

1 Comprehensive comparison of airborne and spaceborne SAR and LiDAR estimates of forest
2 structure in the tallest mangrove forest on Earth

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22

23 **Keywords**

24 AfriSAR; carbon; ALOS; SRTM; TanDEM-X; ICESat-2; GEDI; LVIS; F-SAR; UAVSAR; LVIS;

25 SRTM; GEOCARBON; IPCC

26

27 **Highlights**

- 28 • Comparing height and biomass estimates for 17 products in tallest mangrove forest.
- 29 • Height estimates are inconsistent across sensors in low and tall stature forests.
- 30 • Radar performed well, but calibration with LiDAR improved height estimates.
- 31 • Local calibrated biomass agreed within 15%, but global maps can have >50% error.
- 32 • Future biomass maps should combine local and global calibration strategies.

33 **Abstract**

34 A recent suite of new global-scale satellite sensors and regional-scale airborne campaigns are
35 providing a wealth of remote sensing data capable of dramatically advancing our current
36 understanding of the spatial distribution of forest structure and carbon stocks. However, a
37 baseline for forest stature and biomass estimates has yet to be established for the wide array of
38 available remote sensing products. At present, it remains unclear how the estimates from these
39 sensors compare to one another in terrestrial forests, with a dearth of studies in high
40 carbon density mangrove ecosystems. In the tallest mangrove forest on Earth (Pongara
41 National Park, Gabon), we leverage the data collected during the AfriSAR campaign to evaluate
42 17 state-of-the-art sensor data products across the full range of height and biomass known to
43 exist globally in mangrove forest ecosystems, providing a much-needed baseline for sensor
44 performance. Our major findings are: [1] height estimates are not consistent across products,
45 with opposing trends in relative and absolute errors, highlighting the need for an adaptive
46 approach to constraining height estimates; [2] radar height estimates had the lowest calibration
47 error and bias, with further improvements using LiDAR fusion; [3] biomass variability and
48 uncertainty strongly depends on forest stature, with variation across products increasing with
49 canopy height, while relative biomass variation was highest in low-stature stands; [4] a remote
50 sensing product's sensitivity to variations in canopy structure is more important than the
51 absolute accuracy of height estimates; [5] locally-calibrated area-wide totals are more
52 representative than generalized global biomass models for high-precision biomass estimates.
53 The findings presented here provide critical baseline expectations for height and biomass
54 predictions across the full range of mangrove forest stature, which can be directly applied to
55 current (TanDEM-X, GEDI, ICESat-2) and future (NISAR, BIOMASS) global-scale forest
56 monitoring missions.

57 **1.0 Introduction**

58 Forests hold approximately 45% of the world's active carbon [1]–[3], sequestering approximately
59 32% of anthropogenic emissions every year [4]. Accurate estimates of the distribution and total
60 carbon held in Earth's forests are essential for modelling and monitoring climate change, yet many
61 global maps of carbon storage disagree in critical regions of high carbon density [5]. Mangroves,
62 in particular, have the highest total carbon density of any forest on Earth with a mean of 856
63 Mg/ha, 49-98% of which is stored in the first 3 m of soils [6], [7], with C burial rates of 226 ± 39 g
64 C/m²/yr in comparison to 4 ± 0.5 g C/m²/yr in tropical terrestrial forests [8]. Total mangrove
65 aboveground biomass (AGB) is estimated at 1.75Pg [9], with soil carbon measurements in the
66 range of 1.93 to 6.4 Pg C [10], [11], yielding approximate total carbon estimates of 2.7 – 7.2 Pg
67 C. Additionally, these forests provide valuable ecosystem services – fuel, construction materials,
68 and protection from storms – to local coastal populations [12], [13] and essential habitat for rare
69 and endangered animal species [14]. Despite their importance, anthropogenic-driven loss is
70 occurring on a global scale [15], [16]. In the face of climate change, a detailed understanding of
71 the spatial distribution of carbon storage across the landscape will make future conservation
72 efforts more fruitful [17] and help countries attain Nationally Determined Contribution (NDC)
73 emissions reduction targets over the next half century.

74

75 National-scale carbon inventories contrast in level of detail and uncertainty of estimates. The
76 confidence in reported carbon stocks dictate the three IPCC tiers, corresponding to estimates
77 from [i] Tier 1: a mean ecosystem carbon density, [ii] Tier 2: deploying height-stratified field plots
78 for improved mean carbon density, or [iii] Tier 3: spatially continuous carbon estimates calibrated
79 with field plots and modeled with remote sensing [18]. Global conservation and forest ecology is
80 entering a “golden age” of satellite measurements that stands to significantly improve our current
81 understanding of fine-scale patterns in forest structure and carbon storage [19]. Remote sensing

82 is enabling near-universal Tier 3 carbon estimates, but the long list of mapping products have yet
83 to be directly compared to one another and important sensor-specific differences in forest height
84 and biomass have not been quantified.

85

86 Key missions designed specifically for measuring forest structure include NASA's Global
87 Ecosystem Dynamics Investigation (GEDI), ESA's BIOMASS [20], and NASA-ISRO's Synthetic
88 Aperture Radar (NISAR; [21]). GEDI - a large-footprint waveform LiDAR (Light Detection And
89 Ranging) sensor - solves many of the greatest challenges for creating high-certainty global
90 forest carbon maps [22]. To capture sub-kilometer variations in forest structure, sampling
91 instruments like GEDI or ICESat-2 (dense photon counting LiDAR) must be matched to
92 contiguous height estimates – from either optical or, preferably, radar [23]. DLR's TerraSAR-X
93 add-on for Digital Elevation Measurement (TanDEM-X or TDX) DEM product provides a ~12 m
94 resolution digital surface model (DSM) at a global scale, making it ideal for creating high-
95 resolution spatially continuous forest height products [24], [25]. In tropical forests, upcoming
96 SAR sensors like BIOMASS (P-band) and NISAR (L-band) can penetrate cloud cover, enabling
97 detailed estimates of forest structure that are unreachable with optical and LiDAR remote
98 sensing alone [26]. However, radar backscatter loses sensitivity at high biomass densities -
99 highlighting the need for more direct estimates of structure from SAR (i.e. InSAR) together with
100 multi-faceted fusion approaches as the most viable option for globally consistent estimates of
101 forest structure. In mangroves specifically, Simard et al [9] created a global high-resolution
102 mangrove height and biomass maps using Shuttle Radar Topography Mission (SRTM) data
103 calibrated with NASA Ice, Cloud, and land Elevation Satellite (ICESat) Geoscience Laser
104 Altimeter System (GLAS) and forest inventory data. However, these maps represent the status
105 of global mangrove forest in the year 2000 with 30-meters spatial resolution and limited
106 accuracy at local scales. Given the wealth of current and upcoming near-global remote sensing

107 data capable of estimating forest height and biomass, there is a clear need to evaluate the
108 consistency and differences across sensors in the context of height and biomass.

109
110 In this study, we evaluate 17 forest biomass products from five types of sensors measuring
111 canopy structure ([i] stereo optical photogrammetry, [ii] SAR interferometry, [iii] Polarimetric SAR
112 interferometry, [iv] large-footprint waveform LiDAR and [v] photon counting LiDAR in the tallest
113 known mangrove forest [9], leveraging data collected from the 2016 AfriSAR airborne campaign
114 as well as in situ measurements in Gabon [27], [28]. These products were derived from airborne
115 and spaceborne sensors representing the suite of current and future missions used for measuring
116 forest height, estimating aboveground biomass and ecosystem carbon stocks. The specific
117 objectives of this study are:

- 118 1. Directly compare remotely sensed height products;
- 119 2. Develop and evaluate sensor-specific biomass calibration models from plot data;
- 120 3. Directly compare spatial distributions of locally calibrated and other biomass products;
- 121 4. Directly compare area-wide totals from locally calibrated and other biomass products.

122 The goal of our analysis is to provide a baseline comparison of height and biomass estimates for
123 the most commonly available airborne and spaceborne remote sensing products, providing a
124 much-needed baseline for current and forthcoming sensor performance.

125 **2.0 Methods – 2362**

126 The analysis in this study used field inventory biomass estimates to calibrate a suite of remotely
127 sensed height to generate spatially comprehensive maps of biomass over the study site. The
128 height and biomass maps, and biomass totals are then intercompared.

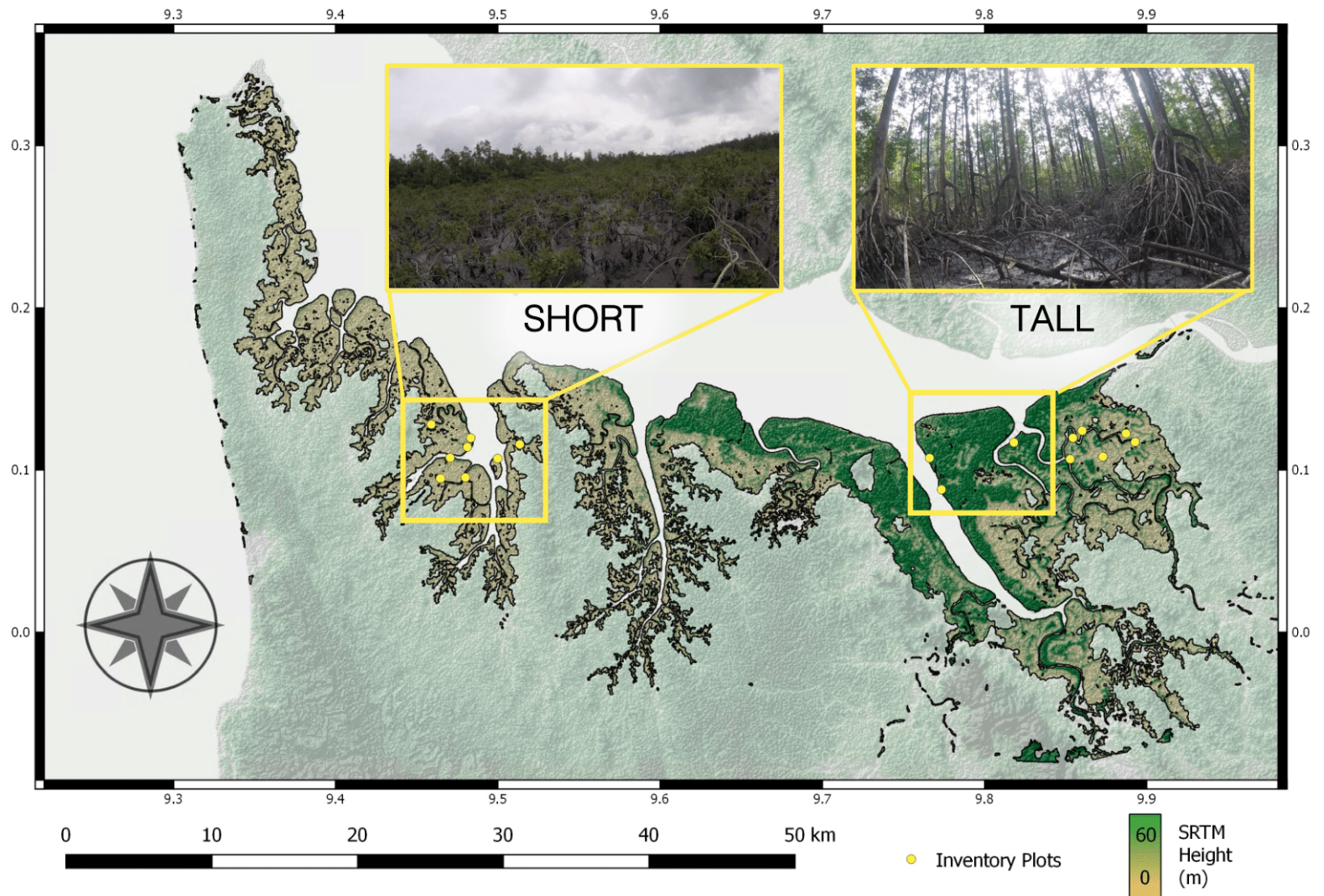
129 2.1 Site Description and Field Data

130 The study site is the Pongara National Park located in Gabon (**Figure 1**). Pongara National Park
131 is located on the southern bank of the Komo Estuary, directly south of Libreville, Gabon's capital
132 city. The Park covers an area of 87,000 ha and is covered primarily by mangroves (52,700 ha)
133 and some *terra firme* rainforests [29]. This site holds both the tallest known mangrove forests on
134 Earth and large areas of short low density stands – an excellent test case for capturing a range
135 in forest stature from 2 m to in excess of 60 m in height [9].

136
137 Circular field plots (n=17) were sampled with a radius between 6 and 12.5 m diameter, with small
138 plots coinciding with short stature forest stands [30]. At each plot, aboveground biomass (AGB)
139 was estimated from stem diameter measurements (0.5 m above the last prop-root). In addition,
140 tree height was estimated using a laser hypsometer. In this study, we chose to use height-based
141 field allometry from Chave et al. [31]'s wet tropical equation as it best related to our remotely-
142 sensed height estimates:

143
144
$$\text{Aboveground Biomass (kg)} = \exp(\alpha + \beta \ln(\rho D^2 H)) \quad [1]$$

145
146 Where α and β are model coefficients derived from least squares regression, ρ is species-
147 specific wood density (0.9 for *Rhizophora* sp.), D is tree diameter, and H is tree height.
148 For reference and to determine the impact of allometric equation selection, we also evaluated the
149 difference in plot-level biomass estimates using allometry relying solely on tree diameter (See
150 Figure S1, Table S1, and Figure S2; [32]).



151

152 **Figure 1:** Map of Pongara National Park with heights from SRTM-based global height product (from Simard et al [9]). Inventory plots
 153 were placed such that canopy heights were sampled proportionally according to the height distribution across the site.

154 2.2 Remote Sensing Datasets

155 We evaluated height and/or biomass products of global and local spatial extent (**Figure 2; Table**
156 **1**). We evaluated five types of sensors measuring canopy structure using [i] stereo optical
157 photogrammetry, [ii] SAR interferometry, [iii] Polarimetric SAR interferometry, [iv] large-footprint
158 waveform LiDAR and [v] photon counting LiDAR. Specific details on products and processing
159 techniques for individual products can be referenced from the associated publications in **Table 1**.
160 Spatially continuous height products were not modified prior to biomass calibration to ensure the
161 generalization of our analysis and results. Measurements from spaceborne LiDAR instruments –
162 ICESat-2 and GEDI were used to calibrate TanDEM-X heights at the corresponding sensor
163 resolution to produce two additional LiDAR-calibrated height and biomass map products.

164 *2.2.1 Brief Sensor Overview*

165 Several of the global sensor products evaluated here were produced in the 2000-2010 timeframe.
166 The Advanced Land Observing Satellite (ALOS) Panchromatic Remote-sensing Instrument for
167 Stereo Mapping (PRISM) is an optical instrument providing a 30 m stereo imagery-based digital
168 surface model (DSM; [33]), also referred to as a Digital Elevation Model (DEM). The Shuttle Radar
169 Topography Mission (SRTM) was a C-Band SAR interferometry mission that flew in February
170 2000 producing a global ~30 m resolution Digital Surface Model (DSM). The SRTM DSM was
171 used in concert with ICESat-1 (Ice, Cloud, and land Elevation Satellite; [34]) canopy height
172 estimates to create local [35] and continental-scale [36] and, more recently, the first global-scale
173 [9] canopy height and biomass models for mangrove forests. TanDEM-X is a high-resolution
174 interferometric SAR mission launched by DLR (German Aerospace Center) to produce ~12 m
175 (commercially available), ~30 m, and ~90 m (freely available) resolution global DSMs [24].

176

177 We evaluated two recently launched global LiDAR sensors for measuring vegetation structure.
178 The ICESat-2 satellite uses a photon-counting LiDAR to measure elevations [37], producing 100
179 m granule with vegetation indices. The low sampling density and polar orbit of ICESat-2 prevents
180 evaluation of a continuous gridded height product. We therefore created continuous ICESat-2
181 mean canopy height product by calibrating 90 m TanDEM-X heights with 100 m ICESat-2 mean
182 canopy height granules using a simple least squares regression model. The Global Ecosystem
183 Dynamics Investigation (GEDI) instrument is a full-waveform LiDAR designed specifically to
184 measure forest structure at a near-global scale using four high-powered (power) and four-low
185 powered (coverage) beams [22]. Aboard the International Space Station (ISS), GEDI produces
186 vegetation metrics at the footprint-level (~25 m) with high vertical resolution. Similar to ICESat-2,
187 we created a continuous GEDI RH100 height product by calibrating 30 m TanDEM-X heights with
188 footprint-level GEDI RH100 heights using a least squares regression model. Geolocation errors
189 are common in the version 001 release of GEDI data and erroneous height measurements often
190 occur in edge areas. A recent simulation study found that the expected GEDI geolocation error of
191 10 m may introduce more than 50% uncertainty into the resulting height estimates [38]. We
192 therefore flagged and removed GEDI shots within 40 m of the forest edge to avoid potential mixed
193 or non-mangrove footprints resulting from geolocation error.

194

195 Local scale sensors in this study were flown as part of the AfriSAR, a joint NASA and ESA
196 (European Space Agency) airborne campaign [27], [28]. The goal of the mission was to fly
197 overlapping airborne sensors analogous to future missions (e.g. ESA BIOMASS [20], NASA-ISRO
198 Synthetic Aperture Radar (NISAR; [21]) and GEDI) to measure forest structure. DLR deployed
199 the airborne F-SAR - a dual band Pol-InSAR instrument analogous to NISAR (L-band) and
200 BIOMASS (P-band) - and covered a small portion of the study area and field plots. JPL's
201 UAVSAR, an L-band SAR covered nearly the entirety of Pongara National Park [39]. The Land

202 Vegetation Ice Sensor (LVIS; [40]) has near-identical technology as GEDI, though with nearly
203 continuous sampling, providing ~25 m footprints of full-waveform LiDAR measurements.

204

205 *2.2.2 ICESat-2- and GEDI-TanDEM-X Fusion*

206 ICESat-2- and GEDI heights were used to calibrate two continuous TanDEM-X canopy height
207 maps covering the entirety of the study area. For both spaceborne LiDAR sensors, we
208 calibrated TanDEM-X heights with a similar procedure: [i] directly matching LiDAR heights to
209 TanDEM-X heights, [ii] developing a calibration model between LiDAR height and TanDEM-X
210 height, and [iii] producing a LiDAR corrected continuous height product covering the entirety of
211 Pongara National Park.

212

213 All available overlapping ICESat-2 data were extracted for the study area comprising 46 total
214 orbits of ATL08 data. All orbits were constrained to the study area and only those classified as
215 mangrove were retained, leaving a total of 597 at 100 m intervals. We used the
216 *h_mean_canopy* variable in our calibration procedure – the mean of canopy heights within a
217 100 m granule. We used the 90 m TanDEM-X geoid-corrected height product [9] to upscale
218 ICESat-2 measurements since the two were closely matched in spatial scale. Next we extracted
219 the 90 m resolution TanDEM-X heights at the overlapping ICESat-2 granules. In an initial
220 assessment between the two heights, we found beam three had consistently anomalous height
221 estimates; We therefore excluded data from this beam entirely. We identified and removed two
222 extreme outliers in the calibration, based on the values exceeding 10x the mean Cooks
223 Distance in the linear model. In total, we built the calibration model on 391 ICESat-2 height
224 measurements.

225

226 Similarly, all available overlapping GEDI data was extracted for the study area – 21 total orbits
227 of L2B data. Again, all orbits were constrained to the study area, retaining only mangrove areas
228 and quality flag 1 data, leaving a total of 3482 canopy height estimates. We used the *rh100*
229 variable in our calibration procedure – the tallest detectable height aboveground.

230

231 As expected, the two height variables for both ICESat-2 and GEDI data were closely and
232 linearly related, so we used a least squares regression to develop the calibration model:

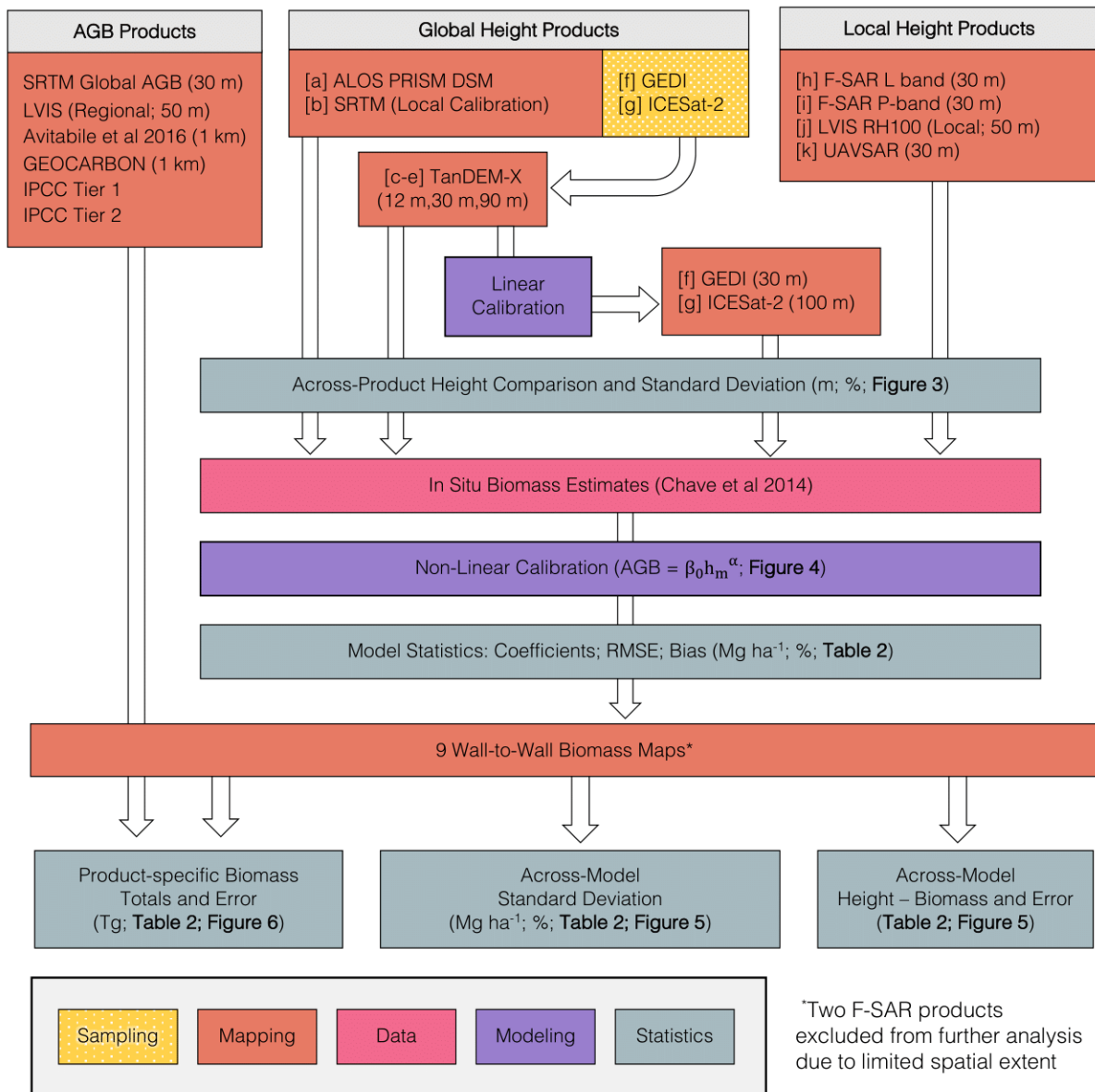
233

$$234 \quad \text{ICESat2 or GEDI Canopy Height (m)} = \beta_0 + \beta_1 h_{TDX}$$

235

236 Where β_0 and β_1 are model coefficients and h_{TDX} is 90 m resolution geoid corrected TanDEM-X
237 height.

238



239
240
241
242

Figure 2: Workflow detailing major processing and analysis steps used to compare 17 area-wide biomass products in Pongara National Park.

243 **Table 1:** Specifications of sensors and products used for local calibration and/or validation in the study.

Extent	Sensor/ Product	Product Resolution	Technology	Acquisition Period	Availability	Variable†	Relevant Publications
Global Height Products	[a] ALOS PRISM DEM	30 m	Stereo Optical	Jan 24, 2006- Apr. 22, 2011	Open	Elevation	[33]
	[b] SRTM	30 m	C-Band SAR Interferometry	Feb 11-22, 2000	Open	Ice-SAT-GLAS- Corrected Mangrove Canopy Height (Hmax)	[9], [41]
	[c] TanDEM-X (12 m)	12 m	X-Band SAR Interferometry	Dec 12, 2010- Jan 16, 2015	Commercial	Geoid corrected height asl	[42]
	[d] TanDEM-X (30 m)	30 m			Commercial		
	[e] TanDEM-X (90 m)	90 m			Open		
	[f] ICESat-2-TanDEM-X	100 m	Photon Counting LiDAR	Sept 15, 2018 - Present	Open	TanDEM-X Elevation corrected with ATL08 98 th percentile heights	[41], [42]
[g] GEDI-TanDEM-X	30 m	Large-Footprint Full- Waveform Spaceborne LiDAR	Mar 25, 2019 - Present	Open	TanDEM-X Elevation corrected with RH100 heights	[22]	
Local Height Products	[h] LVIS	50 m	Large-Footprint Full- Waveform Airborne LiDAR	Mar 3, 2016	Open	RH100	[40]
	[i] F-SAR L band*	30 m	Airborne L-Band PolinSAR	Feb 2016	Open	Modeled Canopy Height	[45], [46]
	[j] F-SAR P-band*	30 m	Airborne P-Band PolinSAR	Feb 2016	Open	Modeled Canopy Height	[45], [46]
	[k] UAVSAR	30 m	Airborne L-Band PolinSAR	Feb 27, 2016	Open	Modeled Canopy Height	[47]
AGB products	LVIS (Regional Calibration)	50 m	Large-Footprint Full- Waveform Airborne LiDAR	Mar 3, 2016	Open	AGBD*	[46], *[47]
	Global SRTM	30 m	C-Band SAR Interferometry	Feb 11-22, 2000	Open	AGBD*	*[9], [36]
	Avitabile <i>et al</i> 2016 GEOCARBON	~1 km	SAR, Optical, Large Footprint LiDAR	2011-2012	Open	AGBD*	*[48], *[49]
	IPCC Tier 1 value: 192 Mg/ha	-	-	Mar 2016	-	IPCC mean mangrove AGBD	*[18]
	IPCC Tier 2 value: 215 Mg/ha					Plot-based	

†The predictor variable matched to plot data used for calibrating the allometric models of aboveground biomass. *Aboveground biomass density estimates derived in the cited study.

*Height-biomass calibration is only evaluated due to limited spatial extent

246 2.3 Height and Biomass Analysis

247 The aim of our analysis was to compare the current available airborne and spaceborne remote
248 sensing products for estimating forest height and biomass (**Figure 2**). Both analyses of height
249 and biomass compared each individual remote sensing product to a mean map (and standard
250 deviation) created from all products. We evaluated the deviation of each product height and
251 biomass from the mean with residual plots. Finally, we compared the total estimated biomass
252 (and uncertainty) across the study site from each remote sensing product, along with six other
253 external AGB maps and IPCC Tier 1 and Tier 2 AGB estimates. In doing so, we provide a basis
254 for comparison and a baseline expectation for height and biomass estimates for all products
255 analyzed in this study.

256 2.3.1 Evaluating Remotely Sensed Mangrove Height

257 Height estimates from nine height products were intercompared at the pixel level over their
258 mutually overlapping area. To evaluate the variability across all height products ($H_{product}$), we
259 calculated the per-pixel (i) average (*mean*) and standard deviation (*sd*) of all height maps to create
260 a single map representing the mean height (H_{mean}) and standard deviation (H_{sd}) of height:

$$261 \quad H_{mean}(i) = mean(H_{product}(i)) \quad [2]$$

$$262 \quad H_{sd}(i) = sd(H_{product}(i) - H_{mean}(i)) \quad [3]$$

263 We evaluated the overall trend in standard deviation from Eq. 3 with respect to 1 m bins of H_{mean} .
264 For individual continuous AGB products ($AGB_{product}$; $n = 9$) by calculating the mean signed
265 deviation (MSD) and standard deviation (H_{sd}) as a function of 1 m bins of H_{mean} , represented as h
266 in equation 4 and 5:

$$267 \quad MSD(h) = mean(H_{product}(i, h) - H_{mean}(i, h)) \quad [4]$$

$$268 \quad H_{sd}(h) = sd(H_{product}(i, h) - H_{mean}(i, h)) \quad [5]$$

269 We also determined how well specific products capture field-measured heights (based on RMSE
270 and bias) by directly comparing the remotely sensed heights to plot-level tree height percentiles.

271 2.3.2 Sensor Aboveground Biomass Calibration and Uncertainty

272 For each of the remote sensing height products, we built a calibration model relating in-situ plot
273 biomass to remotely sensed height using non-linear least squares regression (*nls*; R Core Team
274 2019) with the form:

$$275 \text{ Aboveground Biomass (Mg ha}^{-1}\text{)} = \beta_0 h_m^\alpha \quad [5]$$

276
277 Where β_0 is the scaling coefficient, h_m is a sensor-specific height metric, and α is the scaling
278 exponent. Note, none of the remote sensing products we evaluated had a resolution smaller than
279 the plot size (6-12.5 m diameter), so the value of h_m was simply extracted at the plot location.
280 However, plots were established in ~0.5 ha areas of homogeneous height to mitigate the effects
281 of the smaller plot size (See Trettin et al [53] for details), making the plot data representative for
282 resolutions up to ~70 m. In addition, we used this same model form to evaluate a purely plot-
283 based allometric model (see Supplementary Material Figure S1).

284
285 The precision and accuracy of all locally calibrated predictive biomass models were assessed
286 with a bootstrapped estimate of root mean square error (RMSE) and bias. Over 1000 iterations,
287 a random set of 70% of the plot data was selected for model training using Equation 2. The
288 predicted value was then compared against the independent (measured) plot biomass values
289 using the following equations:

$$291 \text{ RMSE (Mg ha}^{-1}\text{)} = \sqrt{\frac{\sum_{i=1}^n (\text{predicted}_i - \text{measured}_i)^2}{n}} \quad [6]$$

$$292 \text{ RMSE (\%)} = \frac{\text{RMSE}}{\text{measured}} \quad [7]$$

293
$$Bias (Mg ha^{-1}) = \frac{\sum_{i=1}^n predicted_i - measured_i}{n} \quad [8]$$

294
$$Bias (\%) = \frac{Bias}{measured} \quad [9]$$

295

296 Where $\overline{measured}$ is the mean plot-level biomass density estimate across all iterations. In an
 297 additional analysis, we evaluated the role of model uncertainty in calibration coefficient estimates.
 298 A robust parameter-based non-linear pixel-level error estimate for each biomass product was
 299 estimated directly from the non-linear calibration models using the first-order Taylor series method
 300 as implemented in the *errors* package in R [52], [54]. In essence, this method linearizes the
 301 predicted biomass uncertainty for a given height, accounting for the variance and covariance
 302 between model coefficients (See documentation for *errors* package for further details). The
 303 approach simply uses the uncertainty in the calibration model parameters to properly estimate
 304 pixel-level prediction error.

305 *2.3.4 Spatial Patterns and Variability in Biomass*

306 As in our height analysis (Section 2.3.1), AGB estimates from all spatially continuous height
 307 products were directly compared at the pixel level (excluding pixels with missing values from any
 308 sensor). To evaluate the variability across all height products ($AGB_{product}$), we calculated the per-
 309 pixel average and standard deviation of all AGB maps to create a single map representing the
 310 mean AGB (AGB_{mean}) and standard deviation (AGB_{sd}) of height:

311
$$AGB_{mean}(i) = mean(AGB_{product}(i)) \quad [10]$$

312
$$AGB_{sd}(i) = sd(AGB_{product}(i) - AGB_{mean}(i)) \quad [11]$$

313 We evaluated the overall trend in standard deviation with respect to 10 Mg ha⁻¹ bins of AGB_{mean} .
 314 For individual continuous height products ($AGB_{product}$; n = 9), we calculated the mean signed
 315 deviation (MSD) and standard deviation (AGB_{sd}) as a function of 10 Mg ha⁻¹ bins of AGB_{mean} ,
 316 represented as b in equation 4 and 5:

317
$$MSD(b) = mean(AGB_{product}(i, b) - AGB_{mean}(i, b))$$
 [12]

318
$$AGB_{sd}(b) = sd(AGB_{product}(i, b) - AGB_{mean}(i, b))$$
 [13]

319 In addition, we compared the mean biomass predictions (AGB_{mean}) of our area-wide locally
320 calibrated estimates on a per-pixel basis to two contrasting high-resolution independent biomass
321 products based on SRTM (30 m; [9]) and LVIS (50 m; [49]). We quantified systematic deviations,
322 highlighting these differences using residual variation figures.

323 2.3.4 Total Biomass and Uncertainty

324 Total biomass and uncertainty was estimated across the entire study area for all continuous
325 remote sensing biomass products – nine locally calibrated models and six baseline biomass
326 estimates (Figure 2). For the nine locally-calibrated biomass products, we limited the spatial
327 extent to that of the product with lowest spatial coverage (i.e. LVIS). Across this area, we derived
328 the mean biomass prediction and associated uncertainty (determined with pixel-level model
329 parameter-based first-order Taylor series method). The mean and uncertainty estimates were
330 applied across the ~40,000 ha study area for area-wide biomass totals. Uncertainty was
331 propagated using the *errors* package in R [52], [54].

332

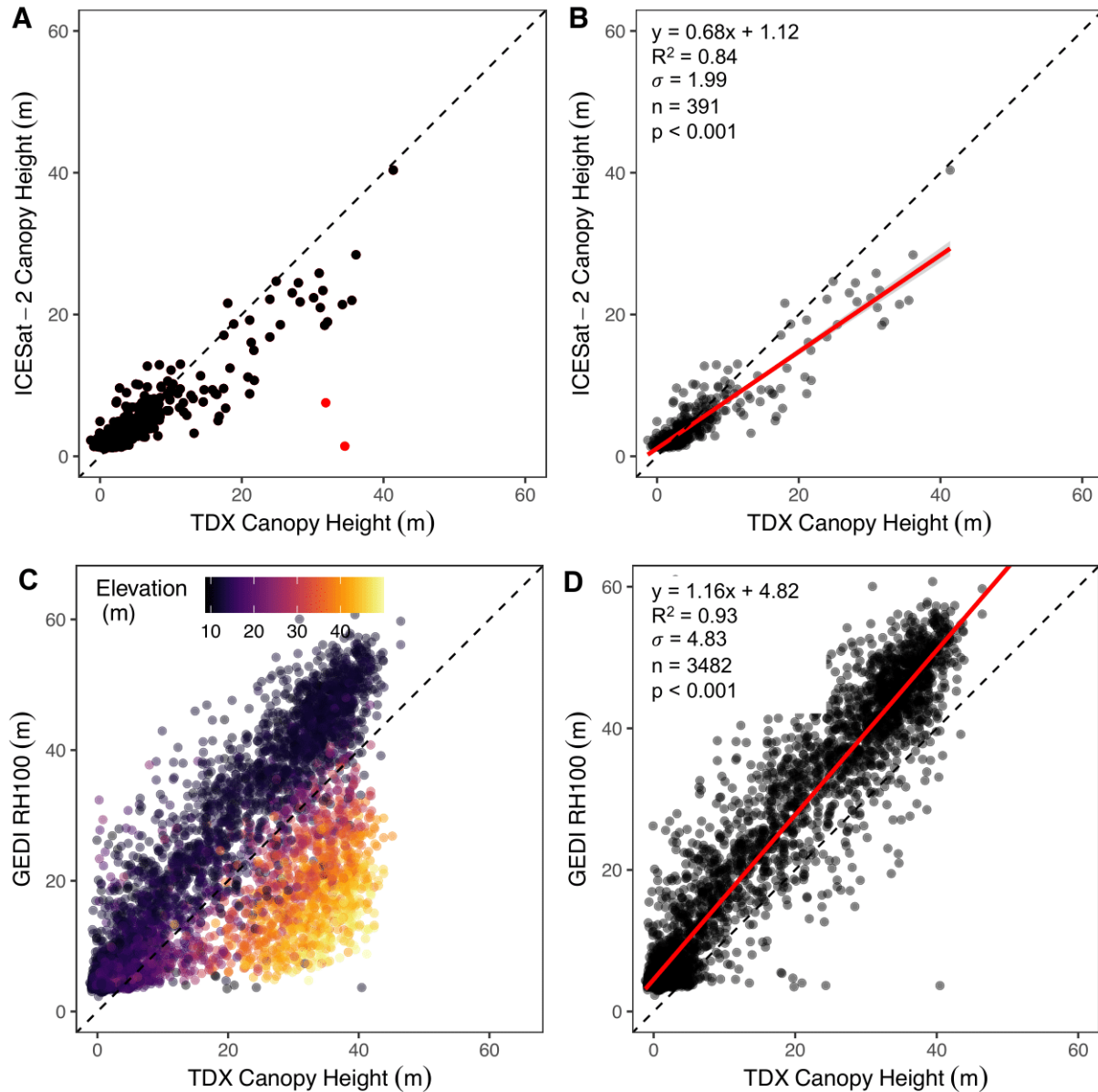
333 We evaluated the totals from six baseline biomass estimates described in Table 1: [i] Global
334 SRTM mangrove biomass [9], [ii] regionally calibrated three-variable LVIS biomass [49], [iii]
335 Avitabile et al [50], [iv] Santoro et al [51] global biomass products (1 km), [v] IPCC Tier 1 values
336 (192 Mg.ha⁻¹), and [vi] IPCC Tier 2 (215 Mg ha⁻¹). The global SRTM mangrove biomass product
337 is based on globally calibrated ICESat-GLAS adjusted SRTM heights (H_{max} variable, [9]). The
338 regional LVIS biomass product was calibrated from all other non-mangrove field data included in
339 the Gabon AfriSAR campaign (see [49]). The Avitabile et al [50] and Santoro et al [51] global
340 biomass products are built from a suite of active and passive remote sensing variables covering
341 the pantropical and global scale, but are not specifically calibrated for mangroves. To ensure our

342 area-wide totals were not biased by product coverage, we limited the extent of each continuous
343 mapped product to the study area, calculated the mean mangrove biomass density estimate of
344 all pixels, and applied the mean values to the entirety of Pongara National Park. For comparison,
345 we included Tier 1 and Tier 2 IPCC-based biomass estimates in our total biomass and
346 assessment. The Tier 1 IPCC estimate was based on the Mangrove Tropical Wet area-based
347 mean (192 Mg ha⁻¹). The field plots were established using probability-based sampling, so for
348 comparison we also derived a Tier 2 IPCC estimate calculated as the mean plot-based biomass
349 (215 Mg ha⁻¹). Both Tier 1 and 2 mean values are not spatially explicit and were simply applied to
350 the total mangrove area used in this study. To ensure a consistent estimate of product uncertainty,
351 baseline product uncertainty was estimated as the standard deviation of the difference between
352 *in-situ* plot biomass and the mapped biomass estimate.

353 **3.0 Results – 915**

354 *3.1 ICESat-2- and GEDI-TanDEM-X Fusion*

355 The ICESat-2 and GEDI height estimates successfully calibrated TanDEM-X heights ($R^2 = 0.84-$
356 0.93). Table S2 provides an overview of the statistics of the final calibration models and Figure 3
357 shows the calibration models, along with anomalous excluded data. The ICESat-2 calibration
358 had an order of magnitude fewer measurements available than GEDI, reducing the power of the
359 calibration model. The ICESat-2 calibration model reduced TanDEM-X heights, while GEDI
360 RH100 calibration increased TanDEM-X heights. The major factor affecting the quality of GEDI
361 height estimates was the ground elevation estimate, which was used as criteria for data quality
362 filtering.



363
 364 **Figure 3:** Calibration models used to create spatially continuous height estimates by fusing
 365 ICESat-2 mean canopy height and GEDI RH100 to TanDEM-X continuous heights. Red points
 366 are removed outliers in the ICESat-2 calibration (anomalous beam 3 data not shown). Colored
 367 points in C show anomalous elevation values that were removed from the final calibration model
 368 (B and D).
 369

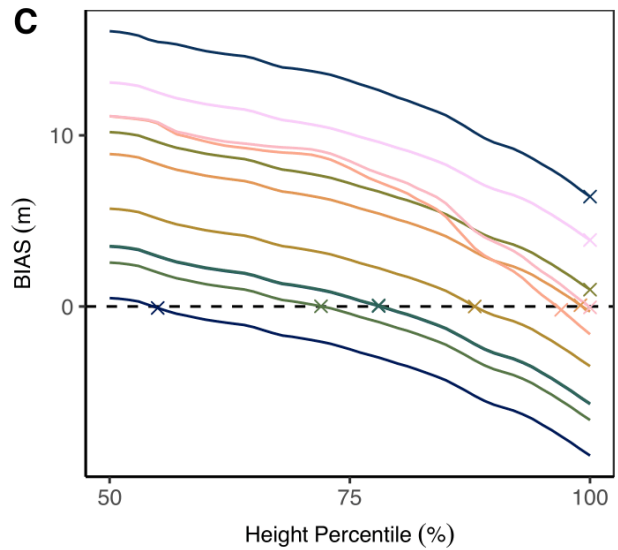
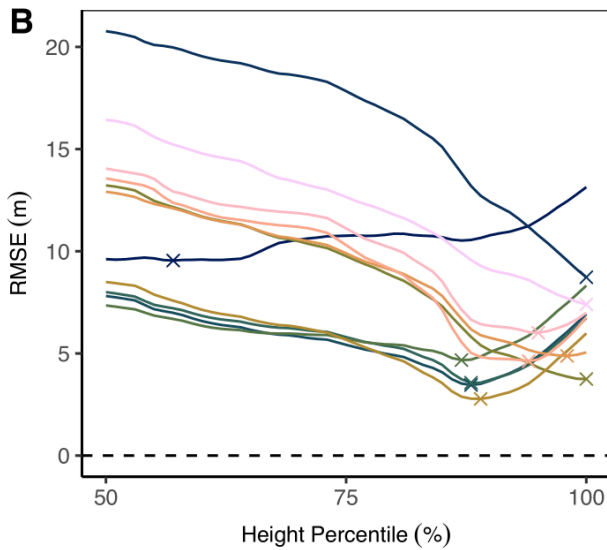
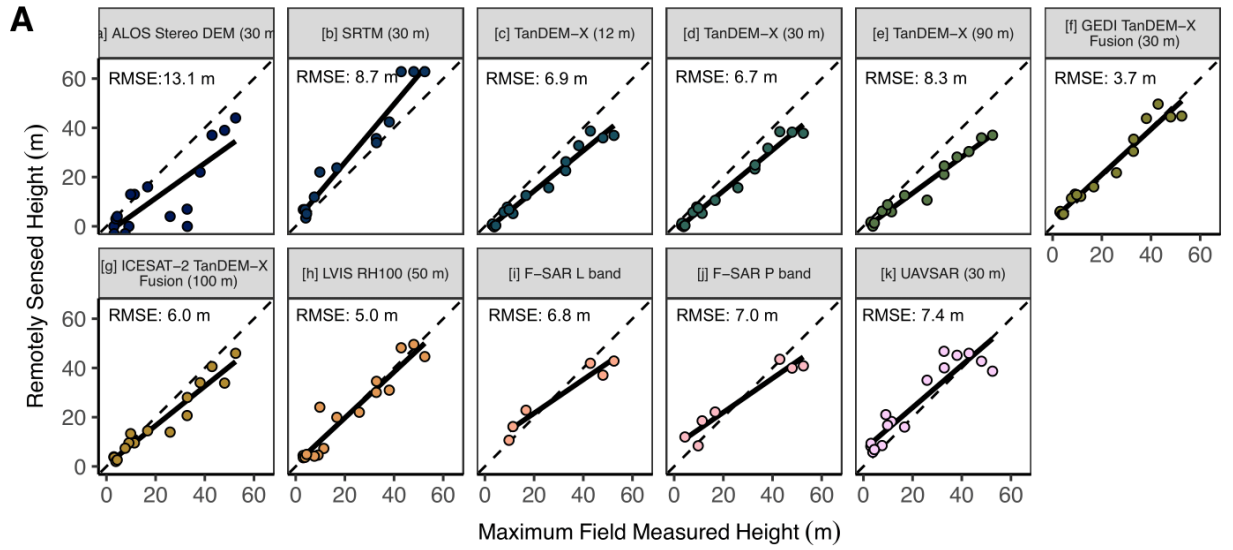
370 3.2 Evaluating Remotely Sensed Mangrove Height

371 Height estimates from nine products were compared to the field height measurements and the
372 mean height map (H_{mean}). Different aspects of canopy height are captured, depending on sensor
373 (Figure 4). TanDEM-X products generally underestimate, LVIS closely estimates, and the SRTM
374 H_{max} product overestimates compared to field height. The ALOS PRISM DEM was highly
375 variable and generally underestimated field height.

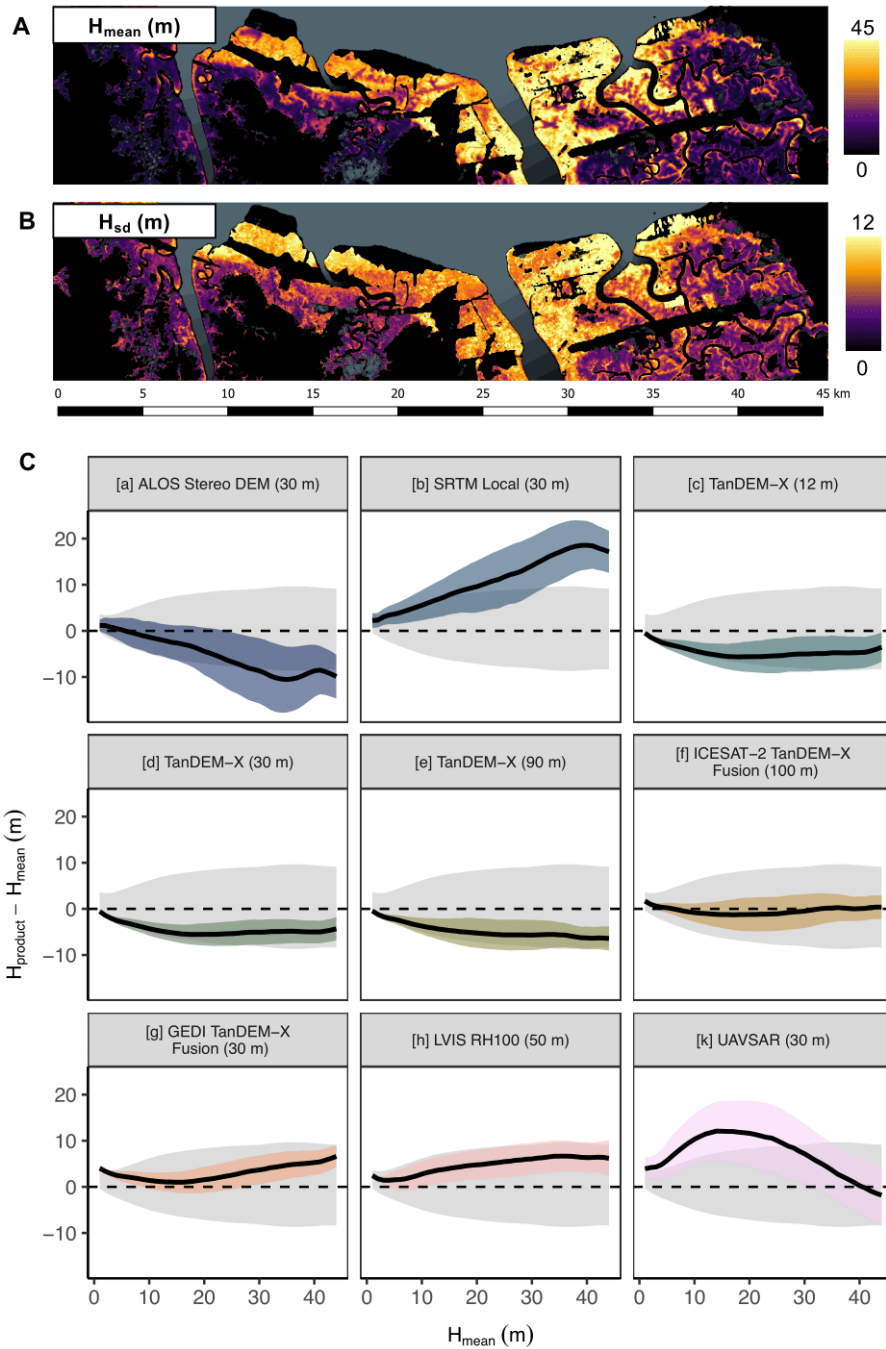
376

377 Most products generally followed a linear trend with the mean height map (H_{mean}) (Figure 5,
378 Figure S4A). ALOS PRISM product had the lowest and SRTM H_{max} had the tallest height
379 estimates covering ~25 m difference in the tallest stature stands, with TanDEM-X-based
380 estimates clustering towards H_{mean} . UAVSAR had the only clear non-linear trend, with a
381 saturating relationship above 20-30 m in mean height. Variability across all sensors increased
382 non-linearly with mangrove stature (Figure 5B and 3D) to a maximum of ~8 m. Variability of
383 individual products with respect to height class displayed a peak of ~2.5-7 m at ~30 m H_{mean}
384 after which remained constant or decreased slightly (Supplementary Material; Figure S3; Figure
385 S4). Relative variation (%; H_{sd} normalized by H_{mean}) universally decreased with increasing H_{mean}
386 (Supplementary Material, Figure S4).

387

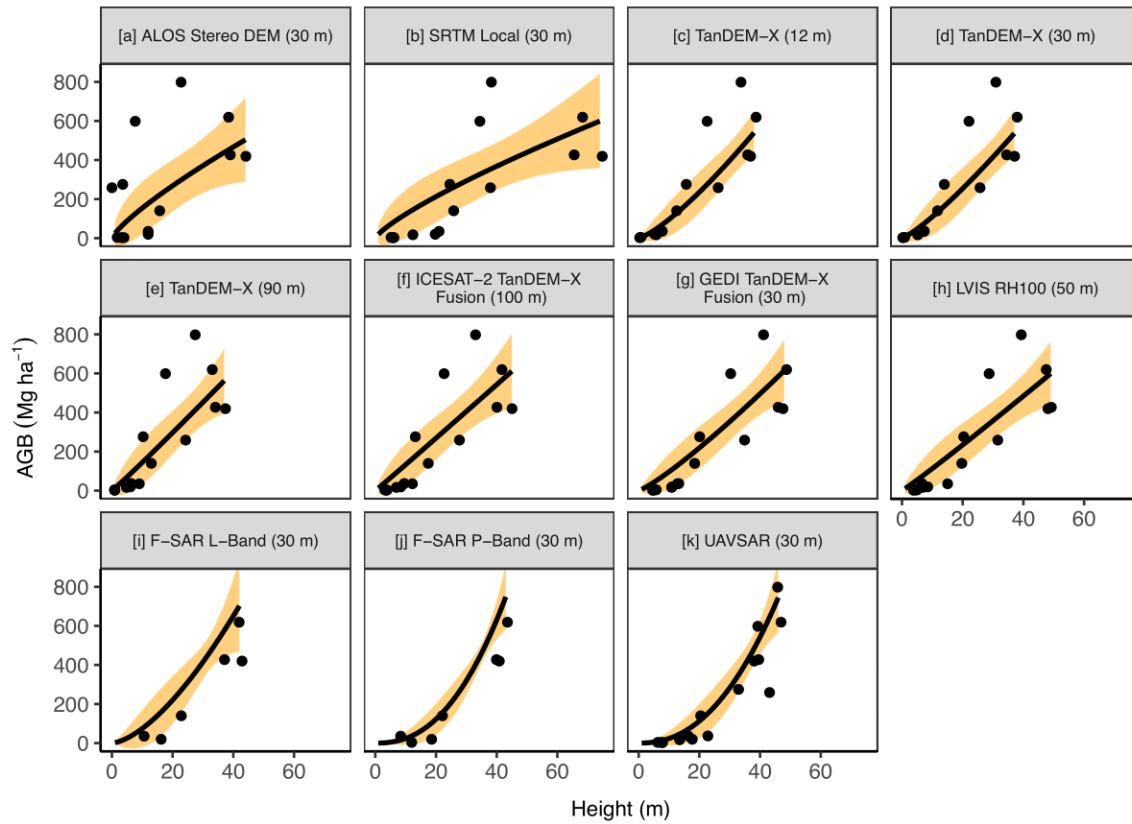


388
 389 **Figure 4:** [A] Comparison between maximum field measured height and remotely sensed
 390 heights (RMSE shown). Several remote sensing products estimate maximum field measured
 391 height, while some represent a specific percentile of field measured tree height. [B] RMSE and
 392 [C] bias in the comparison between field tree height percentiles (50th to 100th) and each remote
 393 sensing product. Colors correspond to point color shown in panel A and X's indicate the
 394 percentile at which RMSE or bias are lowest.



396
 397 **Figure 5:** (A) Example map of mean mangrove canopy height (H_{mean}) and (B) variation across
 398 the 9 sensors compared in this study. Relationship between sitewide mean mangrove canopy
 399 height (H_{mean}) and (C) product heights minus H_{mean} and standard deviation (within a product
 400 (color) and across products (grey)) from 9 remote sensing products. Variability increases with
 401 H_{mean} , while the equivalent relative variation decreases with H_{mean} (See Supplementary Material
 402 Figure S3).

403 3.3 Sensor Aboveground Biomass Calibration and Uncertainty
404 Calibration models scaling coefficients ranged from 0.85 (SRTM Hmax and ALOS PRISM Stereo
405 DEM) to 3.11 (UAVSAR L-band height), with lower scaling coefficients generally having higher
406 calibration uncertainty (Figure 6; Table 2). The locally calibrated AGB products fell into two broad
407 categories: [i] global spaceborne and [ii] local airborne sensors. AGB of spaceborne sensors had
408 higher RMSE (mean = 78%, sd = 14%) than airborne sensors (mean = 62%, sd = 27%). Biomass
409 models using airborne products were ~5% less biased than global spaceborne products. The best
410 performing models for local and global height products were PolInSAR (L-band F-SAR; RMSE=
411 121 Mg ha⁻¹; P-band F-SAR; RMSE= 71 Mg ha⁻¹; UAVSAR; RMSE = 92 Mg ha⁻¹) and X-band
412 interferometry (TanDEM-X; RMSE = 142 Mg ha⁻¹), respectively. The GEDI calibrated TanDEM-
413 X AGB product had lower uncertainty than the TanDEM-X AGB product of the equivalent spatial
414 resolution (RMSE = 139 Mg ha⁻¹; 146 Mg ha⁻¹). ICESat-2 calibration of TanDEM-X heights did not
415 improve model performance (RMSE = 180 Mg ha⁻¹). The F-SAR products covered only 4-5 of the
416 field plots and 20% of the study area, altering mean plot biomass (238-277 Mg ha⁻¹; ~50 Mg ha⁻¹
417 higher than other height products), but, while heights were lower than H_{mean} (Figure S7), AGB
418 predictions remained similar to AGB_{mean} (Figure S8). Given the fact that LiDAR is considered one
419 of the best means of creating areawide biomass maps, the large-footprint waveform LiDAR (LVIS
420 RH100) had a higher than expected RMSE (174 Mg ha⁻¹), putting it on-par with other global
421 spaceborne sensors (e.g. C-band SAR interferometry (SRTM; 190 Mg ha⁻¹) and stereo
422 photogrammetry (ALOS PRISM DEM; 220 Mg ha⁻¹)).
423



424

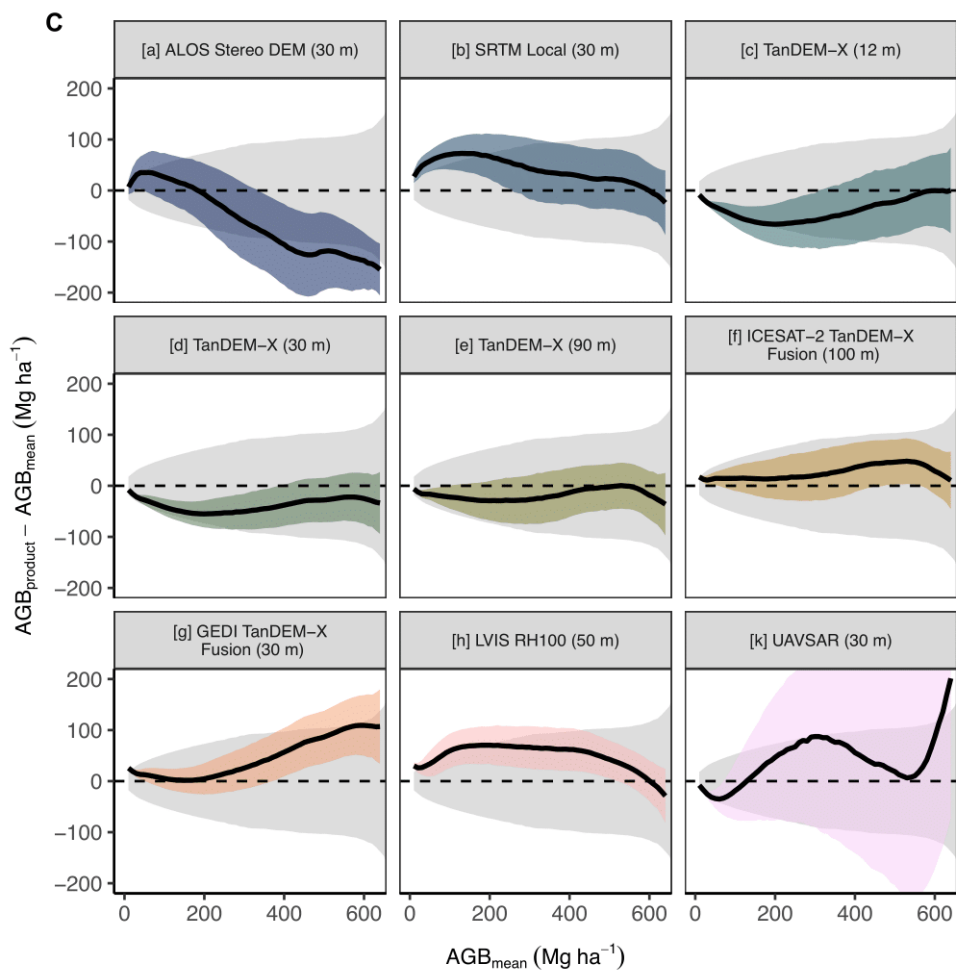
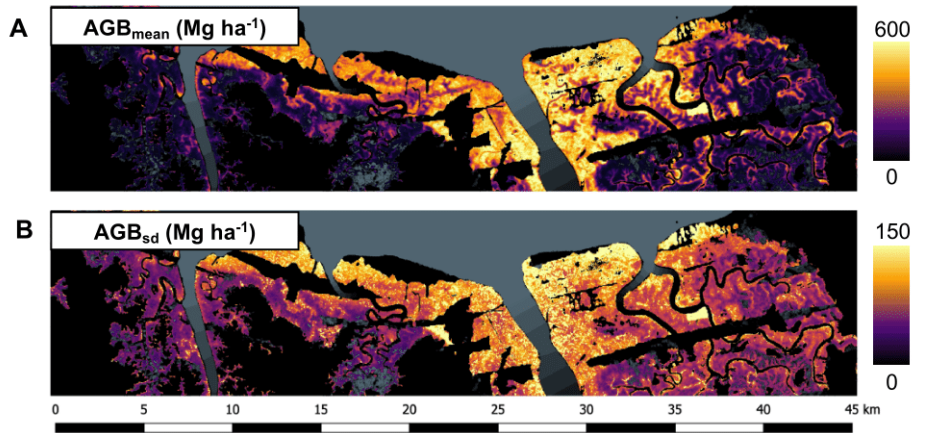
425 **Figure 6:** Non-linear height allometry for 11 remote sensing products. See Table 2 for
 426 corresponding model coefficients and fit and validation statistics.

427

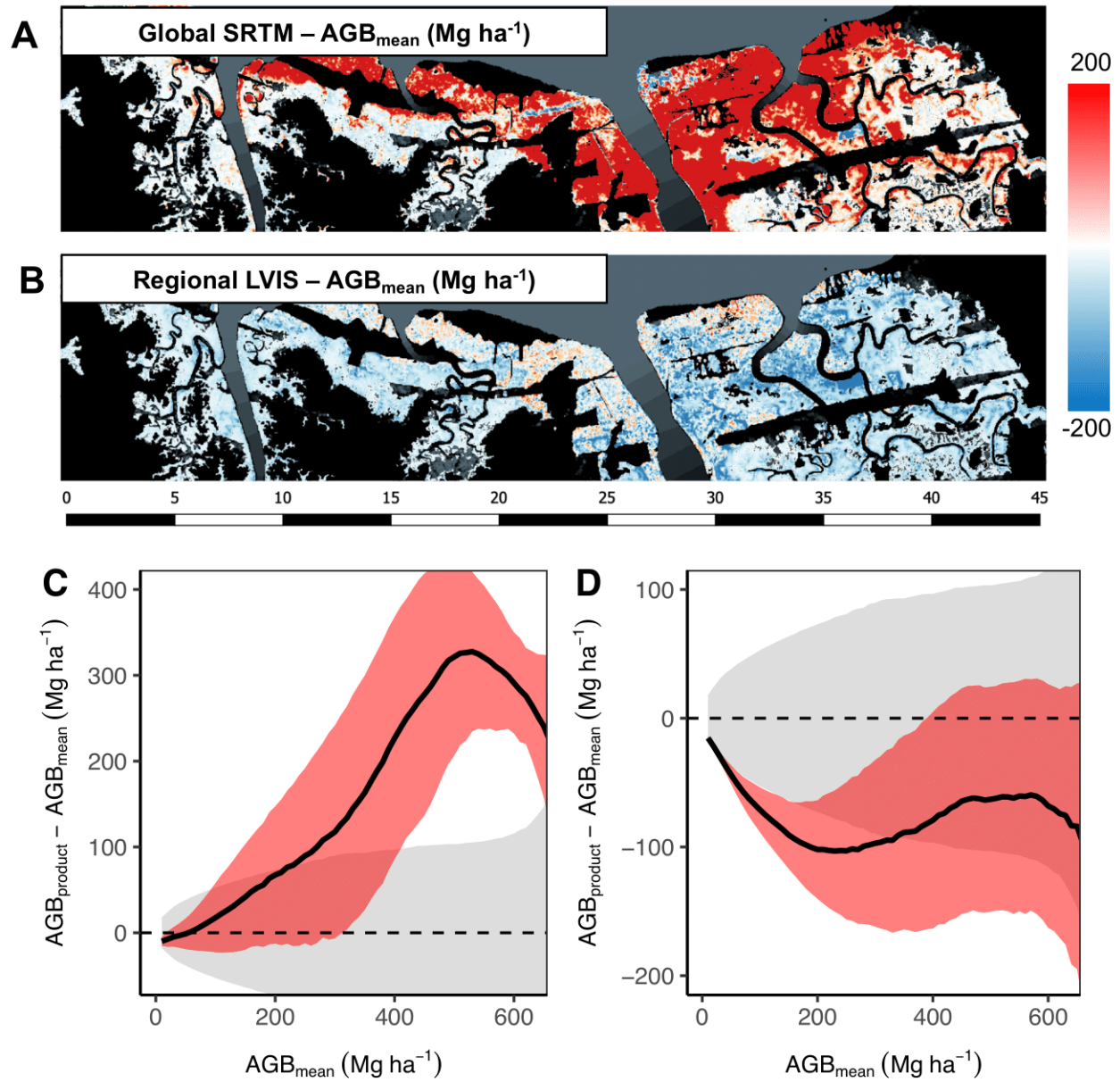
428 3.4 Spatial Patterns and Variability in Biomass

429 Nearly all AGB products approximated AGB_{mean} , with $\sim 60 \text{ Mg ha}^{-1}$ maximum standard deviation
430 across all models above 600 Mg ha^{-1} ($\sim 30 \text{ m H}_{mean}$; grey, Figure 7). All TanDEM-X products
431 clustered towards the AGB_{mean} , with ICESat-2 and GEDI estimate increasing AGB above 400
432 Mg ha^{-1} AGB_{mean} . The ALOS DEM AGB product was lower than average above 200 Mg ha^{-1} ,
433 while SRTM H_{max} AGB was higher than average below 200 Mg ha^{-1} . UAVSAR AGB was highly
434 variable, but predicted consistently higher than AGB_{mean} . AGB_{std} increased with AGB_{mean} class for
435 all AGB products (Figure 7; Supplementary Material; Figure S5; Figure S6). The 30 m global
436 SRTM mangrove biomass product was consistently $\sim 40\%$ higher than AGB_{mean} , while the regional
437 LVIS AGB product was $\sim 200\text{-}400 \text{ Mg ha}^{-1}$ lower compared to AGB_{mean} (Figure 8). Both of these
438 trends are clear in the difference maps (SRTM AGB – AGB_{mean} and LVIS AGB - AGB_{mean} ; Figure
439 8A and 8B).

440



441
 442 **Figure 7:** (A) Example map of mean mangrove biomass (AGB_{mean}) and (B) variation across the
 443 9 sensors compared in this study. Relationship between AGB_{mean} and (C) product biomass
 444 ($AGB_{product}$) minus AGB_{mean} (residuals), with standard deviation (within product (color) and
 445 across all products (grey)) from 9 remote sensing products. AGB variability increases with
 446 AGB_{mean} , while the equivalent relative variation decreases with AGB_{mean} (Supplementary
 447 Material, Figure S5).



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Figure 8: Pixel-level comparison of local mean (across 9 products) aboveground biomass estimates (AGB_{mean}) and the [A] Global SRTM mangrove biomass and [B] regional LVIS biomass products. The residual plots indicate a systematic positive difference in the [C] SRTM-based model, increasing with increasing biomass values, and a systematic negative difference in the [D] regional LVIS biomass model.

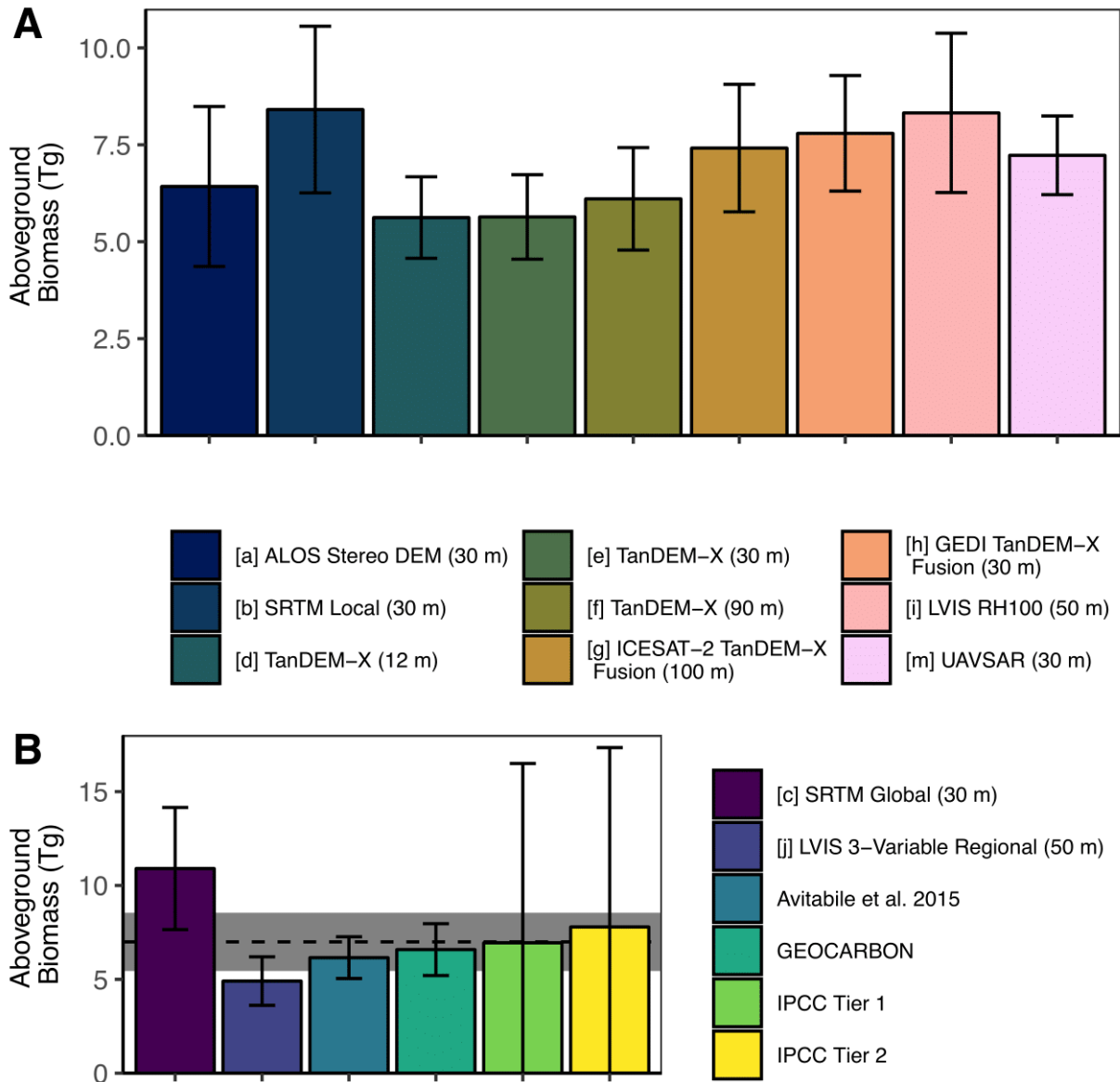
455 3.5 Total Biomass and Uncertainty

456 We compared area-wide totals of nine different locally calibrated AGB products, along with four
457 regional and globally calibrated AGB products, and two IPCC tier-based estimates (**Table 2**;
458 **Figure 9**). All biomass models calibrated with the local plot data predicted similar total biomass
459 for the entirety of Pongara National Park (**Figure 9**) using both global spaceborne (mean=6.8 Tg,
460 sd=1.1 Tg) and local airborne sensors (mean=7.8 Tg, sd=0.8 Tg). The global SRTM-based
461 biomass model predicted ~29% higher total biomass than the locally calibrated SRTM model. The
462 regionally calibrated tropical forest 3-variable LVIS model predicted ~41% lower biomass than the
463 single-variable RH100 model. The local mean predictions were ~19-29% higher than the two 1
464 km global biomass maps [50], [51]. The global models had total uncertainty of 1.1-1.4 Tg or ~18-
465 21%, compared to the 14% total uncertainty in the UAVSAR AGB product. In contrast, the IPCC
466 Tier 1 area-based estimates for Mangrove Tropical Wet forests had 9.6 Tg or 137% total
467 uncertainty and the total predicted biomass was 6.9 Tg. The IPCC Tier 2 estimates increased
468 average biomass from in-situ plot data, resulting in an increase to 7.8 Tg total AGB. The IPCC
469 estimates were less than 0.8 Tg (12%) difference from the average biomass predicted from all
470 locally calibrated high-resolution biomass products.

471 **Table 2:** Summary of comprehensive biomass calibration and predictions for the 17 products evaluated. All aboveground biomass
 472 (AGB) values are in Mg ha⁻¹ unless other units are specified. Uncertainty in the AGB prediction (σ) was derived from area-wide mean
 473 uncertainty from the AGB model parameter fits. Uncertainty in all baseline datasets was derived from an independent validation of the
 474 mapped values with the plot-level AGB estimates.
 475

		Calibration				Validation				Totals				
		Mean AGB	Plots	β	α	RMSE	Bias	Bias (%)	RMSE (%)	Mean AGB	σ	Area (ha)	Total AGB (Tg)	σ
Global	[a] ALOS Stereo DEM (30 m)	215	17	20.21	0.85	220	7	3	102	178	57	34960	6.4	2.1
	[b] SRTM Local (30 m)	226	16	15.89	0.85	190	18	8	84	232	59	37870	8.4	2.1
	[c] TanDEM-X (12 m)	215	17	4.90	1.29	142	16	8	66	155	29	35295	5.6	1.1
	[d] TanDEM-X (30 m)	215	17	6.47	1.21	146	21	10	68	156	30	36350	5.6	1.1
	[e] TanDEM-X (90 m)	215	17	8.33	1.16	165	26	12	77	169	37	37690	6.1	1.3
	[f] ICESAT-2 TanDEM-X Fusion (100 m)	215	17	9.20	1.11	180	26	12	84	205	45	37771	7.4	1.6
	[g] GEDI TanDEM-X Fusion (30 m)	215	17	4.24	1.30	139	11	5	64	215	41	36808	7.8	1.5
Local	[h] LVIS RH100 (50 m)	215	17	11.45	1.01	174	17	8	81	230	57	26650	8.3	2.1
	[i] F-SAR L-Band (30 m)	277	6	2.05	1.56	121	16	6	44	408	50	7322	-	-
	[j] F-SAR P-Band (30 m)	238	7	0.10	2.37	71	-11	-5	30	368	33	7595	-	-
	[k] UAVSAR (30 m)	215	17	0.01	3.11	92	-5	-2	43	200	28	36191	7.2	1.0
Baseline	SRTM Global (30 m)									301	90	31491	10.9	3.3
	LVIS 3-Variable Regional (50 m)									136	36	26864	4.9	1.3
	Avitabile et al. 2015 (1 km)									170	31	83791	6.2	1.1
	GEOCARBON (1 km)									182	38	73880	6.6	1.4
	IPCC Tier 1									192	264	-	6.9	9.6
IPCC Tier 2									215	264	-	7.8	9.6	

476



479 **Figure 9:** Aboveground biomass totals for [A] nine locally calibrated and [B] six regional, global,
 480 and IPCC-based estimates. Totals are based on the mean estimates across the study area
 481 extrapolated via an area-based estimate. Error bars represent the 95% confidence interval of the
 482 total biomass estimate derived from plot based independent validation. Dotted black line and grey
 483 shaded area in B shows the mean and standard deviation of the locally calibrated area-wide
 484 biomass totals. Note: the y-axis scales between A and B are not fixed to highlight differences in
 485 each figure.
 486

487 **4.0 Discussion – 3335**

488 Few previous studies have compared canopy height products from airborne and satellite
489 products for terrestrial and mangrove forests and those that have are limited in the number of
490 datasets [55]–[57]. Recently, new satellite sensors were launched and new overlapping airborne
491 campaigns occurred, collecting a wealth of spatial data. We evaluate the broad spectrum of
492 state-of-the-art sensor data products across the full range of height and biomass known to exist
493 globally in mangrove forest ecosystems, providing a much-needed baseline for sensor
494 performance. Our major findings are as follows:

495 [1] height estimates are not consistent across products, with opposing trends in relative
496 and absolute errors, highlighting the need for an adaptive approach to constraining
497 height estimates, depending on forest stature;

498 [2] radar products had the lowest calibration error and bias, with superior results from
499 airborne instruments and improvements to spaceborne estimates with LiDAR fusion
500 using height alone;

501 [3] AGB variability and uncertainty strongly depends on forest stature, with AGB_{sd}
502 increasing with canopy height, while relative AGB_{sd} variation was highest in low-stature
503 stands, suggesting the greatest improvements may be in low- to mid-biomass density
504 ecosystems;

505 [4] for AGB mapping, a remote sensing product's sensitivity to variations in canopy
506 structure is more important than the absolute accuracy of height estimates;

507 [5] locally-calibrated area-wide totals are more representative than generalized global
508 biomass models for high-precision biomass estimates;

509

510 Here, we first consider the more technical aspects of our results (Section 4.1) and follow with a
511 discussion in the context of scientific and application-focused relevance at local, national, and
512 global spatial scales (Section 4.2).

513 4.1 Evaluating Remotely Sensed Forest Height and Biomass

514 *4.1.1 Evaluating Remotely Sensed Mangrove Height*

515 Mangrove canopy height estimates disagreed substantially across sensor type with absolute
516 errors increasing as a function of canopy height. We attribute these differences almost
517 universally to the sensor measurement approach capturing canopy height [56]. An extreme
518 example of the effect of measurement approach is with the ALOS PRISM product; Though past
519 work highlighted the ALOS PRISM product as capable of capturing broad successional patterns
520 in mangrove stands [58], here, height estimates were insensitive to both low and tall stature
521 forests compared to active remote sensing methods. Our comparison of remotely sensed height
522 to plot-level height percentiles from tree-level inventory measurements highlights major
523 differences in what part of the forest canopy is being measured with each height estimate
524 (**Figure 3**). InSAR instruments are simultaneously sensitive to height and vegetation volume
525 density, which is preferable for biomass modeling. The existing SRTM H_{\max} product is most
526 representative of maximum tree height, but these estimates have high error (RMSE: ~7-8 m).
527 The 12 m TanDEM-X phase center elevation product captures the 75th percentile of tree heights
528 with 50% lower error than SRTM (RMSE: ~4 m) – potentially since TanDEM-X measurements
529 were more closely temporally aligned to the field campaign (5 years vs. 17 years for SRTM). In
530 both cases, field based validation of these remote sensing estimates is key to understanding the
531 specific height attribute represented with a particular remote sensing product [59] and sensor
532 choice should be dependent on the end goal (e.g. height vs. biomass).

533 Radar instrument wavelength and measurement technique reflected specific height anomalies.
534 SRTM heights (C-band PolInSAR) differed positively from average with increasing canopy
535 height, but height estimates would be substantially underestimated without the ICESat GLAS-
536 calibration and is likely affected by secondary structural variables (e.g. canopy cover or basal
537 area). In agreement with past work [39], UAVSAR (L-band) heights were consistent ($\sim\pm 10$ m
538 average difference) until ~ 40 m height, above which heights became shorter than average –
539 evidence of sensor saturation. TanDEM-X offers a precise, high-resolution height product that
540 makes it one of the best options for continuous mapping of mangrove stands at a global scale.
541 With the inclusion of additional height data (e.g ICESat-2 and GEDI), height estimates became
542 less biased (up to 20 m offset in the tallest stands), pointing to a key fusion application in future
543 studies. However, when calibrating continuous height products with sampling instruments,
544 ground surface identification is a major issue in closed canopy systems and is likely
545 exacerbated in the presence of water and dense aboveground mangrove root networks. All of
546 these trends are consistent with our expectations of radar wavelength and forest height. In
547 general, longer wavelengths penetrate further into the canopy, decreasing height estimates from
548 the canopy top, but we expect height estimates are also influenced by canopy density.

549 *4.1.2 Sensor Aboveground Biomass Calibration and Uncertainty*

550 The height-biomass allometry across sensors varied from sublinear to linear with high
551 uncertainty to more power-like models with low calibration uncertainty. In fact, we found a
552 consistent negative trend between the scaling coefficient and model uncertainty. We also found
553 a consistently higher mean biomass prediction with high-uncertainty models with lower scaling
554 coefficient values – suggesting higher uncertainty models may be systematically over predicting
555 biomass density [9]. Sensor measured height was the major factor affecting calibration
556 uncertainty in our analysis, but other factors (e.g. plot size, plot shape, plot sample location,
557 sample size, geolocation errors) can directly impact biomass calibration models and predictions.

558 Future work in mangrove systems that independently evaluate these factors affecting model
559 errors will provide more precise estimates of the spatial distribution of prediction uncertainty.

560

561 In general, radar sensors provided the lowest error and bias biomass calibration of the 11 local
562 models, but the addition of LiDAR-derived canopy heights improved model statistics. InSAR
563 (TanDEM-X) is likely the best available option for developing an updated global mangrove
564 biomass product, evidenced by the low errors in calibration minimally affected by product
565 resolution (e.g. aggregating by a factor of 7.5 inflated RMSE by only 9%). The local scale L-
566 band UAVSAR Polarimetric InSAR product performed even better, likely due to higher
567 sensitivity to canopy cover, trunks and woody components (i.e. basal area). The higher than
568 expected RMSE in the LVIS AGB model, suggests canopy height alone is a less powerful
569 predictor than the phase center height captured with radar instruments. Radar-LiDAR fusion
570 approaches (e.g. GEDI-corrected TanDEM-X heights) improve calibration by reducing bias and
571 RMSE, but even greater benefits are possible in areas with greater topographic relief (e.g. non-
572 mangrove systems), since LiDAR is primarily improving heights through more accurate ground
573 detection. Though our intercomparison provides a robust analysis of height-biomass allometry
574 for calibration of remote sensing datasets, we did not explicitly evaluate the suite of potential
575 multi-variate approaches possible for predicting spatial distributions of AGB (e.g. [49]),
576 especially for LiDAR sensors (e.g. GEDI, ICESat-2, and LVIS). As such, the results presented
577 here do not emphasize the full benefits of using LiDAR-based multivariate models for biomass
578 prediction. Future biomass calibration approaches should incorporate multivariate statistical
579 approaches to take full advantage of the ability of LiDAR to capture internal canopy structure.

580 *4.1.3 Spatial Patterns and Variability in Biomass*

581 Summarizing height-biomass trends from pixel-level predictions highlights the product-specific
582 variations across AGB products in this tall mangrove system. Though plot based calibration

583 models were often super-linear, site scale height-biomass allometry was more linear across all
584 products, with AGB_{mean} increasing by 13 Mg ha^{-1} per unit H_{mean} . The most non-linear calibration
585 model (UAVSAR) deviated most clearly from the general linear trend increasing more rapidly
586 than AGB_{mean} from 0-25 m and increasing less rapidly above 25 m H_{mean} . The observed
587 consistent linear relationship is ideal for cross-calibration, enabling more compatible multi-
588 sensor approaches to biomass monitoring [60]. Variability of a single AGB product was on the
589 same order as the AGB variation across all products. Between 0 and 30 m, the standard
590 deviation in mean biomass across all sensors increased linearly from 20-80 Mg ha^{-1} . Trends in
591 pixel-level prediction uncertainty were similar for all sensors and within the range of AGB
592 variation, increasing from 0-30, leveling off, then increasing to a maximum of 60-90 Mg ha^{-1} .
593 Relative to traditional forest inventory methods, all locally calibrated remote sensing estimates
594 had pixel-level uncertainty that was low, suggesting the use of a remote sensing framework is
595 more important than the choice of sensor itself. For instance, the worst performing product
596 calibration (ALOS PRISM; RMSE =102%) only translated to a marginal site-wide average
597 uncertainty ($\sim 30\text{-}80 \text{ Mg ha}^{-1}$), suggesting a product's sensitivity to variations in canopy structure
598 is more important than the absolute accuracy of height estimates.

599
600 Calibration of remote sensing products should be as local as possible in areas of high biomass
601 density. Our comparison of the global SRTM biomass product to AGB_{mean} highlights the effects
602 of generalized predictive models excluding representative plot data. The global product was
603 systematically $\sim 40\%$ higher than the local predictions, resulting in more than 400 Mg ha^{-1} higher
604 biomass density in some instances of high AGB_{mean} . While the Simard et al. [9] map is unable to
605 accurately capture biomass density in the high biomass areas of Pongara National Park, we
606 believe the core cause is lack of calibration data in these extremely tall stands. Moreover, the
607 map clearly provides the most accurate AGB predictions in mid-stature stands (10-20 m), where
608 nearly all plot-level calibration data and global mangrove canopy heights reside. Alternatively,

609 the regional LVIS AGB product is more precise, but is negatively biased, remaining within 100
610 Mg ha⁻¹ throughout the AGB_{mean} range. The negative bias is similarly related to the product
611 calibration, relying on lower mean wood density forest inventory data, pushing predictions lower
612 than expected in Pongara, where *Rhizophora sp.* have ~0.9 specific gravity. We suggest
613 establishing future field plots and planning airborne campaigns that fill data gaps in high-
614 biomass locations. For example, a targeted approach could use current AGB estimates to
615 identify key areas of high AGB density with few or no available field data. Adding these
616 additional *in-situ* observations will ultimately improve AGB calibration and provide more stable
617 AGB predictions. In summary, these two global and regional products highlight the importance
618 of appropriate plot-level calibration data to ensure both precise and accurate area-wide biomass
619 distributions.

620

621 Two opposing patterns were clear with respect to variation in spatial biomass trends with forest
622 stature: [1] absolute variation increases and [2] relative variation decreases. The extreme, tall
623 forests have ~200 Mg ha⁻¹ (or ~20%) standard deviation on average across sensor predictions.
624 Short forests (0-15 m) disagree by 40 Mg ha⁻¹ (or ~50%), on average. So, where will biomass
625 model improvements be most impactful at the global scale: short or tall stands? We evaluated
626 biomass models in a unique system capturing greater than 60 m of variation in mangrove forest
627 structure, but more than 95% of the worlds mangroves are less than 40 m tall [9], suggesting the
628 greatest benefits may be in low- to mid-biomass density ecosystems. Biomass is an essential
629 biodiversity variable [61], so improved biomass predictions stand to also directly affect biodiversity
630 mapping and conservation efforts. These improvements will help to better capture changes in
631 biomass over time in areas of growth, regeneration, degradation, and loss [62]. With this
632 knowledge, we suggest developing global biomass products that are most precise in low- to mid-
633 stature forests, but identifying and locally calibrating biomass models in tall-stature forests.

634 *4.1.4 Total Biomass and Uncertainty*

635 Locally calibrated biomass products provided similar total area-wide biomass estimates (all 95%
636 confidence intervals overlap), even though biomass distributions often differed depending on
637 sensor choice, having implications for carbon reporting and forest management. Accurate
638 representations of the AGB distribution is key for identifying potential sites for restoration or
639 conservation and carbon accounting priority [63], [64].

640

641 Globally available biomass maps [50], [51] performed well, underestimating total biomass by only
642 ~0.6 Tg (6.2-6.6 Tg totals), with total uncertainty ranging from 18-20%. In contrast, compared to
643 the mean AGB predicted with local models, the global SRTM model [9] over predicted total
644 biomass by ~3.9 Tg or 56%. The overprediction reflects two major issues: [i] the structure of this
645 extremely tall mangrove stand is more closely analogous to a high wood density tropical forest
646 than mangroves and [ii] inclusion of representative plot data is essential when building global
647 biomass products (i.e. predictions outside of observations should be considered with caution).

648

649 From a carbon accounting perspective, the high uncertainty of these predictions substantially
650 reduces their utility in tall forest stands, suggesting these global, coarse resolution generalized
651 models should not be universally relied on for precise and accurate forest carbon estimates. The
652 majority of mangrove calibration data resides in shorter stands [9] and it is here where global
653 biomass maps have less biased carbon estimates. Future global carbon maps should incorporate
654 updated global height datasets that are freely available (e.g. TanDEM-X 90 m resolution), while
655 also addressing the need for recalibration of past datasets as more calibration plot data becomes
656 available. Surprisingly, the average area-based IPCC biomass density produced superior
657 predictions, albeit not spatially explicit, limiting their utility for forest management and
658 conservation.

659

660 4.2 Implications for Multi-scale Forest Structure Applications

661 *4.1 Local Scale*

662 Mangrove forest height is uncertain across the products evaluated, posing a major challenge for
663 incorporating remote sensing products into local forest management schemes. We found the
664 relative uncertainty across height products to be highest in low stature stands (>50% in stands
665 <15 m), while in the tallest stands (~45 m H_{mean}) uncertainty was ~20% or 7-8 m. For context,
666 many forest definitions rely on height thresholds of 5 or 10 m, so the uncertainty in these lower
667 stature forests may impact estimates of forest extent, depending on the product selected [65].
668 Further, canopy height is a major determining factor in selecting harvest or conservation areas
669 and these model errors could potentially lead to misinformed local forest management decisions
670 [66]. In the context of coastal flood protection, mangrove forest density/cover improves flood
671 buffering capacity and consistent monitoring over time will provide consistent and precise
672 estimates for superior disaster planning [67]. The broad relationship between cover and forest
673 height suggests a similar flood buffering capacity may be observed in taller mangrove forests.
674

675 *4.2 National Scale*

676 Inconsistent forest height estimates did not translate to dramatically different estimates of area-
677 wide AGB – an encouraging finding for adopting a diverse array of sensors, depending on data
678 availability, for national carbon accounting [27], [60]. The most important factor to consider at
679 the national scale is the availability of representative forest plot data to ensure the accuracy of
680 remote sensing-based AGB predictions [68]–[73].

681

682 Interestingly, even in the absence of field data or spatially explicit estimates (e.g. IPCC),
683 average mangrove biomass density provided unbiased total AGB estimates (in this case). The
684 accuracy of IPCC estimates is encouraging from a mangrove biomass and accounting

685 perspective (~12% from locally calibrated remote sensing-based totals), particularly with the
686 inclusion of mangroves within payment for ecosystem service (PES) schemes such as REDD+,
687 since nominally attributed values are deemed to be generally representative of reality [74], [75].
688 Middle and low income countries make up the majority of mangrove holding nations and forest
689 area [76], [77], but may be less likely to prioritize expensive field data for improved calibration
690 models. In these cases, the use of the IPCC estimates for regional and national reporting is
691 encouraging, but should be more thoroughly evaluated in other countries. Regardless of the
692 accuracy, the high true uncertainty (based on validation) of IPCC totals (~130%) is still a major
693 barrier limiting their application in the context of PES, which reduce valuations as AGB
694 uncertainty increases [75].

695

696 Spatially explicit estimates made with locally calibrated AGB models were essential to reducing
697 uncertainty in area-wide total mangrove AGB, underscoring the importance of applying remote
698 sensing-based mapping of AGB for carbon accounting, whenever feasible [79]. In contrast, the
699 global mangrove AGB model [9] was ~40% biased in every height class in comparison to the
700 locally calibrated estimates. After a direct comparison of AGB predictions from our local
701 allometric model, we can clearly attribute this consistent bias to a global allometric height-
702 biomass model calibrated without reference data representing the forest heights observed in
703 Pongara National Park (maximum observed plot height in the Simard et al. study was ~40 m;
704 **Figure 10**). In the same respect, application of regional calibration models outside of the
705 specific forest systems can result in bias, altering the total estimates AGB in a forest [5]. In
706 short, both regional and global AGB products must be locally re-calibrated and validated before
707 being taken as “truth” at the local or national scale [80], [81].

708

709 *4.3 Global Scale*

710 The next generation of global mangrove forest structure (height and AGB) products will need to
711 address three major challenges: [1] reducing uncertainty in remotely sensed heights covering
712 the vast majority of mangrove area, [2] ensuring representativeness of sparse plot data and
713 AGB allometry, and [3] understanding of factors controlling secondary structure variables
714 beyond height that directly influence AGB (e.g. basal area).

715

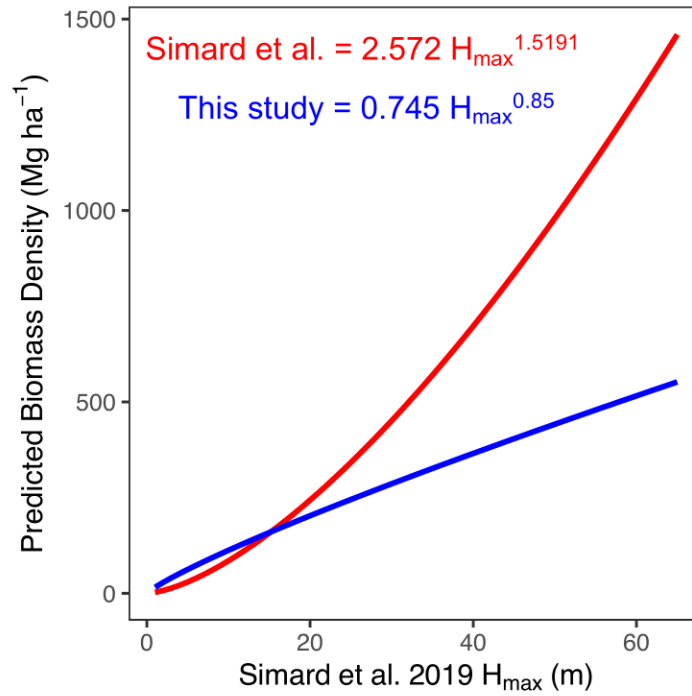
716 Approximately 95% of all mangrove forests are below 40 m in height with a global median of
717 ~13 m – around 50% of global mangrove area has between 50-70% uncertainty in remotely
718 sensed height estimates [9]. In effect, our findings of substantial disagreement in height
719 estimates across sensors for the most common height range of mangroves globally suggests a
720 universal field-based plot height-biomass allometry cannot be confidently applied across
721 sensors. Yet, the comparisons made here provide clear expectations for the biases for each
722 height product and the potential for cross calibration [60]. The near-linear height biomass
723 relationship present across most remote sensing products in this study suggests cross-
724 calibration is possible with a maximum total uncertainty of ~50-100 Mg ha⁻¹. Similar to the
725 disagreements in height, relative variations in AGB predictions across products was high in
726 short stature forests (<15 m), reaching between 50-200%. Given the high cross sensor
727 variability it is critical global continuous height products be created with rigorously validated and
728 spatially continuous height products (e.g. TanDEM-X). Moreover, the key to global validation of
729 canopy height is likely to come with spaceborne LiDAR sensors (e.g. GEDI; [22]) with direct
730 ground detection – a major limitation with PolInSAR height estimates in other forests with
731 topography [23].

732

733 Global AGB calibration datasets are sparse and likely have unrepresentative tree-level
734 allometric estimates of biomass [5], [31], [82] (Figure 10). The most exhaustive remotely sensed

735 mangrove specific AGB map to date used 332 AGB field plots to calibrate ICESat-GLAS
736 adjusted SRTM data [9]. Spatial biases are especially prevalent in the global calibration dataset,
737 with 45 % of plot data from a single country (Bangladesh). Simard et al. [9] compared several
738 regional allometric relationships with significant biases. They did not have allometry for the
739 Atlantic coast of Africa which may be reflected in the observed biases. Improved plot-level
740 calibration data is clearly needed to improve the predictions of global models [83]. Sensor
741 calibration in unique forest ecosystems is limited by a lack of unrepresentative plot-based
742 calibration data (i.e. plot-level biomass estimates may be inaccurate due to biased tree-level
743 allometry; [84], [85]). Here, we constrained our analysis to a single allometric equation [31], but
744 the representativeness of this equation in such an understudied ecosystem remains unknown
745 [86], [87]. Our evaluation of variation due to plot-level allometric biomass estimates using the
746 Komiyama et al. [32] equation the highlights potential for a propagation of changes in plot-based
747 calibration by a simple change in tree-level allometry [73]. In protected and unique systems, as
748 found in Pongara National Park, non-destructive allometric equations with novel technologies
749 (e.g. Terrestrial Laser Scanning; [88]) can bridge a critical gap in our understanding of scaling
750 relationships without detrimentally impacting the study system [82], [89], [90]. Future work
751 should focus on updating these scaling relationships [91] and assessing their impact on sensor
752 calibration [92] to better predict changes in forest biomass over time [62], [93], [94].

753
754



755

756 **Figure 10:** Comparison between Simard et al. [9] global H_{max} biomass predictions (red) and the

757 locally calibrated H_{max} model developed in this study (blue).

758 Major advances in global AGB modeling in mangrove systems will come with the inclusion of
759 measured or modeled secondary structural variables. Only approximately half of the variation in
760 global AGB models can be explained by height alone [9], suggesting secondary axes of
761 variation (e.g. basal area, stem density, regional allometry) will substantially improve mapped
762 AGB. Of the available plot data, efforts to understand the drivers in spatial variability of these
763 secondary structural characteristics will be key in precisely capturing AGB at a global scale.

764

765 Moving forward – especially in the context of this “golden age” of forest-focused active remote
766 sensing – the findings of this study enable sensor cross calibration for consistent monitoring of
767 forest function. Calibrated forest height is a key physiological variable representing organismal
768 function beyond biomass alone (e.g. moderate disturbance: [95]; hurricane damage: [96]-[97];
769 drought susceptibility: [98]-[100]). With the sensor evaluation performed here we gain the ability
770 to monitor three-dimensional structural change across sensors in mangrove forest systems
771 globally by matching past spaceborne missions (e.g. SRTM) with ongoing (e.g. TanDEM-X,
772 GEDI and ICESat-2) and future (e.g. BIOMASS or NISAR) missions.

773

774 **5.0 References**

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- 1072

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1077

1078 **Figure Captions**

1079 **Figure 1:** Map of Pongara National Park with heights from SRTM-based global height product
1080 (from Simard et al [9]). Inventory plots were placed such that canopy heights were sampled
1081 proportionally according to the height distribution across the site.

1082
1083

1084 **Figure 3:** Workflow detailing major processing and analysis steps used to compare 17 area-
1085 wide biomass products in Pongara National Park.

1086

1087 **Figure 3:** Calibration models used to create spatially continuous height estimates by fusing
1088 ICESat-2 mean canopy height and GEDI RH100 to TanDEM-X continuous heights. Red points
1089 are removed outliers in the ICESat-2 calibration (anomalous beam 3 data not shown). Colored
1090 points in C show anomalous elevation values that were removed from the final calibration model
1091 (B and D).

1092

1093 **Figure 4:** [A] Comparison between maximum field measured height and remotely sensed
1094 heights (RMSE shown). Several remote sensing products estimate maximum field measured
1095 height, while some represent a specific percentile of field measured tree height. [B] RMSE and
1096 [C] bias in the comparison between field tree height percentiles (50th to 100th) and each remote
1097 sensing product. Colors correspond to point color shown in panel A and X's indicate the
1098 percentile at which RMSE or bias are lowest.

1099

1100 **Figure 5:** (A) Example map of mean mangrove canopy height (H_{mean}) and (B) variation across
1101 the 9 sensors compared in this study. Relationship between sitewide mean mangrove canopy
1102 height (H_{mean}) and (C) product heights minus H_{mean} and standard deviation (within a product
1103 (color) and across products (grey)) from 9 remote sensing products. Variability increases with
1104 H_{mean} , while the equivalent relative variation decreases with H_{mean} (See Supplementary Material
1105 Figure S3).

1106

1107 **Figure 6:** Non-linear height allometry for 11 remote sensing products. See Table 2 for
1108 corresponding model coefficients and fit and validation statistics.

1109

1110 **Figure 7:** (A) Example map of mean mangrove biomass (AGB_{mean}) and (B) variation across the
1111 9 sensors compared in this study. Relationship between AGB_{mean} and (C) product biomass
1112 ($AGB_{product}$) minus AGB_{mean} (residuals), with standard deviation (within product (color) and
1113 across all products (grey)) from 9 remote sensing products. AGB variability increases with
1114 AGB_{mean} , while the equivalent relative variation decreases with AGB_{mean} (Supplementary
1115 Material, Figure S5).

1116

1117 **Figure 8:** Pixel-level comparison of local mean (across 9 products) aboveground biomass
1118 estimates (AGB_{mean}) and the [A] Global SRTM mangrove biomass and [B] regional LVIS
1119 biomass products. The residual plots indicate a systematic positive difference in the [C] SRTM-
1120 based model, increasing with increasing biomass values, and a systematic negative difference
1121 in the [D] regional LVIS biomass model.

1122

1123 **Figure 9:** Aboveground biomass totals for [A] nine locally calibrated and [B] six regional, global,
1124 and IPCC-based estimates. Totals are based on the mean estimates across the study area
1125 extrapolated via an area-based estimate. Error bars represent the 95% confidence interval of
1126 the total biomass estimate derived from plot based independent validation. Dotted black line and
1127 grey shaded area in B shows the mean and standard deviation of the locally calibrated area-
1128 wide biomass totals. Note: the y-axis scales between A and B are not fixed to highlight
1129 differences in each figure.

1130
1131 **Figure 10:** Comparison between Simard et al. [9] global H_{max} biomass predictions (red) and the
1132 locally calibrated H_{max} model developed in this study (blue).

