

Commercial fishery bycatch risk for large juvenile and adult smalltooth sawfish (*Pristis pectinata*) in Florida waters

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22 **Abstract**

23 1. Incidental catch of marine species can create ecological and economic issues, particularly for
24 endangered species. The smalltooth sawfish (*Pristis pectinata*) is endemic to the Atlantic Ocean
25 and listed as Endangered on the US Endangered Species Act (ESA). One of its major threats is
26 bycatch mortality in commercial fisheries.

27 2. Despite protections of the ESA, smalltooth sawfish are still captured as bycatch in commercial
28 fisheries. Acoustic and satellite tag data collected on 59 sawfish between 2011 and 2019 were
29 analysed to assess commercial fishery bycatch risk for large juveniles and adults off Florida. We
30 focused on three fisheries: shrimp trawl, southeast coastal gillnet, and shark bottom longline, as
31 these were identified in the recovery plan as having the greatest potential threats to recovery.

32 3. Bycatch risk associated with the shrimp trawl fishery was significantly higher than the other
33 fisheries, indicating that this fishery currently poses the greatest threat to recovery.

34 4. Bycatch risk was concentrated in all seasons in the Gulf of Mexico adjacent to the lower
35 Florida Keys for the shrimp trawl fishery, off Cape Canaveral in the southeast coastal gillnet
36 fishery, and in the Atlantic Ocean adjacent to the Florida Keys in the shark bottom longline
37 fishery.

38 5. Tagging location and sex were predictors of bycatch risk. Individuals tagged in Charlotte
39 Harbor had the highest shrimp trawl bycatch risk. Females tagged in south Florida tended to
40 reside in the deepest water, which is where shrimp trawl effort is highest. Therefore, females
41 may be at more risk in these deeper waters.

42 6. Results from this study indicate a year-round closure of waters off southwest Florida to the
43 shrimp trawl fishery between Charlotte Harbor and the western Florida Keys could reduce
44 sawfish bycatch and thus mortality, which is in line with recovery plan goals

45

46 KEYWORDS

47 acoustic monitoring, bycatch, commercial fisheries, conservation, endangered species, satellite
48 telemetry, elasmobranch

49

50 1. INTRODUCTION

51 Bycatch is defined in the United States (US) as the incidental capture and subsequent discard of a
52 non-targeted species (NOAA, 2019). Many marine animals including sea turtles, marine
53 mammals, invertebrates, seabirds, elasmobranchs, and teleosts are incidentally caught in
54 commercial fisheries (Zollett, 2009; Kroetz, Mathers & Carlson, 2020). Bycatch creates both
55 economic and ecological issues including damage to gear, lost income, lost time, and mortality
56 of non-target species. This can create negative ecosystem effects through loss of top predators,
57 removal of large biomasses of important prey taxa, and cryptic mortality of threatened species
58 (Zollett, 2009). Bycatch is of particular conservation concern for species with low intrinsic rates
59 of population growth and small or threatened populations (Dulvy et al., 2008; Northridge et al.,
60 2017).

61 Bycatch mortality is a major threat for many protected marine species and numerous
62 strategies have been used to mitigate this risk (Zollett, 2009). In 1994, amendments were made to
63 the US Marine Mammal Protection Act to mitigate the impacts of bycatch mortality on marine
64 mammals and these protections were successful in ensuring the continued recovery of some

65 threatened species (Johnson et al., 2005). Farmer et al. (2016) evaluated several bycatch
66 mitigation options to reduce entanglement risk of North Atlantic right whales (*Eubalaena*
67 *glacialis*) with black sea bass (*Centropristis striata*) pot gear and ultimately found time-area
68 closures to be a viable option to decrease bycatch mortality. Turtle exclusion devices (TEDs)
69 have led to a significant decrease in bycatch of sea turtles in trawl fisheries worldwide and there
70 is evidence that they may also mitigate bycatch risk for other non-targeted species (Zollett,
71 2009).

72 Sawfishes are among the most endangered elasmobranch families in the world, with all
73 five species listed as Endangered or Critically Endangered on the International Union for the
74 Conservation of Nature Red List of Threatened Species (Dulvy et al., 2016). The smalltooth
75 sawfish (*Pristis pectinata*) is endemic to the Atlantic Ocean, historically occupying subtropical
76 and tropical waters on both sides of the basin. In the western Atlantic, the species inhabited
77 waters along the east coast of the US from Florida at least as far north as North Carolina, the
78 entire Gulf of Mexico, the Caribbean including The Bahamas, and as far south as Uruguay
79 (NMFS, 2009b). Sawfishes are benthic species with long toothed rostra making them prone to
80 entanglement in fishing gear, particularly gear on the bottom. Since the industrial revolution, the
81 range of smalltooth sawfish has declined dramatically due to fishing, habitat loss, and
82 overexploitation (Carlson, Wiley & Smith, 2013). The range has contracted substantially and
83 there are only two known viable ‘lifeboat’ populations remaining (Dulvy et al. 2014). One is
84 centered in southwest Florida waters (NMFS, 2009a; Norton et al., 2012; Brame et al., 2019) and
85 the other is in The Bahamas (Guttridge et al., 2015).

86 In Florida, the smalltooth sawfish is incidentally caught in fisheries in state and federal
87 waters. The smalltooth sawfish was prohibited from harvest in Florida in 1992 and listed as

88 Endangered under the US Endangered Species Act (ESA) in 2003 (NMFS, 2009b). Following
89 the ESA listing, a team of experts was assembled to develop a recovery plan to outline major
90 threats to the species as well as goals and objectives. One of the major goals was to estimate the
91 impact of commercial fisheries on recovery and the feasibility of policy implementation to
92 mitigate fishery threats (NMFS, 2009b). The recovery plan identified the shrimp trawl fishery as
93 the largest source of direct mortality and biggest potential threat to recovery, followed by the
94 southeast coastal gillnet fishery and the shark bottom longline fishery. Like other commercial
95 fisheries, shrimp trawling is prohibited in some State of Florida waters, including Everglades
96 National Park and the Florida Keys National Marine Sanctuary, due to habitat considerations
97 (e.g. to protect seagrass and hardbottom habitats or limits to fishing close to the shoreline) and
98 conflicts with other fisheries (e.g. trap fishery for stone crabs, *Menippe mercenaria*). However,
99 shrimp trawling is currently allowed elsewhere in state and federal waters. All coastal gillnetting
100 was banned in state waters in 1994; longlining is also prohibited in state waters, but both gears
101 are currently allowed in federal waters.

102 The shrimp trawl fishery is one of the most profitable fisheries in the US, but also
103 accounts for a large percentage of incidental catches. According to National Marine Fisheries
104 Service (NMFS) observer data, between 1998 and 2008, trawls were towed for an average of 3.9
105 hr, with some trawls towed as long as 12.8 hr. Shrimp trawling gear is deployed at an average
106 depth of 73 m with some gear being deployed as deep as 540 m. Both penaeid and rock shrimp
107 are targeted by this fishery in the Gulf of Mexico and South Atlantic (Scott-Denton et al., 2012).
108 Harrington et al. (2005) reported that shrimp trawls accounted for nearly half of all fishery
109 bycatch in US waters. For this reason, in 1992 the NMFS Southeast Fisheries Science Center
110 implemented a research plan in collaboration with the Gulf and South Atlantic Fisheries

111 Foundation to collect bycatch data from the fishery (Scott-Denton et al., 2012). However,
112 observer coverage on shrimp trawl vessels in the US is extremely low (1–2% coverage), so
113 bycatch impacts are still largely unknown (Scott-Denton et al., 2012).

114 The southeast coastal gillnet fishery targets sharks and teleosts and uses sink, strike, and
115 drift gillnet gear. According to NMFS observer data gathered between 1998 and 2017,
116 approximately 71% of coastal gillnets deployed were sink, 8% were strike, and 21% were drift.
117 Sawfish are largely benthic, thus the sink gillnets present the biggest threat because they sit on
118 the bottom where sawfish reside. The southeast coastal gillnet fishery targets Spanish mackerel
119 (*Scomberomorus maculatus*), southern kingfish (*Menticirrhus americanus*), spiny dogfish
120 (*Squalus acanthias*), mixed teleosts, and mixed sharks. Depending on target species, nets range
121 from 14 to 3,246 m long with stretch mesh sizes between 3.2 and 38 cm, and are deployed at
122 depths from 1.2 to 110 m for durations between 0.05 and 91 hr (Kroetz, Mathers & Carlson,
123 2020).

124 The shark bottom longline fishery has been monitored by NMFS observers since 1994
125 and approximately 200 fishers have US permits to target sharks in the Atlantic Ocean and Gulf
126 of Mexico (Mathers et al., 2018). The observer coverage goal of this fishery is 5–10%, but there
127 is 100% coverage on the 4–6 commercial shark fishing vessels participating in the shark research
128 fishery programme monitored by NMFS. Based on observer data from vessels not participating
129 in the research programme, on average, mainlines were 7.2 km long (range = 0.9 to 12.0 km),
130 gear was deployed at depths between 3 and 21 m (average = 16.4 m), and had between 47 and
131 401 hooks (average = 289). The majority (63.6%) used 18/0 circle hooks and the average soak
132 time was 7.8 hr. Vessels that participated in the research programme had mainline lengths
133 ranging from 2 to 19.6 km (average = 7.0 km), were deployed at depths between 4 and 158 m

134 (average = 31.4 m), and had between 112 and 300 hooks (average = 247). The majority (51.9%)
135 used 18/0 circle hooks and the average soak time was 5.6 hr (Mathers et al., 2018).

136 For this study, bycatch risk is defined as the probability of commercial fishing occurring
137 in an area at the same time as a sawfish is in that area. Minimizing interaction potential with
138 commercial fisheries is important due to high sawfish mortality rates from incidental catches,
139 particularly in the shrimp trawl fishery (NMFS, 2009b). The toothed rostra of sawfish are prone
140 to entanglement in nets and bringing the entire animal on board to disentangle can be dangerous.
141 This sometimes leads fishers to seriously harm or kill the sawfish. Breaking or removing the
142 rostrum alters a sawfish's behavior and usually leads to death (NMFS, 2009b; Morgan et al.,
143 2016; G. R. Poulakis, unpublished data).

144 Our objective was to use long-term, wide-ranging passive acoustic monitoring and
145 shorter-term satellite telemetry data from large juvenile and adult smalltooth sawfish to
146 determine how movement patterns and habitat use interact with commercial fishing effort of the
147 shrimp trawl, southeast coastal gillnet, and shark bottom longline fisheries. Results can aid
148 resource managers to reduce smalltooth sawfish bycatch and thereby facilitate population
149 recovery.

150

151 **2. METHODS**

152 **2.1 Acoustic receiver networks**

153 Acoustic receivers for monitoring smalltooth sawfish were established within the Charlotte
154 Harbor estuarine system, Everglades National Park, and the Florida Keys. The Charlotte Harbor
155 array contained 51 receivers in the northern portion of the estuary in and around the Peace River
156 as well as 51 receivers in the southern portion of the system in and around San Carlos Bay and

157 the Caloosahatchee River. The array in the Everglades National Park and Florida Keys region
158 contained 26 receivers maintained by co-authors that tagged sawfish. This study also used the
159 Florida Atlantic Coast Telemetry (FACT) (secoora.org/fact), Atlantic Cooperative Telemetry
160 (ACT) (theactnetwork.com), and Integrated Tracking of Aquatic Animals in the Gulf of Mexico
161 (iTAG) (itagscience.com) arrays, which provided access to positive detection data from hundreds
162 of additional receivers along both coasts of Florida (Figure 1). These receivers were maintained
163 by various researchers and institutions so receiver download schedules varied.

164

165 **2.2 Tagging**

166 Sawfish were tagged primarily near where acoustic arrays were maintained for monitoring
167 smalltooth sawfish. Large juveniles (>2 m stretch total length [STL]) and adults (>3.4 m for
168 males; >3.7 m for females; Brame et al., 2019) were captured in Charlotte Harbor with rod and
169 reel and drumlines. Rod and reel used 36–45 kg test braided or monofilament line with 9/0 non-
170 offset circle hooks. Drumlines consisted of 20 kg concrete anchors and 5-m or 10-m gangions
171 with 250 kg test monofilament line and 14/0 non-offset circle hooks. Drumlines soaked for one
172 hr and up to five were set at a time. Rod and reel gear was typically used during the drumline
173 soaks. Sawfish were also tagged in the Florida Keys and portions of Everglades National Park
174 using bottom longlines, almost always set in pairs, of 50 16/0 non-offset circle hooks fished for
175 one hr, rod and reel as described above, and shoreline gillnets 1.5 m deep, between 30.5 and 61
176 m long, with stretch mesh sizes either 7.6 cm or 10.2 cm. Ladyfish (*Elops saurus*) was the
177 primary bait for all baited gears. Two sawfish were opportunistically tagged on the east coast;
178 they were caught in the intake canal net at the Florida Power and Light nuclear power plant in St.
179 Lucie, Florida.

180 Captured sawfish were measured (rostrum length, pre-caudal length, fork length, and
181 STL) and tagged with multiple tag types. External tags included either small rototags (Dalton[®],
182 Newark, UK) or metal-tipped dart tags (FH-69, ©Floy Tag & Mfg., Inc. Seattle, WA, USA)
183 placed on or near a dorsal fin. Sawfish were also injected with a passive integrated transponder
184 (PIT-tag; HPT12; Biomark[®], Inc., Boise, ID, USA) under the skin at the base of a dorsal fin for
185 identifying individuals after external tag loss. Finally, a 69 kHz acoustic transmitter
186 (Vemco/Innovasea V13-1L or a V16-6H) with either an estimated 4-yr or 10-yr battery life was
187 surgically implanted within the body cavity of some sawfish. These tags were programmed to
188 emit unique acoustic sequences on a random delay once every 80 to 180 s (V13) or 70 to 150 s
189 (V16). Surgery involved a 2–4 cm incision on the animal's ventral surface just anterior to the
190 pelvic fins using a sterile, disposable scalpel and 2–3 dissolvable surgical sutures to close the
191 incision after tag placement.

192 Other sawfish were tagged with multiple generations of pop-up archival transmitting
193 (PAT) tags manufactured by Wildlife Computers (i.e. PAT2–4, Mk10-PAT, MiniPAT, PATF).
194 These tags were programmed to pop-off between 60 and 150 days depending on the type. Tags
195 were rigged with either 136 kg monofilament leaders and a Pflieger Institute of Environmental
196 Research nylon “umbrella” dart or a modified harness consisting of 1.8-mm stainless steel cable
197 surrounded by chafe tubing, then clear surgical tubing with polyolefin heat-shrinkable tubing at
198 each end. Umbrella darts were inserted by making a small incision below the middle of the first
199 dorsal fin approximately 5 cm below the fin base and the dart was inserted into the musculature,
200 seating the anchor at a depth of 6–10 cm. For sawfish tagged with the modified harness, a small
201 hole was made through the anterior portion of the base of the first dorsal fin where the free end
202 of the harness assembly was threaded through to the opposite side of the dorsal fin. The free end

203 of steel cable was then inserted into the open sides of two double copperlock crimps, which were
204 closed, and excess cable was removed. The PAT tag trailed just behind the dorsal fin when the
205 sawfish was released.

206

207 **2.3 Data processing**

208 Acoustic data were first processed by removing any single detections within a 24-hr period to
209 avoid including false detections. The data were then binned by day to ensure data were not
210 skewed by a few individuals spending significant time near a single receiver within a single day.
211 Resulting data were used to calculate single band kernel density rasters with a cell size of 0.05
212 decimal degrees and populated by number of sawfish detected per day for each month using the
213 Kernel Density tool in ArcMap (ESRI, 2011 v10.7.1).

214 Satellite data were processed by filtering geolocation point estimates using a maximum
215 travelling speed of 110 km per day, which was based on maximum daily travelling distance
216 calculated from acoustic detections. Papastamatiou et al. (2015) estimated that the average rate
217 of movement of adult smalltooth sawfish actively tracked in Florida Bay was 1.2 km per hr (28.8
218 km per day) and the maximum rate of movement was estimated to be 7.5 km per hr (180 km per
219 day). It was assumed, based on sawfish behaviour, that migrating sawfish likely move faster than
220 the average rate of movement, but it is unlikely that the maximum rate of movement is
221 sustainable for a full day. Thus, the maximum rate of 110 km per day is likely a reasonable proxy
222 for maximum rate of movement over a 24-hr period. All geolocation point estimates on land
223 were also removed. After filtering, the point estimates were binned by month and monthly kernel
224 density rasters were created. To analyse space use, a combined activity raster was created by

225 building a mosaic of the acoustic and satellite data for each month. This was accomplished using
226 the Mosaic to New Raster tool in ArcMap by summing overlapping cells.

227 Smalltooth sawfish vulnerability to bycatch in commercial shrimp trawl, southeast
228 coastal gillnet, and shark bottom longline fisheries was analysed by overlaying movements from
229 acoustic and satellite tag data with fishing effort obtained from NMFS observer programmes.
230 While target observer coverage was only 1–2% for the shrimp trawl fishery, 5–15% for the
231 coastal gillnet fishery, and 5–100% of the total effort for the shark bottom longline fishery
232 (Scott-Denton et al., 2012; Mathers et al., 2017, 2018), these data were more reliable than
233 logbook data. Logbook data are reported by spatial grid and data from Vessel Monitoring
234 Systems (VMS), which makes it difficult to discern whether a vessel is actively fishing or just
235 moving to a new location. Fishing effort was calculated using the number of hours each gear was
236 deployed in a 30.8 km² area, which corresponds to the size of the NMFS’s spatial grids. The
237 shrimp trawl dataset contained 5,789 trawls and approximately 20,837 hr of fishing from 2005 to
238 2018. The southeast coastal gillnet dataset contained 2,480 sets and 7,022 hr of fishing from
239 2005 to 2017. The shark bottom longline fishery dataset contained 8,915 sets and 28,173 hr of
240 fishing from 2005 to 2016.

241 Kernel density rasters were calculated for each fishery to assign a probability of fishing
242 value to each cell. Fishing effort rasters for the shrimp fishery were calculated by creating lines
243 between start and end coordinates of each trawl, and by excluding any trawls that were missing
244 starting or ending coordinates. It is important to note that spatial distribution of shrimping effort
245 can change from year to year and trawling often does not occur in a strictly linear path; however,
246 given the sample size of trawls and the large spatial scale, this method provided an adequate
247 approximation. Trawls were subsampled by month and kernel density rasters with a cell size of

248 0.05 degrees were constructed from the resulting polyline features. For the coastal gillnet fishery,
249 fishing effort rasters were created by subsampling by month and creating kernel density rasters
250 with a cell size of 0.05 degrees from the deployment points. For the longline fishery, the kernel
251 density raster was calculated by using only the starting locations, due to many missing or
252 erroneous ending locations. Data were divided by month and rasters with a cell size of 0.05
253 degrees populated by soak time were created.

254 The relative sawfish-fishery bycatch risk rasters were calculated by multiplying the
255 fishing effort rasters by the sawfish activity rasters to create fishery-specific relative bycatch risk
256 rasters for each month across all years. Bycatch risk is a measure of the probability of a sawfish
257 occurring in the same geographic location that fishing gear is being deployed in any given
258 month. The rasters were normalized and the risk values were assigned to detections in the
259 acoustic dataset for corresponding months using the Extract to Points tool in ArcMap. Average
260 bycatch risk across all individuals was calculated and a series of Kruskal Wallis tests were
261 conducted to analyse the difference in risk across the three fisheries.

262

263 **2.4 Modeling bycatch risk**

264 A linear mixed-effects model, fitted to optimize the Restricted Maximum Likelihood (REML)
265 criterion, was created where the response variable was bycatch risk for a specific fishery (defined
266 above). All possible combinations of the fixed effects stretch total length, sex, and tagging
267 location, were added into the model along with the random effects of individual and month. The
268 change in Akaike information criterion (AICc) values of all potential models for a specific
269 fishery was compared to determine the best model ($\Delta AICc < 2$; Anderson & Burnham, 2002).
270 The AICc comparison was repeated for each of the three fisheries. Because only two sawfish

271 were tagged off the Indian River Lagoon, as compared to 19 in Charlotte Harbor, 10 in
272 Everglades National Park and 11 in the Florida Keys, they were excluded from the model.

273

274 **2.5 Analysis of vertical distribution**

275 Fourteen (7 females and 7 males) of the 17 satellite tags used in this study had viable depth data
276 that could be used for analysis (i.e. daily depth measurements for at least two weeks). Although
277 the maximum number of days depth data were collected on any one tag was 156, this study had
278 coverage across all months when all tags were aggregated. The tags were programmed to record
279 depth readings every 60 s. Data were combined into 4-hr bins distributed in 12 discrete depth
280 bins based on previous vertical distribution data, which were averaged to create histograms
281 showing vertical movement for each sex. Histograms were also made showing vertical space use
282 for each season using data from tags that had depth data for that season. These histograms were
283 compared to seasonal histograms showing fishing depths for each fishery that depth data were
284 recorded for. A linear mixed-effect model fit to maximize REML was run with sex and depth bin
285 as fixed effects, month as a random effect, and percent time as the response variable.

286

287 **3. RESULTS**

288 Fifty-nine large juvenile and adult smalltooth sawfish were tagged in this study. Forty-two were
289 tagged with acoustic tags between 2016 and 2019; 24 were female (mean = 3.13 m STL) and 18
290 were male (mean = 3.09 m STL) (Table 1). Seventeen were tagged with satellite tags between
291 2011 and 2017; 7 were female (mean = 3.43 m STL) and 10 were male (mean = 3.94 m STL)
292 (Table 2). No sawfish were tagged with both tag types.

293

294 **3.1 Acoustic monitoring summary**

295 From May 2016 to September 2019, individuals were detected on 461 acoustic receivers ranging
296 from off the coast of Brunswick, Georgia to the lower Florida Keys and along the Gulf of
297 Mexico to Apalachee Bay, Florida; these receivers were divided into regions (Figure 1; Graham
298 et al., 2021). In general, sawfish moved north from the Keys in spring (March–May) on both
299 Florida coasts and travelled to Charlotte Harbor on the Gulf coast and to Cape Canaveral on the
300 Atlantic coast. Some detections (<1%) were recorded north of these areas in summer (June–
301 August), but most detections occurred south of 27°N latitude on the Gulf coast and south of
302 29°N latitude on the Atlantic coast. Some individuals moved back to the Keys in the fall
303 (September–November) and winter (December–February), while some remained in Charlotte
304 Harbor and the Keys year-round.

305

306 **3.2 Shrimp trawl fishing effort**

307 Shrimp trawl effort varied temporally and spatially within state and federal waters (Figure 2).
308 There was high effort during January, and June through August around the lower Keys and
309 Marquesas Keys, particularly offshore on the Gulf side. There was also high effort between the
310 lower Keys and Charlotte Harbor from January through May and from October through
311 December. On the Atlantic coast, there was high effort off Cape Canaveral during January and
312 north of Cape Canaveral to the Florida-Georgia border in September and November.

313

314 **3.3 Southeast coastal gillnet fishing effort**

315 Southeast coastal gillnet fishing effort occurred in federal waters near Cape Canaveral for most
316 of the year (Figure 2). There was also high effort around the Florida-Georgia border from

317 February through May as well as August. Gulf coast effort was limited to November and
318 December.

319

320 **3.4 Shark bottom longline fishing effort**

321 Longline effort was relatively high year-round in federal waters along both coasts (Figure 2).

322 Gulf coast effort was concentrated in the warmer months and only occasionally extended south
323 of Charlotte Harbor, usually during the winter. On the Atlantic coast, effort was also highest
324 during the warmest months, but extended further south than the Gulf coast to the Florida Keys
325 almost year-round.

326

327 **3.5 Bycatch risk**

328 Bycatch risk for each fishery was examined seasonally (Figure 3). For the shrimp trawl fishery,
329 risk was concentrated year-round off the Gulf side of the lower Florida Keys and Marquesas
330 Keys. Gillnet risk was concentrated off Cape Canaveral for most of the year, but negligible in
331 winter and early spring because the sawfish were overwintering in the Florida Keys during this
332 time. Risk for the longline fishery was concentrated year-round in the Atlantic Ocean adjacent to
333 the Florida Keys. Risk associated with the shrimp trawl fishery was significantly higher than risk
334 associated with the coastal gillnet fishery (Kruskal-Wallis test, $P < 0.001$, $\chi^2 = 4542.5$, $df = 36$)
335 or the longline fishery (Kruskal-Wallis test, $P < 0.001$, $\chi^2 = 68.14$, $df = 305$). Risk for the
336 longline fishery was significantly higher than the gillnet fishery (Kruskal-Wallis test, $P < 0.001$,
337 $\chi^2 = 51810$, $df = 210$).

338

339 **3.6 Modelling bycatch risk**

340 A linear mixed effects model was used to account for individual variation in bycatch risk and
341 determine if there was variation across months. The best fitting models from all three fisheries
342 included sex \times tagging location, length, and the random effects individual and month (Table 3).
343 All three fixed effects variables were included in the best fitting model as well as the interaction
344 between sex and tagging location.

345

346 *3.6.1 Shrimp trawl fishery*

347 Both male and female sawfish tagged in Charlotte Harbor had the highest shrimp trawl bycatch
348 risk, with the risk for males slightly higher (Figure 4). This is likely because all sawfish leaving
349 and returning to this estuary swim through an area that has a high concentration of shrimp trawl
350 effort. Risk was relatively low for sawfish tagged in Everglades National Park, including Florida
351 Bay, and this risk was comparable between sexes. The random effect month showed that
352 February, June, July, and August had higher than average risk. Trawl risk in October was not
353 significantly different from February or June (Tukey, $P = 0.79$, $P = 0.14$), but was significantly
354 higher than all other months (Tukey, $P < 0.02$). February, March, June, and July were not
355 significantly different from each other (Tukey, $P = 1.0$, $P = 0.90$, $P = 0.08$), but risk in February
356 was significantly higher than January, April, May, August, September, November, and
357 December (Tukey, $P < 0.05$). Risk in June was significantly higher than September and August
358 (Tukey, $P = 0.04$, $P = 0.03$). Although risk was higher than average in July, there was no
359 significant difference between risk in July and risk associated with any months with lower-than-
360 average risk (Tukey, $P > 0.30$).

361

362 *3.6.2 Southeast coastal gillnet fishery*

363 Sawfish tagged in the Florida Keys had the highest bycatch risk from the southeast coastal gillnet
364 fishery, with slightly higher risk for females (Figure 4). Sawfish tagged in Charlotte Harbor,
365 Everglades National Park, including Florida Bay, had negligible risk in this fishery because these
366 fish did not travel along the Atlantic coast where this fishery occurs. April, May, June, July,
367 September, November, and December had gillnet bycatch risk and there was no significant
368 difference between these months (Tukey, $P > 0.42$).

369

370 *3.6.3 Shark bottom longline fishery*

371 Average longline bycatch risk was highest for both males and females tagged in the Florida
372 Keys, with both sexes having comparable risk (Figure 4). Risk in this fishery was low for both
373 males and females tagged in Charlotte Harbor and risk was comparable between sexes. Risk was
374 higher for females tagged in Everglades National Park. Males tagged in Florida Bay had slightly
375 higher risk than females. When examining the random effect of month, February, March,
376 November, and December had higher than average risk. December and February had
377 significantly higher risk than all other months except November and March (Tukey, $P < 0.01$).
378 Although November and March had higher than average risk, this risk was not significantly
379 higher than any months with below average risk (Tukey, $P > 0.06$).

380

381 **3.7 Modelling vertical distribution**

382 It is important to consider both the depth that fishing gear is deployed and the depths that sawfish
383 most commonly occupy when assessing bycatch risk. Although sawfish are benthic, they exhibit
384 preferences for areas of certain depths. Therefore, a model was created to analyse the vertical
385 distribution of sawfish activity (Table 2). Percentage time at depth was calculated to examine

386 how the sexes moved along depth gradients and to model the time each sex spent at various
387 depths. Sex was a good predictor of the percentage of time spent at depth (Table 4, Figure 5).
388 Females spent the most time in 0–2 m and 30–100 m depth ranges. Males spent the most time in
389 0–2 m and 30–40 m. Both sexes spent a high percentage of time in the 0, 30 and 40 m depth
390 ranges and a low percentage of time in the 4 and 8 m ranges. Although females spent a high
391 percentage of time at about 100 m, males spent less time at this depth.

392 When analysing the vertical distribution of sawfish and the deployment depth of the gear,
393 it became clear that while bycatch risk for females was highest in the shrimp trawl fishery, risk
394 was not significantly different between the sexes in the other two fisheries (Figure 6). Both sexes
395 spent most of their time in the extremes of their vertical range, remaining either very shallow or
396 venturing deep, though females tended to venture deeper than males. Shrimp trawl effort was
397 highest at depths greater than 100 m and bycatch risk was highest for females that spent more
398 time at these depths than males. Gillnet fishing effort occurred mostly between 4 and 30 m for
399 both sexes and risk was highest between 20 and 30 m. Most of the longline fishing effort
400 occurred between 10 and 30 m and this is also where bycatch risk was highest.

401 We observed elevated bycatch risk for females in the shrimp fishery across seasons
402 (Figure 7). Although the risk was comparable between sexes for the remaining fisheries, risk
403 fluctuated throughout the year. Most of the shrimp trawling effort occurred at depths of 20 m or
404 more, which more heavily affected females. Risk in the shrimp fishery was highest in summer
405 and fall. Risk was highest in spring and summer for the coastal gillnet fishery. Risk in the
406 longline fishery was lowest in fall.

407

408 **4. DISCUSSION**

409

410 **4.1 Implications for Management**

411 This study identifies the spatial and temporal overlap between commercial fishery effort and
412 large juvenile and adult smalltooth sawfish occurrence. Areas and times of overlap represent
413 areas of increased bycatch risk and identify specific locations and times for resource managers to
414 implement conservation measures. Results illustrate minimal overlap in the southeast coastal
415 gillnet fishery, temporally-limited overlap in the shark bottom longline fishery (4 of 12 months),
416 and substantial overlap in the shrimp trawl fishery—both temporally (9 of 12 months) and
417 spatially. Given limited overlap of the southeast coastal gillnet and shark bottom longline
418 fisheries with sawfish occurrence, additional regulations do not appear necessary for these
419 fisheries at this time. In contrast, conservation measures to mitigate bycatch risk in the shrimp
420 trawl fishery appear necessary to promote conservation of this species. Results from this study
421 indicate a year-round closure of waters off southwest Florida to the southeast shrimp trawl
422 fishery between Charlotte Harbor and the western Florida Keys (Figure 8) is warranted to ensure
423 bycatch does not cause population decline.

424 Of the three fisheries examined, the shrimp trawl fishery is most likely to result in both
425 bycatch and mortality of large juvenile and adult smalltooth sawfish. Although uncertainty was
426 very high, in a recent assessment of the shrimp trawl fishery’s effect on smalltooth sawfish,
427 NMFS determined that 1,806 sawfish could be taken as bycatch in this fishery, with 50% of
428 those resulting in mortality, over any running 5-year period (NMFS, 2021). These figures were
429 estimated using current NMFS observer data and estimates of total effort from this fishery.
430 Unfortunately low levels of observer coverage (1–2%) result in high levels of uncertainty, as

431 annual captures from 2008 to 2010 were estimated to be as low as 17 or as high as 162 animals
432 per year (Carlson & Scott-Denton, 2011). Because the assessment based the bycatch value on the
433 highest capture estimate (162 sawfish), it represents a worst-case scenario. To more accurately
434 understand the effect of this fishery on smalltooth sawfish increased observer coverage,
435 especially in high-risk regions, and more information on total fishing effort is needed. Increased
436 observer coverage combined with tagging efforts of released animals could refine bycatch
437 estimates and provide data on post-capture survivorship.

438 Traditionally, fishery observations have been conducted by trained people onboard
439 vessels. However, increasing observer coverage to refine bycatch estimates can be costly,
440 especially for rare captures like smalltooth sawfish. Electronic monitoring techniques, including
441 the use of cameras, are improving and increasingly replacing human observers in some
442 circumstances. For sawfish, electronic monitoring may be a cost-effective complement to
443 onboard observers to help achieve sufficient coverage associated with bycatch reduction goals
444 (Moncrief-Cox et al., 2020).

445 As mentioned, sawfish rostra are easily entangled in nets and are often difficult to
446 disentangle. With shrimp trawl nets, risk to sawfish is exacerbated by relatively long tow times
447 (four hours on average) that result in sawfish being dragged for extended periods. Because of
448 these factors, shrimp trawls have substantially higher sawfish mortality rates than other gears,
449 including hooks and even stationary nets that don't drag the sawfish and allow for faster release.
450 Further study is needed to determine the extent to which tow time restrictions coupled with safe
451 release methods could increase post-release survivorship of sawfish and to evaluate the potential
452 for such measures to facilitate recovery.

453 Bycatch risk varied throughout the year with some months and specific areas having
454 higher associated risks than others. This variation opens the possibility of time-area or seasonal
455 closures. There is evidence that such closures can be an effective management strategy in
456 mitigating bycatch in commercial fisheries with minimal effect on the fisheries (NMFS, 2003;
457 O’Keefe, Cadrin & Stokesbury, 2014). One such success was a closure instituted in the Kuwait
458 shrimp fishery, which significantly decreased bycatch such as sea turtles and marine mammals
459 with a minimal loss of target catch (O’Keefe, Cadrin & Stokesbury, 2014). Closures have also
460 been implemented to assist recovery of other elasmobranch species. For example, a seasonal
461 closure off North Carolina was implemented to protect juvenile dusky (*Carcharhinus obscurus*)
462 and sandbar sharks (*Carcharhinus plumbeus*) (NMFS, 2003). However, closures can cause
463 negative socio-economic impacts on fishers or relocate the problem to another area as fishing
464 efforts shift (O’Keefe, Cadrin & Stokesbury, 2014). Therefore, it is important that managers
465 consider the overlap between target taxa (e.g. shrimp aggregations) and sawfish movements to
466 understand how fishing effort displacement could affect the overall sawfish population.

467 **4.2 Additional Considerations**

468 It is important to address caveats associated with the relative bycatch risk metric and the
469 statistical model used in this study. The sawfish activity raster was driven mostly by positive
470 acoustic data, which are highly dependent on receiver coverage. Therefore, activity estimates
471 were biased towards areas with higher receiver coverage. The satellite tag data may also be
472 biased due to the uneven distribution of tagged males and females; though, by combining these
473 two methods, these biases may have been minimized. Also, the relative risk metric is an estimate
474 of bycatch likelihood and does not necessarily equate to capture or mortality risk. It simply
475 represented the probability that a sawfish was in an area during a given month, multiplied by the

476 probability of fishing occurring in that area during that month. There are other factors that could
477 contribute to whether bycatch occurs, including time of day, tidal cycle, depth of gear
478 deployment, and gear-specific catchability, which were not accounted for. In addition, the
479 differing temporal scales between the fishing effort data and the sawfish activity data was also a
480 source of potential bias. However, we believe the relative risk metric served as an adequate
481 proxy to assess areas that were of highest risk to sawfish even if the true value of that risk was
482 unknown. It is also useful for modelling purposes to determine which sawfish are spending the
483 most time in these high-risk areas and are therefore most likely to interact with the fisheries.

484 Notably, the size distributions of sawfish tagged in Charlotte Harbor, Florida Bay, and
485 the Florida Keys differed. Sawfish tagged in Charlotte Harbor tended to be smaller than the
486 sawfish tagged in the Florida Keys or Florida Bay. There is evidence of ontogenetic shifts in
487 space use, so this skew in size class may have biased the data. However, sawfish larger than two
488 meters STL move from the shallowest waters of the nurseries along mangrove shorelines into
489 deeper waters (> 3 m) in Charlotte Harbor (Poulakis G. R., unpublished data). Thus, the sawfish
490 tagged in Charlotte Harbor spent more time within the estuary and did not move around as much.
491 For this reason, bycatch risk differed between Charlotte Harbor and areas further south.

492 There was a significant difference in movement and associated bycatch risk between
493 males and females depending on where they were tagged. In general, individuals tagged in
494 Charlotte Harbor did not move as much as those tagged in south Florida, but both sexes tagged in
495 Charlotte Harbor had the highest shrimp trawl bycatch risk, with the risk for males being slightly
496 higher. Large females tagged in south Florida tended to reside in the deepest water, which is
497 where shrimp trawl effort was highest. Therefore, females may be more vulnerable than males in
498 the southernmost portions of Florida. We recommend that these sex-specific analyses be

499 revisited as more fish are tagged and analysed, more years of acoustic data are received from the
500 10-yr tags that have been deployed, and sex data are recorded from sawfish caught in shrimp
501 trawls. Consistent funding is needed for acoustic tags, fisheries-independent and fisheries-
502 dependent (e.g. NMFS observers; electronic monitoring) sampling, as well as continuation and
503 expansion of acoustic monitoring, especially in the proposed shrimp trawling closure area.

504 To promote recovery of the smalltooth sawfish population, bycatch fishing mortality rates
505 need to be minimized (NMFS, 2009b). A population viability analysis found that population
506 growth remained stable at low levels (19 females per year) of fishing mortality but, not
507 surprisingly, when fishing mortality levels increased, population growth declined (Carlson &
508 Simpfendorfer, 2015). Increasing observer coverage and acquiring more bycatch and
509 survivability data for sawfish in these fisheries, especially the shrimp trawl fishery, would help
510 managers focus future conservation measures. Regardless, management tools such as the
511 proposed area closure are warranted to mitigate bycatch mortality in the shrimp trawl fishery
512 now. The current study provides baselines for determining which areas and times are of highest
513 risk to sawfish. This information will prove useful as policy makers continue to monitor the
514 smalltooth sawfish population and assess threats to recovery from various fisheries. With
515 effective management practices, the smalltooth sawfish population can grow to eventually reach
516 a healthy population size and expand to its historic range.

517

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682 **TABLE 1** Summary of all acoustic tagged smalltooth sawfish (*Pristis pectinata*) including ID
683 number, sex (F= female, M= male), maturity, stretch total length, tagging location, date tagged
684 (dd/mm/yyyy), date of first detection, date of last detection, days of study, and number of
685 detections. CH = Charlotte Harbor, PP = St. Lucie Power Plant, ENP = Everglades National
686 Park, Keys = Florida Keys

ID	Maturity	Length (m)	Location tagged	Date tagged	Date of first detection	Date of last detection	Days of study	Number of detections
F1	Immature	2.12	CH	15/03/2019	15/03/2019	18/09/2019	188	4639
F2	Immature	2.13	CH	02/08/2018	03/08/2018	03/10/2019	427	31954
F3	Immature	2.16	Keys	10/08/2017	11/03/2018	26/06/2018	108	35
F4	Immature	2.25	ENP	20/06/2016	05/01/2018	04/02/2018	31	53
F5	Immature	2.27	CH	15/03/2019	15/03/2019	14/09/2019	184	2138
F6	Immature	2.34	CH	19/07/2017	21/07/2017	25/06/2019	705	3543
F7	Immature	2.38	CH	25/03/2019	25/03/2019	20/09/2019	180	3808
F8	Immature	2.43	CH	09/07/2018	09/07/2018	10/09/2019	64	8768
F9	Immature	2.46	CH	26/07/2017	26/07/2017	26/04/2018	275	1246
F10	Immature	2.57	CH	26/07/2017	26/07/2017	23/07/2018	363	1927
F11	Immature	2.58	CH	20/03/2019	20/03/2019	29/07/2019	132	5381
F12	Immature	2.69	CH	12/09/2018	12/09/2018	26/12/2018	106	157
F13	Immature	3.18	ENP	30/03/2017	16/11/2017	22/06/2019	584	864
F14	Immature	3.20	CH	11/08/2017	15/08/2017	21/05/2019	645	1940
F15	Immature	3.49	Keys	01/08/2018	27/08/2018	08/06/2019	286	166
F16	Immature	3.55	Keys	11/04/2017	16/04/2017	03/02/2018	294	1279
F17	Mature	3.64	Keys	11/04/2017	27/04/2017	28/03/2019	701	4913
F18	Mature	3.71	PP	02/11/2017	23/11/2017	10/04/2019	504	2069
F19	Mature	3.92	Keys	01/04/2017	01/04/2017	25/05/2019	785	755
F20	Mature	4.26	Keys	01/04/2017	03/04/2017	28/05/2019	786	610
F21	Mature	4.38	Keys	21/05/2016	21/05/2016	01/06/2019	1107	3122
F22	Mature	4.38	ENP	13/09/2016	05/11/2016	04/04/2019	881	1548
F23	Mature	4.42	ENP	02/04/2017	12/05/2017	19/03/2019	677	791
F24	Mature	4.53	ENP	02/04/2017	04/06/2017	06/06/2017	3	27
M1	Immature	2.11	CH	04/06/2018	05/06/2018	27/03/2019	296	5769
M2	Immature	2.35	CH	21/08/2018	21/08/2018	18/09/2019	394	10288

M3	Immature	2.35	CH	26/07/2017	26/07/2017	19/04/2019	633	3118
M4	Immature	2.48	CH	21/08/2018	21/08/2018	14/09/2019	390	12509
M5	Immature	2.59	ENP	09/11/2016	21/01/2018	16/06/2019	512	277
M6	Immature	2.60	CH	23/10/2018	23/10/2018	24/04/2019	184	237
M7	Immature	2.66	CH	18/04/2019	18/04/2019	26/09/2019	162	2615
M8	Immature	2.72	ENP	30/03/2017	26/04/2017	08/06/2019	774	10337
M9	Immature	2.76	CH	24/10/2017	19/07/2017	22/04/2019	643	919
M10	Immature	2.90	CH	12/09/2018	12/09/2018	22/04/2019	223	2229
M11	Immature	2.93	Keys	20/07/2016	22/08/2016	10/06/2019	74	4284
M12	Mature	3.50	PP	17/09/2017	24/09/2017	12/08/2018	323	638
M13	Mature	3.82	ENP	06/04/2019	10/04/2019	15/06/2019	67	25
M14	Mature	3.83	Keys	01/04/2017	01/04/2017	17/06/2019	808	689
M15	Mature	3.98	Keys	15/04/2018	14/02/2018	30/11/2018	290	382
M16	Mature	3.98	ENP	02/04/2017	26/04/2017	07/04/2019	712	388
M17	Mature	3.98	ENP	09/09/2016	12/12/2016	28/05/2019	898	2414
M18	Mature	4.07	Keys	14/04/2017	15/04/2017	02/07/2017	79	1143

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688

689 **TABLE 2** Summary of all satellite tagged smalltooth sawfish (*Pristis pectinata*) including
 690 identification number (ID), sex (F= female, M= male), and stretch total length

ID	Maturity	Length (m)	Used in bycatch analysis	Depth days analyzed
F25	Immature	2.79	No	141
F26	Immature	2.83	No	133
F27	Immature	3.23	No	138
F28	Immature	3.52	Yes	156
F29	Mature	3.68	Yes	84
F30	Mature	3.68	Yes	140
F31	Mature	4.28	Yes	121
M19	Mature	3.65	Yes	N/A
M20	Mature	3.66	Yes	N/A
M21	Mature	3.71	Yes	141
M22	Mature	3.95	Yes	61
M23	Mature	3.95	Yes	62
M24	Mature	3.99	Yes	46
M25	Mature	4.03	Yes	N/A

M26	Mature	4.09	Yes	55
M27	Mature	4.12	Yes	150
M28	Mature	4.27	Yes	151

691
692 **TABLE 3** The two best-fitting bycatch risk models for each fishery with rank, number of
693 parameters (K), $\Delta AICc$, cumulative weight, and model formula. All models include the random
694 effects month and individual

Rank	K	$\Delta AICc$	Cumulative weight	Model
Shrimp trawl				
1	13	0.00	0.98	Av_Risk ~ Sex × Tagging location
2	14	8.11	1.00	Av_Risk ~ Sex × Tagging location + Length
Southeast coastal gillnet				
1	13	0.00	0.89	Av_Risk ~ Sex × Tagging location
2	14	4.28	1.00	Av_Risk ~ Sex × Tagging location + Length
Shark bottom longline				
1	13	0.00	0.98	Av_Risk ~ Sex × Tagging location
2	14	7.37	1.00	Av_Risk ~ Sex × Tagging location + Length

695

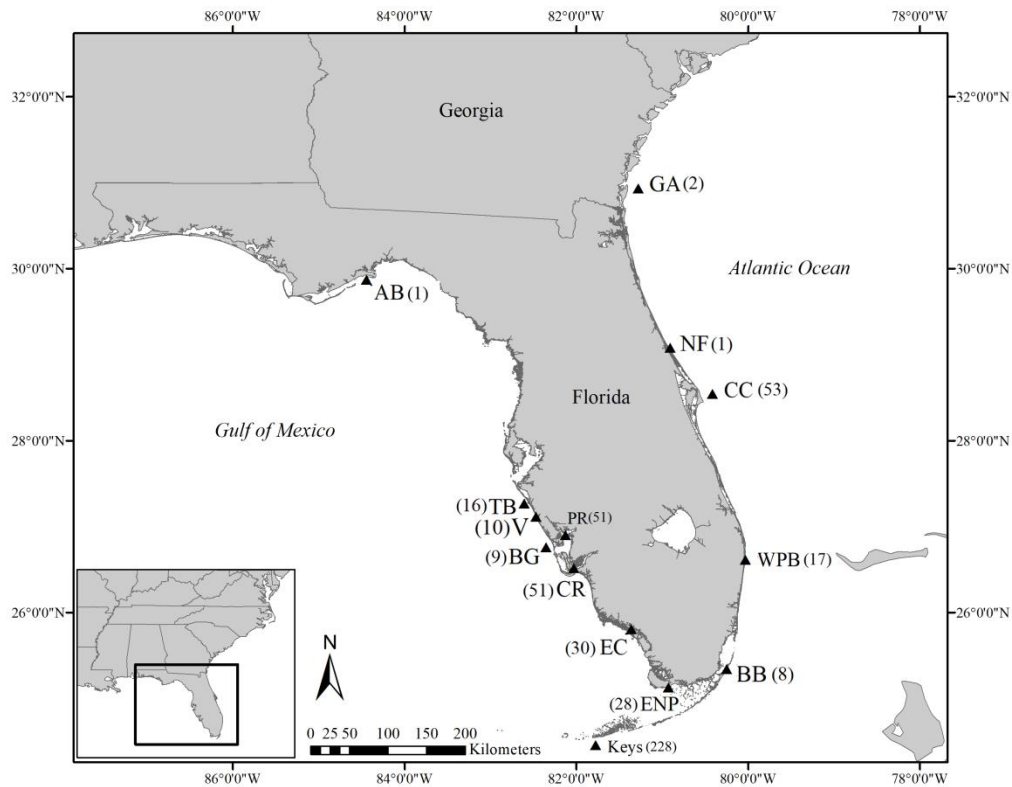
696

697

698 **TABLE 4** The two best models for predicting smalltooth sawfish (*Pristis pectinata*) percent time
 699 at depth with number of parameters (K), $\Delta AICc$, cumulative weight, and model formula

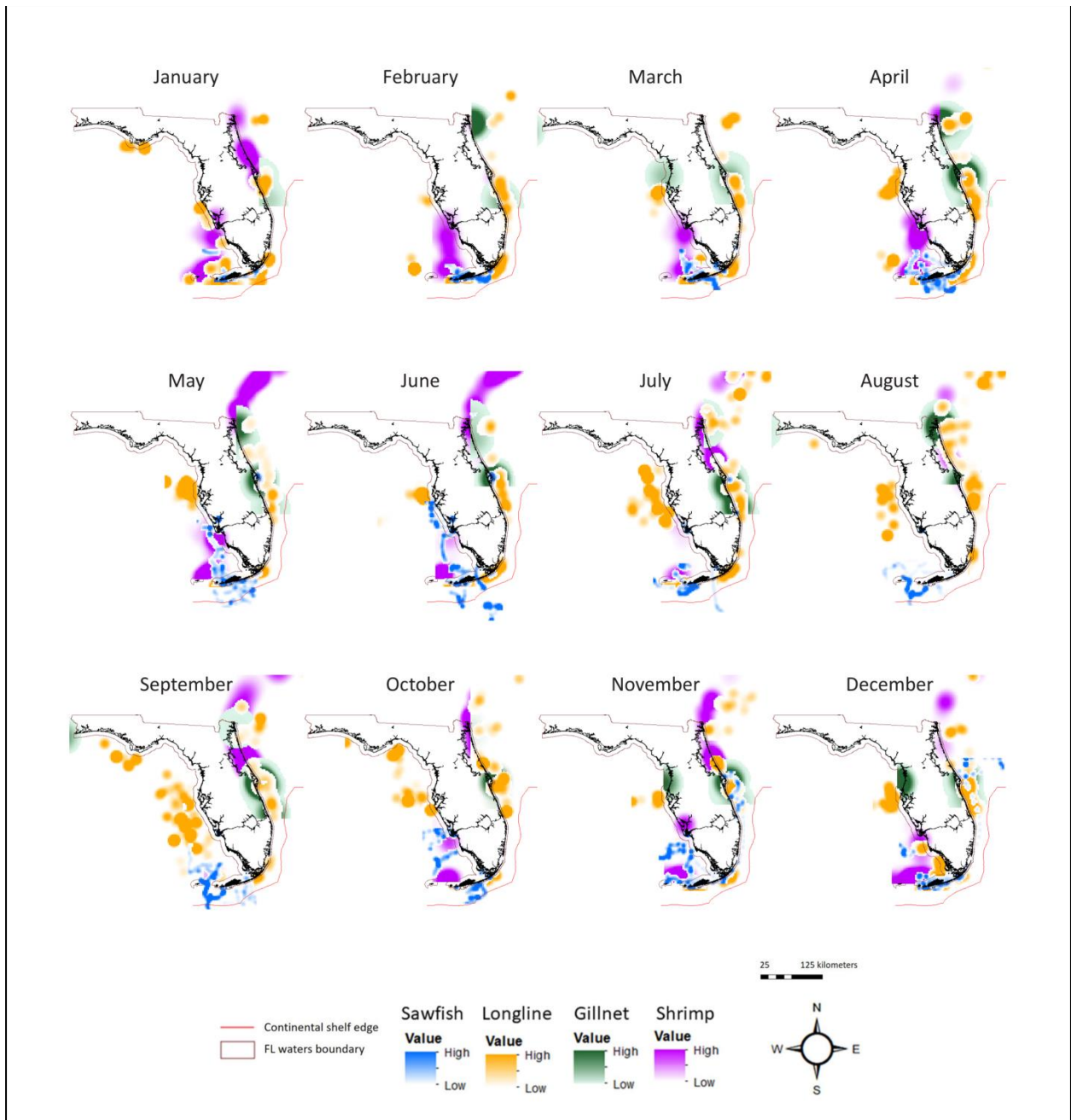
Rank	K	$\Delta AICc$	Cumulative weight	Model
1	26	0.00	1	Percent Time ~ Sex \times Bin + (1 Month)
2	15	45.65	1	Percent Time ~ Sex + Bin + (1 Month)

700
 701



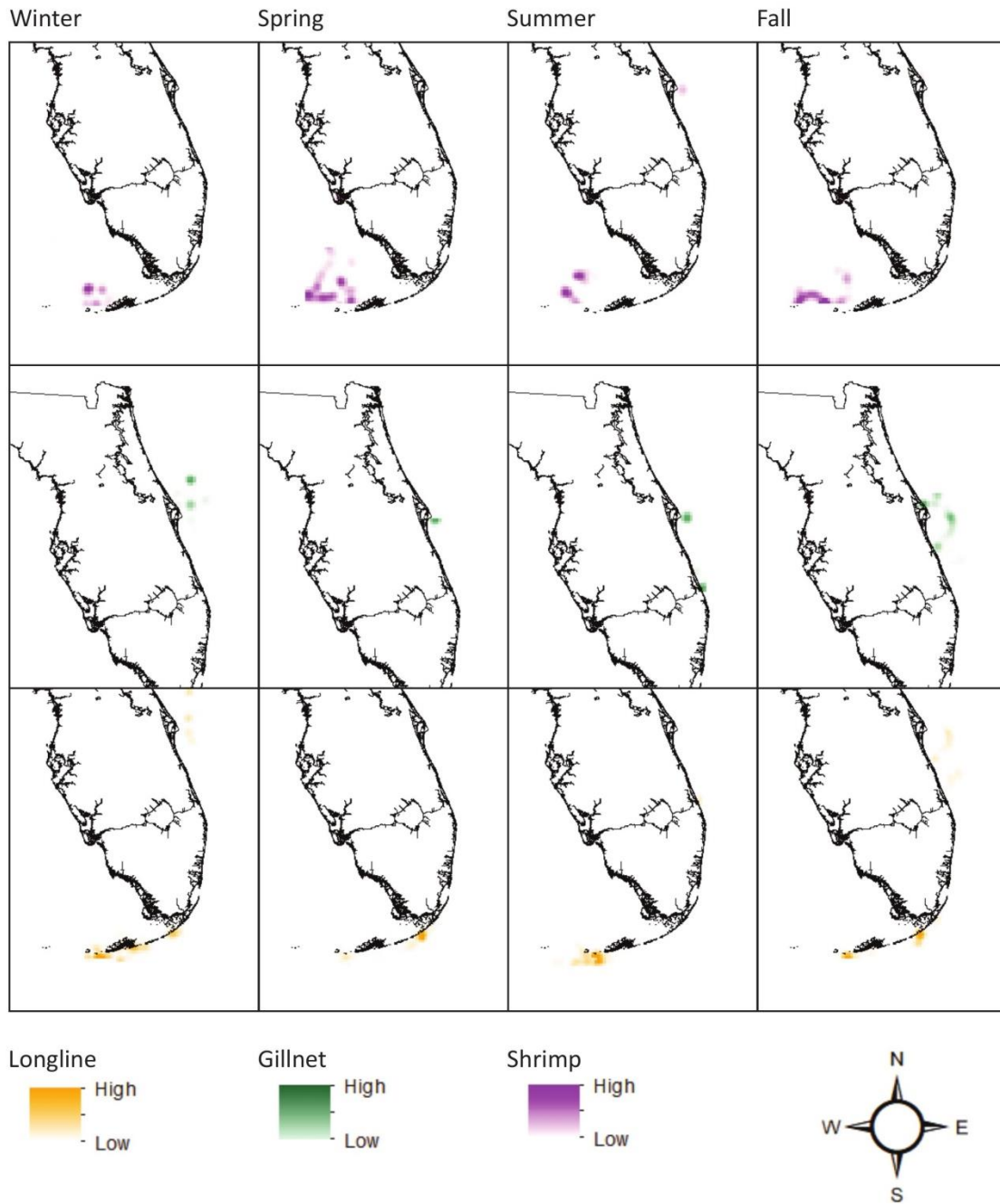
702
 703 **FIGURE 1** Map showing the center of activity for each acoustic receiver region: Apalachee Bay
 704 (AB), Tampa Bay (TB), Venice (V), Peace River (PR), Caloosahatchee River (CR), Boca
 705 Grande (BG), Everglades City (EC), Everglades National Park (ENP), the Florida Keys (Keys),
 706 Biscayne Bay (BB), West Palm Beach (WPB), Cape Canaveral (CC), North Florida (NF),

707 Georgia (GA). The Peace River and Caloosahatchee River regions make up the Charlotte Harbor
708 estuarine system. The number of receivers in each region is shown in parentheses



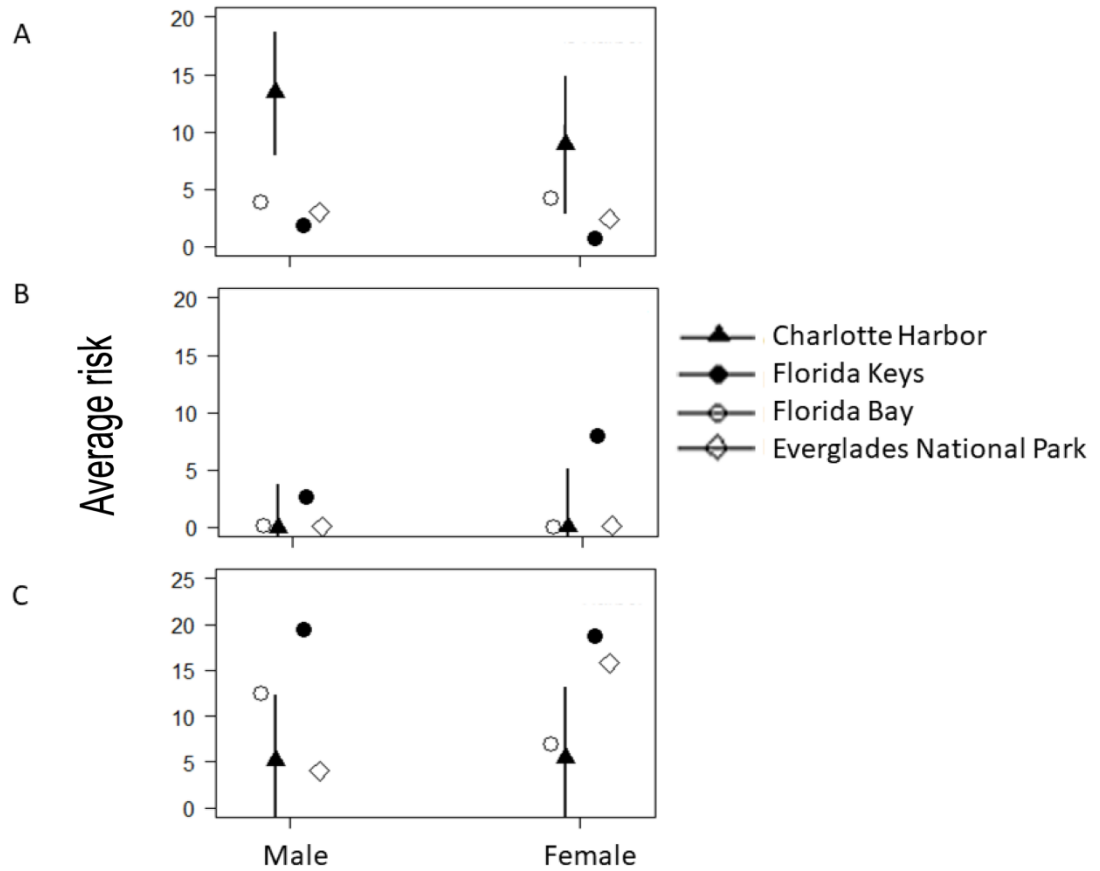
709

710 **FIGURE 2** Smalltooth sawfish (*Pristis pectinata*) activity (blue) and fishing effort rasters for all
 711 three commercial fisheries. The edge of the continental shelf and the state-federal waters
 712 boundary are shown for reference



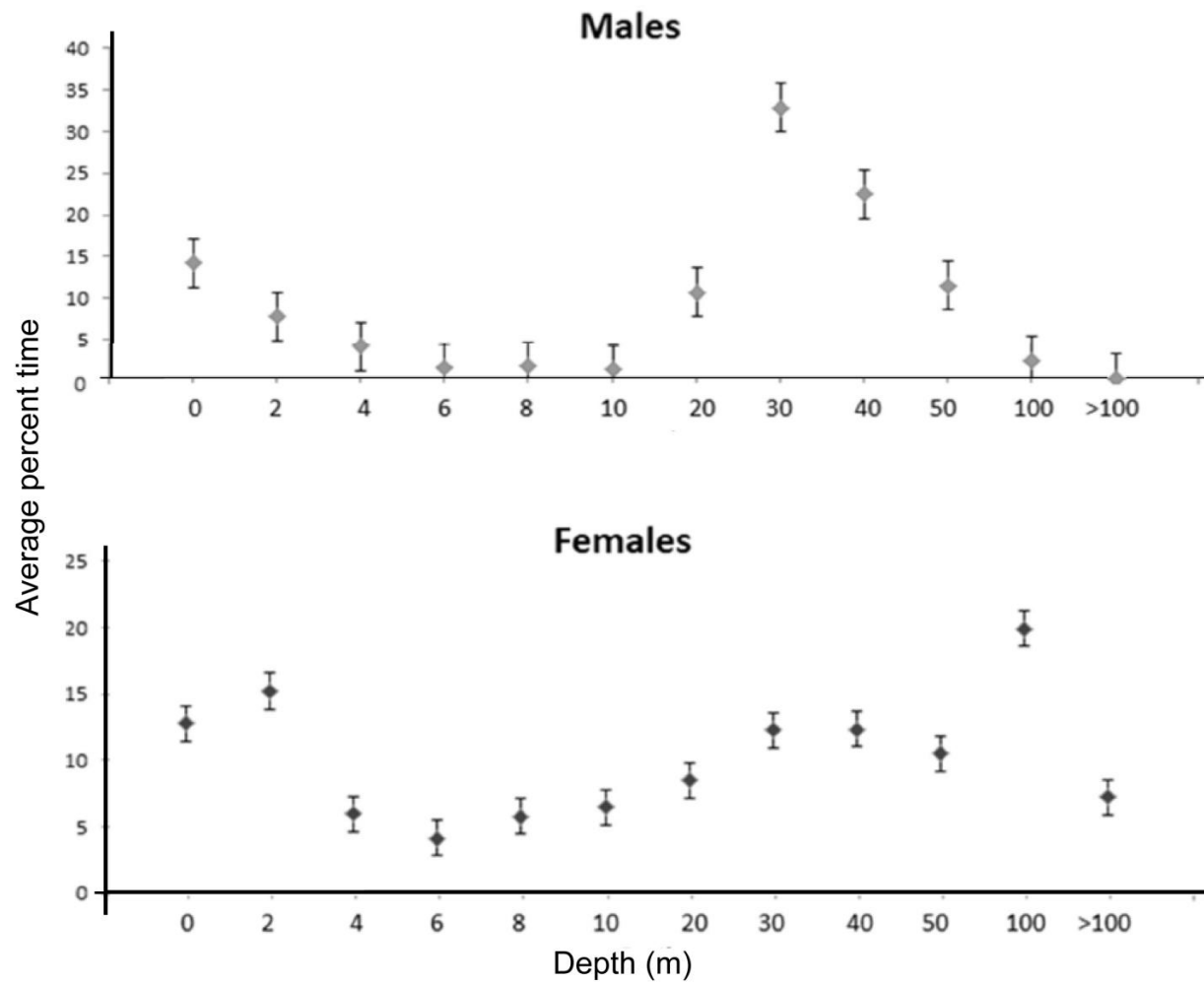
713

714 **FIGURE 3** Shrimp trawl (top row), southeast coastal gillnet (middle row), and shark
 715 longline (bottom row) bycatch risk rasters by season. Darker shades represent higher risk



716

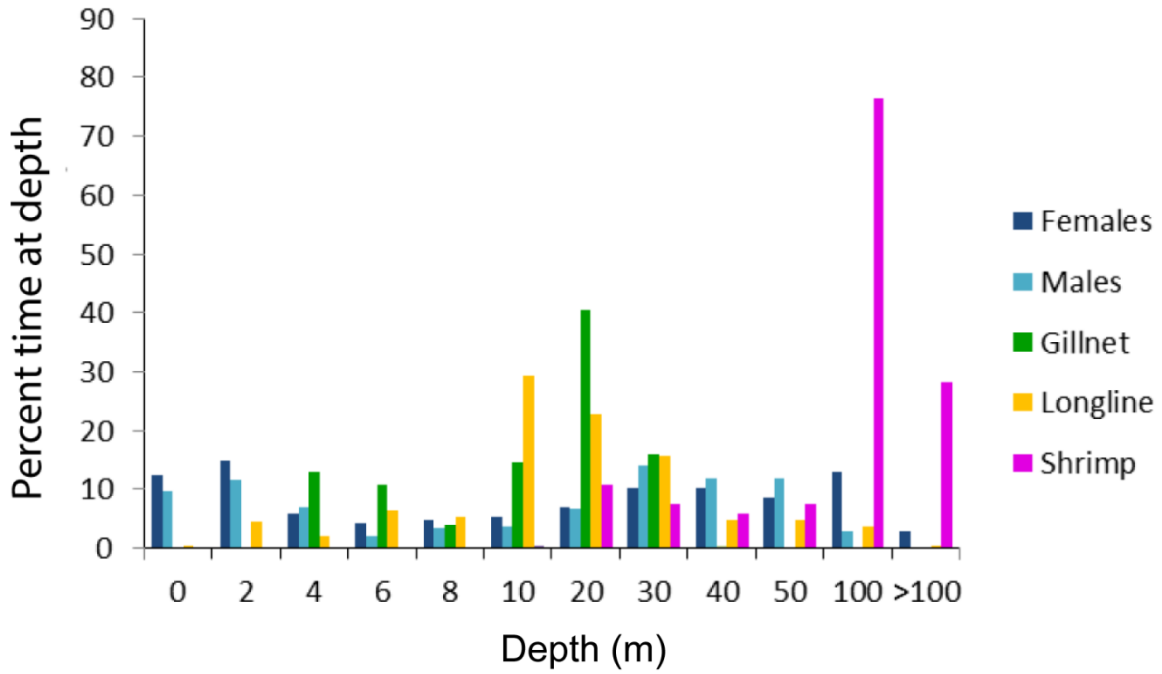
717 **FIGURE 4** Average (A) shrimp trawl, (B) southeast coastal gillnet, and (C) shark bottom
 718 longline bycatch risk as a relative percent probability by sex for acoustic tagged smalltooth
 719 sawfish (*Pristis pectinata*). Bycatch risk was calculated by multiplying the probability of fishing
 720 occurring by the probability of a sawfish occurring in the same area



721

722 **FIGURE 5** Mean percent time (with standard error bars) spent by smalltooth sawfish (*Pristis*
 723 *pectinata*) at 12 depth bins by sex. Note difference in y-axis scales

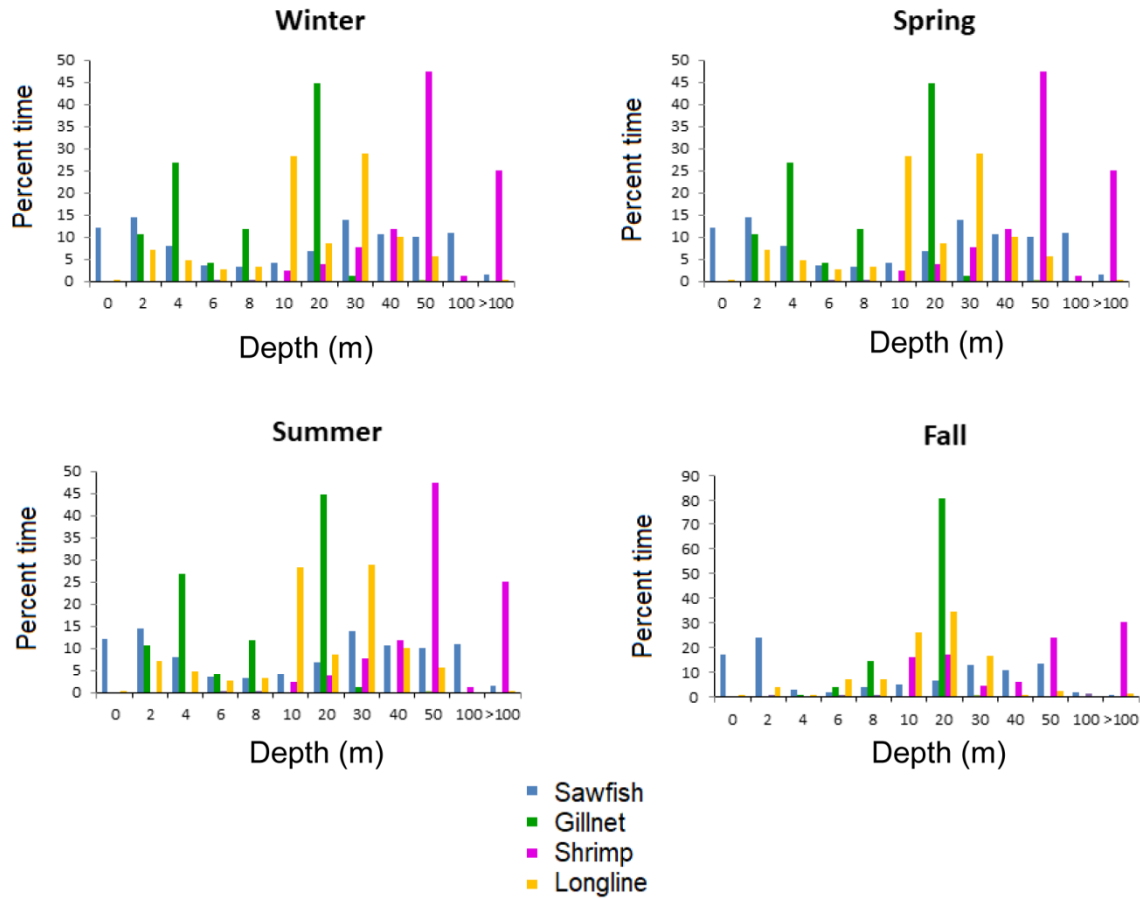
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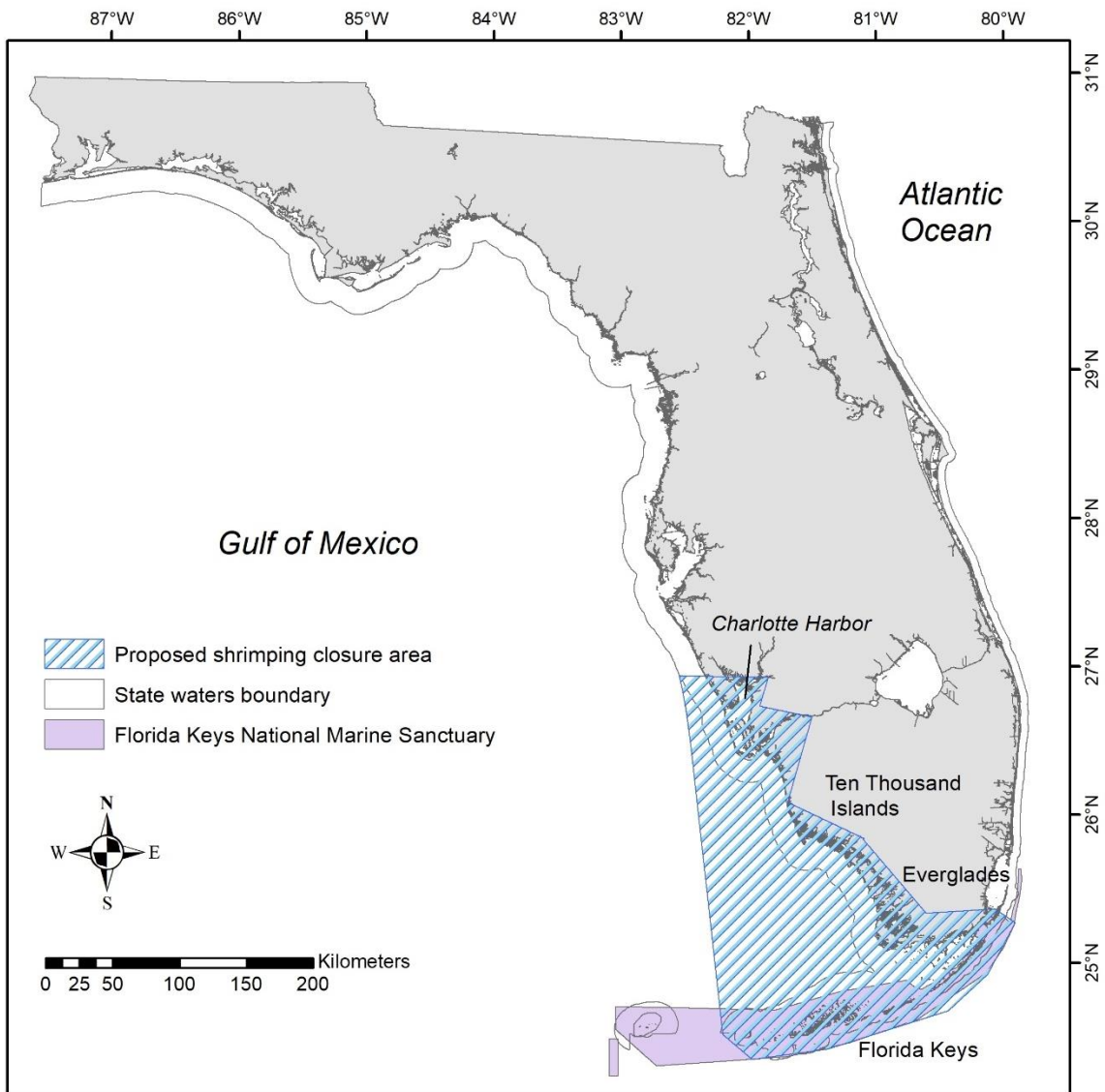
726 **FIGURE 6** Percent time at depth by smalltooth sawfish (*Pristis pectinata*) sex and fishing effort

727 in the shrimp trawl, southeast coastal gillnet, and shark bottom longline fisheries



728

729 **FIGURE 7** Smalltooth sawfish (*Pristis pectinata*) percent time at depth (blue) with shrimp trawl
 730 (purple), southeast coastal gillnet (green), and shark bottom longline (yellow) percent time spent
 731 fishing at depth. Winter = December–February; spring = March–May; summer = June–August;
 732 fall = September–November. Note change in y-axis scale on fall graph



733
 734 **FIGURE 8** Proposed year-round closure area for the shrimp trawl fishery based on our analysis
 735 of where and when large juvenile and adult smalltooth sawfish (*Pristis pectinata*) would most-
 736 likely interact with the fishery