

1 **Bias Correction of Hydrologic Projections Strongly Impacts Inferred Climate Vulnerabilities in**
2 **Institutionally Complex Water Systems**

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11

12 **Abstract**

13 Water-resources planners use regional water management models (WMMs) to identify vulnerabilities to
14 climate change. Frequently, dynamically downscaled climate inputs are used in conjunction with land
15 surface models (LSMs) to provide hydrologic streamflow projections, which serve as critical inputs for
16 WMMs. Here, we show how even modest projection errors can strongly affect assessments of water
17 availability and financial stability for irrigation districts in California. Specifically, our results highlight
18 that LSM errors in projections of flood and drought extremes are highly interactive across timescales,
19 path-dependent, and can be amplified when modeling infrastructure systems (e.g., misrepresenting
20 banked groundwater). Common strategies for reducing errors in deterministic LSM hydrologic
21 projections (e.g., bias correction) can themselves strongly distort projected climate vulnerabilities and
22 misrepresent their inferred financial consequences. Overall, our results indicate a need to move beyond
23 standard deterministic climate projection and error management frameworks that are dependent on single
24 simulated climate change scenario outcomes.

25

26 **Introduction**

27

28 The planning and management of water resources depend heavily on projections of water supply and
29 demand (Loucks and van Beek 2017; Wurbs 1995), strongly shaping water infrastructures and institutions
30 (Malek et al. 2018; Trindade et al. 2019; Yoder et al. 2017). The challenge of infrastructure investment
31 for climate adaptation represents a balance between financial stability and the capacity to meet system
32 demands (Baum et al. 2018; Trindade et al. 2019). Moreover, governments often confront high economic
33 costs, political contention, and social conflicts (Gizelis and Wooden 2010; Petersen-Perlman et al. 2017)
34 as they seek to change water related infrastructures or institutions. These factors promote institutional
35 inertia that favors reactive, post-event responses. Ignoring projections can lead to mal-adaptive and
36 myopic actions that ultimately reduce our ability to respond to changes and reduce the vulnerability of
37 water-dependent sectors to stressors (Lamontagne et al. 2019). Projections of future water resource
38 availability can also shape the perceptions of farmers, irrigation district managers, and water and power
39 utilities about their individual vulnerabilities to climate change, therefore influencing local investment
40 and water-stress hedging decisions (Mase et al. 2017; Mills et al. 2016; Udmale et al. 2014).

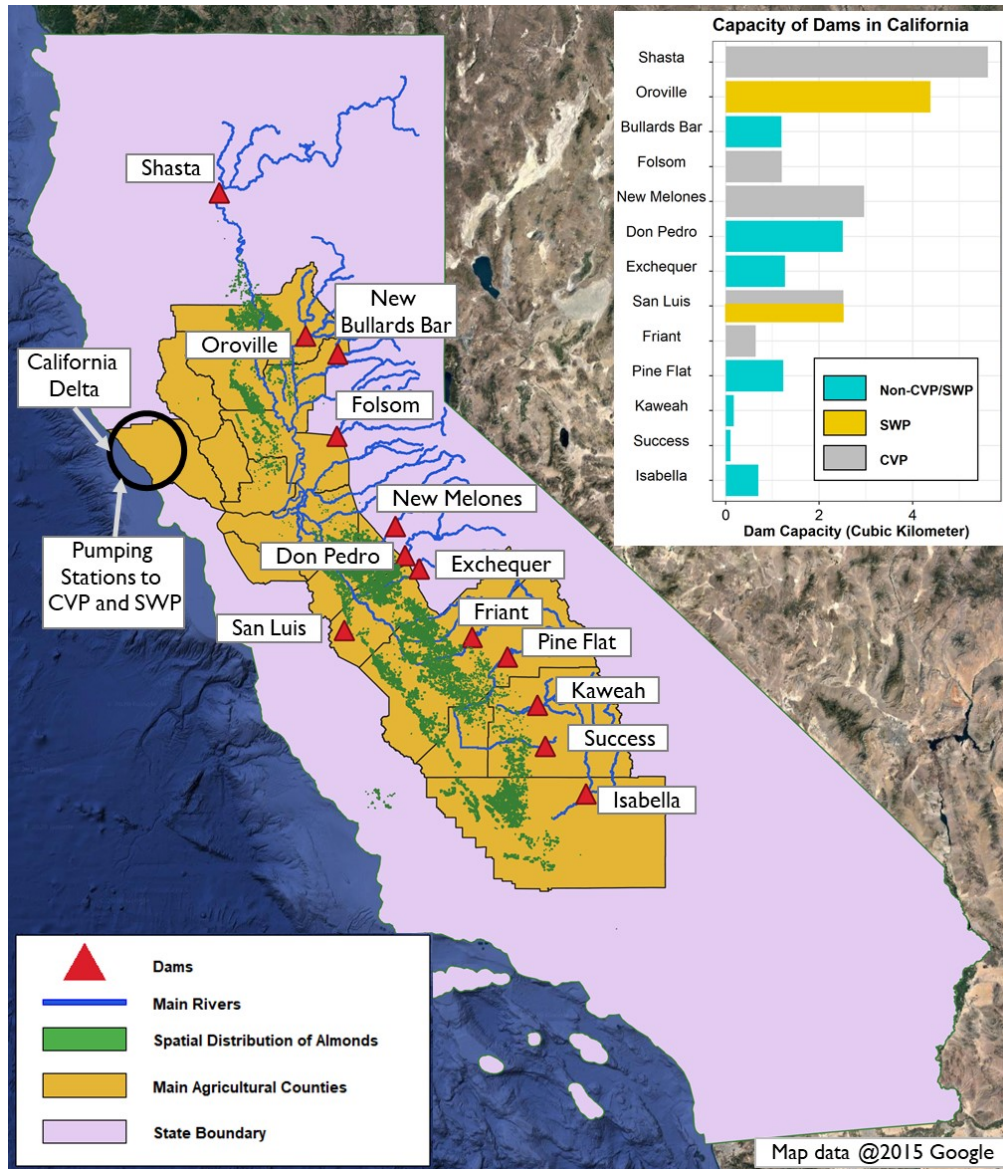
41 Typical model-driven projections of water supply vulnerabilities to climate change consists of: (i)
42 dynamically downscaling climate projections to inform simulation of unregulated streamflow that enters
43 river systems (Clark et al. 2011; Overgaard et al. 2006) using hydrologic land surface models (LSMs) and
44 (ii) the use of the resulting streamflow projections to simulate the allocative water balance dynamics
45 across water-dependent sectors (Wurbs 1995) using water management models (WMMs) (Brown et al.
46 2015). Unregulated streamflow simulations require forcing data from observed meteorological inputs or a
47 combination of global circulation models and regional atmospheric models. Streamflow projections
48 contain errors due to biases in meteorological and soil data as well as model calibration, scale, and limits
49 in process representations (Beven 1993, 2016; Gaganis 2009). A large body of literature has explored
50 how these errors are generated and how they can be categorized (Gupta and Govindaraju 2019; Gupta et
51 al. 2008; Nearing et al. 2016; Refsgaard et al. 2006; Vogel 2017; Wagener et al. 2010). However, it is

52 poorly understood how LSM errors propagate into WMMs, which are themselves subject to errors, and
53 combine to yield biases in our end-point decision-relevant measures of climate vulnerability (e.g., reduced
54 crop yields, water shortages, or financial risks). Recent studies have begun to formally analyze the
55 propagation of uncertainty of inflow water regimes within water management models (e.g., Hassanzadeh
56 et al. 2016; Marton and Paseka 2017; Nazemi and Wheeler 2014; Sordo-Ward et al. 2016). These efforts
57 mainly focus on internal variability or uncertainty that results from ensemble simulations based on
58 synthetically generated streamflow time series. Although understanding the effects of observation record
59 limits and internal variability is important, it is fundamentally different than the error perturbation
60 analyses contributed here. The implications of errors within the broadly used top-down GCM- and LSM-
61 based deterministic simulated streamflow projection products is not well understood in terms of its effects
62 on water management models. It is worth mentioning here that, synthetic generation of streamflow time
63 series is commonly used as an alternative bottom up way of exploring streamflow changes and
64 uncertainty (Borgomeo et al. 2015; Herman et al. 2016; Kirsch et al. 2013; Quinn et al. 2018, 2020;
65 Steinschneider et al. 2015). These methods often employ statistical techniques to construct streamflow
66 timeseries that are non-stationary and more diverse, while, they still maintain a reasonable level of
67 statistical consistency with the past observations. Overall, streamflow scenarios have been used to make
68 up for the lack of long-term streamflow observations. These scenarios also allow us to investigate cases
69 that have not been occurred during the observation periods such as low frequency extreme wet and dry
70 events, and multi-year droughts.

71 Here, we focus on climate-driven vulnerabilities in the California water supply system, which represents
72 one of the most institutionally complex water infrastructure systems in the world. The system (Figure 1)
73 includes thousands of kilometers of conveyance canals and dozens of dams that are operated to satisfy a
74 broad spectrum of objectives, including two state-wide water delivery projects—the State Water Project
75 (SWP) and the Central Valley Project (CVP). California’s water supply is highly dependent on the
76 surface water inflows from the Sierra Nevada mountains into its northern reservoirs. While the state has

77 experienced substantial flood and drought events in the past (Howitt et al. 2014; Mann and Gleick 2015),
78 climate change is expected to worsen the situation (Mann and Gleick 2015; Mote et al. 2005; Tanaka et
79 al. 2006). This vulnerability is motivating a myriad of propositions to improve California's water
80 infrastructures and institutions (Forsythe et al. 2017; Nishikawa 2016; Sandoval-Solis 2020).
81 Groundwater resources and water banks are among the most crucial and vulnerable parts of the water
82 supply in California (Kiparsky et al. 2017; Nishikawa 2016), particularly for the agricultural sector, and
83 are the subject of emerging regulations (Forsythe et al. 2017). A significant portion of California's annual
84 precipitation is generated through atmospheric rivers during the winter and early spring (Dettinger et al.
85 2011), which must be stored to meet summer agricultural demands (Christian-Smith 2013; Kocis and
86 Dahlke 2017). Therefore, water stakeholders in California recharge their groundwater resources during
87 these short-lived extreme events, to use it later when surface water is not sufficient to meet the demand
88 (Ghasemizade et al. 2019; Scanlon et al. 2016). This management regime potentially increases the
89 sensitivity of irrigation focused drought projections to short-term (daily) errors in simulated flood events.
90 To date, the implications of this issue have not been explored in detail.

91 In this study, we show how errors from a well-established coupled atmosphere-land modeling system
92 (WRF-NoahMP; Cai et al. 2014; Skamarock et al. 2005, 2008; Figures S1- S2) propagate into a
93 California-specific WMM (i.e., CALFEWS; Zeff et al. 2021) and impact the simulation of system-wide
94 water supply, groundwater extraction, and annual revenue of irrigation districts in the Central Valley.



95

96 **Figure 1.** Study area of this study (State of California). The map indicates the locations of various dams and
 97 reservoirs in California, the state’s main agricultural areas, and the spatial distribution of almond, one of the most
 98 important crops in California. The figure also shows the capacity of the dams of the two main water delivery
 99 projects in California and the San Luis dam that is shared between the two projects.

100

101 We trace how errors in a single dynamically downscaled deterministic streamflow scenario for a recent
 102 historically observed period can strongly bias the important water balance dynamics for key actors and

103 infrastructure systems. This work highlights the strong interdependence between errors in flood and
104 drought extremes, which are shown to be nonlinear, path-dependent, and amplified in modeled operations
105 of conveyance and storage infrastructures. In other words, the simulation of various system stakeholders
106 depends on the history of exposure of the stakeholder to streamflow errors as well as their flow paths
107 through other system components. Moreover, we show that standard methods for managing and reducing
108 these hydrologic errors exacerbate these water balance distortions as well as associated inferences of
109 climate vulnerabilities for the region.

110

111 **Methods**

112 In this study, we explore how errors in dynamically downscaled projections of surface hydrology impact
113 important California water management systems using the California Food-Energy-Water Systems Model
114 (CALFEWS; Zeff et al. 2021). The model adaptively allocates water across scales and sectors using a
115 detailed representation of the state's infrastructure and institutions. Our analysis compare CALFEWS
116 simulations of critical components of the California water distribution system under four sources of
117 streamflow inputs: i) observed streamflow from the California Department of Water Resources' Data
118 Exchange Center (CDEC); ii) raw WRF-NoahMP streamflow outputs (no groundwater correction;
119 NGW); iii) WRF Noah-MP streamflow outputs with an expert-driven manual removal of groundwater
120 biases (groundwater corrected; CGW); and iv) WRF Noah-MP streamflow data that reduced errors via an
121 automatic bias correction method using quantile mapping (bias-corrected; BC). In this section, we
122 describe the computational framework that was used to conduct the simulation-based analyses that
123 underlie this study (Figure S1 in Supplemental Materials). To that effect, we first introduce the regional
124 atmospheric land-surface model (WRF-Noah-MP) that generated our dynamically downscaled

125 streamflow datasets. We then summarize the water management model used in this study (CALFEWS;
126 Zeff et al. 2021). Finally, we then describe the methods used here to produce our bias-corrected datasets.

127 **WRF-Noah-MP Streamflow Projections:**

128 The input streamflow data to our WMM was generated using the Weather Research and Forecast (WRF)
129 regional climate model (Skamarock et al. 2005, 2008; Tang and Dennis 2014). The version of WRF used
130 to generate the streamflow inputs to CALFEWS is integrated with the Noah-MP land surface model
131 (LSM; Barlage et al. 2015) a mechanistic hydrologic LSM that simulates key surface water and energy
132 fluxes and states required by WRF as a surface boundary condition. Noah-MP also simulates surface
133 runoff and sub-flow, cold season processes, vegetation dynamics, soil water movement, frozen soil, and
134 infiltration processes (Cai et al. 2014; Ingwersen and Streck 2011; Liu et al. 2016).

135 In this study, we compare four sources of streamflow inputs for CALFEWS. The first one is our observed
136 streamflow baseline from the California Department of Water Resources' Data Exchange Center (CDEC).
137 We also considered three variants of WRF-Noah-MP-simulated streamflow scenarios: 1) Raw WRF-
138 Noah-MP (NGW); 2) Groundwater-corrected flow (CGW); and 3) Bias-corrected flow (BC). The first
139 two simulated streamflow scenarios (CGW and NGW) were developed by Holtzman et al. (2020) while
140 the BC scenario was developed in this study (Supplementary Notes S3). To develop NGW and CGW,
141 Holtzman et al. (2020) used two different parameterizations of WRF-Noah-MP. Their baseline WRF
142 setup and parameterization were consistent with Wrzesien et al. (2015), with a spatial resolution of 9
143 kilometers (27-km outer domain). However, Holtzman et al. (2020) showed that the default WRF
144 parameterizations can lead to biased streamflow simulations in California. Therefore, they published a
145 series of modifications to improve the simulated streamflow. The NGW is a direct output of WRF-Noah-
146 MP after improvement of its internal parameterizations. The CGW, on the other hand, was developed by
147 ex-post statistical correction of NGW-simulated streamflow to make up for the lack of a groundwater
148 representation in the original WRF-Noah-MP setup.

149 To develop the NGW streamflow scenario, Holtzman et al. (2020) made the following major
150 modifications: i) The rain-snow partitioning formulation was changed from a function of air temperature
151 to a more-sophisticated WRF microphysics scheme. This allows the model to accumulate more accurate
152 amounts of snow during the winter months. ii) They updated the depth of subsurface runoff generation
153 because WRF's default runoff generation depth was between 1 and 2 meters in all locations. This ignores
154 the fact that, in higher elevations, soil is generally shallower, and assumption of runoff generation from
155 assumed deeper soil layers can potentially lead to unreasonable baseflow generation and biased
156 streamflow timing. To respond to this problem, Holtzman et al. (2020) assumed runoff generation from a
157 shallower layer (10 to 30 cm). iii) Slope to calculate subsurface flow was another parameter that
158 Holtzman et al. (2020) changed to improve the simulation of subsurface flow. The default value of WRF-
159 Noah-MP was 0.1, but they changed this to 0.5. The higher subsurface flow slope was able to improve the
160 simulation of the streamflow amount. iv) Holtzman et al. (2020) also changed the "sand" and "ice" soil
161 types to "sandy loam" to decrease the occurrence of unrealistically large transient soil moisture changes at
162 the beginning of the simulation. v) Soil porosity of "sandy-loam" soil was modified from a default value
163 of 0.434 to 0.52. The reason was that their initial simulations indicated that the water-holding capacity of
164 the default modeled soils was not high enough, which led to earlier streamflow peaks. vi) Finally, they
165 used a constant value for snow capacitance (0.2) of Thompson microphysics scheme (Thompson et al.
166 2008) to ensure a more reasonable simulation of snowflake shape in WRF-Noah-MP.

167 In regions with significant surface water-groundwater interactions (Criss and Davisson 1996; Shaw et al.
168 2014) such as the Sierra-Nevada watersheds, the lack of groundwater representations can lead to biases in
169 simulation of magnitude and timing of runoff and river flow. Because WRF-Noah-MP's NGW setup did
170 not include a mechanistic simulation of groundwater dynamics (Barlage et al. 2015), a post-processing
171 groundwater correction module was utilized in the development of the CGW streamflow scenario. The
172 GW correction was performed using an offline statistical relationship that was utilized to improve the
173 NGW streamflow. The corrected streamflow on a given day was obtained as a weighted sum of three

174 quantities: i) the original NGW daily streamflow, ii) the average NGW streamflow over the past 365 days,
175 and iii) an intercept term which was set so that the correction did not change the overall mean NGW
176 streamflow over the entire simulation period. The weights were constant in time over the simulation
177 period but were allowed to vary between spatial locations. Conceptually, the 365-day running average
178 term represents releases from medium-term groundwater storage, and the intercept represents baseflow
179 due to long-term groundwater storage that is released over a time scale of many years. Including both
180 these terms helped model spatial variation in the residence time of groundwater.

181 Values of the correction weights were obtained separately for each streamflow location using the
182 following procedure: first, both NGW and observed full natural flows (i.e. gauged flows with corrections
183 for upstream human activities) were normalized by dividing by their overall mean value; then, linear
184 regression was used to obtain the weight values that minimized the mean square error between the
185 corrected normalized NGW and the normalized observations. The correction coefficients were fit on
186 normalized flows instead of raw flows because the primary goal of the correction was to remedy errors in
187 the NGW seasonality pattern, not to correct any overall bias. Results presented by Holtzman et al. (2020)
188 suggest that this model substantially improves on the uncorrected Noah-MP results (i.e., the NGW
189 scenario) using a soil-only modeling system. Note that Noah-MP does include an optional groundwater
190 model, but it is often impractical to use because it takes many simulation years to spin up (Niu et al.
191 2007).

192 Note that, there are other approaches that past studies have utilized to improve the representation of
193 groundwater dynamics in their streamflow simulations. For example, past studies have developed and
194 incorporated simple groundwater modules (Niu et al. 2007; Yang and Xie 2003) or dynamically
195 integrated their land surface hydrologic models into well-established groundwater models (Faunt et al.
196 2009; Kim et al. 2008; Molina-Navarro et al. 2019; Xu et al. 2012). A few other studies have used
197 statistical bias correction approaches to match the overall statistical moments of their simulated
198 streamflow with observations which implicitly takes into account groundwater dynamics (Hamlet and

199 Lettenmaier 1999; Tiwari et al. 2021). Finally, there are other methods such as Bayesian filtering methods
200 (Ait-El-Fquih et al. 2016; Panzeri et al. 2014; Rajabi et al. 2018) or offline post-processing procedures
201 (Holtzman et al. 2020; Trabucchi et al. 2021) that implicitly improve the representation of groundwater
202 dynamics and overall quality of streamflow simulations.

203 **California Food-Energy-Water Systems Model (CALFEWS)**

204 We use a Water Management Model (Figure S1) that has been developed to simulate north-central
205 California agro-hydrologic systems. The California Food-Energy-Water Systems model (CALFEWS)
206 model (Zeff et al. 2021) abstracts critical institutional and infrastructure elements (>1000) that capture the
207 complex dynamics for how north-central California's water balance is managed given the region's
208 extreme streamflow variability. CALFEWS simulates the daily timescale operation of dams, water
209 conveyance systems, groundwater banks, and water allocation decisions.

210 CALFEWS exploits state-aware rules that allow it to abstract the highly dynamic and adaptive operational
211 behaviors of the system while complying with the institutional constraints that shape the storage and
212 conveyance of water. More specifically, CALFEWS includes the operation of 12 major reservoirs in
213 north-central California (Figure 1). However, most of the water is conveyed from northern dams such as
214 Shasta and Oroville to central California's agricultural areas. The model mimics the operation of these
215 dams in terms of water storage, flood prevention, and water release for agricultural and environmental
216 services. The dams provide water to a complex transfer system that conveys water to the agricultural and
217 urban areas of California, which are mainly located in the central and southern parts of the state (Figure
218 1). The conveyance systems are based on two state-wide water transfer projects: the State Water Project
219 (SWP) and the Central Valley Project (CVP). Both projects own the storage and conveyance water
220 infrastructures that are included in the CALFEWS model. CALFEWS also takes into account all the
221 major river water rights holders in the Tulare Basin (e.g., Kings, Kaweah, Tule, Kern).

222 The CALFEWS model takes several environmental constraints into account. The model simulates delta-
223 related environmental concerns such as saltwater intrusion, minimum outflow from the delta, and

224 constraints in the old and middle river flow. It also captures other minimum flow regulations in California
225 rivers and their reaches. There are also non-environmental constraints that are enforced in the model, such
226 as pumping limitations, canal capacity limitations, and water rights constraints. The model includes over
227 thirty irrigation districts, ten distinct imported water contract and storage allocations, and nine major
228 water banks in the system. Additionally, the model simulates the water redistribution system in the
229 agricultural areas. For example, it captures direct groundwater banking partnerships and in-lieu
230 exchanges.

231 CALFEWS does not have a physically-based groundwater model that can mechanistically simulate
232 groundwater dynamics, but it does have a water balance accounting model that distributes water to
233 individual irrigation districts and groundwater banks based on surface water allocations, carry-over
234 storage reservations in surface water reservoirs, and the ownership of individual aquifer recharge and
235 recover assets. The model also simulates claims to excess flood water flows based on access and
236 conveyance constraints. The detailed operational rules used within CALFEWS enabling estimation of the
237 annual revenue and financial stability at the irrigation district scale.

238 Although capturing the diverse range of institutional and infrastructure operational considerations that
239 shape water allocation decisions is non-trivial and CALFEWS is subject to representational limits, the
240 model does reasonably capture the complex dynamics of the infrastructure systems and their operations
241 (see more details on the CALFEWS-HIS baselines for major storages and Sacramento-San Joaquin Delta
242 exports in Note S1, Figure S3-S4, and Table S1-S4 in Supplemental Materials, and Figure 2). Also, more
243 details and baseline capabilities of CALFEWS are available in Zeff et al. (2021).

244 **Quantile Mapping-Based Bias Correction**

245 In this study, we used the frequently employed statistical bias-correction technique called quantile
246 mapping (Cannon, Sobie, and Murdock 2015) to remove systematic biases of raw WRF-Noah-MP
247 streamflow data. To do this, we developed and used an R package called “biascorrection” (Supplementary

248 Notes 3) that follows the methodology described by (Hamlet and Lettenmaier 1999). In short, the bias
249 correction module uses the historical observed streamflow to create the monthly flow quantiles of each
250 individual month. After that, it uses the simulated streamflow data to create simulated monthly flow
251 quantiles. Afterwards, the bias correction module creates the monthly bias-corrected flow by swapping
252 each month of the simulated flow with the same quantile from the observed streamflow. Since hydrologic
253 models can simulate the average annual flow reasonably well, after constructing the monthly bias-
254 corrected flow, we adjust them to make sure that their average annual flow is consistent with what the
255 WRF-Noah-MP model has simulated. Finally, we disaggregate the monthly bias-corrected flow to daily
256 by multiplying the raw daily simulated flow of each month by the simulated bias-corrected ratio of that
257 month.

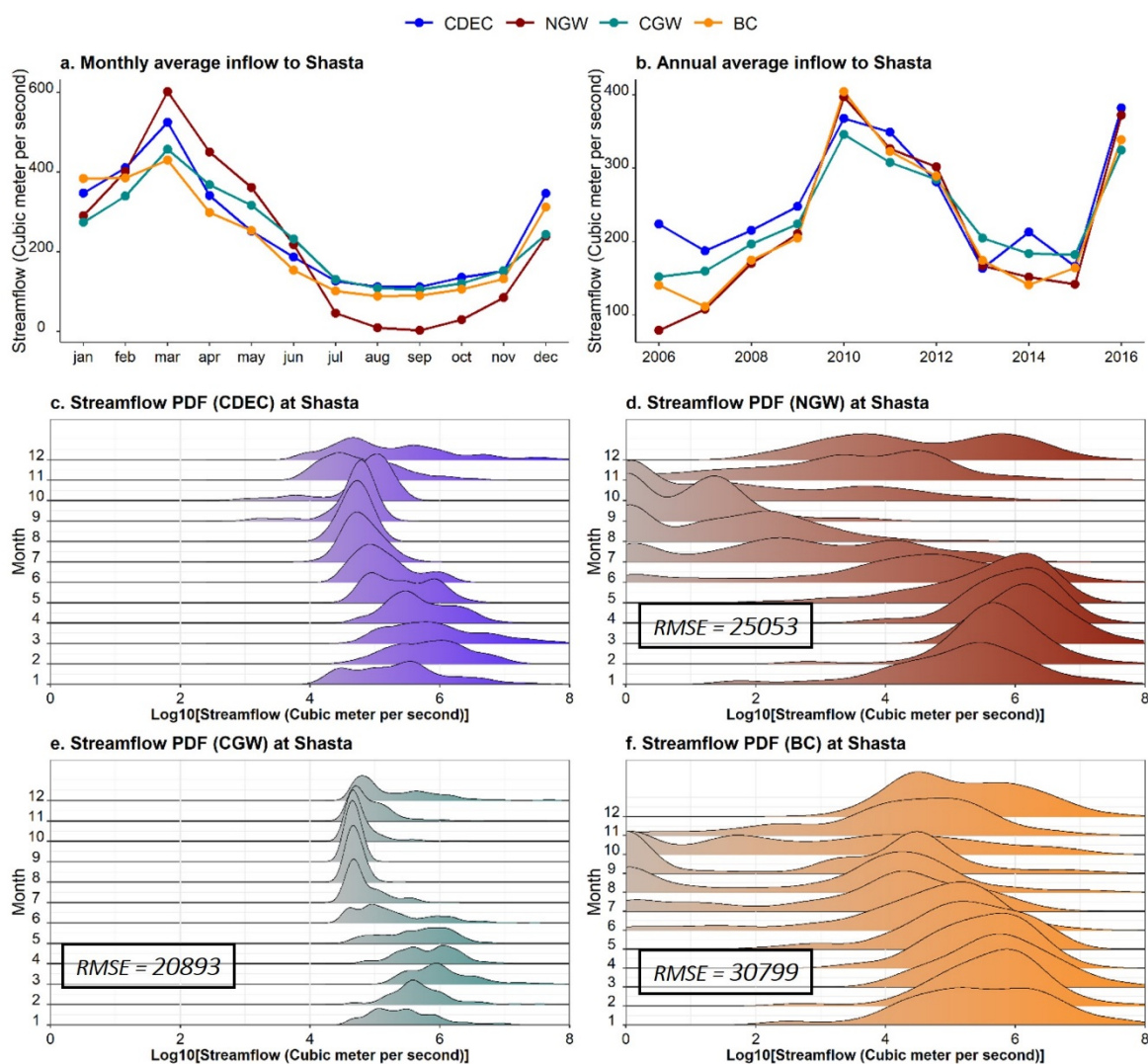
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259 **Results and Discussions**

260 **Diagnosing Streamflow Errors across Timescales**

261 The Shasta reservoir represents a key storage project for the CVP as well as flood control in northern
262 California. As a means of distinguishing floods, seasonal transitions, and drought periods for the Shasta
263 reservoir system, our error analysis is formulated across daily, monthly, and annual timescales (Figure 2
264 and Figure S5-S12 in Supplemental Materials). We show that the raw streamflow output of the WRF-
265 Noah-MP model (NGW scenario) systematically underestimates streamflow during low flow periods
266 (Figure 2 – Panel a). Previous literature has attributed these biases mainly to the significant computational
267 and conceptual constraints associated with representing groundwater processes in Noah-MP (Cai et al.
268 2014; Holtzman et al. 2020). Our results (Figure 2 – Panel a) demonstrate that the groundwater corrected
269 stream flows (CGW) reduce errors during low-flow periods. However, the expert-based CGW calibration
270 (Figure 2 – Panels a and e) yields a consistent underestimation during high-flow periods. More broadly,
271 the distributions of the observed and the simulated streamflow scenarios at the daily time-step (Figure 2 –
272 Panels c, d, e, and f) show that the CGW scenario significantly reduces the range of variability in

273 streamflow and extremes. The water added during the low-flow periods is drawn from the high-flow
 274 periods – more specifically, from extreme flood events such as atmospheric rivers. Atmospheric rivers
 275 (and other extreme flow events) are a crucial component of water availability in California, and the
 276 presence or absence of them is what distinguishes a drought year from a wet year (Diffenbaugh et al.
 277 2015). We also show that the quantile mapping-based bias correction scenario (BC) enhances aggregated
 278 monthly and annual model performance in a manner comparable to the CGW, improving the
 279 representation of streamflow during dry periods (e.g., Figure 2).



280
 281 **Figure 2.** Comparison between the observed (CDEC) and simulated streamflow scenarios at Shasta Dam. The
 282 simulated streamflow scenarios include raw WRF-Noah-MP flow (NGW), WRF Noah-MP groundwater-corrected

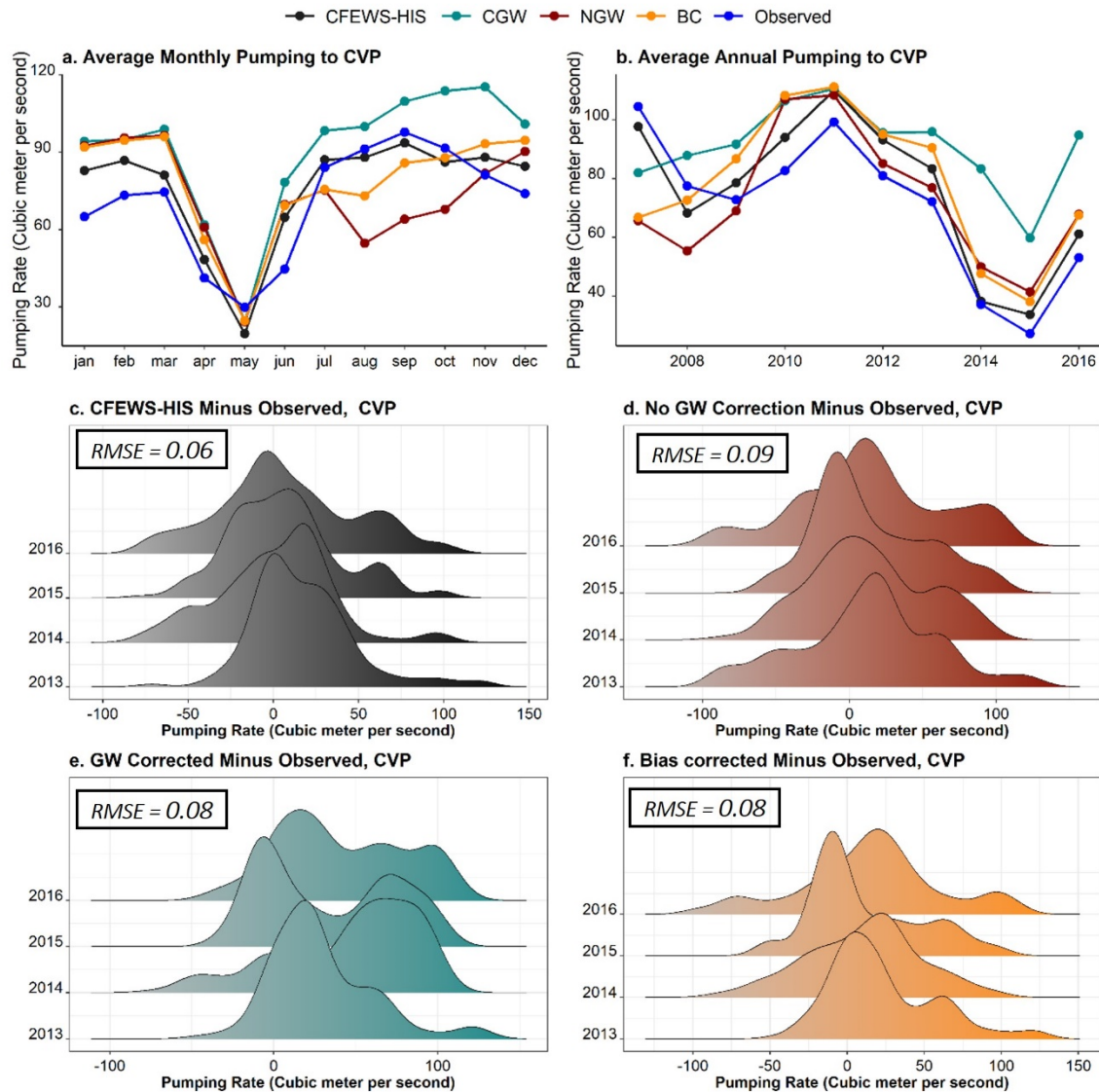
283 flow (CGW), and bias-corrected flow (BC). Panel a. and b. demonstrate the average monthly and average annual
284 streamflow, respectively. Panel c., d., e., and f. show the monthly separated probability density function of daily
285 streamflow for our four flow scenarios (observed, no groundwater correction, groundwater corrected, and bias-
286 corrected).

287
288 However, similar to the CGW scenario, statistical bias correction deteriorates the representation of
289 streamflow during high-flow periods, which dampens inter-seasonal variability. The streamflow error
290 management methods (i.e., BC and CGW) do not improve the entire distribution of flows critical to north-
291 central California. A key concern that emerges from these results is how these streamflow biases could
292 create path-dependent and persistent errors that propagate into the other components of the California
293 water system and affect our perception of downstream, multi-sector climate vulnerabilities. While, we
294 only explain the results for Shasta Dam here, our analysis demonstrate that the simulated inflow time
295 series into other California reservoirs (e.g., Oroville, Folsom, Pine Flat, New Melones, Millerton,
296 Isabella, Don Pedro, and Yuba Dam) are predominantly in agreement with the Shasta dam (Figures S5-
297 S12 in the Supplemental Materials).

298 **Errors in the Main North-to-South Surface Water Transfers**

299 The two major pumping stations at the Sacramento – San Joaquin River Delta play a crucial role in
300 California’s north-to-south water transfer projects. Pumping rates from these stations to the SWP and
301 CVP are among the most important indicators of the system-wide water availability in California,
302 particularly for users in the water-scarce San Joaquin Valley as well as Southern California. Here, we
303 compare CALFEWS simulated pumping rates using the different sources of streamflow inputs with the
304 actual observed historical pumping rates as recorded in CDEC. Our results (Figure 3 – Panels a and b,
305 Figure S13 in Supplemental Materials) show that, in general, the LSM-based streamflow results (CGW,
306 NGW, and BC) introduce significant errors compared to the CFEWS-HIS simulation (CALFEWS
307 simulations under observed streamflow inputs). While, at least in some cases, the baseline (CFEWS-HIS)
308 results do show non-negligible deviations from the observed pumping rates (Figure 3 – Panels a and b),

309 the error distribution is relatively consistent during wet and dry years (Figure 3 – Panel c). It should be
310 noted that capturing the diverse range of institutional and infrastructure operational considerations that
311 shape pumping from the Sacramento – San Joaquin River Delta is non-trivial. As noted above,
312 CALFEWS itself is subject to representational limits. Nonetheless, the CFEWS-HIS results largely
313 capture key trends and dynamics. In the case of the NGW results (raw WRF-Noah-MP streamflow
314 outputs), the underestimation of reservoir inflow during the summer causes a systematic underestimation
315 of the pumping rate to the CVP during that season (Figure 2 – Panels a and d). These errors, which
316 overlap in timing with peak irrigation demand, create consequential biases for projections of agricultural
317 productivity and groundwater extraction.



318

319 **Figure 3.** Pumping rate to Central Valley Project (CVP). This figure compares the “observed” pumping to CVP with
 320 simulations of CALFEWS under different streamflow scenarios (i.e., CDEC [CFEWS-HIS], raw WRF-Noah-MP
 321 output [NGW], groundwater corrected [CGW], and bias-corrected [BC]) across monthly (Panel a), annual (Panel b),
 322 and daily time scales (Panels c-f). In this figure, the unit of Root Mean Square Error (RMSE) is cubic meter per
 323 second.

324

325 Efforts to address these biases in the CGW (groundwater corrected) and BC (bias-corrected) results do
 326 partially address the pumping underestimation issue, at least in some instances (Figure 2 – Panels e and f).

327 However, these scenarios also produce higher pumping biases when estimating the pumping rates to the
328 CVP and SWP (see SWP pumping rate errors in Figure S13 in the Supplemental Materials). These
329 overestimation biases become more pronounced in key CA drought years (e.g., 2014 and 2015). Put
330 simply, the groundwater correction and quantile-mapped bias correction falsely overestimate delta water
331 deliveries in the evaluated drought years.

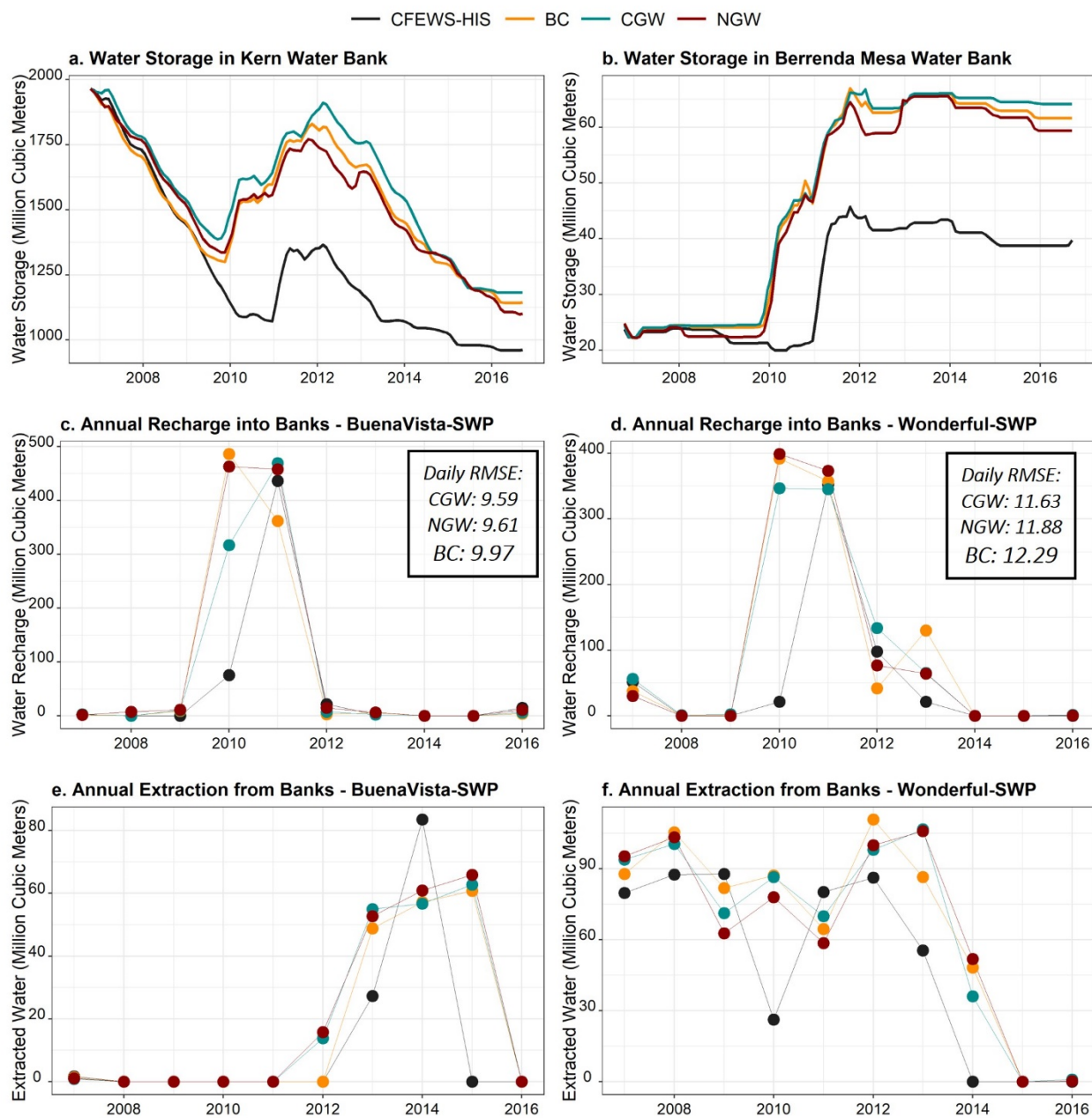
332 The overestimation issue appears more frequently in the CGW case, primarily during high-flow periods in
333 the winter and early spring. This effect is most pronounced in drier years (Figure 3 – Panel e) because the
334 manual deterministic improvements in the representation of dry months can eliminate many low flow
335 days that naturally exist in the observed record (Figure 2 – Panel c. and e.). Also, as the system transitions
336 from the 2013-2015 drought to a wetter year in 2016, the CGW's bias leads to an overly optimistic
337 inference of drought recovery. The BC scenario more closely follows the distribution of the raw simulated
338 results (Figure 3 – Panel f), however, it also amplifies some of the extreme flood events, leading to
339 overestimated project pumping for several periods. Moreover, as discussed before, both the BC and CGW
340 streamflow scenarios tend to underestimate flow during the high-flow periods, which can significantly
341 affect the magnitude and timing of dam storage in the spring and winter. The biases in the delta-to-project
342 deliveries also imply that LSM streamflow errors can significantly influence projections of energy supply
343 and demand in California.

344 **Groundwater Banks**

345 Groundwater banks (GWBs) are critical components of California's water system. In California, GWBs
346 are used as additional sources of storage that help capture excess water during flood events to hedge
347 against droughts (Ghasemizade et al. 2019). For example, from 2012 to 2017, GWBs provided the system
348 with more than 40 km³ of drought relief water (Xiao et al. 2017), playing a key role in California
349 agricultural systems seeking to avoid yield losses and in some cases complete bankruptcy (Diffenbaugh et
350 al. 2015; Sarhadi et al. 2018).

351 Our results indicate that upstream streamflow errors propagate into GWB simulations and significantly
352 degrade the simulated banked storages (Figure 4 – Panels a and b, and Supplemental Materials Table S5),
353 recharge to GWBs (Figure 4 – Panels c and d), and extraction from GWBs (Figure 4 – Panels e and f).
354 For example, the simulated streamflow scenarios (NGW, CGW, and BC) all lead to systematic
355 overestimations of water storage in two groundwater banks of California: Kern Water Bank (Kern) and
356 Berrenda Mesa Project (Berrenda ; Figure 4 – Panels a and b). There are two main factors influencing this
357 overestimation. First, groundwater banks have slower turnover times relative to the other components of
358 the system, allowing water to stay in them for longer periods of time (i.e., higher residence times). This
359 implies that if streamflow inputs have systematic errors in overestimating available water, the errors will
360 not dissipate immediately, and the GWBs can substantially accumulate long-lasting erroneous storage
361 contributions. For example, during the Spring of 2010, our simulated streamflow scenarios (NGW, CGW,
362 and BC) consistently overestimated inflow to upstream reservoirs (Figure S14 in Supplemental
363 Materials). Our results (Figure 4 – Panels c and d) clearly show how a portion of the overestimated water
364 ended up recharging the groundwater system. This erroneous recharge causes a spike in groundwater
365 storage as compared to the CFEWS-HIS baseline (Figure 4 – Panel a and b), and this gap remained to the
366 end of our simulation period six-years later. The second major factor influencing the overestimation of
367 available storages in GWBs within the NGW, BC, and CGW projections is their overestimation of the
368 average annual pumping to the CVP (Figure 2 – Panel b). These overestimation errors ultimately
369 contribute to higher groundwater recharge and lower water deficits and, thus, lower groundwater
370 extraction. Our results emphasize that when evaluating water management options and vulnerabilities in
371 California, drought years and flood years are tightly coupled. This implies that, if a modeling framework
372 struggles to capture floods and wet periods well, it would not be able to capture the dynamic impacts of
373 droughts. These consequential, long-lasting and path-dependent errors also highlight that extra attention
374 should be paid to statistical and deterministic bias correction methods (e.g., BC and CGW) that
375 inadvertently shift the dynamic water balances associated with highly consequential extreme events
376 (Figure 2 – Panels e and f). The overestimation of groundwater bank storage can be also attributed to the

377 fact that the recharge capacities in the groundwater banks are significantly higher than groundwater
 378 extraction capacities. This difference increases the residence time of error in groundwater systems, and
 379 further demonstrate the contrasting sensitivity of the system to errors during wet and dry periods.



380
 381 **Figure 4.** Groundwater storage, recharge, and extraction. This figure shows how different streamflow scenarios (i.e.,
 382 observed [CFEWS-HIS], raw WRF-Noah-MP output [NGW], groundwater corrected [CGW], and bias-corrected

383 [BC]) affect CALFEWS simulation of groundwater banks of the Central Valley. In this figure, the unit of Root
384 Mean Square Error (RMSE) is million cubic meters.

385

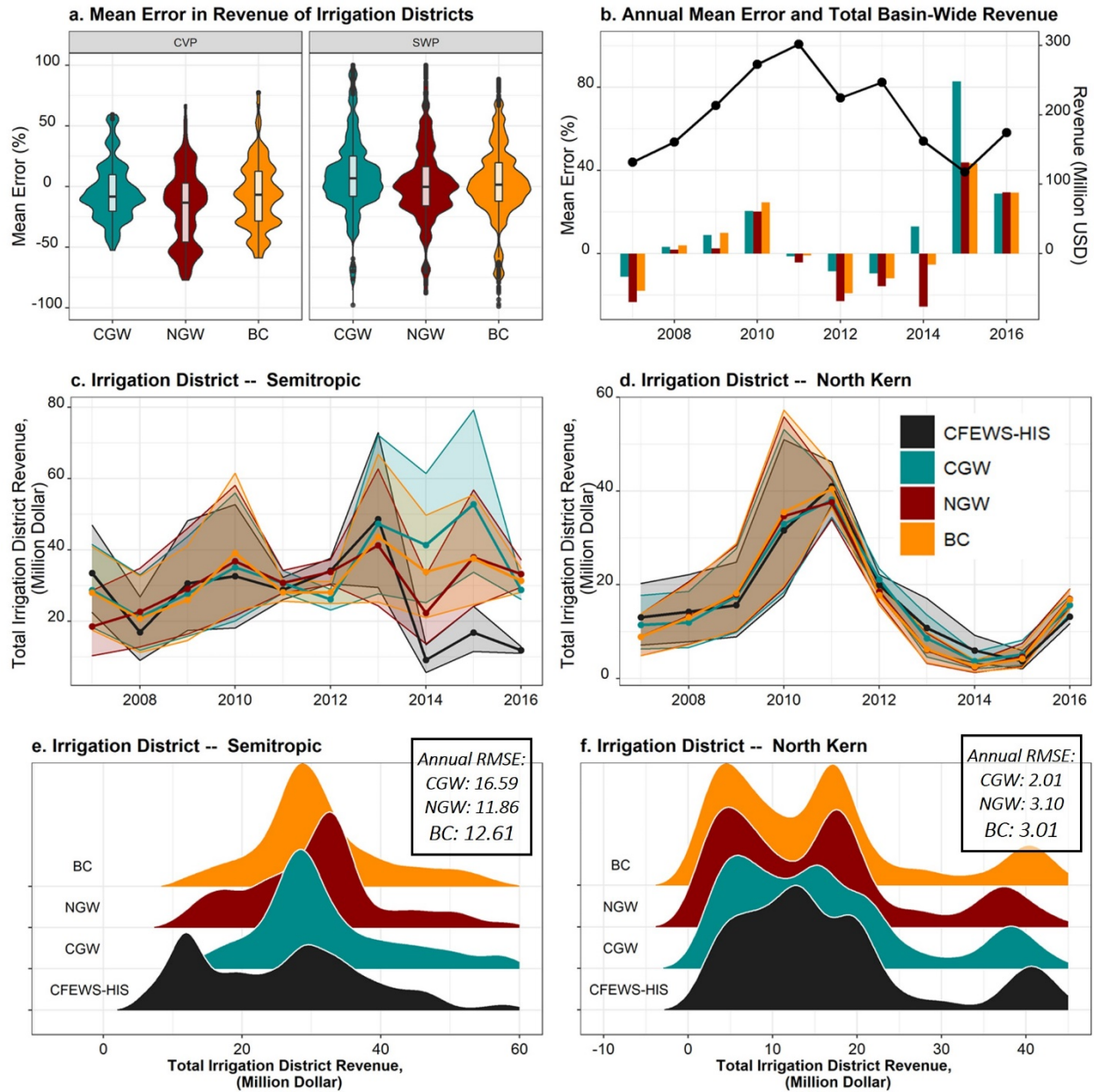
386 Moreover, our results indicate that the water extractions and recharge of various irrigation districts show
387 distinctly different responses to streamflow scenarios. This is due to their unique institutional contexts as
388 defined by their level of water right seniority, contracts, water supply projects (CVP vs. SWP), and
389 geographical location in California (Table S6 in Supplemental Materials). The persistence and path
390 dependence of errors in downscaled hydrologic projections strongly depends on the institutionally
391 complex infrastructure systems of the north-central California water system. Infrastructure elements or
392 users with the most secure water rights, or most advantageous positions within the water distribution
393 network, receive their total water demand more frequently; therefore, an over- or underestimation errors
394 for available inflow to the system are themselves institutionally allocated across the complex network of
395 other water right holders.

396 **The Financial Dynamics of Irrigation Districts**

397 Errors in streamflow projections and the current standard approaches for managing them also strongly
398 shape our ability to infer the financial stability of irrigation districts. Irrigation districts are cooperative
399 water management institutions that facilitate the delivery and storage of water. They are also responsible
400 for the maintenance of water storage and delivery infrastructure. These operational activities are the
401 primary source of irrigation districts' income. Generally, a lower amount of system-wide water supply
402 reduces the total volume of water that they are able to convey and sell to their retail customers leading to
403 lower overall revenues that can cause potential financial instability, higher borrowing costs, lower
404 investment in infrastructure maintenance, and an inability to retain trained staff, all of which have
405 detrimental consequences for the wellbeing of the region's agriculture.

406 Our results show that streamflow errors significantly influence our ability to infer the revenue
407 vulnerabilities of irrigation districts (Figure 5 – Panel a). To estimate these revenues, given the

408 unfortuna te dearth of transparently recorded water price data, we explore here 100 plausible water price
409 scenarios that represent five plausible trajectories of water price change during drought years (Note S4
410 and Figure S15 Supplemental Materials). More specifically, the five baseline trajectories that have been
411 used to generate our 100 synthetic water price scenarios represent -20%, 0%, +20%, +50%, and +80%
412 change in water price during drought years. The biases in revenue vulnerability results stem from
413 different operational activities such as surface water delivery, aquifer recharge and groundwater pumping.
414 Consequently, all of the previously discussed surface- and ground- water sources of errors contribute to
415 the resulting errors for irrigation districts' financial dynamics. We estimate that the combined annual
416 expected costs of the errors among the 26 simulated irrigation districts totals to about \$114-million, \$91-
417 million, and \$81-million US dollars under the CGW, NGW, and BC scenarios, respectively.



418

419 **Figure 5.** Financial stability of irrigation districts. This figure shows how different streamflow scenarios (i.e.,
 420 baseline [CFEWS-HIS], raw WRF-Noah-MP output [NGW], and groundwater corrected WRF-Noah-MP output
 421 [CGW], and bias-corrected WRF-Noah-MP [BC]) affect the simulation of financial stability for the Central Valley's
 422 irrigation districts. In Panel a, c and d, the distributions and uncertainty bounds are generated from our 100 water
 423 price realizations and the solid lines demonstrates the average of all those water price scenarios. Panels e. and f.
 424 show the probability density function of average yearly revenue across different irrigation districts and under the

425 observed, NGW, CGW, BC conditions. In this figure, the unit of Root Mean Square Error (RMSE) is million
426 dollars.

427

428 Such a costly misperception (ranging from underestimation -81% to an overestimation of +111% of
429 average annual revenues among individual districts) of irrigation districts' revenues could lead to
430 infrastructure investment and financial decisions that would likely harm them as well as the broader water
431 dependent north-central California systems. We also highlight that susceptibility of different irrigation
432 districts to streamflow errors depends on the details of their specific institutional contexts (Figure 5 –
433 Panels c and d). For example, our analysis suggests that SWP irrigation districts are more sensitive to
434 streamflow errors (Figure 5 – Panel a), mainly because they tend to rely closely on error-prone water
435 balance dynamics. In addition, various other institutional factors such as water right seniority level of
436 districts, degree of their dependence on ground- versus surface- water systems, and the geographical
437 location of districts contribute to their susceptibility or immunity to headwater streamflow errors.

438 Our analysis indicates that, on average, the expert-based and automatic error management methods (CGW
439 and BC) tend to systematically overestimate irrigation districts' annual revenues (Figure 5 – Panel a and
440 b). The reason is that, under these scenarios, surface water delivery during summertime is generally
441 higher, and higher supply increases the income of irrigation districts. However, as discussed earlier, these
442 errors also compounded with groundwater errors that can stem from failures in capturing key flood
443 events. Given that these groundwater biases have longer residence times, they adversely impact irrigation
444 districts' revenue estimates over the longer term. It is concerning that these biases are very pronounced
445 and more clearly emerge during extreme drought years. For example, the relative error is significantly
446 higher during 2015, which was the most significant drought year in our study period (Figure 5 – Panel c).
447 Furthermore, as the tail of the revenue probability density functions suggest, simulated streamflow
448 scenarios perform exceptionally poorly during extreme low-revenue periods (Figure 5 – Panels e and f,

449 also see Figure S16-S21 in Supplemental Materials). Again, this result is of significant concern because
450 these extreme drought years can trigger major investments or inform planned institutional changes.

451 Finally, our analysis (Figure S22 in Supplemental Materials) suggests that our broad envelope of water
452 pricing scenarios do not substantially modify the core insights from the revenue impacts shown in Figure
453 5. However, water pricing strongly depends on projections of state-wide availability of water (Medellín-
454 Azuara et al. 2012), and is a factor that should be studied closely for its interactions with streamflow error
455 propagation. While fully exploring these dependencies is beyond the scope of this study, future work that
456 employs hydro-economic models that capture the interactions between water supply availability estimates
457 and water rates would provide more comprehensive understanding of the compound dynamics of human-
458 natural system under uncertainty.

459 **Do Streamflow Corrections Increase the Error in Modeled Impacts?**

460 Our results suggest that, at least in some cases, the expert-based manual groundwater bias correction and
461 quantile mapping-based bias correction increase the bias and deteriorate the quality of the CALFEWS
462 simulations. This is slightly counterintuitive, considering the fact that there are severe and well-known
463 biases in the NGW streamflow simulation results from WRF-NoahMP, especially during low-flow
464 periods (Figure 4 – Panels c and e), and the standard aggregated accuracy model performance metrics
465 (e.g., NSE) are higher for the CGW and BC. One reason for the increases in error is that, among the many
466 features of a streamflow time series (including average annual magnitude, average flow magnitude in
467 different seasons, and seasonality), any specific bias correction method will optimize error in terms of
468 only some of those features, while errors in other features may even be increased. Also, capturing the
469 properties of extreme events is very important, as the severity and persistence of streamflow during low-
470 and high-flow periods affect the operation of many components of the north-central California water
471 infrastructure and institutional systems (Hanak et al. 2018; Scanlon et al. 2016). As such, we recommend
472 that future studies claiming to improve simulated representation of hydrologic systems for the purpose of
473 informing water resource decision-making move beyond typical bulk hydrograph metrics (e.g. RMSE,

474 NSE, Kling-Gupta Efficiency) because they do not capture important nonlinear water balance dynamics
475 that shape water resources management. Although these metrics are easy to calculate, our results suggest
476 that they can provide a misleading sense of improvement.

477 Additionally, there is a close relationship between floods and droughts in California's water system.
478 Floodwater is often either stored in surface reservoirs or controlled and diverted toward recharge basins,
479 feeding groundwater banks. Later, the banked/stored water is used by irrigation districts (Dettinger et al.
480 2011; Xiao et al. 2017). Therefore, error generated during high-flow periods will propagate into low-flow
481 years and affect the simulation of system-wide water availability, groundwater extraction, and irrigation
482 district revenue during water shortage periods, when the north-central Californian water system is more
483 vulnerable. In other words, errors across time and space pool, transfer and reside in the institutionally
484 complex infrastructure systems. We use our north-central California example to argue that, in each
485 region, one or more characteristics of flow might be more important to capture, and the interaction of
486 these properties (high- and low- flow periods) must be known in order for a reasonable understanding of
487 the system to be gained. Finally, we warn that the complex institutional and infrastructure contexts of the
488 errors in simulated streamflow projections, are critical to understanding the consequences of any error
489 management strategies. Deterministic bias correction that are commonly used in climate scenario
490 modeling exacerbate this issue, as they ignore the water resources system context in which they are
491 employed. Our results highlight that the impact of changing hydrology on water resources in climate
492 projections cannot be treated as being dominantly a natural systems modeling problem.

493

494 **Conclusions**

495 In this study, we explore how our management of the well-known errors and biases in coupled land-
496 atmosphere modeling systems (e.g., WRF-NoahMP) used to simulate current hydrology (as in this study)
497 and increasingly to project regional climate change impacts (Huang et al. 2018; Musselman et al. 2018;

498 Schwartz et al. 2017; Wrzesien and Pavelsky 2020) can strongly distort our perceptions of vulnerabilities
499 in institutionally complex major global water resources systems such as the north-central California case
500 analyzed in this study. We show how streamflow errors from an atmospheric and land-surface hydrologic
501 model, WRF-Noah-MP, propagate into a water management model, CALFEWS, and affect perceptions of
502 system-wide water supplies, groundwater banking, and the annual revenue of irrigation districts. We show
503 that the north-central California water management infrastructures serve their intended purpose, highly
504 coupling the water balance dynamics of floods and droughts. The infrastructures likewise shape the
505 residence times and conveyance of water balance errors across extreme events. We show that these errors
506 have long, multi-year residence times and become more consequential during severe drought periods.
507 This is concerning because the inferences we draw from simulating extreme drought years are more likely
508 than other years to shape perceptions and trigger institutional and infrastructural changes. We also show
509 that errors and their effects can be unique and path-dependent as illustrated in the north-central California
510 system's dependencies on different major water delivery projects (CVP vs. SWP), the network of water
511 rights, and the complex water portfolios for each irrigation district. We show that ex-post corrections of
512 raw WRF-Noah-MP outputs do not necessarily reduce biases in the simulation of key processes and, in
513 some cases, can strongly degrade system simulations.

514 Finally, our results indicate that the need for future research to more fully engage with how institutional
515 and infrastructure context shapes the efficacy of bias-correction choices in our climate vulnerability
516 assessments for complex water resources systems. We show that they can strongly distort our inferences
517 of climate-driven vulnerabilities given the highly interdependent nature of the human and natural
518 processes that WMMs simulate. The results of this study also highlight the necessity of considering
519 alternative paradigms of water resources vulnerability assessments, such as exploratory modeling (e.g.,
520 (Hadjimichael et al. 2020)), which can more fully incorporate and address the key errors and uncertainties
521 that shape projections of climate change vulnerabilities.

522

523 **Data Availability Statement**

524 All datasets and scripts used in this study are available in [the GitHub repository](#) of the paper. The
525 CALFEWS model is an open-source software and its latest version can be obtained from its doi
526 repository (<https://doi.org/10.31224/osf.io/sqr7e>). Also, the *biascorrection* R package can be found here
527 (<https://github.com/keyvan-malek/biascorrection>).

528

529 **Acknowledgements**

530 This material is based upon work supported by the National Science Foundation (NSF), Innovations at the
531 Nexus of Food- Energy-Water Systems, Track 2 (Award 1639268). The views expressed in this work
532 represent those of the authors and do not necessarily reflect the views or policies of the National Science
533 Foundation.

534 **Author Contribution Statement**

535 Keyvan Malek conducted simulations and wrote the first draft of the manuscript. Patrick Reed
536 supervised the project. Keyvan Malek, Patrick Reed, Harrison Zeff, Andrew Hamilton, Melissa Wrzesien,
537 Natan Holtzman, Scott Steinschneider, Jonathan Herman, and Tamlin Pavelsky participated in
538 development of the paper's central ideas, its study design, and the final version of the manuscript.

539

540 **Conflict of Interest**

541 The authors declare no conflicts of interests.

542

543 **References**

544 Ait-El-Fquih, B., El Gharamti, M., and Hoteit, I. (2016). "A Bayesian consistent dual ensemble Kalman
545 filter for state-parameter estimation in subsurface hydrology." *Hydrology and Earth System*
546 *Sciences*, Copernicus GmbH, 20(8), 3289–3307.

547 Barlage, M., Tewari, M., Chen, F., Miguez-Macho, G., Yang, Z.-L., and Niu, G.-Y. (2015). "The effect of
548 groundwater interaction in North American regional climate simulations with WRF/Noah-MP."
549 *Climatic Change*, 129(3), 485–498.

550 Baum, R., Characklis, G. W., and Serre, M. L. (2018). "Effects of Geographic Diversification on Risk
551 Pooling to Mitigate Drought-Related Financial Losses for Water Utilities." *Water Resources*
552 *Research*, 54(4), 2561–2579.

553 Beven, K. (1993). "Research Perspectives in Hydrology Prophecy, reality and uncertainty in distributed
554 hydrological modelling." *Advances in Water Resources*, 16(1), 41–51.

555 Beven, K. (2016). "Facets of uncertainty: epistemic uncertainty, non-stationarity, likelihood, hypothesis
556 testing, and communication." *Hydrological Sciences Journal*, 61(9), 1652–1665.

557 Borgomeo, E., Farmer, C. L., and Hall, J. W. (2015). "Numerical rivers: A synthetic streamflow generator
558 for water resources vulnerability assessments." *Water Resources Research*, 51(7), 5382–5405.

559 Brown, C. M., Lund, J. R., Cai, X., Reed, P. M., Zagana, E. A., Ostfeld, A., Hall, J., Characklis, G. W., Yu, W.,
560 and Brekke, L. (2015). "The future of water resources systems analysis: Toward a scientific
561 framework for sustainable water management." *Water Resources Research*, 51(8), 6110–6124.

562 Cai, X., Yang, Z.-L., David, C. H., Niu, G.-Y., and Rodell, M. (2014). "Hydrological evaluation of the Noah-
563 MP land surface model for the Mississippi River Basin." *Journal of Geophysical Research:*
564 *Atmospheres*, 119(1), 2013JD020792.

565 Christian-Smith, J. (2013). *Improving Water Management through Groundwater Banking: Kern County*
566 *and the Rosedale-Rio Bravo Water Storage District*. Pacific Institute Farm Water Success Stories:
567 Groundwater Banking, Pacific

568 Clark, M. P., Hendriks, J., Slater, A. G., Kavetski, D., Anderson, B., Cullen, N. J., Kerr, T., Hreinsson, E. Ö.,
569 and Woods, R. A. (2011). "Representing spatial variability of snow water equivalent in hydrologic
570 and land-surface models: A review." *Water Resources Research*, 47(7).

571 Criss, R. E., and Davisson, M. L. (1996). "Isotopic imaging of surface water/groundwater interactions,
572 Sacramento Valley, California." *Journal of Hydrology*, 178(1), 205–222.

573 Dettinger, M. D., Ralph, F. M., Das, T., Neiman, P. J., and Cayan, D. R. (2011). "Atmospheric Rivers, Floods
574 and the Water Resources of California." *Water*, Molecular Diversity Preservation International,
575 3(2), 445–478.

576 Diffenbaugh, N. S., Swain, D. L., and Touma, D. (2015). "Anthropogenic warming has increased drought
577 risk in California." *Proceedings of the National Academy of Sciences*, 112(13), 3931–3936.

578 Faut, C., Hanson, R., Belitz, K., Schmid, W., Predmore, S., Rewis, D., and Mcpherson, K. (2009). "Chapter
579 C. Numerical model of the hydrologic landscape and groundwater flow in California's Central
580 Valley." *USGS professional paper*, in: Faut, C.C., ed.2009, Groundwater Availability of the
581 Central Valley Aquifer, California, 121–212.

582 Forsythe, L. M., Jones, I. M., and Kemp, D. J. (2017). "A Report Card: Progress under California's
583 Sustainable Groundwater Management Act (SGMA)." *U. Denv. Water L. Rev.*, 21, 199.

584 Gaganis, P. (2009). "Model calibration/parameter estimation techniques and conceptual model error."
585 *Uncertainties in Environmental Modelling and Consequences for Policy Making*, NATO Science
586 for Peace and Security Series C: Environmental Security, P. C. Baveye, M. Laba, and J. Mysiak,
587 eds., Springer Netherlands, Dordrecht, 129–154.

588 Ghasemizade, M., Asante, K. O., Petersen, C., Kocis, T., Dahlke, H. E., and Harter, T. (2019). "An
589 Integrated Approach Toward Sustainability via Groundwater Banking in the Southern Central
590 Valley, California." *Water Resources Research*, 55(4), 2742–2759.

591 Gizelis, T.-I., and Wooden, A. E. (2010). "Water resources, institutions, & intrastate conflict." *Political*
592 *Geography*, 29(8), 444–453.

593 Gupta, A., and Govindaraju, R. S. (2019). "Propagation of structural uncertainty in watershed hydrologic
594 models." *Journal of Hydrology*, 575, 66–81.

595 Gupta, H. V., Wagener, T., and Liu, Y. (2008). "Reconciling theory with observations: elements of a
596 diagnostic approach to model evaluation." *Hydrological Processes*, 22(18), 3802–3813.

597 Hadjimichael, A., Quinn, J., Wilson, E., Reed, P., Basdekas, L., Yates, D., and Garrison, M. (2020).
598 "Defining Robustness, Vulnerabilities, and Consequential Scenarios for Diverse Stakeholder
599 Interests in Institutionally Complex River Basins." *Earth's Future*, 8(7), e2020EF001503.

600 Hamlet, A. F., and Lettenmaier, D. P. (1999). "Effects of Climate Change on Hydrology and Water
601 Resources in the Columbia River Basin1." *JAWRA Journal of the American Water Resources
602 Association*, 35(6), 1597–1623.

603 Hanak, E., Jezdimirovic, J., Green, S., and Escriva-Bou, A. (2018). "Replenishing Groundwater in the San
604 Joaquin Valley." *Public Policy Institute of California*, 36.

605 Hassanzadeh, E., Elshorbagy, A., Wheeler, H., Gober, P., and Nazemi, A. (2016). "Integrating Supply
606 Uncertainties from Stochastic Modeling into Integrated Water Resource Management: Case
607 Study of the Saskatchewan River Basin." *Journal of Water Resources Planning and Management*,
608 American Society of Civil Engineers, 142(2), 05015006.

609 Herman, J. D., Zeff, H. B., Lamontagne, J. R., Reed, P. M., and Characklis, G. W. (2016). "Synthetic
610 Drought Scenario Generation to Support Bottom-Up Water Supply Vulnerability Assessments."
611 *Journal of Water Resources Planning and Management*, American Society of Civil Engineers,
612 142(11), 04016050.

613 Holtzman, N. M., Pavelsky, T. M., Cohen, J. S., Wrzesien, M. L., and Herman, J. D. (2020). "Tailoring WRF
614 and Noah-MP to Improve Process Representation of Sierra Nevada Runoff: Diagnostic Evaluation
615 and Applications." *Journal of Advances in Modeling Earth Systems*, 12(3), e2019MS001832.

616 Howitt, R., Medellín-Azuara, J., MacEwan, D., Lund, J., and Sumner, D. (2014). *Economic Analysis of the
617 2014 Drought for California Agriculture*. CCenter for Watershed Sciences, University of
618 California, Davisenter for Watershed Sciences, 20.

619 Huang, X., Hall, A. D., and Berg, N. (2018). "Anthropogenic Warming Impacts on Today's Sierra Nevada
620 Snowpack and Flood Risk." *Geophysical Research Letters*, 45(12), 6215–6222.

621 Ingwersen, J., and Streck, T. (2011). "NOAH-GECROS: A coupled land surface - crop growth model for
622 simulating water and energy exchange between croplands and atmosphere." *AGU Fall Meeting
623 Abstracts*, 1, 0958.

624 Kim, N. W., Chung, I. M., Won, Y. S., and Arnold, J. G. (2008). "Development and application of the
625 integrated SWAT–MODFLOW model." *Journal of Hydrology*, 356(1), 1–16.

626 Kiparsky, M., Milman, A., Owen, D., and Fisher, A. T. (2017). "The Importance of Institutional Design for
627 Distributed Local-Level Governance of Groundwater: The Case of California's Sustainable
628 Groundwater Management Act." *Water*, 9(10), 755.

629 Kirsch, B. R., Characklis, G. W., and Zeff, H. B. (2013). "Evaluating the Impact of Alternative Hydro-
630 Climate Scenarios on Transfer Agreements: Practical Improvement for Generating Synthetic
631 Streamflows." *Journal of Water Resources Planning and Management*, American Society of Civil
632 Engineers, 139(4), 396–406.

633 Kocis, T. N., and Dahlke, H. E. (2017). "Availability of high-magnitude streamflow for groundwater
634 banking in the Central Valley, California." *Environmental Research Letters*, IOP Publishing, 12(8),
635 084009.

636 Lamontagne, J. R., Reed, P. M., Marangoni, G., Keller, K., and Garner, G. G. (2019). "Robust abatement
637 pathways to tolerable climate futures require immediate global action." *Nature Climate Change*,
638 Nature Publishing Group, 9(4), 290–294.

639 Liu, X., Chen, F., Barlage, M., Zhou, G., and Niyogi, D. (2016). "Noah-MP-Crop: Introducing dynamic crop
640 growth in the Noah-MP land surface model." *Journal of Geophysical Research: Atmospheres*,
641 121(23), 2016JD025597.

642 Loucks, D. P., and van Beek, E. (2017). "Water Resource Systems Planning and Management: An
643 Introduction to Methods, Models, and Applications." D. P. Loucks and E. van Beek, eds., Springer
644 International Publishing, Cham, 1–49.

645 Malek, K., Adam, J., Stockle, C., Brady, M., and Rajagopalan, K. (2018). "When Should Irrigators Invest in
646 More Water-Efficient Technologies as an Adaptation to Climate Change?" *Water Resources
647 Research*, 54(11), 8999–9032.

648 Mann, M. E., and Gleick, P. H. (2015). "Climate change and California drought in the 21st century."
649 *Proceedings of the National Academy of Sciences*, 112(13), 3858–3859.

650 Marton, D., and Paseka, S. (2017). "Uncertainty Impact on Water Management Analysis of Open Water
651 Reservoir." *Environments*, Multidisciplinary Digital Publishing Institute, 4(1), 10.

652 Mase, A. S., Gramig, B. M., and Prokopy, L. S. (2017). "Climate change beliefs, risk perceptions, and
653 adaptation behavior among Midwestern U.S. crop farmers." *Climate Risk Management*, Useful
654 to Usable: Developing Usable Climate Science for Agriculture, 15, 8–17.

655 Medellín-Azuara, J., Vergati, J. A., Sumner, D. A., Howitt, R. E., and Lund, J. R. (2012). *Analysis of effects
656 of reduced supply of water on agricultural production and irrigation water use in Southern
657 California*. Working paper. University of California Agricultural Issues Center.

658 Mills, M., Mutafoglu, K., Adams, V. M., Archibald, C., Bell, J., and Leon, J. X. (2016). "Perceived and
659 projected flood risk and adaptation in coastal Southeast Queensland, Australia." *Climatic
660 Change*, 136(3), 523–537.

661 Molina-Navarro, E., Bailey, R. T., Andersen, H. E., Thodsen, H., Nielsen, A., Park, S., Jensen, J. S., Jensen, J.
662 B., and Trolle, D. (2019). "Comparison of abstraction scenarios simulated by SWAT and SWAT-
663 MODFLOW." *Hydrological Sciences Journal*, Taylor & Francis, 64(4), 434–454.

664 Mote, P. W., Hamlet, A. F., Clark, M. P., and Lettenmaier, D. P. (2005). "Declining mountain snowpack in
665 western North America."

666 Musselman, K. N., Lehner, F., Ikeda, K., Clark, M. P., Prein, A. F., Liu, C., Barlage, M., and Rasmussen, R.
667 (2018). "Projected increases and shifts in rain-on-snow flood risk over western North America."
668 *Nature Climate Change*, Nature Publishing Group, 8(9), 808–812.

669 Nazemi, A., and Wheeler, H. S. (2014). "How can the uncertainty in the natural inflow regime propagate
670 into the assessment of water resource systems?" *Advances in Water Resources*, 63, 131–142.

671 Nearing, G. S., Tian, Y., Gupta, H. V., Clark, M. P., Harrison, K. W., and Weijs, S. V. (2016). "A
672 philosophical basis for hydrological uncertainty." *Hydrological Sciences Journal*, 61(9), 1666–
673 1678.

674 Nishikawa, K. (2016). "The End of an Era: California's First Attempt to Manage Its Groundwater
675 Resources Through Its Sustainable Groundwater Management Act and Its Impact on Almond
676 Farmers." *Environmental Claims Journal*, 28(3), 206–222.

677 Niu, G.-Y., Yang, Z.-L., Dickinson, R. E., Gulden, L. E., and Su, H. (2007). "Development of a simple
678 groundwater model for use in climate models and evaluation with Gravity Recovery and Climate
679 Experiment data." *Journal of Geophysical Research: Atmospheres*, 112(D7).

680 Overgaard, J., Rosbjerg, D., and Butts, M. B. (2006). "Land-surface modelling in hydrological
681 perspective ? a review." *Biogeosciences*, 3(2), 229–241.

682 Panzeri, M., Riva, M., Guadagnini, A., and Neuman, S. P. (2014). "Comparison of Ensemble Kalman Filter
683 groundwater-data assimilation methods based on stochastic moment equations and Monte
684 Carlo simulation." *Advances in Water Resources*, 66, 8–18.

685 Petersen-Perlman, J. D., Veilleux, J. C., and Wolf, A. T. (2017). "International water conflict and
686 cooperation: challenges and opportunities." *Water International*, Routledge, 42(2), 105–120.

687 Quinn, J. D., Hadjimichael, A., Reed, P. M., and Steinschneider, S. (2020). "Can Exploratory Modeling of
688 Water Scarcity Vulnerabilities and Robustness Be Scenario Neutral?" *Earth's Future*, 8(11),
689 e2020EF001650.

690 Quinn, J. D., Reed, P. M., Giuliani, M., Castelletti, A., Oyler, J. W., and Nicholas, R. E. (2018). "Exploring
691 How Changing Monsoonal Dynamics and Human Pressures Challenge Multireservoir
692 Management for Flood Protection, Hydropower Production, and Agricultural Water Supply." *Water Resources Research*, 54(7), 4638–4662.

694 Rajabi, M. M., Ataie-Ashtiani, B., and Simmons, C. T. (2018). "Model-data interaction in groundwater
695 studies: Review of methods, applications and future directions." *Journal of Hydrology*, 567, 457–
696 477.

697 Refsgaard, J. C., van der Sluijs, J. P., Brown, J., and van der Keur, P. (2006). "A framework for dealing with
698 uncertainty due to model structure error." *Advances in Water Resources*, 29(11), 1586–1597.

699 Sandoval-Solis, S. (2020). "Water Resources Management in California." *Integrated Water Resource
700 Management: Cases from Africa, Asia, Australia, Latin America and USA*, E. de O. Vieira, S.
701 Sandoval-Solis, V. de A. Pedrosa, and J. P. Ortiz-Partida, eds., Springer International Publishing,
702 Cham, 35–44.

703 Sarhadi, A., Ausín, M. C., Wiper, M. P., Touma, D., and Diffenbaugh, N. S. (2018). "Multidimensional risk
704 in a nonstationary climate: Joint probability of increasingly severe warm and dry conditions." *Science Advances*, American Association for the Advancement of Science, 4(11), eaau3487.

705 Scanlon, B. R., Reedy, R. C., Faunt, C. C., Pool, D., and Uhlman, K. (2016). "Enhancing drought resilience
706 with conjunctive use and managed aquifer recharge in California and Arizona." *Environmental
707 Research Letters*, IOP Publishing, 11(3), 035013.

709 Schwartz, M., Hall, A., Sun, F., Walton, D., and Berg, N. (2017). "Significant and Inevitable End-of-
710 Twenty-First-Century Advances in Surface Runoff Timing in California's Sierra Nevada." *Journal
711 of Hydrometeorology*, American Meteorological Society, 18(12), 3181–3197.

712 Shaw, G. D., Conklin, M. H., Nimz, G. J., and Liu, F. (2014). "Groundwater and surface water flow to the
713 Merced River, Yosemite Valley, California: 36Cl and Cl⁻ evidence." *Water Resources Research*,
714 50(3), 1943–1959.

715 Skamarock, W. C., Klemp, J. B., Dudhia, J., Gill, D. O., Barker, D. M., Wang, W., and Powers, J. G. (2005). *A
716 Description of the Advanced Research WRF Version 2*.

717 Skamarock, W. C., Klemp, J., Dudhia, J., Gill, D. O., Barker, D. M., Duda, M. G., Huang, X.-Y., Wang, W.,
718 and Powers, J. (2008). "A Description of the Advanced Research WRF Version 3." NCAR Technical
719 Note.

720 Sordo-Ward, Á., Granados, I., Martín-Carrasco, F., and Garrote, L. (2016). "Impact of Hydrological
721 Uncertainty on Water Management Decisions." *Water Resources Management*, 30(14), 5535–
722 5551.

723 Steinschneider, S., Wi, S., and Brown, C. (2015). "The integrated effects of climate and hydrologic
724 uncertainty on future flood risk assessments." *Hydrological Processes*, 29(12), 2823–2839.

725 Tanaka, S. K., Zhu, T., Lund, J. R., Howitt, R. E., Jenkins, M. W., Pulido, M. A., Tauber, M., Ritzema, R. S.,
726 and Ferreira, I. C. (2006). "Climate Warming and Water Management Adaptation for California." *Climatic Change*, 76(3–4), 361–387.

728 Tang, C., and Dennis, R. L. (2014). "How reliable is the offline linkage of Weather Research & Forecasting
729 Model (WRF) and Variable Infiltration Capacity (VIC) model?" *Global and Planetary Change*, 116,
730 1–9.

731 Thompson, G., Field, P. R., Rasmussen, R. M., and Hall, W. D. (2008). "Explicit Forecasts of Winter
732 Precipitation Using an Improved Bulk Microphysics Scheme. Part II: Implementation of a New
733 Snow Parameterization." *Monthly Weather Review*, 136(12), 5095–5115.

734 Tiwari, A. D., Mukhopadhyay, P., and Mishra, V. (2021). "Influence of bias correction of meteorological
735 and streamflow forecast on hydrological prediction in India." *Journal of Hydrometeorology*,
736 American Meteorological Society, 1(aop).

737 Trabucchi, M., Fernández-García, D., and Carrera, J. (2021). "Automatic Calibration of Groundwater
738 Models With Bias Correction and Data Filtering: Working With Drawdown Data." *Water*
739 *Resources Research*, 57(3), e2020WR028097.

740 Trindade, B. C., Reed, P. M., and Characklis, G. W. (2019). "Deeply uncertain pathways: Integrated multi-
741 city regional water supply infrastructure investment and portfolio management." *Advances in*
742 *Water Resources*, 134, 103442.

743 Udmale, P., Ichikawa, Y., Manandhar, S., Ishidaira, H., and Kiem, A. S. (2014). "Farmers' perception of
744 drought impacts, local adaptation and administrative mitigation measures in Maharashtra State,
745 India." *International Journal of Disaster Risk Reduction*, 10, 250–269.

746 Vogel, R. M. (2017). "Stochastic watershed models for hydrologic risk management." *Water Security*, 1,
747 28–35.

748 Wagener, T., Sivapalan, M., Troch, P. A., McGlynn, B. L., Harman, C. J., Gupta, H. V., Kumar, P., Rao, P. S.
749 C., Basu, N. B., and Wilson, J. S. (2010). "The future of hydrology: An evolving science for a
750 changing world." *Water Resources Research*, 46(5).

751 Wrzesien, M. L., and Pavelsky, T. M. (2020). "Projected Changes to Extreme Runoff and Precipitation
752 Events From a Downscaled Simulation Over the Western United States." *Frontiers in Earth*
753 *Science*, Frontiers, 7.

754 Wrzesien, M. L., Pavelsky, T. M., Kapnick, S. B., Durand, M. T., and Painter, T. H. (2015). "Evaluation of
755 snow cover fraction for regional climate simulations in the Sierra Nevada." *International Journal*
756 *of Climatology*, 35(9), 2472–2484.

757 Wurbs, R. A. (1995). *Water Management Models: A Guide to Software*. Pearson Education.

758 Xiao, M., Koppa, A., Mekonnen, Z., Pagán, B. R., Zhan, S., Cao, Q., Aierken, A., Lee, H., and Lettenmaier,
759 D. P. (2017). "How much groundwater did California's Central Valley lose during the 2012–2016
760 drought?" *Geophysical Research Letters*, 44(10), 4872–4879.

761 Xu, X., Huang, G., Zhan, H., Qu, Z., and Huang, Q. (2012). "Integration of SWAP and MODFLOW-2000 for
762 modeling groundwater dynamics in shallow water table areas." *Journal of Hydrology*, Hydrology
763 Conference 2010, 412–413, 170–181.

764 Yang, H., and Xie, Z. (2003). "A new method to dynamically simulate groundwater table in land surface
765 model VIC." *Progress in Natural Science*, Taylor & Francis, 13(11), 819–825.

766 Yoder, J., Adam, J., Brady, M., Cook, J., Katz, S., Johnston, S., Malek, K., McMillan, J., and Yang, Q. (2017).
767 "Benefit-Cost Analysis of Integrated Water Resource Management: Accounting for
768 Interdependence in the Yakima Basin Integrated Plan." *JAWRA Journal of the American Water*
769 *Resources Association*, n/a-n/a.

770 Zeff, H. B., Hamilton, A. L., Malek, K., Herman, J. D., Cohen, J. S., Medellin-Azuara, J., Reed, P. M., and
771 Characklis, G. W. (2021). "California's Food-Energy-Water System: An Open Source Simulation
772 Model of Adaptive Surface and Groundwater Management in the Central Valley." *Environmental*
773 *Modelling & Software*, 105052.

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