 million vehicles emitted 58% of the country's total NO₂ emission and
923 34% is emitted by power plants and industries (Amnesty International 2019; Government of
924 Pakistan 2019; UNDP 2019). Another important source of aerosol pollutants, missed by previous
925 studies, is the burning of solid waste and street garbage which is a common practice in Pakistan,
926 even in major urban cities such as Islamabad, Lahore, Rawalpindi, Faisalabad, Gujranwala, Okara,
927 etc. To support this statement, some illustrations with references are provided in the
928 supplementary data (Figure S2). Figures 14a to 14d show that deserts (see Figure 1 for locations)
929 are another source of increasing AOD and PM_x levels in Pakistan. Although local anthropogenic
930 activities are the main source of aerosol pollutants and severe air quality problems in Pakistan,

931 transboundary transport of aerosols may also influence Pakistan's air quality. Contributions of
932 transboundary transport are investigated in section 5.7, using PSCF analyses, integrated with
933 HYSPLIT backward trajectory analysis and ground-based PM_{2.5} measurements.



935 **Figure 14:** Spatial distributions of yearly mean (a) AOD averaged over the years 2003–2017 (b)
936 PM₁ [2003–2020] (c) PM_{2.5} [2003–2020], (d) PM₁₀ [2003–2020], (e) NO₂ [2005–2019], (f) SO₂
937 [2005–2019], (g) Population density [2019], (h) VIIRS Nighttime Lights [2012–2019], (i) Land
938 cover types [2019], and (j) Presumed vegetation fire data [2003–2020].

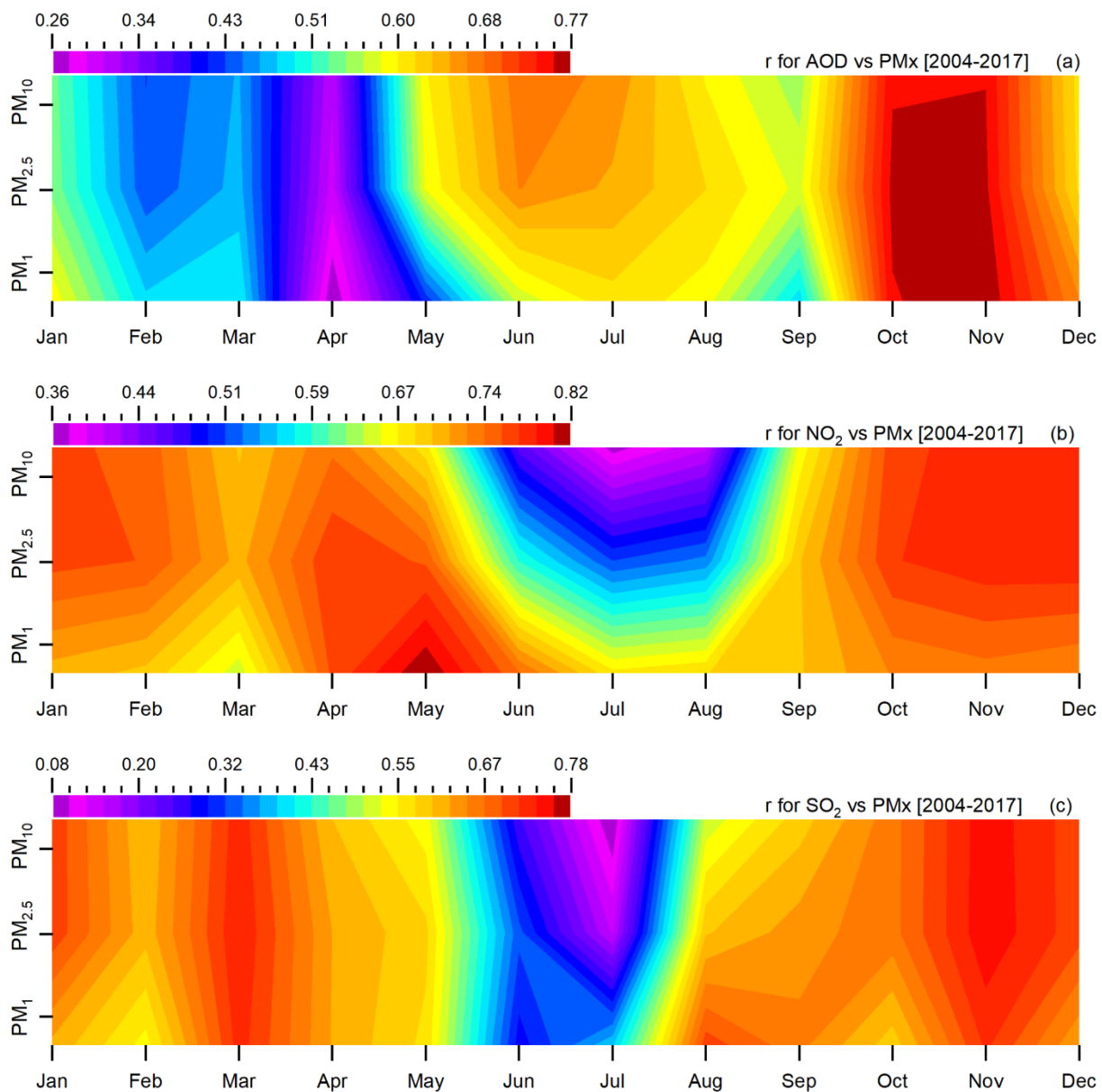
939 **5.5 Relationship of PM_x with AOD, NO₂, and SO₂**

940 AOD provides valuable information about the aerosol loading in the atmospheric column,
941 while the PM_x represents the aerosol concentrations near the ground. This section assesses how
942 well satellite-based AOD describes PM₁, PM_{2.5}, and PM₁₀ by examining the monthly correlation
943 between AOD and PM_x. We have also examined the monthly correlation between PM_x and SO₂
944 and NO₂ to understand the common sources that originated mainly from a combustion process.
945 The relationships between AOD and PM_x vary spatially and temporally and are influenced by
946 several factors such as meteorological variables including boundary layer height and relative
947 humidity, and the vertical distribution of aerosol layer (Li et al. 2016; Mhawish et al. 2021). The
948 linear correlation between AOD and PM_x shows a higher correlation coefficient from October to
949 January (see Figure 15a) when the atmosphere is stably stratified and the boundary layer is
950 shallow. This suggests that the AOD and PM_x variability are well agreed during the stable
951 atmospheric conditions (from Oct to Jan) and AOD can explain > 65% in the PM_x variability. On
952 the other hand, during April and May when the atmosphere is unstable and the boundary layer
953 deeper, the correlation between AOD and PM_x was smaller ($r < 0.4$). In the rainy season (July to
954 August), the correlation coefficient between AOD and PM₁₀ was found higher than PM_{2.5} and PM₁
955 which may be due to the larger contribution of coarse dust particles to the total aerosol loading

956 than $PM_{2.5}$ and PM_1 . The high relative humidity in the summer season enhanced the AOD retrieval
957 due to the hygroscopic growth of aerosol particles. On the other hand, the wash-out of PM_x due
958 to precipitation, deeper boundary layer, and strong convection during rainy months leads to a
959 reduction in the ground-level PM_x concentrations, while the AOD retrieval remains high under
960 cloud-free conditions during the inactive rain phase (Mhawish et al. 2021). The results suggested
961 that using satellite-based AOD to infer the ground-level PM_x variability is limited to specific
962 meteorological conditions such as stable atmospheric conditions and dry seasons. On the other
963 hand, the weak linear relationship between AOD and ground-level PM_x concentrations found
964 during unstable conditions in spring and summer and more influenced by meteorological
965 variables and atmospheric mixing height.

966 Tropospheric NO_2 and SO_2 are precursors for the formation of secondary aerosols which are
967 produced by anthropogenic activities such as fossil fuel burning and power plants. The strong
968 correlation coefficient between PM_x vs. SO_2 and NO_2 in the spring months suggests that
969 photochemical reactions can contribute to the formation of PM_x . The strong correlation in winter
970 suggests that both trace gases NO_2 and SO_2 originated from the same emission sources of PM_x ,
971 mainly domestic heating, industrial activities, and vehicular emissions. While the lower
972 correlation in the summer monsoon may be attributed to the higher contribution of natural
973 sources of PM_x and the deeper boundary layer that enhance the dispersion of air pollutants.

974



975

976

Figure 15: Relationship from PM_x with (a) AOD, (b) NO₂, and (c) SO₂ from 2004-2017.

977

5.6 Trends of aerosol and trace gas concentrations

978

This section presents the annual trends in the six parameters used to assess the air quality in

979

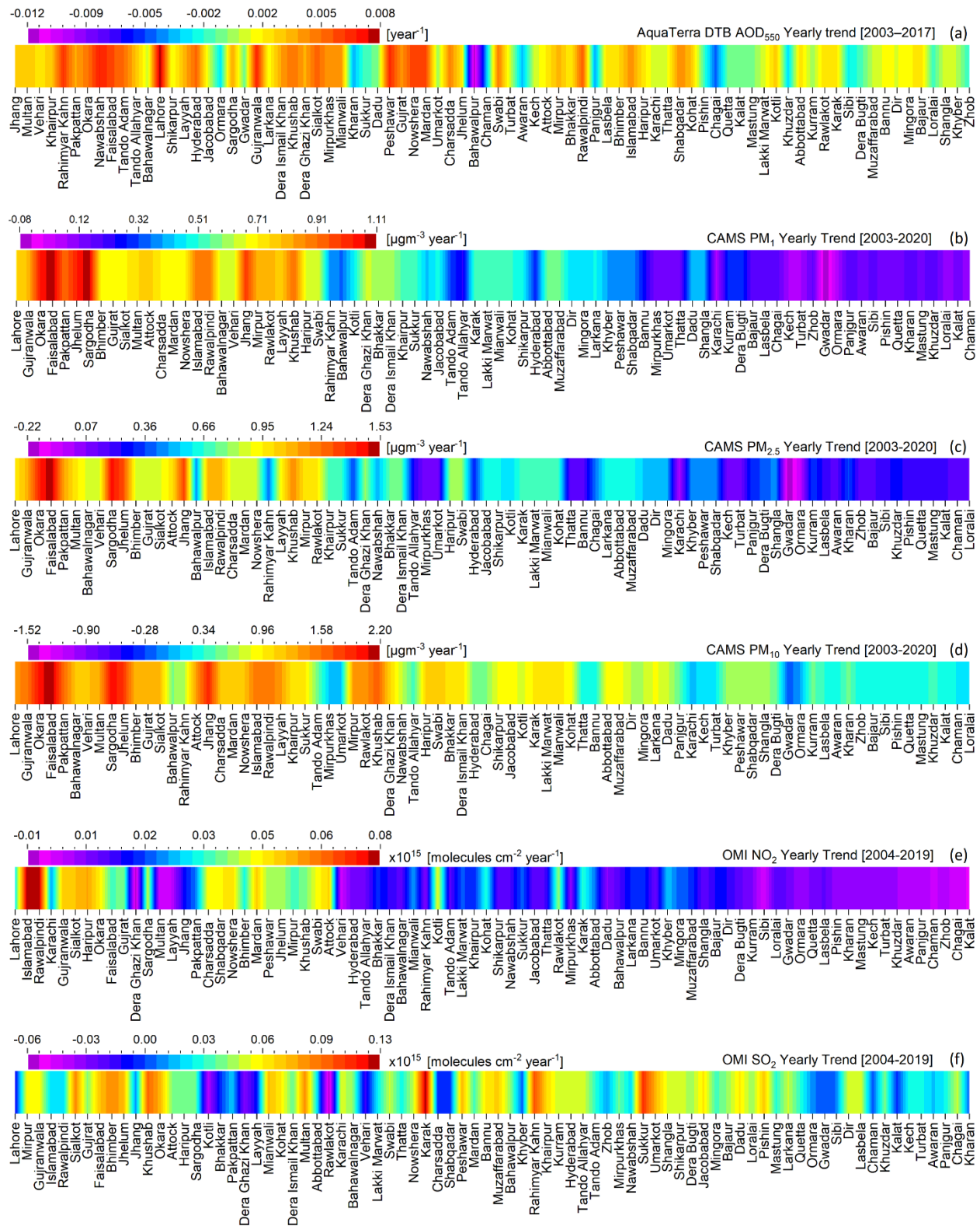
each city of Pakistan. The annual trends were calculated after removing the seasonality from the

980

monthly mean time series data which also accounted for temporal autocorrelation. Figure 16

981 shows the magnitude of the trends as Theil-Sen's slope over each individual city, for the periods
982 indicated at the top of each Figure. A significant positive trend in PM_x was found over most cities,
983 particularly in Punjab, Khyber Pakhtunkhwa, and the Islamabad Capital Territory. The PM_x trends
984 found over cities in Punjab range from $+0.35$ to $+1.10 \mu\text{g}/\text{m}^3 \text{yr}^{-1}$, $+0.42$ to $+1.52 \mu\text{g}/\text{m}^3 \text{yr}^{-1}$ and
985 $+0.57$ to $+2.20 \mu\text{g}/\text{m}^3 \text{yr}^{-1}$ for PM_1 , $PM_{2.5}$ and PM_{10} , respectively. Correspondingly, the AOD trend
986 in Punjab cities was positive, with the strongest increase over Lahore (0.008yr^{-1}). Over cities in
987 Khyber Pakhtunkhwa and Azad Kashmir, the AOD trend was also positive, but smaller than in
988 Punjab. The positive trends in PM_x and AOD, particularly over cities in Punjab, may be due to
989 increasing aerosol emissions and/or secondary aerosol formation. Anthropogenic activities and
990 biomass burning are considered major sources of ultrafine and fine particles (PM_1 and $PM_{2.5}$) over
991 the region (Alam et al. 2015; Stone et al. 2010). Anthropogenic activities also result in the
992 production of NO_2 and SO_2 and $\sim 91\%$, and $\sim 88\%$ of the cities the trends in the NO_2 and SO_2 ,
993 respectively, are positive. This increase in trace gas concentrations would be a further source of
994 increased particulate pollution, as trace gases facilitate secondary aerosol formation via gas-to-
995 particle conversion reactions (Seinfeld and Pandis 1998).

996 In terms of monthly trends, the common feature is that the statistically significant positive
997 trends of PM_x were largest during the cold months (November to February), particularly over
998 major Punjab cities (Lahore, Faisalabad, and Gujranwala) and Islamabad (Figure S3). In contrast,
999 during the summer months, the trends over many cities are negative. The overall positive annual
1000 trends indicate that the increase of the PM_x concentrations in the winter is stronger than the
1001 decrease in the summer. The reasons for these opposing trends are beyond the scope of the
1002 current study and require further, more detailed investigation.



1003

1004 **Figure 16:** Annual mean trend in aerosols and trace gas concentrations for (a) AOD, (b) PM₁, (c)

1005 PM_{2.5}, (d) PM₁₀, (e) NO₂, and (f) SO₂. The trends were calculated over different periods of time,

1006 which are indicated on top of each figure on the right-hand side, together with the type of
1007 species.

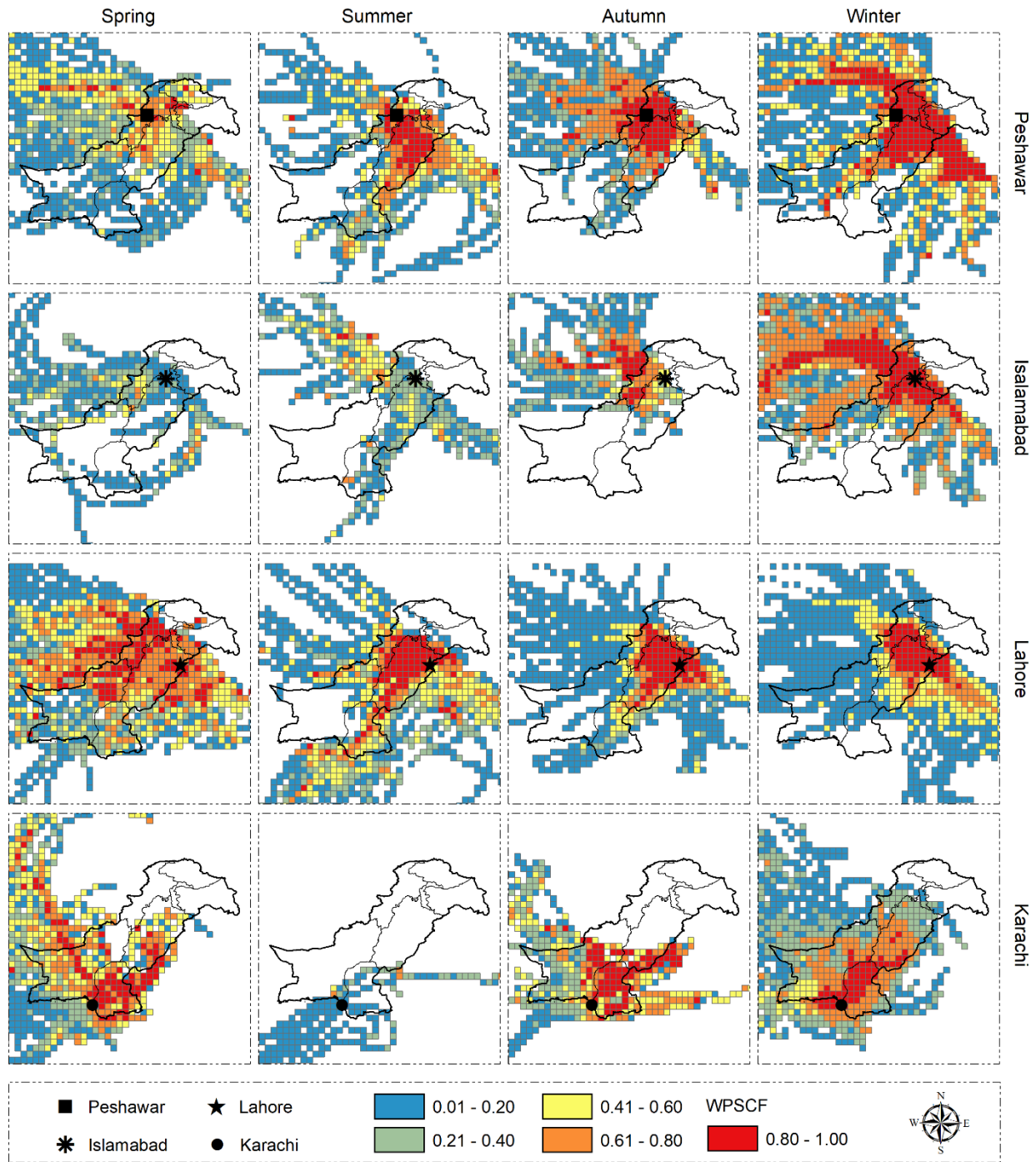
1008 **5.7 Potential Source Contribution Function (PSCF) Analysis**

1009 PSCF analysis was used to identify the potential source areas for PM_{2.5} at four receptor cities:
1010 Peshawar, Islamabad, Lahore, and Karachi, for the period from March 2020 to February 2021. 72
1011 hours HYSPLIT backward trajectories were computed for each receptor site, arriving every 6
1012 hours at the height of 500 m above ground level (AGL). The results were grouped by season as
1013 shown in Figure 17. The results show strong differences between cities and seasons. Starting with
1014 Peshawar, in spring there are some local sources regions around the city, within a few hundreds
1015 of km, but also strong contributions from the WNW (West-NorthWest) in Afghanistan and from
1016 the SE in India. In the summer, the source regions are mostly located in Pakistan, but with a
1017 contribution from sources to the SE (SouthEast), in India. In contrast, in the autumn the
1018 contributions from India are very small but those from Afghanistan, both to the NW (NorthWest)
1019 and W (West) are relatively large. Whereas, in the winter source regions in NW and SE directions
1020 (Afghanistan and India, respectively) are stronger than in other seasons. In Islamabad, not far
1021 from Peshawar, the situation is quite different. In the spring, the source regions have a rather
1022 low PSCF, and are distributed over specific directions to the W (West) into Afghanistan and
1023 toward the SE in India, with few local sources. In the summer, the source regions are similar to
1024 those in Peshawar, but with low PSCF except for the source regions in Afghanistan which seem
1025 to contribute most to the air pollution in Islamabad in the summer, but still with moderate PSCF.
1026 In the autumn sources to the W and N dominate with stronger contributions from Afghanistan
1027 than from the local sources. In the winter, the source regions redistributed over a much larger
1028 area than in other seasons, with strong contributions from both local sources and Afghanistan,
1029 as well as some contributions from India. The situation in Lahore is remarkably different, with
1030 the strongest contributions from sources inside Pakistan (PJ and KP), some contributions from

1031 sources to the SE in India, during all seasons, and in the spring a strong contribution from sources
1032 in Afghanistan. The situation in Karachi is again different, both as regards source regions and
1033 seasonal behaviour. The strongest contributions come from local sources within a few hundreds
1034 of km in Pakistan, except in the summer when all source regions are weak contributors (PSCF
1035 <0.2) and almost all located over the ocean. In the spring, source regions to the NW, reaching far
1036 into Afghanistan, contribute to the PM_{2.5} in Karachi. Oceanic sources also contribute some to
1037 the PM_{2.5} in other seasons

1038 In summary, the values of PSCF indicate the regional transport of aerosol from source regions
1039 in Afghanistan and India (see Figure 1 for locations). Karachi is influenced by fine dust particles
1040 from the Cholistan and Thar deserts (see Figure 1 for locations). Figure 17 shows that the PM_{2.5}
1041 in Lahore, the top polluted city of Pakistan, is mainly influenced by source areas in Pakistan,
1042 during all seasons. This suggests that increases in local anthropogenic activities play an important
1043 role in the worsening of Lahore's air quality. Overall, the higher values of PSCF > 0.6 identify
1044 potential source areas which are located both inside and outside of Pakistan, which indicates that
1045 the air quality in Pakistan is not only influenced by local sources but also influenced by transport
1046 from regional sources areas.

1047



1048

1049 **Figure 17:** Potential source contribution function plots for PM_{2.5} at seasonal scales from March
 1050 2020 to February 2021 for four receptor cities namely, Peshawar, Islamabad, Lahore, and
 1051 Karachi (see legend for identification).

1052 **6. Conclusions**

1053 In this study, long-term (2003–2020) remote sensing, ground-based, and model simulation
1054 datasets were combined to provide the most comprehensive and extensive evaluation ever, of
1055 air quality conditions over Pakistan. Long-term spatio-temporal distributions of aerosol
1056 pollutants and trace gases, recent long-term trends at the city level, ranking of cities in terms of
1057 air pollution levels into three categories (extremely polluted, highly polluted, polluted cities), and
1058 the potential sources of air pollution across Pakistan were reported.

1059 The highest AOD was observed in the summer months (June to August), mainly attributed to
1060 the hygroscopic growth of aerosol particles during the humid summer season. High AOD levels
1061 were also observed during cold months (October to January), mainly over biomass burning
1062 affected regions such as Punjab. For PM_x and trace gases, the highest values were observed
1063 during cold months from October to February, when the atmosphere is stably stratified and the
1064 boundary layer is shallow, and emissions from anthropogenic activities and biomass burning are
1065 higher than in other seasons.

1066 The CAMS PM_{2.5} data are in better agreement with ground-based PM_{2.5} concentrations than
1067 MERRA-2 reanalysis PM_{2.5} data and were therefore used to rank the cities in terms of
1068 concentrations of particulate matter (PM_x). The 18 years average of the PM_{2.5} concentrations for
1069 the 80 cities of Pakistan show that a total of 21 cities fall within the category of extremely polluted
1070 cities ($PM_{2.5} > 92.84$) (namely Punjab: 17, Khyber Pakhtunkhwa: 3, Azad Kashmir: 1), 40 cities fall
1071 within the category of highly polluted cities ($45.69 < PM_{2.5} < 92.84$) (namely 6 in Punjab, 14 in
1072 Sindh, 3 in Balochistan, 13 in Khyber Pakhtunkhwa and 4 in Azad Kashmir); 19 cities fall within

1073 the category of polluted cities ($PM_{2.5} < 45.69$) (16 in Balochistan and 3 in Khyber Pakhtunkhwa).
1074 No single city in Pakistan falls within the $PM_{2.5}$ standards defined by Pak-NEQS and WHO, and the
1075 values of PM_1 and $PM_{2.5}$ for the top 10 cities are 5.6 (8.4) to 9.0 (13.5) times and 7.2 (10.8) to
1076 11.4 (17.1) times larger than the Pak-NEQS (WHO AQG). The map of annual average $PM_{2.5}$ shows
1077 that people in the whole country are exposed to high $PM_{2.5}$ concentrations for many years, with
1078 the annual mean concentrations for all cities exceeding the Pak-NEQS ($<15 \mu\text{g}/\text{m}^3$), and 68, 73,
1079 and 80 cities exceeding the WHO Interim Target-1 ($<35 \mu\text{g}/\text{m}^3$), Target-2 ($<25 \mu\text{g}/\text{m}^3$), and
1080 Target-3 ($<15 \mu\text{g}/\text{m}^3$), respectively. In terms of pollution sources, the study indicates that
1081 biomass (crop residue) burning activities may not be the main source of severe air quality
1082 conditions in Pakistan: the highest PM_x concentrations were observed in December and January
1083 when also the NO_2 TVCD and SO_2 VCD, used as proxies for NO_2 and SO_2 concentrations, were
1084 highest. The emissions of these trace gases are known to be associated with anthropogenic
1085 activities including transport, industrial activities, and power generation. Interestingly, higher
1086 levels of AOD, PM_1 , $PM_{2.5}$, PM_{10} , NO_2 , SO_2 , population density, nighttime lights, and vegetation
1087 fire activities showed the same spatial pattern as cropland: most of the major cities, as well as
1088 rural areas in Pakistan, are surrounded by cropland and transport of pollutants generated from
1089 anthropogenic activities mix with aerosol and trace gases generated from agricultural activities,
1090 biomass burning and natural aerosols such as dust, to produce a rather smooth mixture of the
1091 metrics reported in this study. These findings suggest that Pakistan's extreme air pollution
1092 problems are strongly influenced by anthropogenic activities within Pakistan. This is also
1093 confirmed by the PSCF (> 0.6) analysis based on HYSPLIT air parcel back trajectories and ground-

1094 based PM_{2.5} concentrations. In addition, meteorological factors have a strong influence on the
1095 occurrence of pollution episodes.

1096 Significant positive trends in the concentrations of AOD, PM₁, PM_{2.5}, PM₁₀, NO₂, and SO₂ were
1097 observed from November to February, particularly over Lahore, Islamabad, Gujranwala, and
1098 Faisalabad.

1099 The final remark of this study is that all cities in Pakistan have been exposed to long-term
1100 PM_x, NO₂, and SO₂ concentrations throughout the last two decades. The pollution levels in these
1101 cities imply extremely poor air quality conditions, mainly due to local anthropogenic activities,
1102 which severely threaten human life. This comprehensive study, based on long-term multi-source
1103 information on aerosols and trace gases may be considered a baseline study by the Ministry of
1104 Climate Change, Pakistan, and other policymakers, to mitigate air pollution problems in Pakistan.

1105 **CRedit authorship contribution statement**

1106 **Muhammad Bilal:** Conceptualization, Data curation, Methodology, Formal analysis,
1107 Investigation, Validation, Visualization, Writing - original draft. **Janet E. Nichol:** Supervision,
1108 Investigation, Writing - review & editing. **Zhongfeng Qiu:** Supervision, Investigation, Writing -
1109 review & editing. **Alaa Mhawish:** Data curation, Writing - review & editing. **Majid Nazeer:** Data
1110 curation, Writing - review & editing. **Md. Arfan Ali:** Data curation, Writing - review & editing.
1111 **Gerrit de Leeuw:** Supervision, Writing - review & editing. **Robert C. Levy:** Writing - review &
1112 editing. **Yu Wang:** Data curation. **Yang Chen:** Data curation. **Lunche Wang:** Data curation. **Yuan**

1113 **Shi:** Writing - review & editing. **Max P. Bleiweiss:** Supervision, Investigation, **Luqman Atique:**
1114 Visualization. **Usman Mazhar:** Visualization.

1115 **Declaration of competing interest**

1116 The authors declare that they have no known competing financial interests or personal
1117 relationships that could have appeared to influence the work reported in this paper.

1118 **Acknowledgment**

1119 The authors would like to acknowledge NASA's Level-1 and Atmosphere Archive and
1120 Distribution System (LAADS) Distributed Active Archive Center (DAAC)
1121 (<https://ladsweb.modaps.eosdis.nasa.gov/>) for MODIS data, the Copernicus Atmosphere
1122 Monitoring Service (CAMS) for air quality data (PM₁, PM_{2.5}, and PM₁₀), the Goddard Earth Science
1123 DISC (<https://daac.gsfc.nasa.gov/>) for OMI data, Principal Investigators of Lahore AERONET site,
1124 and the NOAA Air Resources Laboratory (ARL) for the provision of the HYSPLIT air parcel back
1125 trajectories (<https://www.ready.noaa.gov>) used in this publication. The authors are grateful to
1126 Mr. Abid Omar, founder of Pakistan Air Quality Initiative (PAQI), and gratefully acknowledge the
1127 U.S. Department of State for providing the open access ground-based PM_{2.5} data. The authors
1128 also acknowledge the use of data from NASA's Fire Information for Resource Management
1129 System (FIRMS) (<https://earthdata.nasa.gov/firms>), part of NASA's Earth Observing System Data
1130 and Information System (EOSDIS). The authors are thankful to Mr. Pravash Tiwari for helping in
1131 PSCF analysis and Dr. Devin White (Oak Ridge National Laboratory) for MODIS Conversion Tool
1132 Kit (MCTK).

1133

1134 **Funding**

1135 This work was supported by the National Key Research and Development Program of China
1136 (2016YFC1400901), the Special Project of Jiangsu Distinguished Professor (R2018T22), the
1137 National Natural Science Foundation of China (41976165), Jiangsu Technology Project of Nature
1138 Resources (KJXM2019042), and the Startup Foundation for Introduction Talent of NUIST
1139 (2017r107). Additional support came from the New Mexico State University College of
1140 Agriculture Consumer and Environmental Sciences' Agricultural Experiment Station.

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