

1 Combining machine learning and numerical simulation for high-resolution PM_{2.5} concentration
2 forecast

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18

19 Abstract

20 Forecasting ambient PM_{2.5} concentrations with spatiotemporal coverage is key to alerting
21 decision-makers of pollution episodes and preventing detrimental public exposure, especially in
22 regions with limited ground air monitoring stations. The existing methods either rely on chemical
23 transport models (CTMs) to forecast spatial distribution of PM_{2.5} with nontrivial uncertainty or
24 statistical algorithms to forecast PM_{2.5} concentration time-series at air monitoring locations
25 without continuous spatial coverage. In this study, we developed a PM_{2.5} forecast framework by
26 combining the robust Random Forest algorithm with a publicly accessible global CTM forecast
27 product - NASA's Goddard Earth Observing System "Composition Forecasting" (GEOS-CF),
28 providing spatiotemporally continuous PM_{2.5} concentration forecasts for the next five days at a
29 1-km spatial resolution. Our forecast experiment was conducted for a region in Central China
30 including the populous and polluted Fenwei Plain. The forecast for the next two days had overall
31 validation R² of 0.76 and 0.64, respectively; the R² was around 0.5 for the following three
32 forecast days. Spatial cross-validation showed similar validation metrics. Our forecast model,
33 with validation normalized mean bias close to zero, substantially reduced the large biases in
34 GEOS-CF. The proposed framework requires minimal computational resources compared to
35 running CTMs at urban scales, enabling near-real-time PM_{2.5} forecast in resource-restricted
36 environments.

37

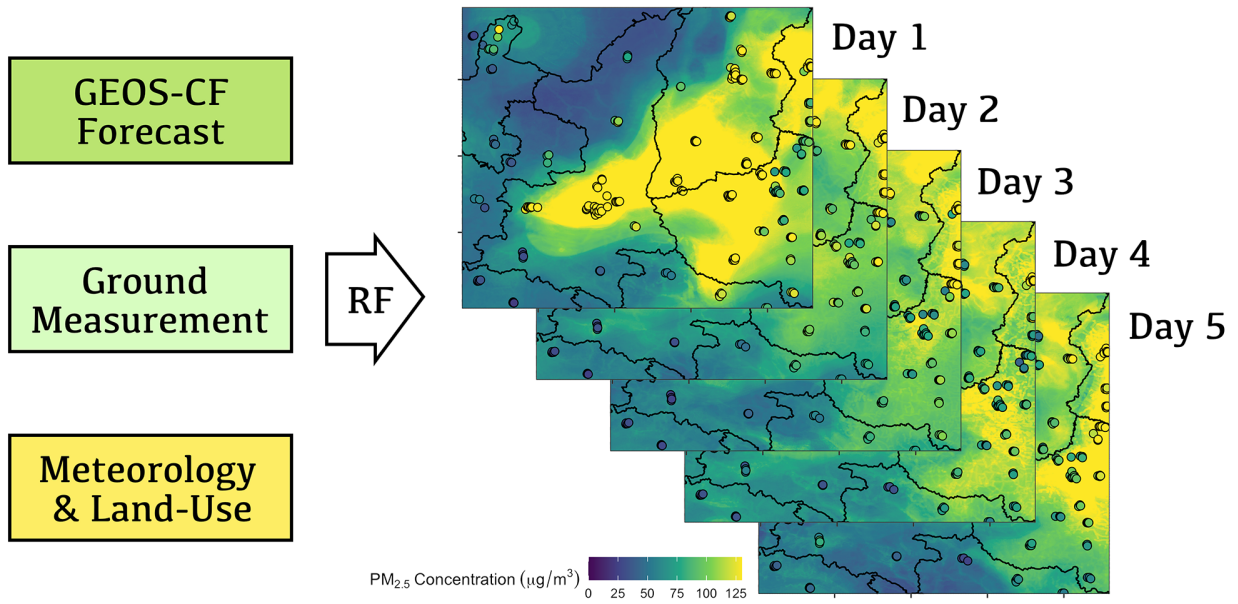
38 Keywords: Air pollution forecast; Chemical transport model; Near-real-time; Near-term;
39 Random Forest; XGBoost

40

41 Synopsis: A spatiotemporal high-resolution model for five-day PM_{2.5} concentration forecast was
42 developed by incorporating chemical transport simulations into a machine learning algorithm.

43 TOC Graphic:

Spatiotemporal Five-Day PM_{2.5} Concentration Forecast



46 1. Introduction

47 Fine particulate matter with an aerodynamic diameter of 2.5 μm or smaller ($\text{PM}_{2.5}$) can be
48 inhaled and deposit in lung alveoli. Epidemiological research has shown that $\text{PM}_{2.5}$ is detrimental
49 and casually associated with morbidity and mortality related to different body systems,
50 especially cardiovascular and respiratory systems.^{1,2} Exposure to ambient $\text{PM}_{2.5}$ was estimated to
51 contribute to 3 million deaths and 83 million disability-adjusted life years (DALYs) globally in
52 2017.³ Countries in Asia, *e.g.*, China and India, are among the regions with the highest ambient
53 $\text{PM}_{2.5}$ concentrations in the world.⁴ While comprehensive control policies have been
54 implemented and air quality has since been improved in China from the early 2010s, ambient
55 $\text{PM}_{2.5}$ concentration levels in some polluted regions are still above China's air quality standards
56 and the World Health Organization (WHO) air quality guidelines.⁵

57
58 Near-term forecast of ambient $\text{PM}_{2.5}$ concentrations is key to alerting decision-makers of
59 potential pollution episodes and preventing detrimental public exposure. Chemical transport
60 models (CTMs) have been widely used to numerically forecast spatiotemporal $\text{PM}_{2.5}$
61 concentrations in the near term - from next hours to days.^{6,7} CTMs forecast $\text{PM}_{2.5}$ concentrations
62 based on estimated emissions and simulated atmospherically physical and chemical processes.

63 Well-known CTM forecast products include those derived from global CTMs, *e.g.*, the
64 Copernicus Atmosphere Monitoring Service (CAMS)⁸ and the National Aeronautics and Space
65 Administration (NASA) Goddard Earth Observing System "Composition Forecasting" (GEOS-
66 CF),⁹ and from regional CTMs, *e.g.*, the Community Multiscale Air Quality Modeling System
67 (CMAQ).^{10,11} However, the CTM forecast products are subject to large biases due to
68 uncertainties in emission inventories, parameterization of physical and chemical processes, and

69 initial and/or boundary conditions.¹² Efforts have been made to improve CTM-based PM_{2.5}
70 forecast data.¹²⁻¹⁹ For instance, the ensemble approach utilizes multiple inputs of emission
71 inventories and meteorological fields or multiple models to reduce random errors in PM_{2.5}
72 simulations.^{12, 19} Assimilation techniques, *e.g.*, the variational (VAR) method (3D- and 4D-VAR)
73 and the Kalman filter, have also been used to incorporate ground truth (*i.e.*, PM_{2.5} observations)
74 to reduce systematic biases in PM_{2.5} simulations.^{13, 15, 17} However, the improved CTM forecast
75 products are still obviously deviated from the ground truth and are usually not able to provide
76 high-resolution forecast data at the scale of a kilometer.^{9, 20, 21} More importantly, the CTM-based
77 methods are computationally intensive and expensive, thus less practical for routine PM_{2.5}
78 forecast in resource-restricted environments.

79
80 Machine learning algorithms, as novel statistical methods, have been increasingly used to
81 forecast near-term PM_{2.5} concentrations. The majority of these algorithms are designed to
82 forecast temporal variations (*i.e.*, time-series) of PM_{2.5} at individual air monitoring sites. A
83 typical example is the recurrent neural network (RNN) and its variant, the long short-term
84 memory network (LSTM).²²⁻²⁴ Unlike the regular neural network, the RNN allows connections
85 between nodes to form a directed graph along a time sequence, therefore to process a time-series
86 of inputs. Other parametric or machine learning algorithms have also been applied in PM_{2.5} time-
87 series forecast.²⁵⁻³² The advantages of machine learning algorithms over the CTM-based methods
88 include higher forecast accuracy and substantially lower computational resources needed.
89 However, PM_{2.5} time-series forecast at air monitoring locations alone is less informative as the
90 monitors, mostly regulatory agency monitors located in urban centers, cannot well represent
91 pollution variations in suburban and rural areas.^{33, 34} Few attempts have been made to forecast

92 spatiotemporal variations of PM_{2.5} based on statistical algorithms.^{35, 36} Specifically, Ma et al³⁵
93 proposed a geo-layer of PM_{2.5} concentrations (a spatially interpolated concentration surface) and
94 integrated it into LSTM to forecast PM_{2.5} with spatiotemporally complete coverage. Lu et al³⁶
95 incorporated PM_{2.5} time-series forecast at monitoring sites from a LSTM model into a 3D-VAR
96 model to spatially extrapolate PM_{2.5} concentrations. However, the existing studies have several
97 limitations. First, the studies tended to forecast spatial PM_{2.5} variations based on inaccurate
98 spatial information, *e.g.*, geographical interpolation or CTM simulations. Exposure modeling
99 studies have shown that statistical prediction of PM_{2.5} based on ground observations and
100 meteorological/land-use predictors can generate more accurate PM_{2.5} spatial distribution than
101 spatial interpolation or chemical simulation.²⁰ Second, the multi-stage modeling process would
102 lead to error propagation which in turn increases overall modeling uncertainty.³⁷ Third, the
103 forecast model training and validation processes were not rigorously designed in these previous
104 studies, in which future PM_{2.5} observations tended to be used to train the forecast model, thus
105 improperly inflating the validation performance. A rigorous validation set should not include
106 ground observations on and after the day for which the forecast is made.

107
108 In this study, we developed a near-term PM_{2.5} forecast framework - with limited computational
109 resources needed - by combining a robust machine learning algorithm with a publicly accessible
110 global CTM forecast product. We aimed to utilize the machine learning framework to improve
111 the CTM forecast product by incorporating ground truth. Given the limitations of the existing
112 methods for PM_{2.5} forecast, we opted to use the Random Forest (RF) algorithm, a widely used
113 machine learning method for spatiotemporal PM_{2.5} prediction,³⁸⁻⁴¹ as our forecast model. Unlike
114 time-series forecast algorithms such as LSTM, we showed that RF can forecast PM_{2.5}

115 concentrations in regions without ground monitors in a unified modeling framework. The
116 proposed framework provided spatiotemporally continuous PM_{2.5} forecast data for the next five
117 days (daily averages) at a spatial resolution of 1 km. We also designed model training and
118 validation processes that can mimic real-world PM_{2.5} forecast to minimize validation biases. We
119 chose a region in Central China with a large population and one of the most polluted city clusters
120 in terms of PM_{2.5}, Fenwei Plain, as our study domain. Unlike other polluted regions in China,
121 *e.g.*, the Beijing-Tianjin-Hebei region, our study domain was less influenced by emergency air
122 pollution response and control actions, hence the proposed forecast framework could be reliably
123 validated.

124

125 2. Data and methods

126 2.1. Study domain and ground PM_{2.5} observations

127 We collected daily PM_{2.5} concentration measurements from regulatory air quality stations of the
128 China National Environmental Monitoring Center (CNEMC, <http://www.cnemc.cn>). We pre-
129 defined a 1-km modeling grid and calculated daily-level, 1-km PM_{2.5} concentrations from the
130 ground measurements by spatial aggregation. Figure 1(a) shows our study domain with the
131 locations of the PM_{2.5} monitoring sites (N of locations = 226). The study domain covered
132 multiple central and western provinces of China, including (alphabetically) Gansu, Hebei,
133 Henan, Hubei, Inner Mongolia, Ningxia, Shaanxi, Shanxi, and Sichuan. The Fenwei Plain was
134 entirely covered. The population within the study domain was estimated to be 150 million in
135 2018 (<https://landscan.ornl.gov/>). There were 97038 daily PM_{2.5} observations at 226 1-km grid
136 cells from January 1st, 2019 to March 14th, 2020. The study domain had a mean PM_{2.5}
137 concentration of 50 µg/m³ (standard deviation = 44 µg/m³) in 2019.

138

139 2.2. CTM-based PM_{2.5} forecast data

140 We acquired PM_{2.5} forecast data from a publicly accessible global CTM database, GEOS-CF, as
141 the baseline forecast data. GEOS-CF is a novel atmospheric composition and meteorology
142 model, providing three-dimensional distributions of hourly-level, five-day forecast PM_{2.5}
143 concentrations at a spatial resolution of 25 km (<https://gmao.gsfc.nasa.gov/>).⁹ While the current
144 version of GEOS-CF is known to have nontrivial systematic bias in PM_{2.5} forecast data due to
145 model representation errors, inaccurate input data (meteorology and emission), and biases in
146 chemical/physical processes, the spatial distribution of PM_{2.5} is reasonably captured.⁹ In this
147 study, we used the surface-level (two-dimensional) GEOS-CF data and calculated daily mean
148 PM_{2.5} concentrations for the five forecast days based on the China Standard Time (CST) and
149 interpolated the concentrations into the pre-defined 1-km grid by ordinary kriging. Due to the
150 difference between CST and Coordinated Universal Time (UTC) based on which GEOS-CF
151 reports the forecast, the first to fourth forecast days had complete 24-hour forecast data while the
152 fifth day had 21-hour forecast data from 12 AM to 8 PM CST.

153

154 2.3. Forecast meteorological data

155 The surface-level meteorological parameters for the five forecast days were acquired from
156 GEOS-CF as well, including total cloud area fraction (unitless), surface pressure (Pa), 10-m
157 specific humidity (kg/kg), 10-m air temperature (K), total precipitation (kg/m²/s), tropopause
158 pressure based on blended estimate (Pa), surface skin temperature (K), 10-m eastward/northward
159 wind (m/s), and planetary boundary layer height (m). We calculated daily averages of the
160 meteorological parameters and interpolated them into the pre-defined 1-km grid by ordinary

161 kriging. These meteorological parameters were used as spatiotemporally varying predictors in
162 our forecast model.

163
164 Prior to the launch of the five-day forecast, GEOS-CF runs a historical segment for the previous
165 24 hours to have the best initial conditions for the forecast. These historical estimates of the
166 recent global atmospheric composition and meteorology are constrained by meteorological
167 observations.⁹ In this analysis, we used the GEOS-CF historical data to build a “now-cast” model
168 for model parameter tuning (see Section 2.5).

169 170 2.4. Land-use data

171 We used land-use parameters as two-dimensional, spatially varying predictors of our forecast
172 model. The parameters included the LandScan ambient population in 2018 at a 900-m resolution
173 (<https://landscan.ornl.gov/>), the Copernicus Climate Change Service (C3S) global land cover
174 (LC) products in 2018 at a resolution of 0.002778° (approximately 300 m)
175 (<https://cds.climate.copernicus.eu/>), and distances to the nearest primary and secondary roads
176 extracted and computed from the OpenStreetMap (OSM) road network data
177 (<https://www.openstreetmap.org/>). The original C3S LC types were reclassified and reprocessed
178 as percentages (%) of vegetation cover, urban areas, bare areas, and water bodies. We under-
179 sampled the parameters into the pre-defined 1-km grid to match with other variables.

180 181 2.5. Forecast model training and prediction

182 Figure 1(b) shows the workflow of our forecast modeling and validation processes. The forecast
183 framework was based on the RF algorithm, a widely used algorithm providing satisfactory PM_{2.5}

184 predictions with little configuration.³⁸⁻⁴¹ The RF algorithm constructs multiple decision trees to
185 recover the non-linear relationships between the PM_{2.5} concentration and its predictors and
186 returns the mean prediction of PM_{2.5} from the individual trees as the final prediction result. We
187 focused on two major hyperparameters of RF: (1) the number of decision trees (n_{tree}) and (2)
188 the number of predictors randomly tried at each split (m_{try}). We built a current-day (“now-cast”)
189 PM_{2.5} prediction model for hyperparameter tuning (we note that this was not a forecast model;
190 this “now-cast” model was only used for hyperparameter tuning). In the current-day model,
191 ground PM_{2.5} observations were used as the dependent variable and the same-day GEOS-CF
192 meteorological variables and temporally invariant land-use parameters were used as predictors.
193 We determined the values of the hyperparameters capable of minimizing the out-of-bag (OOB)
194 error of the current-day model. Specifically, n_{tree} and m_{try} were determined to be 500 and 4,
195 respectively. Following with previous studies,^{39, 40} we relied on RF variable importance for
196 predictor selection. The RF variable importance measures explain the relative importance and
197 contribution of predictors. In this study, we opted to use the permutation variable importance
198 defined to be the decrease in model performance when a single predictor’s values are randomly
199 shuffled. We excluded predictors with importance values close to zero and substantially smaller
200 than other predictors’ values, including percentages of bare areas and water bodies. These two
201 predictors were spatially homogeneous at the monitoring locations within our study domain, thus
202 minimally contributing to model performance. Table S1 lists the final predictors used to build the
203 RF-based forecast model. The RF algorithm was based on the R (Ver. 4.0.2) package “ranger”
204 (Ver. 0.12.1).⁴²
205

206 The forecast model training process should mimic the real-world forecast scenario without future
207 data included as the training sample. Therefore, we built the forecast model for each day
208 individually on a rolling basis (as opposed to merging all training data together in a single
209 model). There are two major forecast model features: (1) the forecast day, *i.e.*, for which day the
210 forecast PM_{2.5} concentrations are generated (from the first to fifth following days), and (2) the
211 rolling period, *i.e.*, how many previous days' training data are included (we tested 10-, 30-, 60-,
212 and 90-day rolling periods). For example, on the current day (Day 0), we aimed to forecast the
213 next day's (Day 1) PM_{2.5} concentrations when the rolling period was set to be 10 days. In this
214 case, for model training, we matched the PM_{2.5} observations on Day 0 with the GEOS-CF PM_{2.5}
215 and meteorological forecast data generated on the previous day (Day -1) for Day 0 and repeated
216 this matching process for the 10-day rolling period from Day -9 to Day 0 (using GEOS-CF PM_{2.5}
217 and meteorological forecast data generated on Day -10 to Day -1); for model prediction, we then
218 used the GEOS-CF PM_{2.5} and meteorological forecast data generated on Day 0 for Day 1 to
219 calculate PM_{2.5} concentrations on Day 1 as the forecast results. The model building process is
220 summarized in Table 1. We determined the rolling period to be 60 days for our forecast model as
221 it allowed the model to have substantially higher forecast performance than those with shorter
222 rolling periods, while the improvement in forecast performance was minimal for a longer rolling
223 period (Tables S2 and S3).

224

225 We spatially interpolated the PM_{2.5} observations on the current day by ordinary kriging to create
226 a PM_{2.5} convolutional layer and treated it as an additional spatiotemporal predictor. The PM_{2.5}
227 convolutional layer is a commonly used predictor of PM_{2.5} exposure in previous modeling
228 studies.^{39, 43} It reflects the interpolated PM_{2.5} concentrations generated with nearby observations,

229 allowing the prediction model to account for spatial autocorrelation of PM_{2.5}. It is worth
230 clarifying that the “convolutional layer” here is different from a similar term in deep
231 convolutional neural networks (CNNs). Instead of a neural-network structure of CNN, our PM_{2.5}
232 convolutional layer is a two-dimensional PM_{2.5} concentration surface generated before the
233 modeling stage and was used as a model predictor. By using the PM_{2.5} convolutional layer, we
234 hypothesized that spatial variations in PM_{2.5} on the current day were correlated with the
235 variations on the forecast day, thus contributing to improved forecast performance.

236

237 Given that PM_{2.5} prediction models based on statistical methods are not designed to predict
238 extreme pollution events originated outside the study domain, *e.g.*, dust storms from northwest
239 China in our case, we removed *a priori* the training and prediction data potentially associated
240 with these extreme events. We adopted the Ultraviolet Aerosol Index (UVAI) from the
241 TROPOspheric Monitoring Instrument (TROPOMI) onboard the Sentinel-5 Precursor satellite
242 (<http://www.tropomi.eu/>) to identify extreme dust events within our study domain which
243 typically occurred in spring. The TROPOMI UVAI is calculated based on wavelength dependent
244 changes in Rayleigh scattering in the UV spectral range where ozone absorption is limited,
245 which is able to track episodic aerosol plumes from dust outbreaks, volcanic ash, and biomass
246 burning.⁴⁴ After checking the UVAI distributions on days with potential dust events, we
247 determined an empirical UVAI threshold level of 0.5 (unitless) and removed the training and
248 prediction data with UVAI values above the threshold (less than 1.2% of the data were
249 removed). A sensitivity analysis for the UVAI threshold (different values around 0.5) showed
250 that the identified training and prediction data associated with extreme dust events were robust
251 (data not shown). Using UVAI to fully identify dust events associated with increased ground

252 PM_{2.5} concentrations is challenging due to two reasons: (1) as TROPOMI only provides a single
253 snapshot of UVAI each day, it is not able to reflect the evolution of dust plumes within a day; (2)
254 UVAI captures aerosol plumes over the entire atmospheric column, which may sometimes be
255 less correlated with ground PM_{2.5}. We identified a substantial dust storm that occurred within our
256 study domain during the week of May 12th, 2019, which was not fully captured by TROPOMI
257 UVAI but significantly affected the performance of our forecast model. Therefore, we removed
258 all training and prediction data on that week (from May 12th to 18th, 2019) from our forecast
259 process.

260

261 We deployed our forecast framework on a personal computing platform with 8 virtual central
262 processing unit (CPU) cores (Intel® Xeon® CPU @ 2.00 GHz). Conducting one-day forecast
263 with a 60-day rolling period in our study domain took approximately 5 seconds, which was
264 negligible compared to generating CTM-based PM_{2.5} forecast.

265

266 2.6. Forecast model validation

267 We validated our forecast model for each day by comparing the forecast predictions with ground
268 PM_{2.5} observations. The validation was out-of-sample because the PM_{2.5} observations on the
269 forecast days were not included in the training process.

270

271 With the out-of-sample validation dataset, we designed three validation schemes: (1) an *overall*
272 validation with all validation sample over the entire modeling period to reflect the overall
273 forecast performance, (2) a *site-specific* validation to summarize forecast performance for

274 individual monitoring sites (at the 1-km grid cells), and (3) a *day-specific* validation to
275 summarize forecast model performance for individual days over the modeling period.
276
277 We used the out-of-sample coefficient of determination (R^2), root-mean-square error (RMSE),
278 mean absolute percentage error (MAPE), and normalized mean bias (NMB) as validation
279 metrics. Eq. S1 to S4 show the formulae of these metrics. R^2 and RMSE are commonly used
280 metrics in $PM_{2.5}$ exposure prediction; reporting them facilitates the comparison of our model
281 performance with other work. MAPE, with standardized values, can improve the comparability
282 of forecast performance among sites and days with different $PM_{2.5}$ concentration levels. NMB
283 can reflect the direction of the forecast bias.

284
285 Additionally, we performed 10-fold spatial cross-validation (CV) to evaluate the forecast
286 performance in regions without ground air monitors. The spatial CV randomly split the ground
287 monitors into 10 approximately equal-sized groups; one group was treated as the test set in
288 which the $PM_{2.5}$ measurements were withheld from the forecast modeling process as well as the
289 calculation of $PM_{2.5}$ convolutional layers, while the other nine groups were treated as the training
290 set. This procedure was repeated 10 times (*i.e.*, for each group). We used the same validation
291 metrics for spatial CV.

292

293 2.7. Auxiliary analyses

294 In addition to forecasting $PM_{2.5}$ concentrations as numerical values, we examined our model's
295 ability to forecast $PM_{2.5}$ pollution categories. Based on China's air quality standards,⁴⁵ we
296 classified the $PM_{2.5}$ pollution categories as clean (24-hour average $< 75 \mu\text{g}/\text{m}^3$), moderate

297 pollution (75-150 $\mu\text{g}/\text{m}^3$), and heavy pollution ($> 150 \mu\text{g}/\text{m}^3$). The categorical forecast was
298 performed by RF with the same set of predictors. We reported the accuracy of the categorical
299 forecast with two metrics, positive predictive value (PPV), *i.e.*, the probability that following a
300 positive forecast result (clean, moderate pollution, or heavy pollution), that day will truly have
301 that specific pollution level, and negative predictive value (NPV), *i.e.*, the probability that
302 following a negative forecast result, that day will truly not have that specific pollution level. PPV
303 and NPV are more intuitive than sensitivity and specificity for the public to understand the
304 categorical forecast accuracy. Eq. S5 and S6 show the formulae of the two metrics.

305

306 Furthermore, we assessed how the spatial resolution of predictors affected our forecast model
307 performance by aggregating the 1-km predictor values to 25-km means (*i.e.*, at the original
308 GEOS-CF resolution) centering around the ground monitoring locations (for model training) and
309 the centers of a 25-km grid we created (for model prediction). We compared the overall
310 validation performance and the forecast predictions of the model with 25-km predictors to those
311 with 1-km predictors. The 1-km and 25-km models shared the same ground $\text{PM}_{2.5}$ measurements
312 as the dependent variable and validation set.

313

314 We also examined another tree-based machine learning algorithm, eXtreme Gradient Boosting
315 (XGBoost), as a reference algorithm. XGBoost has been used in high-resolution $\text{PM}_{2.5}$ exposure
316 prediction with satisfactory prediction accuracy.⁴⁶ We used the same set of predictors for
317 XGBoost. We tuned three major hyperparameters of XGBoost based on cross-validation to
318 obtain an optimal model, including the number of trees, maximum depth of a tree, and learning
319 rate (η). The learning rate is related to a technique to slow down the learning in the boosting

320 process to prevent overfitting, by applying a weighting factor for the residual error corrections by
321 new trees when added to the model. The number of trees, leaning rate, and maximum depth of a
322 tree were determined to be 500, 0.1, and 2, respectively. The algorithm comparison was
323 conducted for November 2019 with the highest forecast performance for both the RF and
324 XGBoost models. The XGBoost algorithm was based on the R package “xgboost” (Ver. 1.4.1.1).

325

326 3. Results

327 3.1. Overall validation and spatial CV performance

328 Table 2 shows the overall model performance for five-day PM_{2.5} forecast over a one-year
329 validation period from March 11th, 2019 to March 10th, 2020. We chose this validation period as
330 it was the only period with a whole calendar year’s data allowing a fair comparison among the
331 five forecast days (*i.e.*, after ruling out the data over the first 60-day rolling period; otherwise, the
332 validation period would be less than a year, which was not representative of annual variations of
333 PM_{2.5}). Table 2 also compares the performance of our RF-based forecast with the original
334 GEOS-CF forecast. In general, our RF-based forecast model outperformed the GEOS-CF model
335 for all five forecast days, in which the first two days had substantially better performance with a
336 validation R² of 0.76 (over 0.56 of GEOS-CF) on the first day and 0.64 (over 0.56) on the second
337 day. Also, even though the original GEOS-CF forecast data had large biases with large validation
338 RMSE, MAPE, and NMB, our RF-based forecast model well corrected the biases with
339 considerably smaller values of the validation metrics. Moreover, as expected, our forecast model
340 performance decreased with smaller R² and larger RMSE, MAPE, and NMB when forecasting
341 the PM_{2.5} concentrations over a longer term.

342

343 Table 3 shows the forecast performance of the 10-fold spatial CV for five-day PM_{2.5} forecast
344 over a one-year validation period from March 11th, 2019 to March 10th, 2020. The spatial CV
345 was based on the same validation dataset as the overall validation. Even though the ground
346 monitoring locations where the validation was performed were withheld, the spatial CV could
347 reach a comparable performance to the overall validation (Table 2) with slightly lower R²,
348 slightly higher RMSE and MAPE, and similar NMB (close to zero). Meanwhile, the spatial CV
349 performance was better than the performance of the original GEOS-CF model for all five
350 forecast days, especially for the first two days.

351
352 Figure 2 summarizes the variable importance values of our forecast models for the five forecast
353 days. The current-day PM_{2.5} convolutional layer and GEOS-CF PM_{2.5} forecast data were the top-
354 two important variables for the first and second forecast days. On the following days, while the
355 GEOS-CF PM_{2.5} was still among the top, the importance of the convolutional layer decreased.
356 The decrease in importance is expected as the current-day PM_{2.5} concentrations tended to have
357 weaker correlations with the concentrations on the following days. We also found that the
358 forecast meteorological variables had higher importance values than the land-use variables,
359 showing the larger contributions of these spatiotemporal variables.

360
361 Figure 3 exemplifies the PM_{2.5} spatial distributions generated by our forecast model. Figure 3(a)
362 shows the PM_{2.5} concentrations from January 25th to 29th forecasted on January 24th, 2020.
363 January 25th, 2020 was the start of the holiday week of the Lunar New Year in China. This date
364 was also right after the lockdown of Wuhan (outside the study domain) due to the outbreak of
365 novel coronavirus “COVID-19”. The first day of the Lunar New Year holiday week, January

366 25th, appeared to have higher PM_{2.5} concentrations possibly associated with increased human
367 activities and fireworks. The concentration levels then decreased over the following days. As
368 expected, the high-level concentrations tended to be in and around the populous Fenwei Plain.
369 Compared to the ground observations, our forecast data were shown to well capture the
370 spatiotemporal variations of PM_{2.5} over the period. Figure 3(b) shows the PM_{2.5} concentrations
371 from May 12th to 16th forecasted on May 11th, 2019. This is a negative example as the forecast
372 data were not able to capture the strong dust storm event that occurred in the western part of our
373 study domain (Gansu, Ningxia, and Shaanxi) during the period, when the ground observations
374 appeared to be high. This example illustrates our forecast model's limitation to capture sudden
375 extreme events that originated outside the domain.

376

377 3.2. Site-specific and day-specific validation performance

378 The site- and day-specific validation performance for the five forecast days, with the comparison
379 to the GEOS-CF forecast performance, is shown in Figure 4. Tables S4 and S5 summarize the
380 site- and day-specific validation metrics, respectively. Our forecast model was shown to
381 substantially improve the forecast accuracy and precision of the original GEOS-CF forecast data.

382

383 For the site-specific validation (Figure 4(a)), our forecast model had higher R² for the first two
384 forecast days (with a median > 0.7 for the first day and > 0.6 for the second day) over the GEOS-
385 CF forecast model (with medians around 0.5); the validation R² values of the two models were
386 comparable for the following days (with medians around or below 0.5). The interquartile ranges
387 (IQR) of R² of our model were narrower for all five forecast days, indicating the robustness of
388 the model. Our forecast model, with considerably lower RMSE (with medians < 30 µg/m³),

389 MAPE (with medians < 50%), and NMB (with medians around zero), corrected the large biases
390 in the GEOS-CF data.

391
392 For the day-specific validation (Figure 4(b)), our forecast model had higher R^2 than the GEOS-
393 CF forecast model for all five forecast days. The day-specific validation had wider IQRs of R^2
394 than the site-specific validation, indicating a greater challenge of our model to forecast spatial
395 variability of $PM_{2.5}$ than its temporal variability, aligning with previous RF-based “now-cast”
396 models.^{39, 40} As in the site-specific validation, our model, with substantially lower RMSE (with
397 medians < 20 $\mu\text{g}/\text{m}^3$), MAPE (with medians < 50%), and NMB (with medians close to zero),
398 corrected the large biases in the GEOS-CF data.

399
400 Figure 5(a) shows the day-specific validation MAPE values with daily-mean $PM_{2.5}$
401 concentrations (using the first forecast day as an example). The daily MAPE variation displayed
402 a pattern: MAPE tended to increase right after a sudden decrease in $PM_{2.5}$ concentrations.
403 Figures 5(b) and (c) show the GEOS-CF wind speeds and directions (wind roses) within the
404 study domain on days with validation R^2 above its 95th percentile (*i.e.*, good forecast
405 performance) and below its 5th percentile (*i.e.*, poor forecast performance), respectively. The
406 wind roses indicate that when the forecast models had an unsatisfactory performance, the
407 dominant wind direction tended to be northeast with higher wind speeds. In comparison, there
408 was not an obvious dominant wind direction when the models had a good performance. The
409 association between sudden decreases in $PM_{2.5}$ and gusts of high-speed, northeast winds
410 indicates that the northeast winds might bring relatively clean air to the study domain, therefore
411 rapidly and temporarily eliminating $PM_{2.5}$ pollution. This result reflects a reduced forecast ability

412 of our framework for sudden decreases in $PM_{2.5}$ resulting from wind elimination originated
413 outside the domain.

414

415 3.3. Categorical forecast performance

416 Table 4 shows the forecast performance for categorical pollution levels (clean, moderate
417 pollution, and heavy pollution) as well as the comparison between the original GEOS-CF and
418 our RF-based forecast models. The original GEOS-CF model could not forecast well both the
419 moderate and heavy pollution categories due to their large biases (with extremely low PPVs). In
420 comparison, our RF-based data substantially improved the forecast accuracy for both categories
421 with higher PPVs. The corresponding NPVs for the clean category increased as well. The clean
422 category had the largest number of training sample ($N = \sim 66000$; the number varied for different
423 forecast days) with high PPVs ($\sim 90\%$) and NPVs ($\sim 70 - 80\%$) for all five forecast days. With
424 considerably fewer training sample, the moderate- ($N = \sim 9800$) and heavy-pollution ($N = \sim 2300$)
425 categories had lower PPVs with decreased performance for longer forecast days. The NPVs for
426 these two pollution categories were above 90% for all five forecast days.

427

428 3.4. Spatial resolution of predictors

429 Table S6 shows the overall model performance for five-day $PM_{2.5}$ forecast with 25-km predictors
430 over a one-year validation period from March 11th, 2019 to March 10th, 2020. Figure S1 shows
431 an example of spatial forecast concentrations derived with 1-km and 25-km predictors (the next-
432 day $PM_{2.5}$ concentrations forecasted on January 24th, 2020). The overall validation performance
433 was not substantially affected by the coarser 25-km resolution, with slightly lower R^2 and higher
434 RMSE, MAPE, and NMB than the 1-km metrics. This comparison indicates that the original

435 resolution of the spatiotemporal GEOS-CF variables might limit our model performance even
436 after we interpolated them to 1-km. However, the forecast concentration surfaces exhibited
437 different spatial patterns, where the 1-km concentration surface reflected substantially finer
438 details of PM_{2.5} distribution (associated with elevation, traffic, *etc.*) because the model took
439 much greater advantage of high-resolution land-use information.

440

441 3.5. Comparison with XGBoost

442 Table S7 compares the forecast performance of the RF and XGBoost models in November 2019.
443 Both models had similar validation R² and MAPE. Although RF slightly outperformed XGBoost,
444 the differences between the two algorithms were not meaningful. Therefore, we expect these two
445 tree-based algorithms can be interchangeable for our proposed forecast framework. We opted to
446 use the RF algorithm due to its easy configuration with fewer major hyperparameters and its
447 ability to provide robust predictions without much tuning effort.

448

449 4. Discussion

450 In this study, we proposed a RF-based framework for the near-term (next five days), daily-mean
451 PM_{2.5} concentration forecast at a 1-km spatial resolution. We also designed model training and
452 validation processes that can mimic the real-world forecast scenario to minimize validation
453 biases. All input data of our forecast framework are publicly accessible, including ground PM_{2.5}
454 observations, GEOS-CF PM_{2.5} and meteorological forecast data, and land-use parameters. The
455 forecast framework requires minimal computational resources and can be deployed in personal
456 computing platforms. While the framework was evaluated in China with a satisfactory number of
457 regulatory air monitoring stations in this study, we expect that it can also be deployed in

458 resource-restricted environments in conjunction with a growing number of ground measurements
459 from low-cost air quality monitors (when rigorously calibrated).³⁴ We note that our framework
460 provides near-real-time rather than real-time forecast as the forecast product is generated at the
461 end of each day when ground observations are collected and reported. However, as our
462 framework provides rapid five-day forecast (at a scale of seconds for our study domain), the
463 potential influence on heavy pollution awareness and response due to this level of delay is
464 minimal.

465

466 The GEOS-CF PM_{2.5} forecast data had large systematic biases as shown in this study (Table 2)
467 and a previous evaluation study for a number of reasons, including model representation errors,
468 uncertainties in the meteorology, and biases arising from errors in the treatment of emissions,
469 deposition, or atmospheric chemistry.⁹ The overall, site-specific, and day-specific validations
470 showed that our statistical framework substantially improved the GEOS-CF data and generated
471 acceptable PM_{2.5} forecast concentrations, especially for the first two forecast days (Table 2 and
472 Figure 4). While the third to fifth forecast days had comparable validation R² with the original
473 GEOS-CF model, the large biases in the GEOS-CF data were well corrected. The spatial CV
474 showed similar validation metrics to the overall validation (Table 3), indicating that our forecast
475 framework was able to provide reliable forecast results in regions without ground monitors. This
476 is a unique advantage of our RF-based framework over the widely adopted time-series forecast
477 methods such as RNN and LSTM, which perform air pollution forecast only at ground
478 monitoring locations based on their historical measurements. Intuitively, our modeling
479 framework can be seen as a statistical “calibration” for the GEOS-CF forecast product by
480 building a statistical model with “gold-standard” PM_{2.5} observations as the dependent variable

481 and the uncertain GEOS-CF forecast data as an independent variable with additional
482 meteorological and land-use parameters as covariates, the concept of which is similar to low-cost
483 air monitor calibration.⁴⁷

484
485 The variable importance rankings suggested that the GEOS-CF PM_{2.5} forecast data were always
486 among the top important predictors (Figure 2), indicating that this product, although biased, was
487 the key input of our forecast model as it provided meaningful information regarding PM_{2.5} spatial
488 distribution. We opted to use the forecast data from the GEOS-CF model with a relatively coarse
489 spatial resolution because it was openly accessible with global coverage. We showed that the
490 forecast concentrations greatly benefited from the interpolated GEOS-CF predictors at a higher
491 spatial resolution, where detailed spatial patterns of PM_{2.5} could be reflected more clearly (Figure
492 S1). Meanwhile, we also expect that a regional CTM model at a higher spatial resolution, with
493 proper boundary conditions, may further improve our forecast performance. According to the
494 importance rankings, the current-day PM_{2.5} convolutional layer contributed to an improved
495 forecast performance, especially for the first two forecast days. This finding proves that the
496 PM_{2.5} convolutional layer is not only informative for the same-day prediction as shown in the
497 previous studies,^{39, 43} but for the near-term forecast as well (due to the correlations between the
498 current-day and future PM_{2.5} concentrations).

499
500 Categorical pollution levels are more intuitive than continuous concentrations for public
501 awareness and emergency response to air pollution. Although the original GEOS-CF product had
502 large biases in forecasting categorical PM_{2.5} levels (clean, moderate pollution, and heavy
503 pollution), our forecast model was able to substantially improving the forecast, especially for the

504 first two to three forecast days (Table 4). The clean-day forecast had the highest accuracy for all
505 five forecast days (with PPVs close or greater than 90%) possibly because the majority of the
506 training data were in this category. Similarly, the moderate- and heavy-pollution forecast had
507 higher NPVs than PPVs because of the fewer training data in these categories. While the heavy-
508 pollution forecast had a PPV of ~70% on the first forecast day, the forecast accuracy decreased
509 quickly on the fourth and fifth days. This pattern indicates a greater challenge of our framework
510 to forecast high-level pollution over a longer term, which is worth further improvements.

511
512 The rolling period, *i.e.*, the number of previous days on which the PM_{2.5} measurements are
513 included as the training sample, was a key forecast model feature. We found that although a
514 longer rolling period was associated with an increased forecast performance (Table S2), the
515 increase was marginal when the rolling period was greater than 60 days (Table S3). Hence, we
516 suggest that the two-month rolling period was optimal for our study domain and period, offering
517 satisfactory forecast performance while minimizing the number of training data included. When
518 applying the framework to other regions and periods, the optimal value of the rolling period
519 should be re-evaluated according to forecast accuracy.

520
521 CTMs, though with higher uncertainties resulting from inaccurate emission inventories,
522 atmospherically physical and chemical processes, and initial and boundary conditions, have been
523 the dominant tool to forecast near-term PM_{2.5} concentrations with spatiotemporal complete
524 coverage.^{6, 7} Few statistical efforts have been made to build more accurate spatiotemporal
525 forecast models based on the ground truth (*i.e.*, PM_{2.5} observations). Recently, Ma et al³⁵ and Lu
526 et al³⁶ proposed statistical/empirical methods to forecast spatiotemporally complete PM_{2.5}

527 concentrations. The advantages of our proposed forecast framework over these studies are two-
528 fold. First, our machine learning framework can generate more reliable spatial distributions of
529 PM_{2.5} than the distributions generated by spatial interpolation (*e.g.*, the geo-layer in Ma et al ³⁵)
530 and the spatial information provided by CTM (*e.g.*, the simulation method used in Lu et al ³⁶).
531 This advantage has been proven in previous PM_{2.5} prediction studies using statistical
532 algorithms.²⁰ Second, we proposed a more rigorous model training process by building daily
533 forecast models on a rolling basis. This strategy guaranteed that no future PM_{2.5} observations
534 (*i.e.*, observations beyond the current day when the forecast is conducted) were included as the
535 training sample. In contrast, if the observations across the entire period are randomly separated
536 into a training set and a test set, the training set is very likely to include some same-day
537 observations from the test set. In that case, the validation performance may be improperly
538 inflated as the same-day observations are likely to be informative of the forecast on this day
539 (even if the same-day training and validation samples are not at the same monitoring locations,
540 the training locations can still be informative if they are geographically proximate to the
541 validation locations).

542

543 The major limitation of our forecast framework is the limited ability to forecast PM_{2.5} associated
544 with out-of-domain factors, *e.g.*, extreme dust storms from the desert regions north/northwest to
545 the domain and the sudden pollution elimination process associated with strong northeast winds.
546 Without proper indicators of these out-of-domain factors, statistical models alone can hardly
547 capture the associated pollution variations.⁴⁸ While CTM is supposed to have the ability to
548 forecast these physical processes, the global GEOS-CF model was shown to unsatisfactorily
549 simulate these processes in this study. We expect the forecast data from regional CTMs at a finer

550 spatial resolution with better emission information and more accurate physical simulation
551 processes, *e.g.*, CMAQ, may help our framework better capture and forecast these sudden events.
552 It is also worth exploring the use of outputs from trajectory models, *e.g.*, the Hybrid Single-
553 Particle Lagrangian Integrated Trajectory (HYSPLIT) model,^{49, 50} in improving the forecast of
554 sudden events with our framework. Additionally, the spatial interpolation of GEOS-CF PM_{2.5}
555 and meteorological forecast parameters based on ordinary kriging (to oversample them to the 1-
556 km resolution) may not accurately reflect small-scale, terrain-related variations in the
557 parameters, especially in mountainous areas. However, the potential interpolation uncertainty
558 should have a limited influence on the PM_{2.5} forecast as the uncertainty is likely to be
559 substantially smaller than the CTM-related uncertainty in these parameters. Additional effort is
560 needed to further reduce the potential interpolation uncertainty.

561
562 In summary, this study is among the first to generate high-resolution (1-km), near-term (next five
563 days), and near-real-time PM_{2.5} forecast data based on a robust machine learning framework.
564 While we showcased the forecast ability of our framework in a populated region of Central
565 China with high-level PM_{2.5} pollution, we expect that the framework can be generalized to other
566 regions and for other air pollutants, *e.g.*, ozone and nitrogen dioxide, based on the same input
567 data sources.⁵¹ Our proposed framework with near-real-time forecast products holds promise for
568 improved public awareness, policy development, and emergency response regarding detrimental
569 air pollution exposure.

570
571 Supporting Information

572 Six equations, seven tables, and a figure, providing additional information regarding PM_{2.5}
573 forecast model evaluation methods and results.

574

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- 580 1. Bourdrel, T.; Bind, M.-A.; Béjot, Y.; Morel, O.; Argacha, J.-F., Cardiovascular effects of
581 air pollution. *Archives of Cardiovascular Diseases* **2017**, *110*, (11), 634-642.
- 582 2. Lu, F.; Xu, D.; Cheng, Y.; Dong, S.; Guo, C.; Jiang, X.; Zheng, X., Systematic review
583 and meta-analysis of the adverse health effects of ambient PM_{2.5} and PM₁₀ pollution in the
584 Chinese population. *Environmental Research* **2015**, *136*, 196-204.
- 585 3. Bu, X.; Xie, Z.; Liu, J.; Wei, L.; Wang, X.; Chen, M.; Ren, H., Global PM_{2.5}-attributable
586 health burden from 1990 to 2017: Estimates from the Global Burden of disease study 2017.
587 *Environmental Research* **2021**, *197*, 111123.
- 588 4. Lim, C.-H.; Ryu, J.; Choi, Y.; Jeon, S. W.; Lee, W.-K., Understanding global PM_{2.5}
589 concentrations and their drivers in recent decades (1998–2016). *Environment International* **2020**,
590 *144*, 106011.
- 591 5. Liang, F.; Xiao, Q.; Huang, K.; Yang, X.; Liu, F.; Li, J.; Lu, X.; Liu, Y.; Gu, D., The 17-
592 y spatiotemporal trend of PM_{2.5} and its mortality burden in China. *Proceedings of the National*
593 *Academy of Sciences* **2020**, *117*, (41), 25601-25608.
- 594 6. Cheng, X. H.; Liu, Y. L.; Xu, X. D.; You, W.; Zang, Z. L.; Gao, L. N.; Chen, Y. B.; Su,
595 D. B.; Yan, P., Lidar data assimilation method based on CRTM and WRF-Chem models and its
596 application in PM_{2.5} forecasts in Beijing. *Science of the Total Environment* **2019**, *682*, 541-552.
- 597 7. Wu, C. B.; Li, K.; Bai, K. X., Validation and Calibration of CAMS PM_{2.5} Forecasts
598 Using In Situ PM_{2.5} Measurements in China and United States. *Remote Sensing* **2020**, *12*, (22),
599 19.
- 600 8. Flemming, J.; Benedetti, A.; Inness, A.; Engelen, R. J.; Jones, L.; Huijnen, V.; Remy, S.;
601 Parrington, M.; Suttie, M.; Bozzo, A.; Peuch, V. H.; Akritidis, D.; Katragkou, E., The CAMS
602 interim Reanalysis of Carbon Monoxide, Ozone and Aerosol for 2003–2015. *Atmos. Chem.*
603 *Phys.* **2017**, *17*, (3), 1945-1983.
- 604 9. Keller, C. A.; Knowland, K. E.; Duncan, B. N.; Liu, J.; Anderson, D. C.; Das, S.;
605 Lucchesi, R. A.; Lundgren, E. W.; Nicely, J. M.; Nielsen, E.; Ott, L. E.; Saunders, E.; Strode, S.
606 A.; Wales, P. A.; Jacob, D. J.; Pawson, S., Description of the NASA GEOS Composition
607 Forecast Modeling System GEOS-CF v1.0. *Journal of Advances in Modeling Earth Systems*
608 **2021**, *13*, (4), e2020MS002413.
- 609 10. Cheng, F. Y.; Feng, C. Y.; Yang, Z. M.; Hsu, C. H.; Chan, K. W.; Lee, C. Y.; Chang, S.
610 C., Evaluation of real-time PM_{2.5} forecasts with the WRF-CMAQ modeling system and
611 weather-pattern-dependent bias-adjusted PM_{2.5} forecasts in Taiwan. *Atmos Environ* **2021**, *244*,
612 17.
- 613 11. Sayeed, A.; Lops, Y.; Choi, Y.; Jung, J.; Salman, A. K., Bias correcting and extending
614 the PM forecast by CMAQ up to 7 days using deep convolutional neural networks. *Atmos*
615 *Environ* **2021**, *253*, 9.
- 616 12. Zhang, H.; Wang, J.; García, L. C.; Ge, C.; Plessel, T.; Szykman, J.; Murphy, B.; Spero,
617 T. L., Improving Surface PM_{2.5} Forecasts in the United States Using an Ensemble of Chemical
618 Transport Model Outputs: 1. Bias Correction With Surface Observations in Nonrural Areas. *J*
619 *Geophys Res-Atmos* **2020**, *125*, (14), e2019JD032293.
- 620 13. Feng, S. Z.; Jiang, F.; Jiang, Z. Q.; Wang, H. M.; Cai, Z.; Zhang, L., Impact of 3DVAR
621 assimilation of surface PM_{2.5} observations on PM_{2.5} forecasts over China during wintertime.
622 *Atmos Environ* **2018**, *187*, 34-49.

- 623 14. June, N.; Vaughan, J.; Lee, Y.; Lamb, B. K., Operational bias correction for PM_{2.5} using
624 the AIRPACT air quality forecast system in the Pacific Northwest. *Journal of the Air & Waste*
625 *Management Association* **2021**, *71*, (4), 515-527.
- 626 15. Kong, Y. W.; Sheng, L. F.; Li, Y. P.; Zhang, W. H.; Zhou, Y.; Wang, W. C.; Zhao, Y. H.,
627 Improving PM_{2.5} forecast during haze episodes over China based on a coupled 4D-LETKF and
628 WRF-Chem system. *Atmos. Res.* **2021**, *249*, 14.
- 629 16. Liang, Y. F.; Zang, Z. L.; Liu, D.; Yan, P.; Hu, Y. W.; Zhou, Y.; You, W., Development
630 of a three-dimensional variational assimilation system for lidar profile data based on a size-
631 resolved aerosol model in WRF-Chem model v3.9.1 and its application in PM_{2.5} forecasts
632 across China. *Geosci. Model Dev.* **2020**, *13*, (12), 6285-6301.
- 633 17. Peng, Z.; Liu, Z. Q.; Chen, D.; Ban, J. M., Improving PM_{2.5} forecast over China by the
634 joint adjustment of initial conditions and source emissions with an ensemble Kalman filter.
635 *Atmos Chem Phys* **2017**, *17*, (7), 4837-4855.
- 636 18. Zheng, H. T.; Liu, J. G.; Tang, X.; Wang, Z. F.; Wu, H. J.; Yan, P. Z.; Wang, W.,
637 Improvement of the Real-time PM_{2.5} Forecast over the Beijing-Tianjin-Hebei Region using an
638 Optimal Interpolation Data Assimilation Method. *Aerosol Air Qual Res* **2018**, *18*, (5), 1305-
639 1316.
- 640 19. Ge, C.; Wang, J.; Reid, J. S.; Posselt, D. J.; Xian, P.; Hyer, E., Mesoscale modeling of
641 smoke transport from equatorial Southeast Asian Maritime Continent to the Philippines: First
642 comparison of ensemble analysis with in situ observations. *J Geophys Res-Atmos* **2017**, *122*,
643 (10), 5380-5398.
- 644 20. Chu, Y.; Liu, Y.; Li, X.; Liu, Z.; Lu, H.; Lu, Y.; Mao, Z.; Chen, X.; Li, N.; Ren, M.; Liu,
645 F.; Tian, L.; Zhu, Z.; Xiang, H., A Review on Predicting Ground PM_{2.5} Concentration Using
646 Satellite Aerosol Optical Depth. *Atmosphere-Basel* **2016**, *7*, (10), 129.
- 647 21. Lightstone, S. D.; Moshary, F.; Gross, B., Comparing CMAQ Forecasts with a Neural
648 Network Forecast Model for PM_{2.5} in New York. *Atmosphere-Basel* **2017**, *8*, (9), 161.
- 649 22. Huang, C. J.; Kuo, P. H., A Deep CNN-LSTM Model for Particulate Matter (PM_{2.5})
650 Forecasting in Smart Cities. *Sensors* **2018**, *18*, (7), 22.
- 651 23. Liu, H.; Duan, Z.; Chen, C., A hybrid multi-resolution multi-objective ensemble model
652 and its application for forecasting of daily PM_{2.5} concentrations. *Inf. Sci.* **2020**, *516*, 266-292.
- 653 24. Zhou, Y.; Chang, F.-J.; Chang, L.-C.; Kao, I. F.; Wang, Y.-S., Explore a deep learning
654 multi-output neural network for regional multi-step-ahead air quality forecasts. *J. Clean Prod.*
655 **2019**, *209*, 134-145.
- 656 25. Mahajan, S.; Chen, L. J.; Tsai, T. C., Short-Term PM_{2.5} Forecasting Using Exponential
657 Smoothing Method: A Comparative Analysis. *Sensors* **2018**, *18*, (10), 15.
- 658 26. Niu, M.; Wang, Y.; Sun, S.; Li, Y., A novel hybrid decomposition-and-ensemble model
659 based on CEEMD and GWO for short-term PM_{2.5} concentration forecasting. *Atmos Environ*
660 **2016**, *134*, 168-180.
- 661 27. Qin, S.; Liu, F.; Wang, J.; Sun, B., Analysis and forecasting of the particulate matter
662 (PM) concentration levels over four major cities of China using hybrid models. *Atmos Environ*
663 **2014**, *98*, 665-675.
- 664 28. Bai, Y.; Li, Y.; Zeng, B.; Li, C.; Zhang, J., Hourly PM_{2.5} concentration forecast using
665 stacked autoencoder model with emphasis on seasonality. *J. Clean Prod.* **2019**, *224*, 739-750.
- 666 29. Franceschi, F.; Cobo, M.; Figueredo, M., Discovering relationships and forecasting
667 PM₁₀ and PM_{2.5} concentrations in Bogotá, Colombia, using Artificial Neural Networks,

668 Principal Component Analysis, and k-means clustering. *Atmos. Pollut. Res.* **2018**, *9*, (5), 912-
669 922.

670 30. Sun, W.; Li, Z. Q., Hourly PM_{2.5} concentration forecasting based on feature extraction
671 and stacking-driven ensemble model for the winter of the Beijing-Tianjin-Hebei area. *Atmos.*
672 *Pollut. Res.* **2020**, *11*, (6), 110-121.

673 31. Ventura, L. M. B.; Pinto, F. D.; Soares, L. M.; Luna, A. S.; Gioda, A., Forecast of daily
674 PM_{2.5} concentrations applying artificial neural networks and Holt-Winters models. *Air Qual.*
675 *Atmos. Health* **2019**, *12*, (3), 317-325.

676 32. Zhou, Y. L.; Chang, F. J.; Chang, L. C.; Kao, I. F.; Wang, Y. S.; Kang, C. C., Multi-
677 output support vector machine for regional multi-step-ahead PM_{2.5} forecasting. *Science of the*
678 *Total Environment* **2019**, *651*, 230-240.

679 33. Bi, J.; Stowell, J.; Seto, E. Y. W.; English, P. B.; Al-Hamdan, M. Z.; Kinney, P. L.;
680 Freedman, F. R.; Liu, Y., Contribution of low-cost sensor measurements to the prediction of
681 PM_{2.5} levels: A case study in Imperial County, California, USA. *Environmental Research* **2020**,
682 *180*, 108810.

683 34. Bi, J.; Wildani, A.; Chang, H. H.; Liu, Y., Incorporating Low-Cost Sensor Measurements
684 into High-Resolution PM_{2.5} Modeling at a Large Spatial Scale. *Environ Sci Technol* **2020**, *54*,
685 (4), 2152-2162.

686 35. Ma, J.; Ding, Y.; Cheng, J. C. P.; Jiang, F.; Wan, Z., A temporal-spatial interpolation and
687 extrapolation method based on geographic Long Short-Term Memory neural network for PM_{2.5}.
688 *J. Clean Prod.* **2019**, *237*, 117729.

689 36. Lu, X. C.; Sha, Y. H.; Li, Z. N.; Huang, Y. Q.; Chen, W. Y.; Chen, D. H.; Shen, J.; Chen,
690 Y.; Fung, J. C. H., Development and application of a hybrid long-short term memory - three
691 dimensional variational technique for the improvement of PM_{2.5} forecasting. *Science of the*
692 *Total Environment* **2021**, *770*, 10.

693 37. Pu, Q.; Yoo, E. H., Ground PM_{2.5} prediction using imputed MAIAC AOD with
694 uncertainty quantification. *Environmental Pollution* **2021**, *274*, 9.

695 38. Brokamp, C.; Jandarov, R.; Hossain, M.; Ryan, P., Predicting Daily Urban Fine
696 Particulate Matter Concentrations Using a Random Forest Model. *Environ Sci Technol* **2018**, *52*,
697 (7), 4173-4179.

698 39. Hu, X.; Belle, J. H.; Meng, X.; Wildani, A.; Waller, L. A.; Strickland, M. J.; Liu, Y.,
699 Estimating PM_{2.5} Concentrations in the Conterminous United States Using the Random Forest
700 Approach. *Environ Sci Technol* **2017**, *51*, (12), 6936-6944.

701 40. Bi, J.; Belle, J. H.; Wang, Y.; Lyapustin, A. I.; Wildani, A.; Liu, Y., Impacts of snow and
702 cloud covers on satellite-derived PM_{2.5} levels. *Remote Sens Environ* **2019**, *221*, 665-674.

703 41. Huang, K.; Bi, J.; Meng, X.; Geng, G.; Lyapustin, A.; Lane, K. J.; Gu, D.; Kinney, P. L.;
704 Liu, Y., Estimating daily PM_{2.5} concentrations in New York City at the neighborhood-scale:
705 Implications for integrating non-regulatory measurements. *Science of the Total Environment*
706 **2019**, *697*, 134094.

707 42. Wright, M. N.; Ziegler, A., ranger: A Fast Implementation of Random Forests for High
708 Dimensional Data in C++ and R. *Journal of Statistical Software* **2017**, *77*, (1), 1-17.

709 43. Di, Q.; Kloog, I.; Koutrakis, P.; Lyapustin, A.; Wang, Y.; Schwartz, J., Assessing PM_{2.5}
710 Exposures with High Spatiotemporal Resolution across the Continental United States. *Environ*
711 *Sci Technol* **2016**, *50*, (9), 4712-4721.

- 712 44. Kooreman, M. L.; Stammes, P.; Trees, V.; Sneep, M.; Tilstra, L. G.; de Graaf, M.; Stein
713 Zweers, D. C.; Wang, P.; Tuinder, O. N. E.; Veefkind, J. P., Effects of clouds on the UV
714 Absorbing Aerosol Index from TROPOMI. *Atmos. Meas. Tech.* **2020**, *13*, (12), 6407-6426.
- 715 45. Wang, S.; Hao, J., Air quality management in China: Issues, challenges, and options.
716 *Journal of Environmental Sciences* **2012**, *24*, (1), 2-13.
- 717 46. Xiao, Q.; Chang, H. H.; Geng, G.; Liu, Y., An Ensemble Machine-Learning Model To
718 Predict Historical PM_{2.5} Concentrations in China from Satellite Data. *Environ Sci Technol* **2018**,
719 *52*, (22), 13260-13269.
- 720 47. Barkjohn, K. K.; Gantt, B.; Clements, A. L., Development and application of a United
721 States-wide correction for PM_{2.5} data collected with the PurpleAir sensor. *Atmos. Meas. Tech.*
722 **2021**, *14*, (6), 4617-4637.
- 723 48. Liou, N. C.; Luo, C. H.; Mahajan, S.; Chen, L. J., Why is Short-Time PM_{2.5} Forecast
724 Difficult? The Effects of Sudden Events. *IEEE Access* **2020**, *8*, 12662-12674.
- 725 49. Stein, A.; Draxler, R. R.; Rolph, G. D.; Stunder, B. J.; Cohen, M.; Ngan, F., NOAA's
726 HYSPLIT atmospheric transport and dispersion modeling system. *B Am Meteorol Soc* **2015**, *96*,
727 (12), 2059-2077.
- 728 50. Li, Y.; Tong, D. Q.; Ngan, F.; Cohen, M. D.; Stein, A. F.; Kondragunta, S.; Zhang, X.;
729 Ichoku, C.; Hyer, E. J.; Kahn, R. A., Ensemble PM_{2.5} Forecasting During the 2018 Camp Fire
730 Event Using the HYSPLIT Transport and Dispersion Model. *J. Geophys. Res.-Atmos.* **2020**, *125*,
731 (15), 19.
- 732 51. Malings, C.; Knowland, K. E.; Keller, C. A.; Cohn, S. E., Sub-City Scale Hourly Air
733 Quality Forecasting by Combining Models, Satellite Observations, and Ground Measurements.
734 *Earth and Space Science* **2021**, *8*, (7), e2021EA001743.
- 735

736 Table 1: The forecast model building process by matching ground PM_{2.5} observations with the PM_{2.5} convolutional layer and GEOS-
 737 CF forecast data (PM_{2.5} pollution and meteorology forecast) as training and prediction data. N is the rolling period (N = 60 days). Day
 738 0 is the present day, Day 1 is the next day, *etc.* The CTM running date is the day on which the CTM is run. The CTM forecast date is
 739 the day the forecast is made for.

Forecast Day	Date of PM _{2.5} Convolutional Layer	Training		Prediction	
		CTM Running Date	PM _{2.5} Observation & CTM Forecast Date	CTM Running Date	CTM Forecast Date
Day 1	Day 0	Day -N to -1	Day -(N-1) to 0	Day 0	Day 1
Day 2		Day -(N+1) to -2			Day 2
Day 3		Day -(N+2) to -3			Day 3
Day 4		Day -(N+3) to -4			Day 4
Day 5		Day -(N+4) to -5			Day 5

740

741 Table 2: The overall validation performance of our forecast framework (RF + GEOS-CF) and the
 742 original CTM forecast model (GEOS-CF) for the validation period of March 11th, 2019 to March
 743 10th, 2020. The rolling period was 60 days.

Forecast Day	N of Test Sample	R ²	RMSE (µg/m ³)	MAPE (%)	NMB
RF + GEOS-CF					
Day 1	78378	0.76	18.70	34.3	0.003
Day 2	78398	0.64	23.07	43.2	0.008
Day 3	78158	0.56	25.48	46.8	0.005
Day 4	78372	0.51	26.70	48.9	-0.001
Day 5	78372	0.47	27.81	52.4	-0.003
GEOS-CF					
Day 1	78378	0.56	140.76	278.1	2.23
Day 2	78398	0.56	141.46	277.3	2.224
Day 3	78158	0.53	145.55	279.4	2.246
Day 4	78372	0.50	144.09	280.0	2.216
Day 5	78372	0.45	141.87	278.8	2.139

744

745 Table 3: The 10-fold spatial CV performance of our forecast framework (RF + GEOS-CF) for
 746 the validation period of March 11th, 2019 to March 10th, 2020. The rolling period was 60 days.

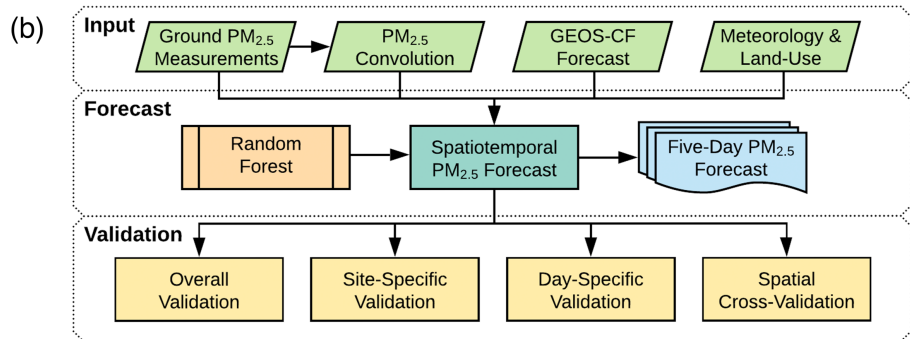
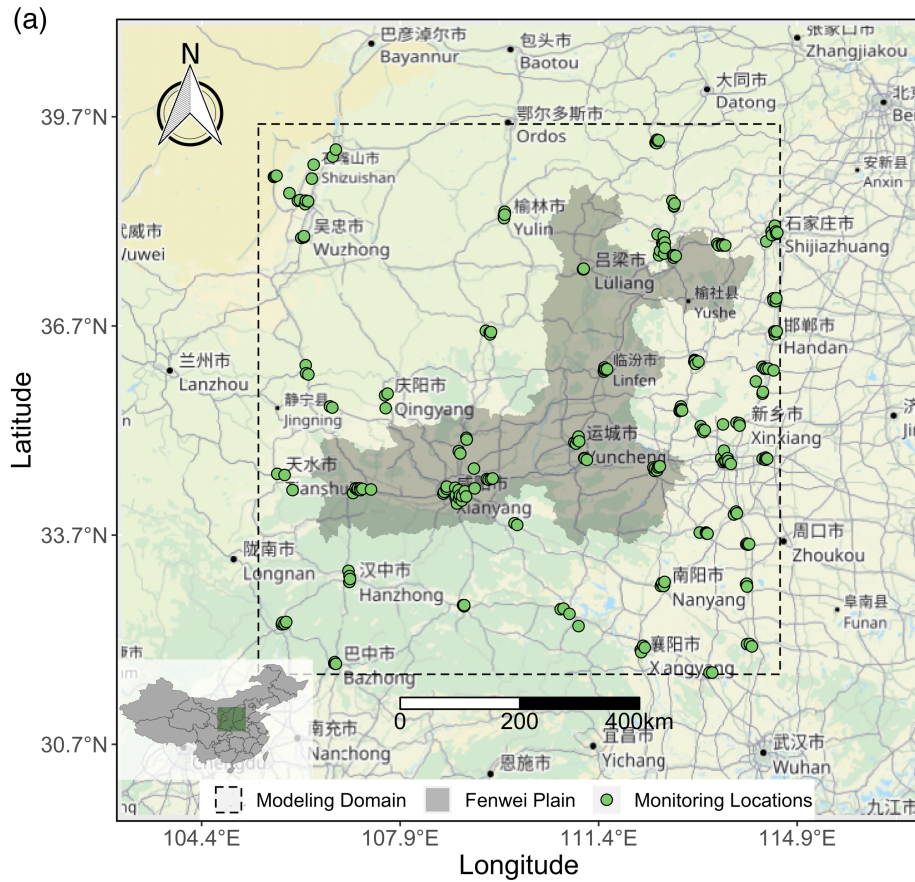
Forecast Day	N of Test Sample	R ²	RMSE ($\mu\text{g}/\text{m}^3$)	MAPE (%)	NMB
Day 1	78378	0.74	19.47	36.7	0.001
Day 2	78398	0.63	23.53	45.0	0.009
Day 3	78158	0.55	25.82	48.4	0.006
Day 4	78372	0.50	27.02	50.5	0.001
Day 5	78372	0.46	28.08	54.0	0

747

748 Table 4: The categorical forecast performance metrics of our forecast framework (RF + GEOS-
749 CF) and the original CTM forecast model (GEOS-CF), including positive predictive value (PPV)
750 and negative predictive value (NPV), for clean (N of training sample = ~66000, varied on
751 different forecast days), moderate-pollution (N of training sample = ~9800), and heavy-pollution
752 categories (N of training sample = ~2300).

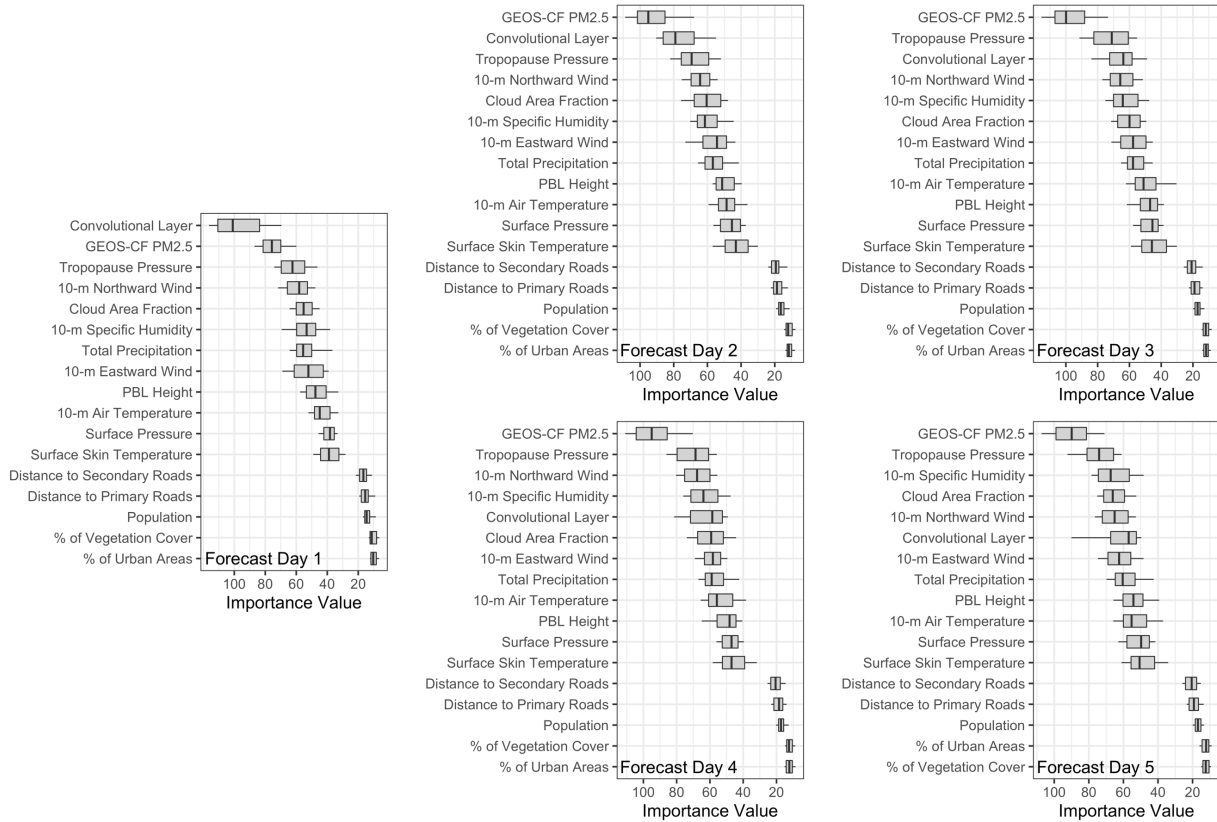
Forecast Day	Clean		Moderate Pollution		Heavy Pollution	
	PPV	NPV	PPV	NPV	PPV	NPV
RF + GEOS-CF						
Day 1	94.1%	81.6%	64.9%	94.0%	71.5%	98.2%
Day 2	92.2%	77.2%	56.9%	92.3%	54.5%	97.7%
Day 3	91.5%	72.7%	53.8%	92.0%	44.3%	97.4%
Day 4	90.5%	71.1%	52.3%	91.3%	43.1%	97.5%
Day 5	89.6%	66.8%	48.4%	90.6%	33.0%	97.3%
GEOS-CF						
Day 1	98.8%	20.6%	3.4%	82.3%	7.4%	99.7%
Day 2	98.9%	20.7%	3.1%	81.9%	7.7%	99.7%
Day 3	98.8%	20.6%	3.2%	81.6%	7.9%	99.8%
Day 4	98.4%	20.5%	3.3%	81.6%	7.8%	99.7%
Day 5	98.1%	20.8%	4.5%	82.4%	8.1%	99.7%

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 755 Figure 1: (a) Our study domain (the dashed box) with the locations of PM_{2.5} monitoring sites (at
 756 the 1-km grid cells; N = 226); the shadow region shows the municipality boundary of the Fenwei
 757 Plain. (b) The workflow of our PM_{2.5} forecast modeling and validation processes.

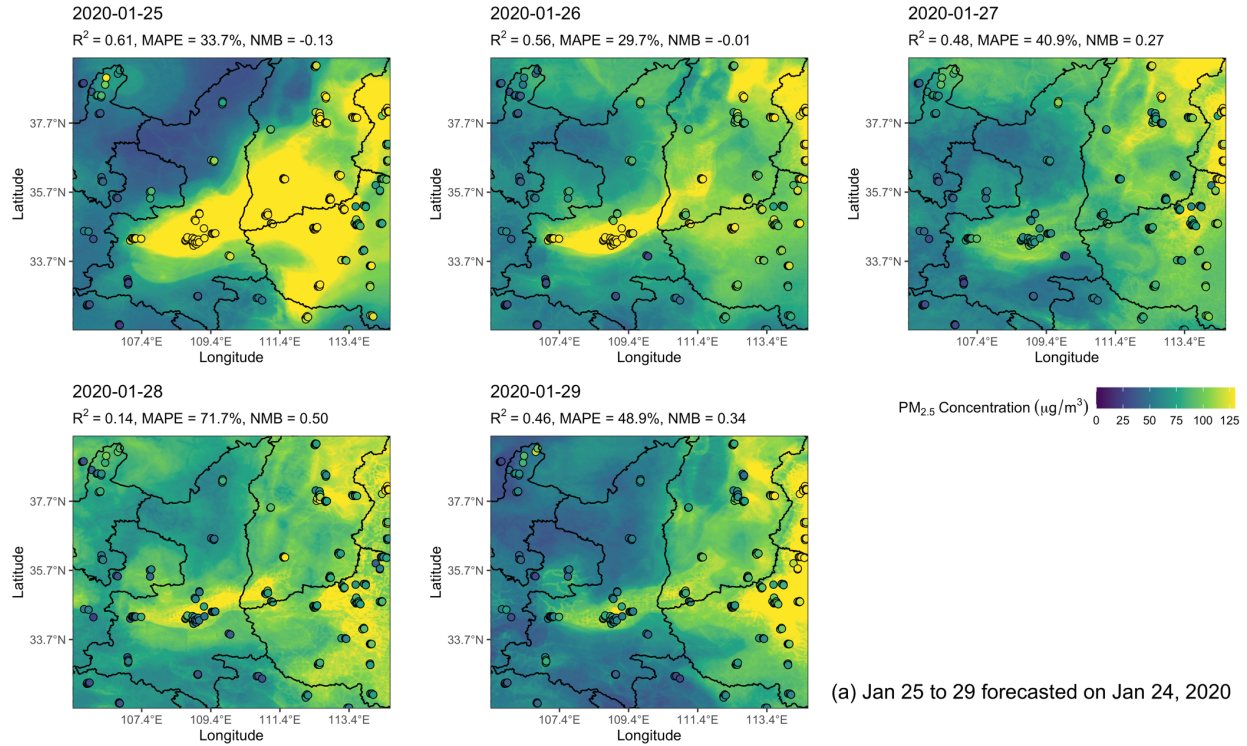
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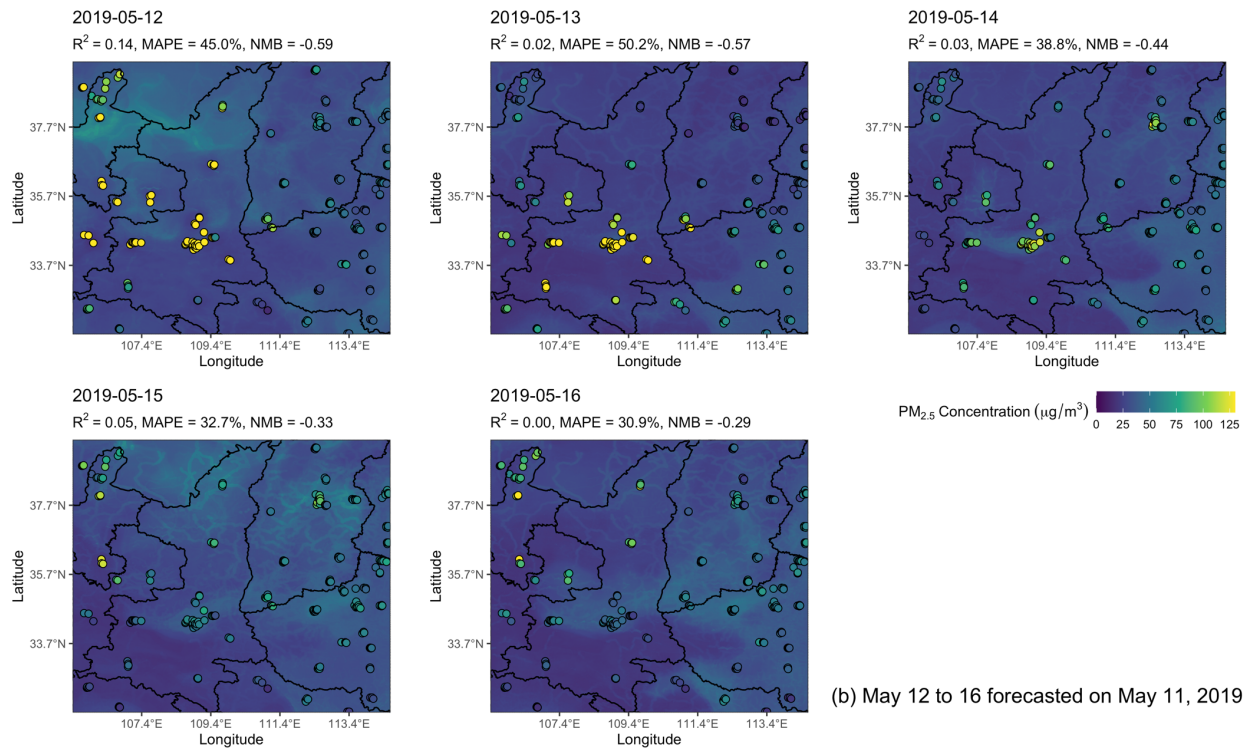
759

760 Figure 2: RF variable importance values for the five forecast days. The box plots summarize the
 761 importance values of the daily models from March 11th, 2019 to March 10th, 2020. The boxes
 762 represent the 25th and 75th percentile ranges; the whiskers represent the 10th and 90th percentile
 763 ranges; the bars within the boxes represent the 50th percentiles.

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(a) Jan 25 to 29 forecasted on Jan 24, 2020



(b) May 12 to 16 forecasted on May 11, 2019

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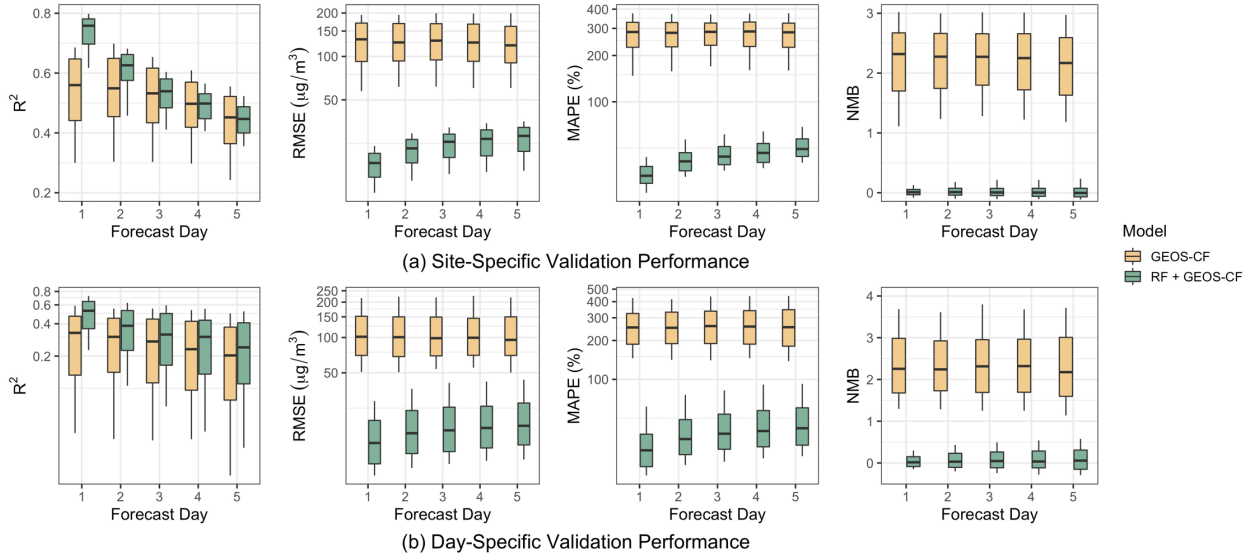
768

Figure 3: Spatial PM_{2.5} forecast concentrations in two example periods: (a) January 25th to 29th

forecasted on January 24th, 2020, and (b) May 12th to 16th forecasted on May 11th, 2019. The

769 colored dots show the observed PM_{2.5} concentrations at the monitoring locations, which share the
770 same color scheme with the forecast concentrations.

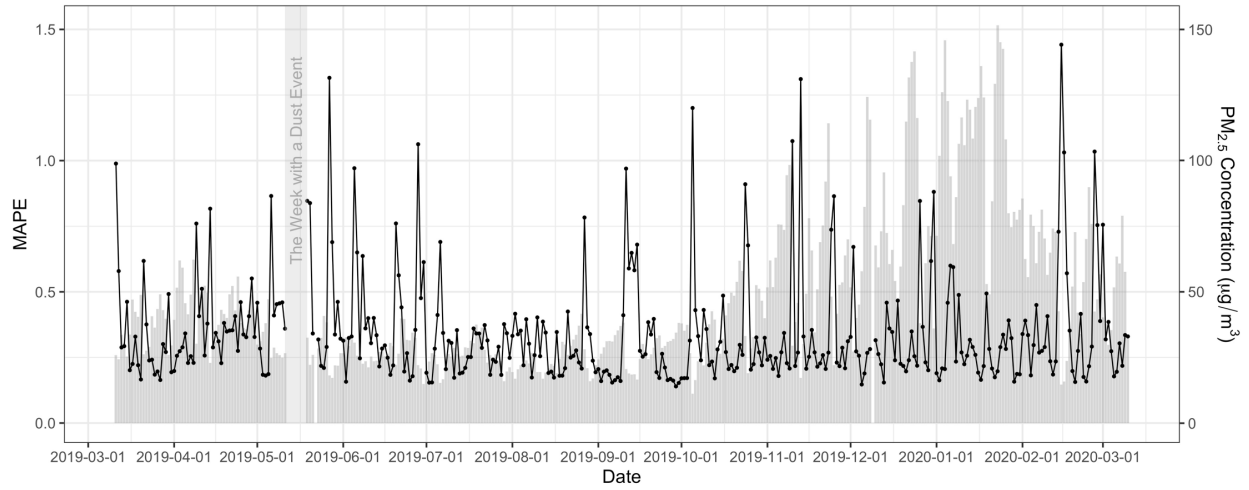
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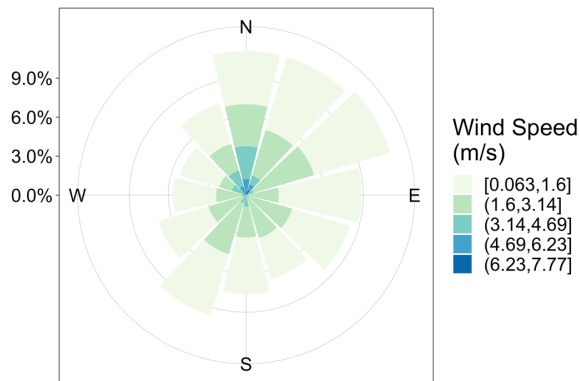
772

773 Figure 4: The (a) site-specific and (b) day-specific validation performance from March 11th,
 774 2019 to March 10th, 2020. The boxes represent the 25th and 75th percentile ranges; the whiskers
 775 represent the 10th and 90th percentile ranges; the bars within the boxes represent the 50th
 776 percentiles. The RMSE and MAPE plots are on the log scale.

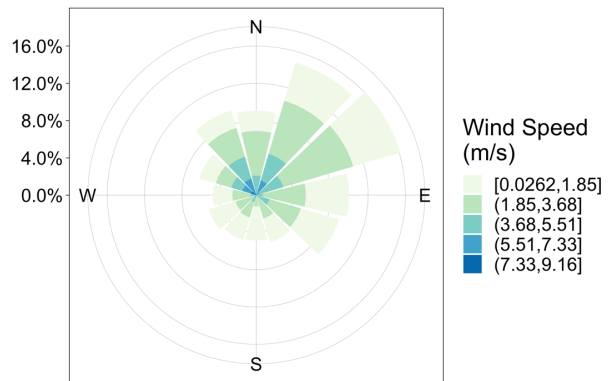
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(a) Daily MAPE with PM_{2.5} Concentrations (First Forecast Day)



(b) GEOS-CF Wind Rose (Good Forecast Performance)



(c) GEOS-CF Wind Rose (Poor Forecast Performance)

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Figure 5: (a) Daily validation MAPE values (black dots) with domain-average PM_{2.5} concentrations (grey bars) using the first forecast day as an example; (b) GEOS-CF wind rose on days with validation $R^2 >$ its 95th percentile (good forecast performance); (c) GEOS-CF wind rose on days with validation $R^2 <$ its 5th percentile (poor forecast performance).