

1 Evaluation of remote sensing-based evapotranspiration products at low-latitude eddy covariance
2 sites

3

4 Diego Salazar-Martínez^a, Friso Holwerda^{b,*}, Thomas R.H. Holmes^c, Enrico A. Yépez^d,
5 Christopher R. Hain^c, Susana Alvarado-Barrientos^f, Gregorio Ángeles-Pérez^g, Tulio Arredondo-
6 Moreno^h, Josué Delgado-Balbuenaⁱ, Bernardo Figueroa-Espinoza^j, Jaime Garatuza-Payán^d,
7 Eugenia González del Castillo^b, Julio C. Rodríguez^k, Nidia E. Rojas-Robles^d, Jorge M. Uuh-
8 Sonda^j, Enrique R. Vivoni^l

9

10 ^a Posgrado en Ciencias de la Tierra, Universidad Nacional Autónoma de México, Ciudad de México, Mexico

11 ^b Instituto de Ciencias de la Atmósfera y Cambio Climático, Universidad Nacional Autónoma de México, Ciudad de México,
12 Mexico

13 ^c Hydrological Sciences Lab, NASA Goddard Space Flight Center, Greenbelt, MD, USA

14 ^d Departamento de Ciencias del Agua y Medio Ambiente, Instituto Tecnológico de Sonora, Ciudad Obregón, Sonora, Mexico

15 ^e Earth Science Office, NASA Marshall Space Flight Center, Huntsville, AL, USA

16 ^f Red de Ecología Funcional, Instituto de Ecología, A.C., Xalapa, Veracruz, Mexico

17 ^g Posgrado en Ciencias Forestales, Campus Montecillo, Colegio de Postgraduados, Texcoco, Estado de México, Mexico

18 ^h División de Ciencias Ambientales, Instituto Potosino de Investigación Científica y Tecnológica, A.C., San Luis Potosí, Mexico

19 ⁱ Centro Nacional de Investigación Disciplinaria Agricultura Familiar, Instituto Nacional de Investigaciones Forestales, Agrícolas
20 y Pecuarias, Ojuelos de Jalisco, Jalisco, Mexico

21 ^j Laboratorio de Ingeniería y Procesos Costeros, Instituto de Ingeniería, Universidad Nacional Autónoma de México, Sisal,
22 Mexico

23 ^k Departamento de Agricultura y Ganadería, Universidad de Sonora, Hermosillo, Sonora, Mexico

24 ^l School of Earth and Space Exploration & School of Sustainable Engineering and the Built Environment, Arizona State
25 University, Tempe, AZ, USA

26

27 * Corresponding author.

28 E-mail address: friso.holwerda@gmail.com (F. Holwerda).

29

30 Abstract

31 Remote sensing-based evapotranspiration (ET) products have been evaluated primarily using
32 data from northern middle latitudes; therefore, little is known about their performance at low
33 latitudes. To address this bias, an evaluation dataset was compiled using eddy covariance data
34 from 40 sites between latitudes 30° S and 30° N. The flux data were obtained from the emerging
35 network in Mexico (MexFlux) and from openly available databases of FLUXNET, AsiaFlux, and
36 OzFlux. This unique reference dataset was then used to evaluate remote sensing-based ET
37 products in environments that have been underrepresented in earlier studies. The evaluated
38 products were: MODIS ET (MOD16, both the discontinued collection 5 (C5) and the latest
39 collection (C6)), Global Land Evaporation Amsterdam Model (GLEAM) ET, and Atmosphere-
40 Land Exchange Inverse (ALEXI) ET. Products were compared with unadjusted fluxes (ET_{orig})
41 and with fluxes corrected for the lack of energy balance closure (ET_{ebc}). Three common
42 statistical metrics were used: coefficient of determination (R^2), root mean square error (RMSE),
43 and percent bias (PBIAS). The effect of a vegetation mismatch between pixel and site on product
44 evaluation results was investigated by examining the relationship between the statistical metrics
45 and product-specific vegetation match indexes. Evaluation results of this study and those
46 published in the literature were used to examine the performance of the products across latitudes.
47 Differences between the MOD16 collection 5 and 6 datasets were generally smaller than
48 differences with the other products. Performance and ranking of the evaluated products depended
49 on whether ET_{orig} or ET_{ebc} was used. When using ET_{orig} , GLEAM generally had the highest R^2 ,
50 smallest PBIAS, and best RMSE values across the studied land cover types and climate zones.
51 Neither MOD16 nor ALEXI performed consistently better than the other. When using ET_{ebc} ,
52 none of the products stood out in terms of both low bias and strong correlations. The use of ET_{ebc}

53 instead of ET_{orig} affected the biases more than the correlations. The product evaluation results
54 showed no significant relationship with the degree of match between the vegetation at the pixel
55 and site scale. The latitudinal comparison showed tendencies of lower R^2 (all products) but better
56 PBIAS and normalized RMSE values (MOD16 and GLEAM) for forests at low latitudes than for
57 forests at northern middle latitudes. For non-forest vegetation, the products showed no clear
58 latitudinal differences in performance.

59

60 Keywords: MOD16; GLEAM; ALEXI; tropics; subtropics

61

62 1. Introduction

63 The low latitudes (30° S–30° N) are characterized by large contrasts in terrestrial
64 evapotranspiration (ET). They are home to tropical rainforests and other ecosystems with
65 abundant rainfall where energy (radiation) is the main constraint to ET (see, for example, Fisher
66 et al., 2009; Bruijnzeel et al., 2011). They are also home to tropical and subtropical arid and
67 semi-arid ecosystems where ET is limited by water supply (see, for example, Eamus et al., 2013;
68 Delgado-Balbuena et al., 2019). Estimates of ET for this region are needed for local and regional
69 applications such as water resource management and drought monitoring and for global
70 applications such as climate change studies (Wang and Dickinson, 2012; Fisher et al., 2017). The
71 spatial and temporal scale of these applications require other methods than those used to study
72 ET at the plot to ecosystem scale (i.e., lysimeter, sap flow, and micrometeorological methods).
73 Recently, the potential of remote sensing-based ET estimates for these purposes has been
74 recognized (Dolman et al., 2014; Fisher et al., 2017; Sheffield et al., 2018).
75 Since the 1990s, numerous remote sensing-based ET models have been developed (see Ke Zhang
76 et al., 2016 for an overview). These models can be broadly divided into three categories (in no
77 specific order): models based on the (1) Penman-Monteith (Monteith, 1965) or (2) Priestley-
78 Taylor (Priestley and Taylor, 1972) equations and (3) models that determine the sensible heat
79 flux (H) and calculate ET (or latent heat flux, LE) as the residual of the surface energy balance
80 (i.e., so-called SEB models; Wang and Dickinson, 2012; Ke Zhang et al., 2016; Chen and Liu,
81 2020). Of interest to the user community is the development of global ET products from these
82 models that are readily available to the public and regularly updated to include the latest data.
83 Two such datasets have been produced since the early 2010s: 1) the MODIS ET product based
84 on the MOD16 algorithm (Penman-Monteith type model; Mu et al., 2007, 2011; Running et al.,
85 2019); and 2) the ET product from the Global Land Evaporation Amsterdam Model (GLEAM;

86 Priestley-Taylor type model; Miralles et al., 2011; Martens et al., 2017). More recently, global
87 ET datasets based on the SEB model of Senay et al. (2013, 2020) and the Penman-Monteith-
88 Leuning (PML) model of Yongqiang Zhang et al. (2016, 2019) have become available. Efforts
89 are also underway to develop a global ET product based on the Atmosphere-Land Exchange
90 Inverse (ALEXI) model (another SEB-based approach; Anderson et al., 2011; Hain and
91 Anderson, 2017; Holmes et al., 2018).

92 Measurements of ET from eddy covariance flux towers have been used as the standard reference
93 data against which remote sensing-based ET products are evaluated (Miralles et al., 2011; Mu et
94 al., 2011; Holmes et al., 2018; Yongqiang Zhang et al., 2019; Senay et al., 2020), despite the
95 problems related to the lack of energy balance closure observed at eddy covariance sites and the
96 scale difference between the flux footprint and the model pixels (see below). Broadly speaking,
97 two different types of evaluation studies can be distinguished: 1) those that evaluate the
98 published ET datasets (hereafter referred to as product evaluation studies); and 2) those that
99 evaluate the performance of the underlying models (model evaluation studies). In the latter type
100 of study, all models are run with the same input data to isolate the effect of different modeling
101 approaches from differences in forcing data (Vinukollu et al., 2011a, b; McCabe et al., 2016;
102 Michel et al., 2016; Melo et al., 2021). Because remote sensing ET models are sensitive to
103 changes in input data (Vinukollu et al., 2011b; Badgley et al., 2015), the results of model
104 evaluation studies do not necessarily apply to the actual products.

105 The performance of remote sensing-based ET products at low latitudes is largely unknown
106 because most evaluation studies have focused on the northern middle latitudes (30° N–60° N;
107 Miralles et al., 2011; Mu et al., 2011; Kim et al., 2012; Hu et al., 2015; Velpuri et al., 2013; Tang
108 et al., 2015; Reitz et al., 2017; Holmes et al., 2018; Khan et al., 2018). The few studies that
109 evaluated ET products at low latitudes did this at a small number (two to five) of eddy

110 covariance sites (Ruhoff et al., 2013; Ramoelo et al., 2014; Aguilar et al., 2018; Souza et al.,
111 2019). The bias toward the northern middle latitudes can be explained by geographic differences
112 in the availability of eddy covariance data (Schimel et al., 2015; Villareal and Vargas, 2021).
113 Because of the lack of evaluation results from the low latitudes, it is unknown whether global
114 remote sensing-based ET products perform equally well at all latitudes. One can think of several
115 reasons why this might not be the case. For example, the MOD16 ET algorithm was calibrated
116 using eddy covariance data from sites located primarily in the US and Canada (Mu et al., 2011).
117 Hence, it is possible that the model is less accurate in other regions of the world, including the
118 low latitudes (Kun Zhang et al., 2019). Similarly, GLEAM uses constant values for the Priestley-
119 Taylor coefficient (α ; Miralles et al., 2011), while α varies with climate (Shuttleworth, 1993) and
120 forest type (Komatsu, 2005). Because the distribution of climate and forest types is related to
121 latitude, the use of constant values for α may result in (apparent) latitude-dependent biases in ET.
122 Latitudinal differences in product performance can also be caused by regional differences in
123 input data quality (Vinukollu et al., 2011b) or cloud cover (Running et al., 2019).
124 While eddy covariance observations of ET are probably the best option to evaluate remote
125 sensing datasets, there are two problems to consider: 1) the energy balance observed at eddy
126 covariance sites is usually not closed; and 2) the footprint of the eddy covariance observations
127 and the pixels of the ET products have different spatial scales. The degree of energy balance
128 closure is quantified by the energy balance ratio (EBR), which is the ratio of turbulent energy
129 fluxes ($H + LE$) to available energy, A (Wilson et al., 2002). Available energy is the difference
130 between net radiation (R_n) and changes in energy storage. The average EBR observed at eddy
131 covariance sites is about 0.8 (Wilson et al., 2002; Stoy et al., 2013). While the cause of the
132 energy imbalance is still being investigated, there are several plausible explanations, including
133 the systematic underestimation of the eddy covariance fluxes (Frank et al., 2016; Gao et al.,

134 2017; Mauder et al., 2020). As a practical solution to the closure problem, the energy surplus is
135 added to H and LE. Because it is unknown in what proportion the energy should be divided
136 between the fluxes (Mauder et al., 2020), the surplus is usually distributed in proportion to the
137 magnitude of H and LE, which preserves the Bowen ratio (Twine et al., 2000). Comparisons
138 with independent estimates of ET have yielded contrasting results, with some studies finding
139 better agreement for energy balance closure-corrected ET (Barr et al., 2012; Mauder et al., 2018)
140 and others for unadjusted ET (Denager et al., 2020). Although the energy balance closure
141 problem has been recognized for many years (Wilson et al., 2002; Foken et al., 2011), its effect
142 on the evaluation results of remote sensing-based ET products or models has rarely been
143 examined (Michel et al., 2016; Melton et al., 2021).

144 The evaluation results can also be affected by the scale difference between the footprint of the
145 eddy covariance observations and the pixels of the ET products. The flux footprint is typically
146 smaller than 1 km² (Chu et al., 2021), while the pixel sizes of ET products are as small as 0.25
147 km² (MOD16) and as large as 750 km² (GLEAM). The scale difference can result in a mismatch
148 in vegetation between pixel and site (Hobeichi et al., 2018; Jiménez et al., 2018). Such a
149 mismatch may also result from errors in the vegetation input data used by the models (due to, for
150 example, incorrect classification). Because most models calculate ET using land cover-specific
151 parameters (Anderson et al., 2007; Miralles et al., 2011; Mu et al., 2011), a mismatch between
152 the actual vegetation of the observation site and that detected in the model pixel could potentially
153 affect the evaluation results (Hu et al., 2015). However, the few studies that have examined this
154 issue found no clear effect (Hobeichi et al., 2018; Jiménez et al., 2018).

155 The objectives of this study were to: 1) evaluate the performance of the MOD16 and GLEAM
156 global ET products as well as of ET based on the ALEXI model at 40 eddy covariance sites in
157 the low latitudes; 2) examine the effect of the energy balance closure problem on product

158 evaluation results; 3) examine the dependence of product evaluation results on the vegetation-
159 match between pixel and site; and 4) investigate potential latitudinal dependence of product
160 performance. The MOD16 and GLEAM products were chosen because they are the longest
161 regularly produced remote sensing-based ET datasets. From MOD16, both the discontinued
162 collection 5 (C5) and the latest collection (C6) were evaluated (Mu et al., 2011; Running et al.,
163 2019). In the case of GLEAM, the v3.3a dataset was evaluated (Martens et al., 2017). While
164 most applications of ALEXI have focused on the continental US, recent efforts have paved the
165 way for routine global implementation of ALEXI (Hain and Anderson, 2017). The reference
166 dataset compiled in this study provides an excellent opportunity to evaluate the performance of
167 ALEXI at low latitudes. The products were evaluated using a reference dataset of eddy
168 covariance observations, including data from the emerging flux network in Mexico (MexFlux;
169 Vargas et al., 2013; Delgado-Balbuena et al., 2018) and from openly available databases of
170 FLUXNET (Pastorello et al., 2020), AsiaFlux, and OzFlux (Beringer et al., 2016).

171

172 2. Methods

173 2.1. Data

174 The remote sensing-based ET products evaluated in this study have different spatial and
175 temporal resolutions (Table 1). The comparisons with the eddy covariance ET observations were
176 made at the original spatial resolution of each product, except in the case of MOD16 C6 for
177 which the 500-m data were resampled to a 1-km resolution to match MOD16 C5. Using the
178 original spatial resolution is the common practice when evaluating these products against eddy
179 covariance data (see references in Table S4). An exception was made for MOD16 C6 to allow
180 for a more direct comparison with the previous C5 version. The effect of the scale mismatch
181 between product pixel and flux footprint on the evaluation results was examined using the

182 vegetation match index (Section 2.3). For each product, ET data were obtained from the pixels
183 matching the location of the flux towers (Velpuri et al., 2013; Hu et al., 2015). To evaluate all
184 products at the same temporal resolution (some performance statistics depend on the temporal
185 resolution of the data), the daily GLEAM and ALEXI data were averaged over the 8-day MODIS
186 interval. This was the highest common temporal resolution possible among the evaluated
187 datasets. Likewise, the eddy covariance data were averaged to yield mean daily ET for each
188 MODIS interval (Section 2.1.4).

189 The remote sensing ET products were evaluated by grouping the data by land cover type and
190 climate zone (Section 2.2). The eddy covariance data from the various sites were collected
191 during different periods between 2000 and 2019, with the length of the data records ranging from
192 1 to 11 years (Table 2). Hence, the flux datasets for a given land cover type or climate zone may
193 not coincide in time. In addition, data availability varied among the evaluated products. MOD16
194 C5 was discontinued in 2015 and GLEAM data for 2019 were not available at the time of
195 download (Table 1). For GLEAM and ALEXI, seven and four sites, respectively, were omitted
196 from the analysis because the fraction of open water in the corresponding pixels was too high
197 due to proximity to the coast (Sections 2.1.2 and 2.1.3). This problem did not affect MOD16
198 because of the smaller pixel size. As a result, the amount of data available for each of the
199 comparisons by land cover type and climate zone often varied from product to product (Table 3).
200 Ideally, one would compare the products using a common reference dataset (i.e., same sites and
201 same MODIS intervals). However, this would reduce the amount of available data by about one-
202 third (12 fewer sites and about 36% fewer MODIS intervals). Therefore, it was decided to
203 perform the regression analysis of observations versus product estimates (Section 2.2.1) and the
204 comparison of the performance statistics by land cover type and climate zone (Sections 2.2.2 and
205 2.2.3) using the complete dataset. The extent to which the two approaches (all data or a common

206 reference dataset) may have influenced the results was examined through a sensitivity analysis
207 (Section 2.2.4). The seasonal trend analysis (Section 2.2.5) was performed using the common
208 reference dataset.

209 The MOD16 and GLEAM ET data were extracted from the published global ET datasets.
210 Because detailed information about the models and datasets used to generate these products can
211 be found in the references listed in Table 1, only a brief explanation is provided below. The
212 ALEXI ET data were calculated specifically for this study. The methodology is described in
213 Anderson et al. (2011) and Hain and Anderson (2017). For completeness, the main features of
214 the model and the specific input datasets used are briefly described below.

215

216 <Table 1>

217

218 2.1.1. MODIS ET data

219 The MOD16 ET product is derived using a three-source Penman-Monteith model, which
220 estimates ET as the sum of evaporation from the dry canopy (transpiration), wet canopy
221 (interception loss), and soil (Mu et al., 2007, 2011). Separate calculations are performed for the
222 day and night. The model uses MODIS retrievals of: albedo (for the calculation of R_n); fraction
223 of absorbed photosynthetically active radiation, FPAR (to partition R_n between canopy and soil);
224 land cover type (to assign the physiological parameters needed to calculate the leaf stomatal and
225 aerodynamic resistances); and leaf area index (to calculate the bulk canopy resistances). The land
226 cover-specific parameters in the MOD16 algorithm were obtained by comparison with eddy
227 covariance flux data from 46 sites (located primarily in the US and Canada). MOD16 C5 used
228 C4 MOD12Q1 Land Cover Type 2 data, while MOD16 C6 uses the MCDLCHKM product
229 (Running et al., 2019). The meteorological data (incoming shortwave radiation and air

230 temperature and humidity) are obtained from reanalysis products (Modern-Era Retrospective
231 analysis for Research and Applications version 2 (MERRA-2) for C5 and Goddard Earth
232 Observing System Model Version 5 (GEOS-5) for C6). Transpiration and soil evaporation are
233 constrained by vapor pressure deficit, VPD (i.e., there is no soil moisture control). The C5 data
234 were downloaded from the University of Montana’s Numerical Terradynamic Simulation Group
235 (NTSG) website (<https://www.ntsg.umt.edu/project/modis/mod16.php>). The C6 data were
236 obtained from NASA’s Land Processes Distributed Active Archive Center (LP DAAC)
237 (<https://lpdaac.usgs.gov/>). The gap-filled version of the C6 dataset was used. The gap-filling
238 method is the same as that used for MOD16 C5 (Running et al., 2019). Apart from the difference
239 in resolution, the C5 and C6 datasets were produced using different reanalysis datasets and
240 different MODIS vegetation and albedo products.

241

242 2.1.2. GLEAM ET data

243 In GLEAM, ET is defined as the sum of the following processes: transpiration from short and
244 tall vegetation, bare soil evaporation, rainfall interception loss from tall vegetation, open water
245 evaporation, and snow sublimation (Miralles et al., 2011; Martens et al., 2017). The rainfall
246 interception loss module is based on the Gash (1979) analytical rainfall interception model
247 (Miralles et al., 2010). GLEAM v3.3a used the MEaSUREs VCF5KYRv001 product (Hansen et
248 al., 2018) to determine the fractions of bare soil, short vegetation, and tall vegetation. The model
249 first calculates potential ET with the Priestley-Taylor equation using R_n and air temperature from
250 reanalysis data (ERA-Interim). For bare soil and short vegetation, the typical value of 1.26 is
251 used for the Priestley-Taylor α coefficient, while for tall vegetation $\alpha = 0.97$ (Martens et al.,
252 2017). Actual ET is calculated by multiplying potential evaporation with land cover-dependent
253 stress functions. The stress functions simulate soil water constraints on transpiration and soil

254 evaporation. Soil water content is estimated using a multilayer running water balance model that
255 uses a merged precipitation product, ET from the previous time step, and microwave surface soil
256 moisture as the main inputs. The soil is divided in three layers: shallow (0–10 cm); intermediate
257 (10–100 cm); and deep (100–250 cm). Tall vegetation can extract water from all three layers,
258 short vegetation can extract water from the shallow and intermediate layers, and for bare soil
259 evaporation only water from the shallow layer is available. The stress functions for vegetation
260 also simulate the effect of phenology using microwave vegetation optical depth. The data were
261 accessed through the GLEAM website (<https://www.gleam.eu>). GLEAM pixels containing more
262 than 20% open water were excluded (this concerned a total of seven sites; Table S2). The open
263 water fraction (OWF) was obtained from the MOD44B product (Section 2.3). This filtering was
264 performed only for GLEAM. In the case of ALEXI, sites affected by the presence of open water
265 were filtered out during production of the dataset (Section 2.1.3), while in the case of MOD16,
266 no sites were affected because of the smaller pixel size.

267

268 2.1.3. ALEXI ET data

269 The ALEXI algorithm consists of a two-source SEB model coupled with an atmospheric
270 boundary layer model (Anderson et al., 1997, 2007). The latent heat flux is calculated separately
271 for the canopy and soil. An initial estimate of the canopy LE is obtained using the Priestley-
272 Taylor equation with $\alpha = 1.26$ (assuming potential transpiration). Next, the soil LE is calculated
273 as the residual of the energy balance. If the resulting soil LE is negative, the actual canopy LE
274 must be less than the potential value (which may indicate an effect of soil water limitation on
275 transpiration). The α coefficient is then reduced until the residual soil LE is non-negative. The
276 calculated LE represents the instantaneous flux at approximately one hour before local noon.
277 This time corresponds to the end of the time span over which H is calculated (see below). The

278 instantaneous latent heat fluxes are extrapolated to daily ET values by multiplying by the ratio of
279 daily total to instantaneous shortwave radiation and dividing by the latent heat of vaporization.
280 The ALEXI algorithm calculates H from the morning rise in the radiometric surface temperature
281 (Hain and Anderson, 2017). By using the temporal change in surface temperature, the effect of
282 bias in the temperature retrievals on H is minimized. This ALEXI implementation uses the
283 MODIS land surface temperature product (MYD11C1), retrieved using a generalized split-
284 window atmospheric compensation technique (Wan, 2004). The composite values of surface
285 temperature are partitioned between canopy and soil using estimates of vegetation cover fraction
286 from leaf area index. The leaf area index data were obtained from the 8-day MODIS MOD15A3
287 product (Myneni et al., 2002). Instead of using absolute values of air temperature, ALEXI uses
288 the slope of the vertical temperature profile (lapse rate) in the boundary layer. The lapse rate
289 profile, as well as the surface longwave radiation flux and wind speed were obtained from the
290 NCEP Climate Forecast System Reanalysis product (CFS-R, CFSRv2; Saha et al., 2010).
291 Incoming shortwave radiation fluxes were obtained from the CERES SYN1deg product
292 (Doelling, 2012). Soil heat flux is calculated as a diurnal varying function of net radiation
293 (Santanello and Friedl, 2003). The ALEXI model uses land cover data to assign canopy
294 parameters such as canopy height (to calculate the aerodynamic resistances to H) and leaf
295 absorptivity (to estimate R_n for the canopy and soil). The land cover data were obtained from the
296 MODIS MCD12C1 product (Land Cover Type 2). Since the thermal infrared based surface
297 temperature observations are only available during clear sky conditions, ALEXI employs a gap-
298 filling technique to generate estimates of weekly totals. The clear-sky fraction of actual ET to
299 incoming radiation is interpolated to a daily record and then multiplied by the daily incoming
300 radiation to generate a complete record. Along the coast the coarse-scale meteorological inputs
301 result in limited retrievals; this is why four coastal sites (Table S2) are not included in the

302 ALEXI dataset.

303

304 2.1.4. Eddy covariance ET data

305 Data from four different flux networks (MexFlux, FLUXNET, AsiaFlux, OzFlux) were used to
306 evaluate the ET products. The data from MexFlux were obtained directly from the site PIs (12
307 sites) because they were not available through a repository. The data from the other networks
308 were obtained through the respective web-based portals. FLUXNET data available under the
309 open data policy (tier 1) of the FLUXNET2015 dataset were used (Pastorello et al., 2020). This
310 dataset includes a total of 28 sites between latitudes 30° S and 30 °N. From OzFlux and
311 AsiaFlux, openly available data from sites not included in FLUXNET2015 were considered
312 (three and nine sites, respectively). Prior to the more extensive data quality control (see below),
313 sites for which the data record was too short (< 1 year), latent heat flux data were not available,
314 or the degree of energy balance closure was too low ($EBR < 0.5$) were excluded (one site from
315 MexFlux, four sites from FLUXNET, and seven sites from AsiaFlux). This left a total of 40 sites
316 for further analysis (Figure 1, Table 2). Information needed for the correction of the soil heat flux
317 (G) data or for the calculation of the sensible and latent heat storage terms, S (see below) was
318 obtained from the metadata accompanying the datasets, from articles or other publications, or
319 directly from the site PIs.

320 The remote sensing ET products were evaluated using the mean daily eddy covariance ET (mm
321 day^{-1}) calculated for each MODIS interval. The comparisons were made using the unadjusted
322 eddy covariance fluxes (ET_{orig}) and those corrected for the lack of energy balance closure
323 (ET_{ebc}). FLUXNET2015 includes corrected fluxes (Pastorello et al., 2020) but the datasets from
324 the other networks do not. For consistency, the fluxes were corrected using the same method for
325 all datasets (including FLUXNET2015). After filling the missing half-hourly or hourly values

326 (see below), a correction factor was calculated for each MODIS interval as $A/(H + LE)$, where
327 each term is the average daytime flux in $W m^{-2}$ (see above for definition of terms). Daytime was
328 defined as having solar radiation $> 10 W m^{-2}$. This method is based on the assumption that H and
329 LE were underestimated by the same percentage (Twine et al., 2000). The available energy was
330 calculated as $R_n - G - S$. The correction was only applied to the daytime data because, in
331 absolute terms, the missing energy is small during the night (Stoy et al., 2013; Mauder et al.,
332 2020) so that the correction will have little effect on total daily ET. In addition, this eliminated
333 the need to ensure the completeness and consistency of the energy balance data for the nighttime
334 period. The daytime and nighttime LE as well as the other energy balance terms (only daytime
335 data) were converted from energy units ($W m^{-2}$) to millimetres (mm) using a constant value for
336 the latent heat of vaporization ($2.45 MJ kg^{-1}$). The unadjusted nighttime fluxes were added to
337 daytime ET_{orig} and ET_{ebc} to give daily ET_{orig} and ET_{ebc} .

338 Data on G were available for 24 of the 40 sites. At all these sites G was measured using the soil
339 heat flux plate method (Sauer, 2002). For six sites, the measurements were not corrected for heat
340 storage above the plates (Mayocchi and Bristow, 1995). This correction was applied
341 retrospectively using the method of Wang and Bou-Zeid (2012). This method calculates G at the
342 soil surface (which is required in the energy balance calculations) from the time series of G at
343 any depth. It requires the thermal diffusivity of the soil, which was calculated as the ratio of soil
344 thermal conductivity to soil volumetric heat capacity. The thermal conductivity was calculated
345 following Lu et al. (2014) using site-specific soil physical data. The volumetric heat capacity was
346 calculated from soil bulk density and soil moisture. For sites without data on G but with data on
347 soil temperature (seven sites), G was estimated using the method of Hsieh et al. (2009).
348 Estimates of G derived from temperature measurements at depths > 2 cm were corrected for heat
349 storage using the method of Wang and Bou-Zeid (2012). For the remaining nine sites, G was

350 estimated using the method of Mu et al. (2011), using in situ air temperature and R_n , and
351 vegetation cover estimated from the MODIS FPAR product (MCD15A2H; see Section 2.1.5 for
352 more details about this dataset).

353 The sensible and latent heat storage terms are generally not included in the flux datasets (Stoy et
354 al., 2013; Pastorello et al., 2020). In this study, S was estimated from the half-hourly changes in
355 air temperature and humidity measured at the reference level (Brutsaert, 1982). This estimate did
356 not include heat storage in the vegetation biomass.

357 The data from the 40 sites were carefully screened for inconsistencies. These checks were in
358 addition to those performed by the site PIs/teams and by some of the networks (FLUXNET,
359 Pastorello et al., 2014, 2020; OzFlux, Isaac et al., 2017). For the daytime period, all data needed
360 for the energy balance calculations were checked. For the nighttime period, only the LE data
361 were screened. The quality checks were similar to those performed by Pastorello et al. (2014) for
362 FLUXNET2015. Where possible, errors in the radiation, air temperature, and relative humidity
363 data were corrected with the help of the site PIs, using calculated clear-sky radiation (in the case
364 of the radiation data), or using data from another sensor or from a nearby station (Allen, 2008;
365 Pastorello et al., 2014). No attempts were made to correct questionable eddy covariance flux data
366 (H , LE) or soil data (G , temperature, moisture).

367 Gap-filling of the data was carried out in two steps. In the first step, gaps in the half-hourly or
368 hourly data were filled on a daily basis. For the daytime period, missing values of any energy
369 balance term, x (where $x = LE, H, G, S$) on a particular day were filled using $\overline{(x/R_n)}R_{n,i}$, where
370 $\overline{(x/R_n)}$ is the average daytime ratio of x to R_n and $R_{n,i}$ the net radiation during time step i with
371 missing data. For any x , the maximum allowed percentage of missing values was 30%. For the
372 nighttime period, missing values of LE on a particular day were replaced by the mean nighttime
373 LE for that day (also using an upper threshold of 30% for the percentage of missing data). For

374 consistency, the same method was used for all datasets (i.e., the gap-filled data in
375 FLUXNET2015 were not used). In the second step, missing daily values of ET_{orig} and ET_{ebc} were
376 replaced by the mean of the available observations for individual MODIS intervals. The
377 maximum allowed percentage of missing values was 25% (i.e., two days for an 8-day MODIS
378 interval) (Hu et al., 2015).

379 Energy balance closure was analyzed for each site individually by summing the 8-day mean
380 daytime totals of $H + ET_{orig}$ and A and calculating the energy balance ratio as: $EBR = \sum(H +$
381 $ET_{orig})/\sum A$, with all terms in mm (Wilson et al., 2002). Energy balance closure was also analyzed
382 by grouping the data according to land cover type and climate zone (Section 2.2). For the pooled
383 data in each group, the 8-day mean daytime totals of $H + ET_{orig}$ were plotted against A and the
384 corresponding linear regression line and EBR were calculated.

385 Geographic coordinates and land cover type data for each site were obtained from the metadata
386 accompanying the datasets or from the literature (Table 2). The classification scheme of the
387 International Geosphere-Biosphere Programme (IGBP) was followed. This classification system
388 is adopted by most flux networks. It is also used in most evaluation studies of remote sensing ET
389 models (see, e.g., Velpuri et al., 2013; McCabe et al., 2016; Michel et al., 2016).

390 For each site, the evaporative fraction (EF) was calculated as (Shuttleworth et al., 1989): $EF =$
391 $\sum ET_{orig}/\sum(H + ET_{orig})$, where ET_{orig} and H are the mean daytime latent and sensible heat fluxes
392 for each MODIS interval, not corrected for the lack of energy balance closure. The obtained
393 values are listed in Table 2.

394

395 <Figure 1>

396

397 <Table 2>

398

399 2.1.5. Other datasets

400 The Köppen-Geiger climate class of each site was obtained using the 1-km resolution global map
401 of Beck et al. (2018). The map was downloaded from www.gloh2o.org/koppen/. Each site was
402 assigned the climate class of the pixel where the flux tower was located. The 40 sites represented
403 a total of 10 different climate classes (Table 2). For the evaluation of the remote sensing ET
404 products, these were grouped into four main climate zones (Section 2.2). For each of these
405 climate zones, the average EF was calculated using the site-specific values listed in Table 2.
406 To investigate the match between the actual vegetation type at the flux tower site and the
407 vegetation class or category used in the remote sensing ET models (Section 2.3), the yearly
408 MODIS land cover (MCD12Q1; 500 m resolution) and vegetation cover (MOD44B; 250 m
409 resolution) products were used. The data were downloaded from the NASA LP DAAC website.
410 From MCD12Q1, the Land Cover Type 2 data were used. From MOD44B, the data layers
411 containing percent tree cover and percent nontree vegetation were used. For each site, the
412 following three subsets were generated for the years with eddy covariance data: Subset 1) four
413 pixels of MCD12Q1 data corresponding to the 1-km MOD16 pixel; Subset 2) all pixels of
414 MOD44B data falling within the 0.25° GLEAM pixel; and Subset 3) all pixels of MCD12Q1
415 data corresponding to the 5-km ALEXI pixel. These subsets were used in the analysis described
416 in Section 2.3.

417 Finally, FPAR data from the MCD15A2H product were used to calculate G with the method of
418 Mu et al. (2011) (Section 2.1.4). This product is an 8-day composite dataset with a spatial
419 resolution of 500 m. The data were again obtained from NASA's LP DAAC. The pixels
420 matching the location of the flux towers were used. Data with a cloud flag or retrieved by the
421 backup algorithm were replaced by interpolated values (Zhao et al., 2005).

422

423 2.2. Evaluation of product performance

424 The remote sensing ET products were evaluated by grouping the data by IGBP land cover type
425 and Köppen-Geiger climate zone (Velpuri et al., 2013; McCabe et al., 2016). To avoid groups
426 with only one site, the woody savanna site was included in the group with the savanna sites and
427 the closed shrubland site was left out of the evaluations by land cover type (but included in the
428 evaluations by climate zone). This resulted in the following five groups of vegetation cover
429 types: evergreen broadleaf forest (EBF); deciduous broadleaf forest (DBF); evergreen needleleaf
430 forest (ENF); savanna (SAV); and grassland (GRA).
431 Likewise, the sites were grouped into the following four main climate zones: i) Af, Am: tropical
432 fully humid and tropical monsoon, respectively (from now on referred to as tropical wet); ii) Aw:
433 tropical savanna; iii) B: dry; and iv) C: mild temperate. Sites assigned the mild temperate (C)
434 climate were either located on tropical or subtropical mountains (five sites) or in lowland areas in
435 the subtropics (three sites) (see also Richter, 2016). Table 3 shows the number of sites and the
436 number of site years available in the complete dataset for each product by land cover type and
437 climate zone.

438

439 <Table 3>

440

441 2.2.1. Scatter plots and regression analysis

442 Scatter plots allow visual evaluation of the match between the remote sensing-based and the
443 observed ET data (Velpuri et al., 2013; McCabe et al., 2016; see also Chang and Hanna, 2004).
444 In addition, the slope, intercept, and coefficient of determination (R^2) of the fitted linear
445 regression line provide a quantitative way to evaluate product performance (Willmott, 1982;

446 Velpuri et al., 2013; McCabe et al., 2016). Most studies evaluating remote sensing ET products
447 perform the regression analysis with the product estimates on the y -axis and the observations on
448 the x -axis (see, e.g., Mu et al., 2011; Velpuri et al., 2013; McCabe et al., 2016). However,
449 Piñeiro et al. (2008) showed that this can lead to erroneous estimates of the regression
450 coefficients. Therefore, in this study the observations were used as the y variable and the product
451 estimates as the x variable (Piñeiro et al., 2008). For each land cover type and climate zone in
452 Table 3, the eddy covariance observations were plotted against the ET estimates of each product
453 and the corresponding linear regression lines were calculated, using the pooled data from the
454 different sites in each group. This analysis was performed using both ET_{orig} and ET_{ebc} .

455

456 2.2.2. Statistical performance metrics

457 In addition to visual inspection of the scatter plots and examination of the regression results,
458 three commonly used statistics in evaluation studies of remote sensing ET products were
459 calculated: root mean square error (RMSE), percent bias (PBIAS), and the coefficient of
460 determination (R^2) (see references in Table S4). The use of these common statistics allowed for
461 comparison with evaluation results from other latitudes (Section 2.4). The selected metrics
462 provide complementary information about product performance. The RMSE is a measure of total
463 error (i.e., both random and systematic errors) and is defined by:

464

$$465 \text{RMSE} = \sqrt{N^{-1} \sum [\text{ET}(\text{Prod}) - \text{ET}(\text{Obs})]^2} \quad (1)$$

466

467 where $\text{ET}(\text{Prod})$ is the product ET, $\text{ET}(\text{Obs})$ the eddy covariance ET, and N the total number of
468 data points (i.e., the number of MODIS intervals).

469 The PBIAS is the systematic (bias) error in percent of the average of the observations:

470

$$471 \quad \text{PBIAS} = \frac{N^{-1} \sum [\text{ET}(\text{Prod}) - \text{ET}(\text{Obs})]}{N^{-1} \sum \text{ET}(\text{Obs})} \times 100 \quad (2)$$

472

473 Third, in addition to the R^2 calculated from the pooled data (Section 2.2.1), the R^2 of the linear
474 regression between product ET and observed ET was calculated for each site separately. Besides
475 being a measure of correlation, R^2 indicates how much of the variation in observed ET is
476 explained by the product ET.

477 The three metrics were calculated using both ET_{orig} and ET_{ebc} . Averages of both sets of RMSE,
478 PBIAS, and R^2 values were calculated for each land cover type and climate zone in Table 3. The
479 average metrics by land cover type and climate zone were displayed graphically in plots for each
480 product (McCabe et al., 2016) for ET_{orig} and ET_{ebc} . The results for the individual sites can be
481 found in Tables S2 and S3.

482

483 2.2.3. Combining the different performance metrics into a single score

484 To facilitate comparison of the overall performance of the different ET products, the individual
485 metrics (R^2 , RMSE, PBIAS) were combined into the Ideal Point Error (IPE) score (Elshorbagy et
486 al., 2010; Dawson et al., 2012). The IPE score takes values between 0 and 1, with 0 indicating
487 perfect performance (i.e., all metrics are at their optimum values) and 1 being assigned to the
488 worst performing product. In practice, no product (and no observation) is without error.
489 Therefore, the best performing product will usually have an IPE greater than 0. The IPE values
490 were calculated for each of the comparisons by land cover type and climate zone. The calculation
491 of IPE consists of two steps. In the first step, each performance metric is standardized to the
492 worst score for that metric. Dawson et al. (2012) provides expressions for this standardization
493 step for different categories of performance measures (denoted by S1–S5; their Table 1). PBIAS

494 is not listed in this table. However, as mentioned by Dawson et al. (2012), the flexibility of this
495 method allows other metrics to be included. PBIAS classifies as an S4 category metric and was
496 standardized using the corresponding expression. In the second step, the IPE is calculated from
497 the standardized metrics using Equation (2) in Dawson et al. (2012). The results were plotted as
498 heatmaps for ET_{orig} and ET_{ebc} .

499

500 2.2.4. Sensitivity to the choice of reference dataset

501 The statistical metrics (R^2 , RMSE, PBIAS) and the IPE scores were calculated as explained
502 above but now using the common reference dataset. This direct comparison approach included
503 12 fewer sites and about 36% fewer MODIS intervals than when using all data (see Table S1 for
504 the number of sites and site years by land cover type and climate zone). The sensitivity analysis
505 was performed for ET_{orig} only. Differences in the ranking of products for each of the
506 comparisons by land cover type and climate zone were determined by comparing the IPE scores
507 from both approaches. Changes in ranking were indicated by adding an asterisk to the IPE scores
508 in the heatmap for ET_{orig} .

509

510 2.2.5. Evaluation of seasonal trends in ET from products

511 The ability of the products to capture seasonal changes in ET was examined by plotting the
512 average monthly ET for each product together with the average monthly ET_{orig} and ET_{ebc} . This
513 was again done for each land cover type and climate zone in Table 3. To account for the different
514 timing of the rainy seasons, separate plots were made for sites located in the northern and
515 southern hemispheres.

516

517 2.3. Vegetation match index (VMI) and open water fraction (OWF)

518 The effect of a mismatch between the vegetation at the flux tower site and that detected in the
519 model pixel on the product evaluation results was examined by calculating a vegetation match
520 index (VMI). The models underlying the investigated ET products differ in the level of detail
521 with which they distinguish between different vegetation types. Both MOD16 and ALEXI assign
522 land cover-specific parameters to a wide range of cover types, while GLEAM only considers two
523 vegetation categories (i.e., tall and short vegetation). However, also for MOD16 and ALEXI the
524 largest differences between the land cover-specific parameters occur between tall and short (or
525 forest and non-forest) vegetation types (Anderson et al., 2007; Mu et al., 2011). Therefore, for all
526 three products, VMI was calculated based on these two vegetation categories.

527 The datasets used to calculate the VMIs are described in Section 2.1.5. As explained in Section
528 1, a mismatch in vegetation can be caused by scale differences or inaccuracies in the vegetation
529 input data. To account for the latter, vegetation data were selected that were as similar as
530 possible to those used to generate the products (Sections 2.1.1–2.1.3). For MOD16 and ALEXI,
531 MCD12Q1 Land Cover Type 2 data were used (Subsets 1 and 3, respectively). The data from
532 Subsets 1 and 3 were aggregated into forest and non-forest categories. For sites with a forest land
533 cover (EBF, DBF, ENF; Table 2), VMI_{MOD16} or VMI_{ALEXI} was calculated as the proportion of
534 forest vegetation present in the 1-km MOD16 or 5-km ALEXI pixel. For sites with a non-forest
535 land cover (SAV, GRA), the VMIs were calculated as the proportion of non-forest vegetation.

536 For GLEAM, MOD44B vegetation cover data were used (Subset 2). These data were assumed to
537 be similar to those of the VCF5KYR product (used as input to GLEAM v3.3a; Section 2.1.2).

538 The VCF5KYR product is based on AVHRR observations calibrated with MODIS data (Hansen
539 et al., 2018). In each data layer of the MOD44B product, pixels with water are masked out with a
540 fill value of 200. Hence, VMI_{GLEAM} was calculated as either the average percent tree cover (for
541 sites with forest vegetation) or the average percent nontree vegetation (for sites with non-forest

542 vegetation) multiplied by the fraction of land pixels. In addition, the open water fraction (OWF)
543 was calculated. This index was used to filter out sites for which the pixel contained more than
544 20% water (Section 2.1.2).

545 The dependence of product performance on the vegetation-match between pixel and site was
546 examined by plotting the performance metrics (R^2 , RMSE, PBIAS) against VMI. Individual site
547 values for the metrics were bin-averaged into four evenly spaced intervals of 0.25 VMI units
548 wide in the case of GLEAM and ALEXI or for each of the five discrete VMI values in the case
549 of MOD16. For each metric-VMI combination, the linear regression line was calculated. In
550 addition to visual inspection of the scatter plots, the p -values of the calculated regression slopes
551 were used to evaluate whether there was a relationship between VMI and product performance.
552 For this analysis, performance statistics obtained for ET_{orig} were used.

553

554 2.4. Latitudinal comparison of product performance

555 To investigate latitudinal dependence of the performance of the ET products examined here, a
556 literature search was conducted to find studies that evaluated these products. To allow for direct
557 comparison, only studies that evaluated the products with eddy covariance-based ET were
558 considered. Furthermore, a study needed to report at least one of the three performance metrics
559 used in this study (R^2 , RMSE, PBIAS) or provide the data from which these metrics could be
560 calculated. The performance statistics depend on the averaging time used. Hence, ideally, only
561 studies using the same time average as used here (8-day) should be considered. This would,
562 however, drastically reduce the number of evaluation results available. Therefore, studies using
563 daily or monthly time averages were also included in the initial search.

564 In the end, a total of 15 studies were found (Table S4). As will be shown below, the evaluation
565 results were different for ET_{orig} and ET_{ebc} . Of the 15 studies found in the literature, 13 used ET_{orig}

566 and only two used ET_{ebc} . No studies were found that used both. For the final analysis, only
567 studies using ET_{orig} were considered. The study of Miralles et al. (2011) was also excluded
568 because: i) GLEAM ET was calculated using in situ measured R_n ; ii) comparisons were made
569 using modelled ET for the vegetation type (i.e., tall or short vegetation) matching that at the
570 tower site; and iii) days with rainfall were excluded. Likewise, the study of Mu et al. (2011) was
571 excluded because their evaluation results are in fact calibration results. This yielded a total of 12
572 studies, including the current one. Most studies used MOD16 C5 because MOD16 C6 was only
573 recently released. Therefore, the results obtained here for C5 were used. To account for
574 latitudinal differences in ET, RMSE was normalized by mean ET_{orig} (NRMSE). Not all studies
575 reported ET_{orig} (Table S4). The studies evaluated product performance at a minimum of two eddy
576 covariance sites to a maximum of 119 sites. Most studies reported evaluation results for
577 individual sites but some reported averages for land cover classes (e.g., Velpuri et al., 2013;
578 Reitz et al., 2017). The latter were treated as if they were results for individual sites. Performance
579 results were grouped into results for forest and non-forest vegetation; there were not enough
580 performance data available to create more specific subgroups. The results were further grouped
581 into three latitudinal bands: southern low latitudes ($30^\circ S-0^\circ$); northern low latitudes ($0^\circ-30^\circ N$);
582 and northern middle latitudes ($30^\circ N-60^\circ N$). For latitudes outside these regions, there were not
583 enough data available (Table S4). Table 4 summarizes the number of evaluation results (N_{ER})
584 available, broken down into results for R^2 , NRMSE, and PBIAS, and grouped by latitudinal
585 zone, product, and vegetation category. Averages of each performance metric for each product-
586 vegetation category combination were plotted as a function of latitude.

587

588 <Table 4>

589

590 3. Results

591 3.1. Energy balance closure of eddy covariance data

592 Table 2 shows the daytime energy balance ratio (EBR) for each of the 40 individual sites. The
593 average daytime EBR for the 40 sites was 0.83, with a standard deviation (SD) of 0.10, and with
594 values ranging from 0.63 to 1.03. Figures 2 and 3 show scatter plots between the sum of the
595 daytime turbulent heat fluxes and available energy grouped by land cover type and climate zone,
596 respectively. The daytime EBR values calculated from the pooled data were similar across land
597 cover types (ranging from 0.79 to 0.87) and climate zones (ranging from 0.77 to 0.85). The
598 slopes and intercepts of the regression lines ranged from 0.67 to 0.80 and -0.01 to 0.67 mm
599 day^{-1} , respectively, across land cover types and from 0.72 to 0.81 and 0.00 to 0.69 mm day^{-1} ,
600 respectively, across climate zones. The coefficient of determination (R^2) ranged from 0.64 and
601 0.82 across land cover types and from 0.59 and 0.82 across climate zones.

602

603 <Figure 2>

604

605 <Figure 3>

606

607 3.2. Evaluation of ET products by land cover type

608 Figures 4 and S1 show scatter plots comparing eddy covariance-based and remote sensing-based
609 ET by land cover type for each of the evaluated products for ET_{orig} and ET_{ebc} , respectively. First,
610 the results for ET_{orig} will be examined. Although the scatter plots and the regression results for
611 MOD16 C5 and MOD16 C6 show some differences, these were generally smaller than the
612 differences with the other products (see also below). Hence, from now on the two collections
613 will be referred to as MOD16. When necessary, a distinction will be made between the two.

614 Overall, GLEAM ET showed the best agreement with ET_{orig} . This follows from the results of the
615 regression analysis (i.e., slope closer to 1, intercept closer to 0, higher R^2) and can be observed
616 visually as a narrower distribution of data points around the 1:1 line. For DBF and SAV, the
617 correlations between GLEAM ET and ET_{orig} were strong (R^2 of 0.81 and 0.73, respectively). A
618 weak correlation was observed for EBF ($R^2 = 0.32$). The agreement with ET_{orig} was generally
619 poorer for MOD16 and ALEXI. Neither of these products consistently outperformed the other.
620 The scatter plots show a clear overestimation of ET_{orig} by MOD16 for ENF and a clear
621 underestimation for SAV. Although both products showed weaker correlations with ET_{orig} than
622 GLEAM, this was most pronounced for ALEXI. Also MOD16 and ALEXI had the strongest
623 correlations for DBF and SAV and the weakest for EBF. When evaluating the products with
624 ET_{ebc} , the regression slopes and intercepts increased, while changes in R^2 were generally small
625 (Figure S1).

626

627 <Figure 4>

628

629 Figure 5 shows the mean R^2 , RMSE, and PBIAS by land cover type for each of the evaluated
630 products for ET_{orig} and ET_{ebc} . Again, the results for ET_{orig} will be examined first. As already
631 observed in Figure 4, the mean performance statistics show that the differences between MOD16
632 C5 and MOD16 C6 are generally smaller than the differences with the other products. One
633 exception is ENF; this group, however, included a relatively small number of sites and site years
634 (Table 3), which may have affected the comparisons. Figure 5 confirms the superior performance
635 of GLEAM. Overall, GLEAM had the strongest correlations, the lowest RMSEs, and the
636 smallest PBIAS values. In agreement with the graphical analysis, neither MOD16 nor ALEXI
637 was second best over all land covers. Both GLEAM and MOD16 tended to overestimate ET_{orig}

638 for forest vegetation and underestimate ET_{orig} for non-forest vegetation; however, biases were
639 smaller for GLEAM. ALEXI tended to overestimate ET_{orig} for all land cover types. The variation
640 in PBIAS across land cover types was smaller for GLEAM and ALEXI than for MOD16. As
641 seen in the scatter plots, ALEXI had the weakest correlations with ET_{orig} . All ET products had
642 the strongest correlations for DBF and SAV and the weakest for EBF.

643

644 <Figure 5>

645

646 As expected, PBIAS shifted to more negative values when the products were evaluated with
647 ET_{ebc} (Figure 5). Depending on whether PBIAS decreased or increased, the corresponding
648 RMSE became smaller or larger (although not for all products; see below). The use of ET_{ebc}
649 generally had little effect on the correlations (as also seen in the scatter plots). For GLEAM,
650 PBIAS values were negative for all land cover types when using ET_{ebc} and were generally
651 greater in absolute terms than when using ET_{orig} . As a result, the RMSEs were larger (and closer
652 to those of the other products) when using ET_{ebc} than when using ET_{orig} . For MOD16, PBIAS
653 values were also negative for most land cover types when using ET_{ebc} . The corresponding
654 RMSEs were either somewhat larger (e.g., SAV, GRA) or smaller (e.g., EBF, ENF) than when
655 using ET_{orig} . In the case of ALEXI, PBIAS values decreased for all land cover types except SAV.
656 However, only in the case of EBF this was accompanied by a decrease in RMSE. For ENF, DBF
657 and GRA, the RMSE actually increased. A partial explanation for this is the tendency of ALEXI
658 to overestimate low ET_{ebc} and underestimate high ET_{ebc} (DBF, GRA; Figure S1).

659 Figure 6 shows the IPE scores for the different ET products by land cover type as obtained using
660 ET_{orig} or ET_{ebc} for evaluation. The IPE scores confirm that GLEAM ET best matched ET_{orig}
661 across all land cover types. The IPE scores also support the earlier observations that: i) the

662 differences in performance between the two MOD16 collections were generally smaller than the
663 differences with the other products; and ii) neither MOD16 nor ALEXI consistently
664 outperformed the other. Figure 6 shows that the IPE values of the products converged when
665 using ET_{ebc} for evaluation. This largely reflects the changes in PBIAS and RMSE mentioned
666 above. When using ET_{ebc} there is no product that stands out in terms of both small PBIAS and
667 high R^2 across all land cover types.

668

669 <Figure 6>

670

671 Figure 7 compares the seasonal trends in ET from the products with those from the observations
672 by land cover type by hemisphere. Note that these curves were calculated using the common
673 reference dataset. Only curves calculated with data from at least two sites are shown. Clear
674 differences in the seasonality and timing of rainfall can be observed. In both hemispheres, ET of
675 EBF was characterized by weak seasonality, with constant high values throughout the year. Yet,
676 MOD16 C5 seemed to capture the small variations in ET quite well. This was also the case for
677 GLEAM, except during the wet season in the southern hemisphere when it showed a strong
678 positive bias. A closer look at the data showed that this involved the two Brazilian rainforests
679 (Table 2; Figure 1). Similarly, ALEXI had a strong positive bias at the end of the dry season in
680 the southern hemisphere. This could be traced mainly to EBF in northeastern Australia (Table 2;
681 Figure 1). For ENF, all products seemed to represent the observed seasonal trend in ET fairly
682 well. For DBF, GLEAM closely followed the observed seasonal trend in ET. MOD16 had a
683 negative bias during the dry season. Conversely, ALEXI had a positive bias during the transition
684 from the wet to dry season. For SAV and GRA, both GLEAM and MOD16 had a strong negative
685 bias during the dry season. Conversely, ALEXI seemed to have a positive bias during the dry

686 period in these cover types.

687

688 <Figure 7>

689

690 3.3. Evaluation of ET products by climate zone

691 The performance of the ET products was also examined across four main climate zones (Table
692 3). For each climate zone, an average evaporative fraction (EF) was calculated from the site-
693 specific values in Table 2, yielding (ranked from wet to dry): 0.73 ± 0.04 (SD) for Af, Am
694 (tropical wet); 0.60 ± 0.10 for C (mild temperate); 0.50 ± 0.11 for Aw (tropical savanna); and
695 0.35 ± 0.11 for B (dry). The tropical wet climate zone included mainly EBF sites (seven in total;
696 Table 3). The mild temperate climate zone included all ENF sites and for the rest mainly EBF
697 sites. The savanna and dry climate zones included mostly SAV, DBF, and GRA sites. The results
698 of this analysis were presented in the same way as in the previous section, i.e., scatter plots
699 (Figures 8 and S2), average performance statistics (R^2 , RMSE, PBIAS; Figure 9), heatmaps of
700 IPE scores (Figure 6), and average seasonal trends in ET (Figure 10).

701 The comparisons by climate zone confirmed many of the findings in the previous section. Again,
702 the differences in performance between the two MOD16 ET collections were usually smaller
703 than the differences with the other products (Figures 8 and 9). Furthermore, the performance and
704 ranking of the products depended on whether ET_{orig} or ET_{ebc} was used for evaluation. When
705 using ET_{orig} , GLEAM again showed the strongest correlations and best agreement (i.e., closeness
706 to observations) (Figures 8 and 9). As a result, GLEAM had the smallest RMSEs and best IPE
707 scores across all climate zones (Figures 9 and 6, respectively). Again, neither MOD16 nor
708 ALEXI performed consistently better than the other. That is, MOD16 showed better agreement
709 with ET_{orig} for the Aw climate zone, whereas ALEXI gave better results for the B and C climate

710 zones (as summarized by the IPE scores in Figure 6). Finally, there was no clear ranking among
711 the products when ET_{ebc} was used for evaluation (Figure 6). This mainly reflected the
712 underestimation of GLEAM ET with respect to ET_{ebc} , leading to higher (i.e., more negative)
713 PBIAS values and larger RMSEs than when using ET_{orig} (Figure 9). For MOD16, PBIAS and
714 RMSE values both decreased (e.g., C climate zone) and increased (Aw climate zone). For
715 ALEXI, PBIAS decreased to values close to zero (C, Aw, B); however, instead of decreasing, the
716 corresponding RMSEs increased. The use of ET_{ebc} generally had little effect on the correlations
717 (Figures 8, S2, and 9).

718

719 <Figure 8>

720

721 <Figure 9>

722

723 All products had the weakest correlations in the wet tropical climate zone and the strongest in the
724 tropical savanna and dry climate zones (Figure 9). This is consistent with the results in Section
725 3.2 (weakest correlations for EBF and strongest correlations for SAV and DBF). Overall, ALEXI
726 had again the weakest correlations. The biases of MOD16 ET varied markedly across climate
727 zones (Figure 9). When compared with ET_{orig} , MOD16 tended to overestimate ET in the wet
728 tropical and mild temperate climate zones and underestimate ET in the dry climate zone. This
729 result is consistent with the positive biases observed in Figure 5 for forest vegetation (dominating
730 the wet tropical and mild temperate climate zones) and the negative biases for non-forest
731 vegetation (dominating the dry climate zone). Biases in GLEAM showed the same tendency but
732 were generally much smaller in size. For ALEXI, the bias with respect to ET_{orig} was practically
733 zero in the wet tropical climate zone but positive in the other climate zones.

734 The seasonal trend analysis (Figure 10) revealed the same patterns as found earlier in Section
735 3.2, again reflecting the close correspondence between climate and vegetation. For the tropical
736 wet climate zone, MOD16 C5 ET closely followed the observed seasonal changes in ET. This
737 was also the case for GLEAM, except for the positive bias during the wet season in the southern
738 hemisphere (traced back mainly to the Brazilian rainforests; Section 3.2). The positive bias of
739 ALEXI ET at the end of the dry season in the southern hemisphere can also be observed again
740 (traced back mainly to EBF in northeastern Australia; Section 3.2). In addition, ALEXI ET
741 showed large, seemingly erratic, variations in the northern hemisphere. For the mild temperate
742 climate zone, all products represented the observed seasonal trend in ET fairly well. For the
743 tropical savanna climate, both MOD16 and GLEAM had a strong negative bias during the dry
744 season, which was also observed in the plots for GRA and SAV in Figure 7. The positive bias for
745 ALEXI during the dry period can also be observed again. For the dry climate zone, GLEAM ET
746 closely followed the observed seasonal trend in ET. ALEXI had again a positive bias during the
747 dry period. MOD16 had a strong negative bias during the wet season in the southern hemisphere.
748

749 <Figure 10>

750

751 3.4. Sensitivity to the choice of reference dataset

752 The IPE scores based on the common reference dataset (Figure S3) show similar results to those
753 obtained using all data (Figure 6, ET_{orig}). For both the comparisons by land cover type and
754 climate zone, GLEAM ET generally showed the best agreement with ET_{orig} . Furthermore, the
755 differences among the two MOD16 collections and ALEXI were generally too small to identify a
756 second best performing product. Differences in ranking results between the two approaches were
757 observed only for MOD16 and ALEXI (Figure 6).

758

759 3.5. Product performance versus VMI

760 Figure 11 shows binned scatter plots between the performance metrics (R^2 , RMSE, PBIAS) and
761 the vegetation match index for the different products. In addition, the regression lines and the p -
762 values indicating the statistical significance of the regression slopes are shown. The VMIs for
763 each individual site are given in Table S2. The average VMI was 0.77 ± 0.41 for MOD16, $0.71 \pm$
764 0.40 for ALEXI, and 0.51 ± 0.23 for GLEAM. These results indicate a decreasing vegetation-
765 match between pixel and site with increasing pixel size, although the VMIs of GLEAM and the
766 other products cannot be directly compared as they are based on different data. For none of the
767 products there was an improvement in performance (i.e., increasing R^2 or decreasing RMSE or
768 PBIAS) with increasing VMI (Figure 11). Moreover, for none of the regressions the slope was
769 statistically significant.

770

771 <Figure 11>

772

773 3.6. Latitudinal comparison of product performance

774 Figure 12 shows zonal averages (southern and northern low latitudes and northern middle
775 latitudes) of the performance metrics grouped by forest and non-forest vegetation for MOD16
776 and GLEAM and ALEXI. The averages were calculated using evaluation results from this study
777 and from the literature (Tables S2 and S4). For ALEXI, no data on NRMSE and PBIAS were
778 available for the northern middle latitudes (Table 4). Figure 12 should be interpreted with
779 caution because the number of evaluation results (N_{ER}) available varied considerably among
780 latitudinal zones, products, and vegetation categories (Table 4). In the case of forest vegetation,
781 correlations (all products) seem to be weaker while PBIAS and NRMSE scores (MOD16 and

782 GLEAM) seem to be better at low latitudes than at northern middle latitudes. Both MOD16 and
783 GLEAM seem to overestimate ET_{orig} in all latitudinal zones. In contrast, in the case of non-forest
784 vegetation the performance metrics show no clear variation with latitude. Moreover, both
785 MOD16 and GLEAM seem to underestimate ET_{orig} in all latitudinal zones.

786

787 <Figure 12>

788

789 4. Discussion

790 4.1. The effect of the energy balance closure problem on product evaluation results

791 The average energy balance ratio for the 40 sites in this study (0.83) is nearly identical to that
792 reported by Stoy et al. (2013) for 173, mainly mid-latitude, FLUXNET sites (0.84). When
793 grouped by land cover type or climate zone, the ranges of EBR values were fairly small (0.79–
794 0.87 or 0.77–0.85, respectively). A greater range was observed for the 173 FLUXNET sites
795 grouped by land cover type (0.70–0.94; Table 2 in Stoy et al., 2013).

796 As mentioned in Section 1, the reasons for the energy balance closure problem and the extent to
797 which it affects the ET fluxes are not yet clear. Despite being long recognized, the effect of the
798 energy balance closure problem on the evaluation results of remote sensing-based ET products
799 has rarely been examined (Michel et al., 2016; Melton et al., 2021). This study found that the
800 performance and ranking of the evaluated products depended on whether the unadjusted or the
801 energy balance closure corrected ET fluxes were used. When using ET_{orig} , GLEAM showed the
802 best overall performance with the strongest correlations and smallest biases. However, when
803 using ET_{ebc} , none of the products was superior to the others. Not surprisingly, the use of ET_{ebc}
804 instead of ET_{orig} affected the product biases more than the correlations. Overall, MOD16 and
805 GLEAM underestimated ET compared to ET_{ebc} . For ALEXI, PBIAS decreased when using ET_{ebc}

806 instead of ET_{orig} , but the corresponding RMSEs tended to increase rather than decrease. For most
807 SEB approaches used in the OpenET project, cumulative totals of ET over the growing season or
808 water year also agreed better with ET_{ebc} than with ET_{orig} (Melton et al., 2021).

809 Both MOD16 and GLEAM include parameters that were calibrated using field observations of
810 ET. MOD16 was calibrated using ET obtained from eddy covariance-based estimates of water
811 use efficiency (WUE) and MODIS-based gross primary production (GPP), with WUE being
812 calculated as the ratio between GPP and ET fluxes not corrected for energy balance closure (Mu
813 et al., 2011). The use of ET_{orig} could possibly explain the negative bias of MOD16 with respect
814 to ET_{ebc} (Michel et al., 2016). However, the GPP fluxes may have been underestimated for the
815 same reason as ET_{orig} (Foken et al., 2011). In that case, the estimated WUE would not (or only
816 partly) be affected. In GLEAM, fixed values are used for the Priestley-Taylor coefficient for
817 short ($\alpha = 1.26$) and tall ($\alpha = 0.97$) vegetation (Martens et al., 2017). These values are averages
818 of α values published in the literature, which in turn were obtained by comparing field
819 measurements of ET under well-watered conditions with potential ET. Some of the α values
820 were derived with ET_{orig} , but others were obtained using ET estimates based on other methods,
821 such as the weighing lysimeter and bowen ratio energy balance techniques (see references cited
822 in Martens et al., 2017). Hence, also the negative bias error of GLEAM with respect to ET_{ebc}
823 cannot be directly linked to calibration with ET_{orig} . ALEXI ET had smaller PBIAS when using
824 ET_{ebc} than when using ET_{orig} . In contrast to the other models, ALEXI is not calibrated with field
825 data. However, no conclusions can be drawn from this observation without a better
826 understanding of the effect of the energy balance closure problem on ET_{orig} . In addition, the
827 RMSEs of ALEXI tended to increase rather than decrease when using ET_{ebc} .

828 The literature review showed that most studies evaluated the products using ET_{orig} (Table S4). It
829 is recommended to use both ET_{orig} and ET_{ebc} as long as the effect of the energy balance closure

830 problem on ET_{orig} is not clear. The remainder of the discussion will focus on the results obtained
831 with ET_{orig} to facilitate comparisons with the literature.

832

833 4.2. Relative performance of the evaluated products

834 Similar results were obtained when grouping the data by land cover type or climate zone,
835 showing the close relationship between the two (see, for example, Cui et al., 2021). Therefore,
836 the results of these two analyses will be discussed together and interchangeably. As explained in
837 Section 1, a distinction should be made between product and model evaluation studies. The
838 former evaluate the published ET products while the latter evaluate the performance of the
839 underlying models using a common input dataset. Because modeled ET is sensitive to the input
840 data, the results of the model evaluation studies do not necessarily apply to the final ET products
841 (see Section 1 for references).

842 Overall, GLEAM had the best performance across different land cover types and climate zones
843 in the low latitudes; neither MOD16 nor ALEXI could be identified as the second best
844 performing product. These results were obtained regardless of whether the comparisons were
845 made using all data or a common reference dataset. There are very few product evaluation
846 studies that have compared the performance of the products assessed in this study. Khan et al.
847 (2018, 2020) compared the performance of GLEAM and MOD16 at nine and five eddy
848 covariance sites, respectively; in both studies, about one third of sites were located in the low
849 latitudes and about two thirds in the middle latitudes (see also Table S4). Khan et al. (2018) used
850 ET_{orig} to evaluate the products, while Khan et al. (2020) used ET_{ebc} . Consistent with the results of
851 this study, Khan et al. (2018) found that GLEAM generally corresponded better with ET_{orig} than
852 MOD16. Also when using ET_{ebc} , Khan et al. (2020) found that GLEAM outperformed MOD16
853 (as opposed to this study where differences were small when using ET_{ebc}). In the absence of

854 other comparative studies, the compilation of product evaluation results for the latitudinal
855 analysis (Figure 12, Table S4) allows for an indirect comparison of product performance.
856 Focusing only on the results for the northern middle latitudes (to exclude the evaluation data
857 from this study), the overall better match of GLEAM with ET_{orig} as compared to MOD16 is also
858 evident from the studies included in this analysis (see Table S4 for references). The stronger
859 correlations of GLEAM as compared to ALEXI are also noticeable when comparing the R^2
860 values from studies that evaluated these products separately.

861 An overall better performance of GLEAM as compared to MOD16 was also observed in the
862 model evaluation studies by McCabe et al. (2016) and Michel et al. (2016). GLEAM also
863 performed better than the surface energy balance approach evaluated in these studies (i.e., the
864 SEBS model of Su, 2002). Similarly, Vinukollu et al. (2011a) obtained better performance
865 results for a Priestley-Taylor approach developed by NASA's Jet Propulsion Laboratory (PT-
866 JPL; Fisher et al., 2008) than for MOD16 and SEBS. However, all these studies concluded that
867 no single model was superior in all cases. The same conclusion was reached in a recent model
868 evaluation study for South America (Melo et al., 2021). Such a conclusion is not supported by
869 the results of this study as GLEAM performed better than MOD16 and ALEXI in all land cover
870 types and climate zones. It is not known whether this is a feature of the low latitudes (i.e., the
871 other studies focused mainly on the middle latitudes), whether it is related to the differences
872 between product and model evaluation studies discussed earlier, or whether it is a result specific
873 to the products studied here.

874 All products had the weakest correlations in the wet tropical climate zone (dominated by EBF)
875 and the strongest in the tropical savanna and dry climate zones (dominated by DBF and SAV).
876 For the most part this reflects differences in seasonality (i.e., the greater the variation in ET, the
877 stronger the correlations; Miralles et al., 2011; Yilmaz et al., 2014) rather than differences related

878 to the performance of the products. ALEXI generally had the weakest correlations of all
879 products. Comparing the R^2 values from the studies used for the latitudinal analysis shows values
880 for ALEXI between those of MOD16 and GLEAM (northern middle latitudes, Figure 12).
881 Although this is an indirect comparison (because it involves studies that evaluated the products
882 separately), it suggests that the low correlations observed in this study for ALEXI are not a
883 general feature of the product. A known challenge for thermal-based approaches is the filtering
884 of cloud-contaminated data and the resulting gaps between successful retrievals. Failure to detect
885 cloud-contaminated data can lead to large errors in ALEXI ET estimates (Anderson et al., 2007;
886 Yilmaz et al., 2014). The uncertainty in gap-filled ALEXI ET can be twice as large as that in ET
887 generated by the algorithm under clear-sky conditions (Anderson et al., 2007). These cloud-
888 related problems could be responsible for the weak correlations of ALEXI, but that still does not
889 explain the difference with the northern middle latitudes where the data are also affected by
890 clouds. A possible approach to solving these problems is to use cloud-tolerant microwave-based
891 land surface temperature in ALEXI (Holmes et al., 2018). Finally, Holmes et al. (2018) found
892 that averaging 0.05° ALEXI ET estimates to 0.25° spatial resolution generally improved
893 correlations with flux tower data. They attributed this to the presence of noise in the MODIS
894 land surface temperature data that outweighed the benefits of the higher resolution compared to a
895 0.25° average. However, the overall effect reported in that study is too small to explain the
896 differences in correlation with the other products observed here. More work is needed to
897 understand the reasons for the low correlations of ALEXI observed at the low-latitude sites
898 studied here.

899 Both MOD16 and GLEAM had a positive bias for forest vegetation and a negative bias for non-
900 forest vegetation. No such land cover type-dependent biases were observed for ALEXI. The
901 biases of MOD16 and GLEAM are also evident from the evaluation results of other studies

902 (northern middle latitudes, Figure 12). Both Kun Zhang et al. (2019) and Brust et al. (2021)
903 showed that the biases in MOD16 can be significantly reduced when calibrating the algorithm
904 with more and a greater diversity of sites than used in the original calibration. Brust et al. (2021)
905 found that the accuracy of MOD16 can also be improved by including the effect of soil moisture
906 on ET. Although the apparent vegetation type-dependent biases were less pronounced in
907 GLEAM, more work is needed to understand the causes. In a model evaluation study, Miralles et
908 al. (2016) found the opposite pattern, i.e., MOD16 and GLEAM underestimated ET (determined
909 from rainfall and streamflow data) in wet regions (dominated by forest vegetation) and
910 overestimated ET in dry regions (dominated by non-forest vegetation types). A similar pattern
911 was observed in the model evaluation study by Michel et al. (2016) (using ET_{orig} as reference
912 data). In the model evaluation study for South America, GLEAM underestimated ET in both wet
913 and dry regions (Melo et al., 2021). The biases of MOD16 were small in that study. This shows
914 again that the results of model evaluation studies do not necessarily apply to the actual ET
915 products.

916 None of the products were able to correctly represent the seasonal trend in ET in all land cover
917 types and climate zones. Detailed analyses such as this one can help identify the causes of the
918 biases discussed above. For example, the negative biases of MOD16 and GLEAM in GRA and
919 SAV seemed to occur mainly during the dry season. This may indicate an overestimation of the
920 effect of water stress on ET. In contrast, ALEXI seemed to overestimate ET during the dry
921 season. In some cases, the differences between the product-based and the observed trends could
922 be traced to individual sites. The overestimation of ET of Brazilian rainforest by GLEAM during
923 the wet season was also observed by Chen et al. (2022). These authors suggested the lack of an
924 atmospheric moisture control on transpiration as a possible cause of this overestimation. It is
925 likely that the erratic variation observed in the ALEXI data for the tropical wet sites was caused

926 by the cloud-related problems discussed above. However, to correctly identify possible seasonal
927 biases, a more comprehensive analysis that includes more sites and site years is needed.
928 The results showed that the differences between the MOD16 C5 and C6 products were generally
929 smaller than the differences with the other products. Differences between C5 and C6 were to be
930 expected because of differences in input data and spatial resolutions (Mu et al., 2013; Running et
931 al., 2019). The differences persisted when using the common reference dataset (Figure S3).
932 Future work can focus on quantifying the level of consistency between these two collections.

933

934 4.3. Latitudinal comparison of product performance

935 The literature review revealed that remote sensing-based ET products have been evaluated
936 primarily in the northern middle latitudes. The bias is the result of geographic differences in the
937 availability of eddy covariance data due to uneven distribution of flux towers (see, for example,
938 Schimel et al., 2015) and regional differences in data sharing (Villareal and Vargas, 2021). With
939 the results of this study, the availability of evaluation data for the low latitudes was significantly
940 improved. This allowed a comparison of product performance across latitudes. The results of this
941 analysis should be interpreted with caution, however, because the number of evaluation results
942 available varied considerably among latitudes, products, and vegetation categories (Table 4).
943 Smaller normalized RMSEs and smaller PBIAS values suggest better performance of MOD16
944 and GLEAM for low-latitude forests than for northern mid-latitude forests. The weaker
945 correlations at low latitudes are thought to be the result of differences in seasonality rather than
946 differences in performance (see below). The similarity between the latitudinal trends in NRMSE
947 and PBIAS of MOD16 and GLEAM is striking considering the different approaches, forcing
948 data, and resolutions of the underlying models. More work is needed to understand the causes of
949 the apparent latitudinal dependence of these products. In the case of non-forest vegetation, none

950 of the performance metrics showed a clear trend with latitude. Noteworthy is that both MOD16
951 and GLEAM seem to overestimate ET of forest vegetation and underestimate ET of non-forest
952 vegetation in all latitudinal bands (see also discussion above).

953 A limitation of the current analysis is that regional differences were not detected because of the
954 broad zonal bands used. For example, NRMSEs were considerably larger for seasonally dry DBF
955 (0.36 and 0.84 for GLEAM and MOD16, respectively; data not shown) than for wet tropical EBF
956 (0.27 and 0.46, respectively). Similarly, Vinukollu et al. (2011b) and Miralles et al. (2016) found
957 higher relative uncertainties for the subtropics than for the tropics. In these studies, relative
958 uncertainty was estimated from the spread between different model outputs.

959 The weaker correlations for low-latitude forests are most likely explained by the small seasonal
960 variation in ET of EBF. The seasonal variation in ET of temperate forests is much greater due to
961 stronger seasonal variations in radiation and temperature (Baldocchi and Ryu, 2011). Again,
962 however, differences among forests in the low latitudes were large. For example, the R^2 values
963 for DBF were about twice as high as those for EBF (Figure 5). There were no clear latitudinal
964 differences in R^2 for non-forest vegetation. At low latitudes, non-forest vegetation occurs mainly
965 in regions with high seasonality of rainfall (e.g., savanna regions) and thus large variations in
966 ET. Likewise, temperate non-forest vegetation types such as grass and crops show large
967 variations in ET due to seasonal variation in radiation and temperature (e.g., Monteith and Moss,
968 1977).

969

970 4.4. Product performance versus vegetation-match between pixel and site

971 The linear regression analyses across all 40 sites showed that there was no relationship between
972 the product evaluation results and the vegetation-match between pixel and site. Indirect evidence
973 for this was also provided by the finding that the product with the largest pixel size and the

974 lowest average VMI (GLEAM) performed best overall. Similar results were obtained by
975 Hobeichi et al. (2018) and Jiménez et al. (2018). Hobeichi et al. (2018) investigated the effect of
976 a vegetation mismatch between pixel and site on the performance of a merged ET product. For
977 this they divided the eddy covariance sites in two groups, those for which the IGBP land cover
978 type was the same as that of the pixel and those for which it was not. They used MODIS land
979 cover data at the same spatial resolution (0.5°) as the merged ET product. No clear differences in
980 the performance of the product were observed between the two groups of sites. Jiménez et al.
981 (2018) investigated the effect of a vegetation mismatch between pixel and site on the
982 performance of the GLEAM, PT-JPL, and MOD16 algorithms. The models were run with a
983 common input dataset at a spatial resolution of 0.25° . For all three models a single vegetation
984 match index was used (called homogeneity index). This index was calculated using MODIS
985 IGBP land cover data (MCD12Q1) and MODIS vegetation cover data (MOD44B). Also in their
986 study, no significant relationships were found between model performance and the homogeneity
987 index. A challenge is to correctly define the vegetation match index (Hobeichi et al., 2018). In
988 this study, only two vegetation categories were considered (forest and non-forest vegetation), as
989 the land cover type-dependent parameters in MOD16 and ALEXI can be broadly grouped into
990 these two categories (and GLEAM only distinguishes between these two categories). In the other
991 studies, a match was only obtained if the specific IGBP land cover type corresponded. This may
992 be too stringent if the parameters are similar among certain cover types. Understanding the
993 sensitivity of the model outputs to the land cover type-specific parametrizations can help
994 determine which of these approaches is more adequate. Nevertheless, the results obtained so far
995 suggest that errors other than those caused by a vegetation mismatch between pixel and site are
996 more important (Jiménez et al., 2018).

997

998 5. Conclusions

999 There is a geographical bias in the availability of evaluation data for remote sensing-based ET
1000 products in favor of the northern middle latitudes. To address this bias, three products (GLEAM,
1001 MOD16, ALEXI) were evaluated at 40 eddy covariance sites in the low latitudes. From MOD16,
1002 the discontinued collection 5 (C5) and the latest collection (C6) were evaluated. Two potential
1003 problems need to be considered when using eddy covariance observations of ET as reference
1004 data. First, eddy covariance data suffer from uncertainties related to the energy balance closure
1005 problem. Second, scale differences and classification errors can lead to a mismatch in vegetation
1006 between pixel and site (which in turn can complicate the comparisons). Because of the
1007 geographical bias in evaluation studies, it is unknown whether the products perform equally well
1008 at all latitudes.

1009 The differences between MOD16 C5 and C6 were generally smaller than the differences with the
1010 other products. More work is needed, however, to determine the degree of consistency between
1011 the two collections.

1012 Performance and ranking of the evaluated products depended on whether or not the eddy
1013 covariance ET data were corrected for the lack of energy balance closure. When using the
1014 unadjusted fluxes (ET_{orig}), GLEAM showed the best overall performance across the studied land
1015 cover types and climate zones, with the strongest correlations and smallest biases. Neither
1016 MOD16 nor ALEXI consistently outperformed the other. When using the corrected fluxes
1017 (ET_{ebc}), there was no product that stood out in terms of both low bias and strong correlations.

1018 The uncertainty associated with the energy balance closure problem affected the product biases
1019 more than the correlations. Most product evaluation studies use ET_{orig} as reference data. Use of
1020 both ET_{orig} and ET_{ebc} is recommended until a better understanding of the effect of the energy
1021 balance closure problem on ET is obtained.

1022 Few studies have compared the performance of the products examined here. However, a
1023 comparison of results from studies that evaluated these products separately seems to confirm that
1024 GLEAM generally outperforms the other products (when using ET_{orig} as reference data).
1025 Latitudinal dependence of product performance was examined using the results of this study and
1026 those published in the literature. The comparison suggests that MOD16 and GLEAM perform
1027 better for low-latitude forests than for northern mid-latitude forests. However, regional
1028 differences, such as between the tropics and subtropics, can be large and were not detected
1029 because of the broad zonal bands used in this analysis. In the case of non-forest vegetation, the
1030 products show no clear latitudinal differences in performance.
1031 No relationship was found between the product evaluation results and the degree of match
1032 between the vegetation at the flux tower site and that detected in the model pixel. More work is
1033 needed to understand the effect of a vegetation mismatch between pixel and site on product
1034 performance.

1035

1036 CRediT authorship contribution statement

1037 **Diego Salazar-Martínez:** Formal analysis, Investigation, Visualization, Writing - Original
1038 Draft. **Friso Holwerda:** Conceptualization, Data curation, Formal analysis, Investigation,
1039 Methodology, Supervision, Writing - Original Draft, Writing - Review & Editing. **Thomas R.H.**
1040 **Holmes:** Conceptualization, Investigation, Writing - Review & Editing. **Enrico A. Yépez:**
1041 Investigation, Resources, Writing - Review & Editing. **Christopher R. Hain:** Investigation,
1042 Writing - Review & Editing. **Susana Alvarado-Barrientos:** Investigation, Writing - Review &
1043 Editing. **Rest of the authors (alphabetical):** Investigation.

1044

1045 Data availability

1046 The 8-day mean ET_{orig} and ET_{ebc} from the MexFlux sites (Dataset S1), as well as the 8-day mean
1047 ALEXI ET estimates for all but four eddy covariance sites (Dataset S2; see Section 2.1.3 and
1048 Table S2 for more details regarding the missing data) can be accessed in the supporting
1049 information. The rest of the data used in this study can be accessed through the open data portals
1050 as explained in Section 2.

1051

1052 Declaration of Competing Interest

1053 The authors declare that they have no known competing financial interests or personal
1054 relationships that could have appeared to influence the work reported in this paper.

1055

1056 Acknowledgements

1057 This work used eddy covariance data acquired and shared by the FLUXNET community,
1058 including these networks: AmeriFlux, AfriFlux, AsiaFlux, CarboAfrica, CarboEuropeIP,
1059 CarboItaly, CarboMont, ChinaFlux, Fluxnet-Canada, GreenGrass, ICOS, KoFlux, LBA, NECC,
1060 OzFlux-TERN, TCOS-Siberia, and USCCC. The FLUXNET eddy covariance data processing
1061 and harmonization was carried out by the ICOS Ecosystem Thematic Center, AmeriFlux
1062 Management Project and Fluxdata project of FLUXNET, with the support of CDIAC, and the
1063 OzFlux, ChinaFlux and AsiaFlux offices.

1064 We are grateful to Dr. Takashi Hirano and Dr. Shih-Chieh Chang for their permission to use data
1065 from the Palangkaraya Drained Forest (PDF) and Chi-Lan Mountain (CLM) sites, respectively
1066 (AsiaFlux), and to Dr. Michael Liddell for making available data from the Cape Tribulation,
1067 Cow Bay, and Robson Creek sites (OzFlux).

1068 D.S.M. was supported by a graduate scholarship from CONACYT (number 595847). E.A.Y.
1069 acknowledges support from CONACYT (grant number 221014) and PROFAPI-ITSON.

1070 N.E.R.R. was supported by a graduate scholarship from CONACYT (number 278991). G.A.P.
1071 acknowledges support from CONACYT PN (grant number 2017-6231) and USAID-USFS (grant
1072 number 12-IJ-11242306-033). E.G.C. acknowledges support from the Inter-American Institute
1073 for Global Change Research (grant numbers CRN2, CRN3-025) and CONACYT (grant number
1074 175725). J.D.B. and T.A. acknowledge support from SEMARNAT-CONACYT (grant numbers
1075 108000, CB 2008- 01 102855, CB 2013 220788). J.D.B. was supported by a graduate
1076 scholarship from CONACYT (number 211819). F.H. acknowledges support from CONACYT
1077 (grant number 187646) and UNAM-PAPIIT (grant number IB100113). S.A.B. acknowledges
1078 support from CONACYT (grant number INFR-2016-01-269269). B.F.E and J.U.S. acknowledge
1079 support from LANRESC-CONACYT (grant number 271544) and UNAM-PINCC-2020. J.U.S.
1080 was supported by a graduate scholarship from CONACYT (number 415123). J.G.P.
1081 acknowledges support from PROFAPI-ITSON.
1082 Finally, we are grateful to the two anonymous reviewers whose comments and suggestions led to
1083 significant improvements in this work.

1084

1085 Supplementary materials

1086 Supporting information associated with this article can be found, in the online version, at...

1087

1088

1089 References

1090

1091 Aguilar, A.L., Flores, H., Crespo, G., Marín, M.I., Campos, I., Calera, A., 2018. Performance assessment
1092 of MOD16 in evapotranspiration evaluation in Northwestern Mexico. *Water* 10, 901.

1093 <https://doi.org/10.3390/w10070901>.

1094

1095 Allen, R.G., 2008. Quality assessment of weather data and micrometeorological flux—impacts on
1096 evapotranspiration calculation. *Journal of Agricultural Meteorology* 64 (4), 191–204.

1097 <https://doi.org/10.2480/agrmet.64.4.5>.

1098

1099 Alvarado-Barrientos, M.S., López-Adame, H., Lazcano-Hernández, H.E., Arellano-Verdejo, J.,
1100 Hernández-Arana, H.A., 2021. Ecosystem- Atmosphere Exchange of CO₂, Water, and Energy in a Basin

1101 Mangrove of the Northeastern Coast of the Yucatan Peninsula. *Journal of Geophysical Research*

1102 *Biogeosciences* 126 (2), e2020JG005811. <https://doi.org/10.1029/2020JG005811>.

1103

1104 Anderson, M.C., Norman, J.M., Diak, G.R., Kustas, W.P., Mecikalski, J.R., 1997. A two-source time-
1105 integrated model for estimating surface fluxes using thermal infrared remote sensing. *Remote Sensing of*

1106 *Environment* 60(2), 195–216. [https://doi.org/10.1016/S0034-4257\(96\)00215-5](https://doi.org/10.1016/S0034-4257(96)00215-5).

1107

1108 Anderson, M.C., Norman, J.M., Mecikalski, J.R., Otkin, J.A., Kustas, W.P., 2007. A climatological study
1109 of evapotranspiration and moisture stress across the continental United States based on thermal remote

1110 sensing: 1. Model formulation. *Journal of Geophysical Research Atmospheres* 112, DD10117.

1111 <https://doi.org/10.1029/2006JD007506>.

1112

1113 Anderson, M.C., Kustas, W.P., Norman, J.M., Hain, C.R., Mecikalski, J.R., Schultz, L., González-Dugo,

1114 M.P., Cammalleri, C., d'Urso, G., Pimstein, A., Gao, F., 2011. Mapping daily evapotranspiration at field

1115 to continental scales using geostationary and polar orbiting satellite imagery, *Hydrology and Earth*
1116 *System Sciences* 15, 223–239. <https://doi.org/10.5194/hess-15-223-2011>.

1117

1118 Ardö, J., Mölder, M., El-Tahir, B.A., Elkhidir, H.A.M., 2008. Seasonal variation of carbon fluxes in a
1119 sparse savanna in semi arid Sudan. *Carbon Balance and Management* 3, 7. <https://doi.org/10.1186/1750->
1120 0680-3-7.

1121

1122 Badgley, G., Fisher, J.B., Jiménez, C., Tu, K.P., Vinukollu, R., 2015. On uncertainty in global
1123 terrestrial evapotranspiration estimates from choice of input forcing datasets.
1124 *Journal of Hydrometeorology* 16 (4), 1449–1455. <http://dx.doi.org/10.1175/JHM-D-14-0040.1>.

1125

1126 Baldocchi, D.D., Ryu, Y., 2011. A synthesis of forest evaporation fluxes—from days to years—as
1127 measured with eddy covariance. In: Levia, D.F., Carlyle-Moses, D. (Eds.), *Forest Hydrology and*
1128 *Biogeochemistry: Synthesis of Past Research and Future Directions*. Springer Verlag, New York, pp.
1129 101–116.

1130

1131 Barr, A.G., Van der Kamp, G., Black, T.A., McCaughey, J.H., Nesic, Z., 2012. Energy balance closure at
1132 the BERMS flux towers in relation to the water balance of the White Gull Creek watershed 1999–2009.
1133 *Agricultural and Forest Meteorology* 153, 3–13. <https://doi.org/10.1016/j.agrformet.2011.05.017>.

1134

1135 Beck, H.E., Zimmermann, N.E., McVicar, T.R., Vergopolan, N., Berg, A., Wood, E.F., 2018. Present and
1136 future Köppen-Geiger climate classification maps at 1-km resolution. *Scientific Data* 5, 180214.
1137 <https://doi.org/10.1038/sdata.2018.214>.

1138

1139 Beringer, J., Hutley, L.B., Tapper, N.J., Cernusak, L.A., 2007. Savanna fires and their impact on net
1140 ecosystem productivity in North Australia. *Global Change Biology* 13 (5), 990–1004.

1141 <https://doi.org/10.1111/j.1365-2486.2007.01334.x>.

1142

1143 Beringer, J., Hacker, J., Hutley, L.B., Leuning, R., Arndt, S.K., Amiri, R., Bannehr, L., Cernusak, L.A.,
1144 Grover, S., Hensley, C., Hocking, D., Isaac, P., Jamali, H., Kanniah, K., Livesley, S., Neininger, B., Paw
1145 U, K.T., Sea, W., Straten, D., Tapper, N., Weinmann, R., Wood, S., Zegelin, S., 2011. SPECIAL—
1146 Savanna Patterns of Energy and Carbon Integrated across the Landscape. *Bulletin of the American*
1147 *Meteorological Society* 92 (11), 1467–1485. <https://doi.org/10.1175/2011BAMS2948.1>.

1148

1149 Beringer, J., Livesley, S.J., Randle, J., Hutley, L.B., 2013. Carbon dioxide fluxes dominate the
1150 greenhouse gas exchanges of a seasonal wetland in the wet–dry tropics of northern Australia. *Agricultural*
1151 *and Forest Meteorology* 182–183, 239–247. <https://doi.org/10.1016/j.agrformet.2013.06.008>.

1152

1153 Beringer, J., Hutley, L.B., McHugh, I., Arndt, S.K., Campbell, D., Cleugh, H.A., Cleverly, J., Resco de
1154 Dios, V., Eamus, D., Evans, B., Ewenz, C., Grace, P., Griebel, A., Haverd, V., Hinko-Najera, N., Huete,
1155 A., Isaac, P., Kanniah, K., Leuning, R., Liddell, M.J., Macfarlane, C., Meyer, W., Moore, C., Pendall, E.,
1156 Phillips, A., Phillips, R.L., Prober, S.M., Restrepo-Coupe, N., Rutledge, S., Schroder, I., Silberstein, R.,
1157 Southall, R., Yee, M.S., van Gorsel, E., Vote, C., Walker, J., Wardlaw, T., 2016. An introduction to the
1158 Australian and New Zealand flux tower network–OzFlux. *Biogeosciences* 13, 5895–5916.
1159 <https://doi.org/10.5194/bg-13-5895-2016>.

1160

1161 Bonal, D., Bosc, A., Ponton, S., Goret, J.-Y., Burban, B., Gross, P., Bonnefond, J.-M., Elbers, J.,
1162 Longdoz, B., Epron, D., Guehl, J.-M., Granier, A., 2008. Impact of severe dry season on net ecosystem
1163 exchange in the Neotropical rainforest of French Guiana. *Global Change Biology* 14, 1917–1933.
1164 <https://doi.org/10.1111/j.1365-2486.2008.01610.x>.

1165

1166 Bristow, M., Hutley, L.B., Beringer, J., Livesley, S.J., Edwards, A.C., Arndt, S.K., 2016. Quantifying the

1167 relative importance of greenhouse gas emissions from current and future savanna land use change across
1168 northern Australia. *Biogeosciences* 13, 6285–6303. <https://doi.org/10.5194/bg-13-6285-2016>.
1169
1170 Bruijnzeel, L.A., Mulligan, M., Scatena, F.N., 2011. Hydrometeorology of tropical montane cloud
1171 forests: emerging patterns. *Hydrological Processes* 25, 465–498. <https://doi.org/10.1002/hyp.7974>.
1172
1173 Brust, C., Kimball, J.S., Maneta, M.P., Jencso, K., He, M., Reichle, R.H., 2021. Using SMAP Level-4
1174 soil moisture to constrain MOD16 evapotranspiration over the contiguous USA. *Remote Sensing of*
1175 *Environment* 255, 112277. <https://doi.org/10.1016/j.rse.2020.112277>.
1176
1177 Brutsaert, W., 1982. *Evaporation Into the Atmosphere: Theory, History, and Applications*. Springer,
1178 Dordrecht. <https://doi.org/10.1007/978-94-017-1497-6>.
1179
1180 Cernusak, L.A., Hutley, L.B., Beringer, J., Holtum, J.A.M., Turner, B.L., 2011. Photosynthetic
1181 physiology of eucalypts along a sub-continental rainfall gradient in northern Australia. *Agricultural and*
1182 *Forest Meteorology* 151 (11), 1462–1470. <https://doi.org/10.1016/j.agrformet.2011.01.006>.
1183
1184 Chang, J.C., Hanna, R.S., 2004. Air quality model performance evaluation. *Meteorology and*
1185 *Atmospheric Physics* 87, 167–196. <https://doi.org/10.1007/s00703-003-0070-7>.
1186
1187 Chen, J.M., Liu, J., 2020. Evolution of evapotranspiration models using thermal and shortwave remote
1188 sensing data. *Remote Sensing of Environment* 237, 111594. <https://doi.org/10.1016/j.rse.2019.111594>.
1189
1190 Chen, H., Zhu, G., Shang, S., Qin, W., Zhang, Y., Su, Y., Zhang, K., Zhu, Y., Xu, C., 2022. Uncertainties
1191 in partitioning evapotranspiration by two remote sensing-based models. *Journal of Hydrology* 604,
1192 127223. <https://doi.org/10.1016/j.jhydrol.2021.127223>.

1193

1194 Chu, H.-S., Chang, S.-C., Klemm, O., Lai, C.-W., Lin, Y.-Z., Wu, C.-C., Lin, J.-Y., Jiang, J.-Y., Chen, J.,
1195 Gottgens, J.F., Hsia, Y.-J., 2014. Does canopy wetness matter? Evapotranspiration from a subtropical
1196 montane cloud forest in Taiwan. *Hydrological Processes* 28, 1190–1214.
1197 <https://doi.org/10.1002/hyp.9662>.

1198

1199 Chu, H., Luo, X., Ouyang, Z., Chan, W.S., Dengel, S., Biraud, S.C., Torn, M.S., Metzger, S., Kumar, J.,
1200 Arain, M.A., Arkebauer, T.J., Baldocchi, D., Bernacchi, C., Billesbach, D., Black, T.A., Blanken, P.D.,
1201 Bohrer, G., Bracho, R., Brown, S., Brunzell, N.A., Chen, J., Chen, X., Clark, K., Desai, A.R., Duman, T.,
1202 Durden, D., Fares, S., Forbrich, I., Gamon, J.A., Gough, C.M., Griffis, T., Helbig, M., Hollinger, D.,
1203 Humphreys, E., Ikawa, H., Iwata, H., Ju, Y., Knowles, J.F., Knox, S.H., Kobayashi, H., Kolb, T., Law,
1204 B., Lee, X., Litvak, M., Liu, H., Munger, J.W., Noormets, A., Novick, K., Oberbauer, S.F., Oechel, W.,
1205 Oikawa, P., Papuga, S.A., Pendall, E., Prajapati, P., Prueger, J., Quinton, W.L., Richardson, A.D.,
1206 Russell, E.S., Scott, R.L., Starr, G., Staebler, R., Stoy, P.C., Stuart-Haëntjens, E., Sonnentag, O., Sullivan,
1207 R.C., Suyker, A., Ueyama, M., Vargas, R., Wood, J.D., Zona, D., 2021. Representativeness of Eddy-
1208 Covariance flux footprints for areas surrounding AmeriFlux sites. *Agricultural and Forest Meteorology*
1209 301-302, 108350. <https://doi.org/10.1016/j.agrformet.2021.108350>.

1210

1211 Cleverly, J., Boulain, N., Villalobos-Vega, R., Grant, N., Faux, R., Wood, C., Cook, P.G., Yu, Q., Leigh,
1212 A., Eamus, D., 2013. Dynamics of component carbon fluxes in a semi-arid Acacia woodland, central
1213 Australia. *Journal of Geophysical Research Biogeosciences* 118 (3), 1168–1185.
1214 <https://doi.org/10.1002/jgrg.20101>.

1215

1216 Cleverly, J., Eamus, D., Isaac, P., 2016. FLUXNET2015 AU-TTE Ti Tree East. Australia: N. p.
1217 <https://doi.org/10.18140/FLX/1440205>.

1218

1219 Cui, D., Liang, S., Wang, D., 2021. Observed and projected changes in global climate zones based on
1220 Köppen climate classification. *WIREs Clim Change* 12, e701. <https://doi.org/10.1002/wcc.701>.
1221
1222 Dawson, C.W., Mount, N.J., Abrahart, R.J., Shamseldin, A.Y., 2012. Ideal point error for model
1223 assessment in data-driven river flow forecasting. *Hydrology and Earth System Sciences* 16, 3049–3060.
1224 <https://doi.org/10.5194/hess-16-3049-2012>.
1225
1226 Delgado-Balbuena, J., Yépez, E.A., Paz-Pellat, F., Ángeles-Pérez, G., Aguirre-Gutiérrez, C., Alvarado-
1227 Barrientos, M.S., Arredondo, T., Ayala-Niño, F., Bullock, S.H., Castellanos, A.E., Cueva, A., Figueroa-
1228 Espinoza, B., Garatuza-Payán, J., González-del Castillo, E., González-Sosa, E., Guevara-Escobar, A.,
1229 Hinojo-Hinojo, C., Paw U, K.-T., Lizárraga-Celaya, C., Maya-Delgado, Y., Oechel, W., Pérez-Ruiz, E.R.,
1230 Quesada-Avendaño, M., Robles-Zazueta, C.A., Rodríguez, J.C., Rojas-Robles, N.E., Tarin-Terrazas, T.,
1231 Troyo-Diéguez, E., Uuh-Sonda, J., Vargas-Terminel, M.L., Vargas, R., Vega-Puga, M.G., Verduzco,
1232 V.S., Vivoni, E.R., Watts, C.J., 2018. Base de datos de flujos verticales de dióxido de carbono en
1233 ecosistemas terrestres y costeros en México. *Elementos para Políticas Públicas* 2 (2), 93–108.
1234
1235 Delgado- Balbuena, J., Arredondo, J.T., Loescher, H.W., Pineda- Martínez, L.F., Carbajal, J.N., Vargas,
1236 R., 2019. Seasonal precipitation legacy effects determine the carbon balance of a semiarid grassland.
1237 *Journal of Geophysical Research Biogeosciences* 124, 987–1000. <https://doi.org/10.1029/2018JG004799>.
1238
1239 Denager, T., Looms, M.C., Sonnenborg, T.O., Jensen, K.H., 2020. Comparison of evapotranspiration
1240 estimates using the water balance and the eddy covariance methods. *Vadose Zone Journal* 19, e20032.
1241 <https://doi.org/10.1002/vzj2.20032>.
1242
1243 Doelling, D., 2012. CERES Level 3 SYN1DEG-DAYTerraCAqua netCDF file – Edition 3A. NASA
1244 Langley Atmospheric Science Data Center DAAC.

1245 https://doi.org/10.5067/TERRA+AQUA/CERES/SYN1DEGDAY_L3.003A.

1246

1247 Dolman, A.J., Miralles, D.G., De Jeu, R.A.M., 2014. Fifty years since Monteith's 1965 seminal paper: the
1248 emergence of global ecohydrology. *Ecohydrology* 7, 897–902. <https://doi.org/10.1002/eco.1505>.

1249

1250 Drake, B., Hinkle, R., 2003–2006. FLUXNET2015 US-KS2 Kennedy Space Center (scrub oak). Dataset.
1251 <https://doi.org/10.18140/FLX/1440075>.

1252

1253 Eamus, D., Cleverly, J., Boulain, N., Grant, N., Faux, R., Villalobos-Vega, R., 2013. Carbon and water
1254 fluxes in an arid-zone Acacia savanna woodland: An analyses of seasonal patterns and responses to
1255 rainfall events. *Agricultural and Forest Meteorology* 182–183, 225– 238.
1256 <http://dx.doi.org/10.1016/j.agrformet.2013.04.020>.

1257

1258 Elshorbagy, A., Corzo, G., Srinivasulu, S., Solomatine, D.P., 2010. Experimental investigation of the
1259 predictive capabilities of data driven modeling techniques in hydrology - Part 1: Concepts and
1260 methodology. *Hydrology and Earth System Sciences* 14, 1931–1941. [https://doi.org/10.5194/hess-14-](https://doi.org/10.5194/hess-14-1931-2010)
1261 [1931-2010](https://doi.org/10.5194/hess-14-1931-2010).

1262

1263 Figueroa-Espinoza, B., Sánchez-Mejía, Z., Uuh-Sonda, J.M., Salles, P., Méndez-Barroso, L., Gutiérrez-
1264 Jurado, H.A., 2021. Friction Velocity estimation using a 2D Sonic Anemometer in Coastal Zones.
1265 *Atmósfera*. <https://doi.org/10.20937/ATM.52960>.

1266

1267 Fisher, J.B., Tu, K.P., Baldocchi, D.D., 2008. Global estimates of the land–atmosphere water flux based
1268 on monthly AVHRR and ISLSCP-II data, validated at 16 FLUXNET sites. *Remote Sensing of*
1269 *Environment* 112, 901 – 919. <https://doi.org/10.1016/j.rse.2007.06.025>.

1270

1271 Fisher, J.B., Malhi, Y., Bonal, D., Da Rocha, H.R., De Araújo, A.C., Gamo, M., Goulden, M.L., Hirano,
1272 T., Huete, A.R., Kondo, H., Kumagai, T., Loescher, H.W., Miller, S., Nobre, A.D., Nouvellon, Y.,
1273 Oberbauer, S.F., Panuthai, S., Rouspard, O., Saleska, S., Tanaka, K., Tanaka, N., Tu, K.P., Von Randow,
1274 C., 2009. The land–atmosphere water flux in the tropics. *Global Change Biology* 15, 2694–2714.
1275 <https://doi.org/10.1111/j.1365-2486.2008.01813.x>.
1276
1277 Fisher, J.B., Melton, F., Middleton, E., Hain, C., Anderson, M., Allen, R., McCabe, M.F., Hook, S.,
1278 Baldocchi, D., Townsend, P.A., Kilic, A., Tu, K., Miralles, D.D., Perret, J., Lagouarde, J.-P., Waliser, D.,
1279 Purdy, A.J., French, A., Schimel, D., Famiglietti, J.S., Stephens, G., Wood, E.F., 2017. The future of
1280 evapotranspiration: Global requirements for ecosystem functioning, carbon and climate feedbacks,
1281 agricultural management, and water resources, *Water Resources Research* 53, 2618–2626.
1282 <https://doi.org/10.1002/2016WR020175>.
1283
1284 Foken, T., Aubinet, M., Finnigan, J.J., Leclerc, M.Y., Mauder, M., Paw U, K.T., 2011. Results of a panel
1285 discussion about the energy balance closure correction for trace gases. *Bulletin of the American*
1286 *Meteorological Society* 92 (4), ES13–ES18.
1287
1288 Frank, J.M., Massman, W.J., Swiatek, E., Zimmerman, H.A., Ewers, B.E., 2016. All sonic anemometers
1289 need to correct for transducer and structural shadowing in their velocity measurements. *Journal of*
1290 *Atmospheric and Oceanic Technology* 33 (1), 149–167. <https://doi.org/10.1175/JTECH-D-15-0171.1>.
1291
1292 Gao, Z., Liu, H., Katul, G.G., Foken, T., 2017. Non-closure of the surface energy balance explained by
1293 phase difference between vertical velocity and scalars of large atmospheric eddies. *Environmental*
1294 *Research Letters* 12, 034025. <https://doi.org/10.1088/1748-9326/aa625b>.
1295
1296 Gash, J.H.C., 1979. An analytical model of rainfall interception by forests. *Quarterly Journal of the Royal*

1297 Meteorological Society 105, 43–55. <https://doi.org/10.1002/qj.49710544304>.

1298

1299 González del Castillo, E., Paw U, K.T., Sánchez-Azofeifa, A., 2018. Turbulence scales for eddy
1300 covariance quality control over a tropical dry forest in complex terrain. *Agricultural and Forest*
1301 *Meteorology* 249, 390–406. <https://doi.org/10.1016/j.agrformet.2017.11.014>.

1302

1303 Goulden, M. 2000–2004. FLUXNET2015 BR-Sa3 Santarem-Km83-Logged Forest, Dataset.
1304 <https://doi.org/10.18140/FLX/1440033>.

1305

1306 Hansen, M.C., Song, X., DiMiceli, C., Carroll, M., Sohlberg, R., Kim, D.-H., Townshend, J., 2018.
1307 MEaSURES Vegetation Continuous Fields ESDR Algorithm Theoretical Basis Document (ATBD)
1308 Version 2.0. University of Maryland, College Park, MD. Available at:
1309 https://lpdaac.usgs.gov/documents/144/VCF5KYR_ATBD.pdf (last access: 29 August 2021).

1310

1311 Hidalgo-Sánchez, M., Ángeles-Pérez, G., Yépez, E.A., Plascencia-Escalante, F.O., Delgado-Balbuena, J.,
1312 González-Martínez, T.M., 2021. Evapotranspiration and energy exchange in a temperate forest in Mexico.
1313 *Tecnología y Ciencias del Agua* 12(2), 490–537. <https://doi.org/10.24850/j-tyca-2021-02-11>.

1314

1315 Hirano, T., Kusin, K., Limin, S., Osaki, M., 2015. Evapotranspiration of tropical peat swamp forests.
1316 *Global Change Biology* 21, 1914–1927. <https://doi.org/10.1111/gcb.12653>.

1317

1318 Hobeichi, S., Abramowitz, G., Evans, J., Ukkola, A., 2018. Derived Optimal Linear Combination
1319 Evapotranspiration (DOLCE): a global gridded synthesis ET estimate. *Hydrology and Earth System*
1320 *Sciences* 22, 1317–1336. <https://doi.org/10.5194/hess-22-1317-2018>.

1321

1322 Hain, C.R., Anderson, M.C., 2017. Estimating morning change in land surface temperature from MODIS

1323 day/night observations: Applications for surface energy balance modeling. *Geophysical Research Letters*
1324 44, 9723–9733. <https://doi.org/10.1002/2017GL074952>.

1325

1326 Holmes, T.R.H., Hain, C.R., Crow, W.T., Anderson, M.C., Kustas, W.P., 2018. Microwave
1327 implementation of two-source energy balance approach for estimating evapotranspiration. *Hydrology and*
1328 *Earth System Sciences* 22, 1351–1369. <https://doi.org/10.5194/hess-22-1351-2018>.

1329

1330 Holwerda, F., Alvarado-Barrientos, M.S., González-Martínez, T.M., 2016. Surface energy exchange in a
1331 tropical montane cloud forest environment: flux partitioning, and seasonal and land cover-related
1332 variations. *Agricultural and Forest Meteorology* 228, 13–28.
1333 <https://doi.org/10.1016/j.agrformet.2016.06.011>.

1334

1335 Holwerda, F., Meesters, A.G.C.A., 2019. Soil evaporation in a shaded coffee plantation derived from
1336 eddy covariance measurements. *Journal of Geophysical Research Biogeosciences* 124 (6), 1472–1490.
1337 <https://doi.org/10.1029/2018JG004911>.

1338

1339 Hsieh, C.-I., Huang, C.-W., Kiely, G., 2009. Long-term estimation of soil heat flux by single layer soil
1340 temperature. *International Journal of Biometeorology* 53, 113–123. [https://doi.org/10.1007/s00484-008-](https://doi.org/10.1007/s00484-008-0198-8)
1341 0198-8.

1342

1343 Hu, G., Jia, L., Menenti, M., 2015. Comparison of MOD16 and LSA-SAF MSG evapotranspiration
1344 products over Europe for 2011. *Remote Sensing of Environment* 156, 510–526.
1345 <http://dx.doi.org/10.1016/j.rse.2014.10.017>.

1346

1347 Hutley, L.B., Beringer, J., Isaac, P.R., Hacker, J.M., Cernusak, L.A., 2011. A sub-continental scale living
1348 laboratory: Spatial patterns of savanna vegetation over a rainfall gradient in northern Australia.

1349 Agricultural and Forest Meteorology 151, 1417–1428. <https://doi.org/10.1016/j.agrformet.2011.03.002>.
1350

1351 Isaac, P., Cleverly, J., McHugh, I., Van Gorsel, E., Ewenz, C., Beringer, J., 2017. OzFlux data: network
1352 integration from collection to curation. Biogeosciences 14, 2903–2928. [https://doi.org/10.5194/bg-14-](https://doi.org/10.5194/bg-14-2903-2017)
1353 2903-2017.
1354

1355 Jiménez, C., Martens, B., Miralles, D.M., Fisher, J.B., Beck, H.E., Fernández-Prieto, D., 2018. Exploring
1356 the merging of the global land evaporation WACMOS-ET products based on local tower measurements.
1357 Hydrology and Earth System Sciences 22, 4513–4533. <https://doi.org/10.5194/hess-22-4513-2018>.
1358

1359 Khan, M.S., Liaqat, U.W., Baik, J., Choi, M., 2018. Stand-alone uncertainty characterization of GLEAM,
1360 GLDAS and MOD16 evapotranspiration products using an extended triple collocation approach.
1361 Agricultural and Forest Meteorology 252, 256–268. <https://doi.org/10.1016/j.agrformet.2018.01.022>.
1362

1363 Khan, M.S., Baik, J., Choi, M., 2020. Inter-comparison of evapotranspiration datasets over heterogeneous
1364 landscapes across Australia. Advances in Space Research 66 (3), 533–545.
1365 <https://doi.org/10.1016/j.asr.2020.04.037>.
1366

1367 Kim, H.W., Hwang, K., Mu, Q., Lee, S.O., Choi, M., 2012. Validation of MODIS 16 global terrestrial
1368 evapotranspiration products in various climates and land cover types in Asia. KSCE Journal of Civil
1369 Engineering 16 (2), 229–238. <https://doi.org/10.1007/s12205-012-0006-1>.
1370

1371 Komatsu, H., 2005. Forest categorization according to dry-canopy evaporation rates in the growing
1372 season: comparison of the Priestley–Taylor coefficient values from various observation sites.
1373 Hydrological Processes 19, 3873–3896. <https://doi.org/10.1002/hyp.5987>.
1374

1375 Liddell, M., 2013. Cape Tribulation OzFlux tower site OzFlux: Australian and New Zealand Flux
1376 Research and Monitoring hdl: 102.100.100/14242.
1377
1378 Lu, Y., Lu, S., Horton, R., Ren, T., 2014. An empirical model for estimating soil thermal conductivity
1379 from texture, water content, and bulk density. *Soil Science Society of America Journal* 78, 1859–1868.
1380 <https://doi.org/10.2136/sssaj2014.05.0218>.
1381
1382 Martens, B., Miralles, D.G., Lievens, H., Van der Schalie, R., De Jeu, R.A.M., Fernández-Prieto, D.,
1383 Beck, H.E., Dorigo, W.A., Verhoest, N.E.C., 2017. GLEAM v3: satellite-based land evaporation and
1384 root-zone soil moisture. *Geoscientific Model Development* 10, 1903–1925. [https://doi.org/10.5194/gmd-](https://doi.org/10.5194/gmd-10-1903-2017)
1385 [10-1903-2017](https://doi.org/10.5194/gmd-10-1903-2017).
1386
1387 Mauder, M., Genzel, S., Fu, J., Kiese, R., Soltani, M., Steinbrecher, R., Zeeman, M., Banerjee, T., De
1388 Roo, F., Kunstmann, H., 2018. Evaluation of energy balance closure adjustment methods by independent
1389 evapotranspiration estimates from lysimeters and hydrological simulations. *Hydrological Processes* 32,
1390 39–50. <https://doi.org/10.1002/Hyp.11397>.
1391
1392 Mauder, M., Foken, T., Cuxart, J., 2020. Surface-energy-balance closure over land: a review. *Boundary*
1393 *Layer Meteorology* 177, 395–426. <https://doi.org/10.1007/s10546-020-00529-6>.
1394
1395 Mayocchi, C.L., Bristow, K.L., 1995. Soil surface heat flux: some general questions on measurements.
1396 *Agricultural and Forest Meteorology* 75, 43–50. [https://doi.org/10.1016/0168-1923\(94\)02198-S](https://doi.org/10.1016/0168-1923(94)02198-S).
1397
1398 McCabe, M.F., Ershadi, A., Jimenez, C., Miralles, D.G., Michel, D., Wood, E.F., 2016. The GEWEX
1399 LandFlux project: evaluation of model evaporation using tower-based and globally gridded forcing data.
1400 *Geoscientific Model Development* 9, 283–305. <https://doi.org/10.5194/gmd-9-283-2016>.

1401

1402 Melo, D.C.D., Anache, J.A.A., Borges, V.P., Miralles, D.G., Martens, B., Fisher, J.B., Nóbrega, R.L.B.,
1403 Moreno, A., Cabral, O.M.R., Rodrigues, T.R., Bezerra, B., Silva, C.M.S., Meira Neto, A.A., Moura,
1404 M.S.B., Marques, T.V., Campos, S., Nogueira, J.S., Rosolem, R., Souza, R.M.S., Antonino, A.C.D., Holl,
1405 D., Galleguillos, M., Perez-Quezada, J.F., Verhoef, A., Kutzbach, L., Lima, J.R.S., Souza, E.S., Gassman,
1406 M.I., Perez, C.F., Tonti, N., Posse, G., Rains, D., Oliveira, P.T.S., Wendland, E., 2021. Are remote
1407 sensing evapotranspiration models reliable across South American ecoregions? *Water Resources*
1408 *Research* 57, e2020WR028752. <https://doi.org/10.1029/2020WR028752>.

1409

1410 Melton, F.S., Huntington, J., Grimm, R., Herring, J., Hall, M., Rollison, D., Erickson, T., Allen, R.,
1411 Anderson, M., Fisher, J.B., Kilic, A., Senay, G.B., Volk, J., Hain, C., Johnson, L., Ruhoff, A., Blankenau,
1412 P., Bromley, M., Carrara, W., Daudert, B., Doherty, C., Dunkerly, C., Friedrichs, M., Guzman, A.,
1413 Halverson, G., Hansen, J., Harding, J., Kang, Y., Ketchum, D., Minor, B., Morton, C., Ortega-Salazar, S.,
1414 Ott, T., Ozdogan, M., ReVelle, P.M., Schull, M., Wang, C., Yang, Y., Anderson, R.G., 2021. OpenET:
1415 Filling a critical data gap in water management for the western United States. *Journal of the American*
1416 *Water Resources Association* 1–24. <https://doi.org/10.1111/1752-1688.12956>.

1417

1418 Merbold, L., Ardö, J., Arneith, A., Scholes, R.J., Nouvellon, Y., de Grandcourt, A., Archibald, S.,
1419 Bonnefond, J.M., Boulain, N., Brueggemann, N., Bruemmer, C., Cappelaere, B., Ceschia, E., El-Khidir,
1420 H.A.M., El-Tahir, B.A., Falk, U., Lloyd, J., Kergoat, L., Le Dantec, V., Mougín, E., Muchinda, M.,
1421 Mukelabai, M.M., Ramier, D., Rouspard, O., Timouk, F., Veenendaal, E.M., Kutsch, W.L., 2009.
1422 *Precipitation as driver of carbon fluxes in 11 African ecosystems. Biogeosciences* 6 (6), 1027–1041.
1423 <https://doi.org/10.5194/bg-6-1027-2009>.

1424

1425 Michel, D., Jiménez, C., Miralles, D.G., Jung, M., Hirschi, M., Ershadi, A., Martens, B., McCabe, M.F.,
1426 Fisher, J.B., Mu, Q., Seneviratne, S.I., Wood, E.F., Fernández-Prieto, D., 2016. The WACMOS-ET

1427 project – Part 1: Tower-scale evaluation of four remote-sensing-based evapotranspiration algorithms.
1428 Hydrology and Earth System Sciences 20, 803–822. <https://doi.org/10.5194/hess-20-803-2016>.
1429
1430 Miralles, D.G., Gash, J.H., Holmes, T.R.H., De Jeu, R.A.M., Dolman, A.J., 2010. Global canopy
1431 interception from satellite observations. *Journal of Geophysical Research Atmospheres* 115, D16122.
1432 <https://doi.org/10.1029/2009JD013530>.
1433
1434 Miralles, D.G., Holmes, T.R.H., De Jeu, R.A.M., Gash, J.H., Meesters, A.G.C.A., Dolman, A.J., 2011.
1435 Global land-surface evaporation estimated from satellite-based observations. *Hydrology and Earth*
1436 *System Sciences* 15, 453–469. <https://doi.org/10.5194/hess-15-453-2011>.
1437
1438 Miralles, D.G., Jiménez, C., Jung, M., Michel, D., Ershadi, A., McCabe, M.F., Hirschi, M., Martens, B.,
1439 Dolman, A.J., Fisher, J.B., Mu, Q., Seneviratne, S.I., Wood, E.F., Fernández-Prieto, D., 2016. The
1440 WACMOS-ET project – Part 2: Evaluation of global terrestrial evaporation data sets. *Hydrology and*
1441 *Earth System Sciences* 20, 823–842. <https://doi.org/10.5194/hess-20-823-2016>.
1442
1443 Monteith, J.L., 1965. Evaporation and environment. *Symposia of the Society for Experimental Biology*
1444 19, 205–234.
1445
1446 Monteith, J.L., Moss, C.J., 1977. Climate and the efficiency of crop production in Britain.
1447 *Philosophical Transactions of the Royal Society B* 281, 277–294. <https://doi.org/10.1098/rstb.1977.0140>.
1448
1449 Mu, Q., Heinsch, F.A., Zhao, M., Running, S.W., 2007. Development of a global evapotranspiration
1450 algorithm based on MODIS and global meteorology data. *Remote Sensing of Environment* 111, 519–536.
1451 <https://doi.org/10.1016/j.rse.2007.04.015>.
1452

1453 Mu, Q., Zhao, M., Running, S.W., 2011. Improvements to a MODIS global terrestrial evapotranspiration
1454 algorithm. *Remote Sensing of Environment* 115, 1781–1800. <https://doi.org/10.1016/j.rse.2011.02.019>.
1455

1456 Mu, Q., Zhao, M., Running, S.W., 2013. MODIS Global Terrestrial Evapotranspiration (ET) Product
1457 (NASA MOD16A2/A3). Algorithm Theoretical Basis Document, Collection 5.
1458

1459 Myneni, R.B., Hoffman, S., Knyazikhin, Y., Privette, J.L., Glassy, J., Tian, Y., Wang, Y., Song, X.,
1460 Zhang, Y., Smith, G.R., Lotsch, A., Friedl, M., Morisette, J.T., Votava, P., Nemani, R.R., Running, S.W.,
1461 2002. Global products of vegetation leaf area and fraction absorbed PAR from year one of MODIS data.
1462 *Remote Sensing of Environment* 83 (1–2), 214–231. [https://doi.org/10.1016/S0034-4257\(02\)00074-3](https://doi.org/10.1016/S0034-4257(02)00074-3).
1463

1464 MY-PSO, 2003–2009. FLUXNET2015 MY-PSO Pasoh Forest Reserve (PSO), Dataset.
1465 <https://doi.org/10.18140/FLX/1440240>.
1466

1467 Pastorello, G., Agarwal, D., Papale, D., Samak, T., Trotta, C., Ribeca, A., Poindexter, C., Faybishenko,
1468 B., Gunter, D., Hollowgrass, R., Canfora, E., 2014. Observational data patterns for time series data
1469 quality assessment. In: 2014 IEEE 10th International Conference on e-Science, pp. 271–278.
1470 <https://doi.org/10.1109/eScience.2014.45>.
1471

1472 Pastorello, G., et al., 2020. The FLUXNET2015 dataset and the ONEFlux processing pipeline for eddy
1473 covariance data. *Scientific Data* 7, 225. <https://doi.org/10.1038/s41597-020-0534-3>.
1474

1475 Pérez-Ruiz, E.R., Vivoni, E.R., Yépez, E.A., Rodríguez, J.C., Gochis, D.J., Robles-Morua, A., Delgado-
1476 Balbuena, J., Adams, D.K., 2021. Landscape controls on water-energy-carbon fluxes across different
1477 ecosystems during the North American monsoon. *Journal of Geophysical Research Biogeosciences* 126,
1478 e2020JG005809. <https://doi.org/10.1029/2020JG005809>.

1479

1480 Piñeiro, G., Perelman, S., Guerschman, J.P., Paruelo, J.M., 2008. How to evaluate models: Observed vs.
1481 predicted or predicted vs. observed? *Ecological modelling* 216, 316–322.

1482 <https://doi.org/10.1016/j.ecolmodel.2008.05.006>.

1483

1484 Posse, G., Lewczuk, N., Richter, K., Cristiano, P., 2016. Carbon and water vapor balance in a subtropical
1485 pine plantation. *iForest - Biogeosciences and Forestry* 9 (5), 736–742. <https://doi.org/10.3832/ifor1815->
1486 009.

1487

1488 Priestley, C.H.B., Taylor, R.J., 1972. On the assessment of surface heat flux and evaporation using large-
1489 scale parameters. *Monthly Weather Review* 100, 81–92.

1490

1491 Ramoelo, A., Majazi, N., Mathieu, R., Jovanovic, N., Nickless, A., Dziki, S., 2014. Validation of global
1492 evapotranspiration product (MOD16) using flux tower data in the African Savanna, South Africa. *Remote*
1493 *Sensing* 6, 7406–7423. <http://dx.doi.org/10.3390/rs6087406>.

1494

1495 Reitz, M., Senay, G.B., Sanford, W.E., 2017. Combining remote sensing and water-balance
1496 evapotranspiration estimates for the conterminous United States. *Remote Sensing* 9, 1181.

1497 <https://doi.org/10.3390/rs9121181>.

1498

1499 Rojas- Robles, N.E., Garatuza- Payán, J., Álvarez- Yépiz, J.C., Sánchez- Mejía, Z.M., Vargas, R.,
1500 Yépez, E.A., 2020. Environmental controls on carbon and water fluxes in an old- growth tropical dry
1501 forest. *Journal of Geophysical Research Biogeosciences* 125, e2020JG005666.

1502 <https://doi.org/10.1029/2020JG005666>.

1503

1504 Ruhoff, A.L., Paz, A.R., Aragao, L.E.O.C., Mu, Q., Malhi, Y., Collischonn, W., Rocha, H.R., Running,

1505 S.W., 2013. Assessment of the MODIS global evapotranspiration algorithm using eddy covariance
1506 measurements and hydrological modelling in the Rio Grande basin. *Hydrological Sciences Journal* 58 (8),
1507 1658–1676. <http://dx.doi.org/10.1080/02626667.2013.837578>.
1508
1509 Running, S.W., Mu, Q., Zhao, M., Moreno, A., 2019. MODIS Global Terrestrial Evapotranspiration (ET)
1510 Product (MOD16A2/A3 and Year- end Gap- filled MOD16A2GF/A3GF) NASA Earth Observing
1511 System MODIS Land Algorithm (for Collection 6). Version 2.0, Jan 3rd, 2019. [https://modis-](https://modis-land.gsfc.nasa.gov/pdf/MOD16UsersGuideV2.022019.pdf)
1512 [land.gsfc.nasa.gov/pdf/MOD16UsersGuideV2.022019.pdf](https://modis-land.gsfc.nasa.gov/pdf/MOD16UsersGuideV2.022019.pdf).
1513
1514 Saha, S., Moorthi, S., Wu, X., Wang, J., Nadiga, S., Tripp, P., Behringer, D., Hou, Y.-T., Chuang, H.-Y.,
1515 Iredell, M., Ek, M., Meng, J., Yang, R., Mendez, M.P., van den Dool, H., Zhang, Q., Wang, W., Chen,
1516 M., Becker, E., 2011. NCEP Climate Forecast System Version 2 (CFSv2) 6-hourly Products. Research
1517 Data Archive at the National Center for Atmospheric Research, Computational and Information Systems
1518 Laboratory. <https://doi.org/10.5065/D61C1TXF>.
1519
1520 Saleska, S., 2002–2011. FLUXNET2015 BR-Sa1 Santarem-Km67-Primary Forest, Dataset.
1521 <https://doi.org/10.18140/FLX/1440032>.
1522
1523 Santanello, J.A., Jr., Friedl, M.A., 2003. Diurnal covariation in soil heat flux and net radiation. *Journal of*
1524 *Applied Meteorology* 42 (6), 851–862.
1525
1526 Sauer, T.J., 2002. 5.5 Heat flux density. In: Dane, J.H., Topp, G.C. (Eds.), *Methods of Soil Analysis: Part*
1527 *4 Physical Methods*. Soil Science Society of America, Madison, WI, pp. 1233–1248.
1528 <https://doi.org/10.2136/sssabookser5.4.c52>.
1529
1530 Schimel, D., Pavlick, R., Fisher, J.B., Asner, G.P., Saatchi, S., Townsend, P., Miller, C., Frankenberg, C.,

1531 Hibbard, K., Cox, P., 2015. Observing terrestrial ecosystems and the carbon cycle from space. *Global*
1532 *Change Biology* 21, 1762–1776. <https://doi.org/10.1111/gcb.12822>.
1533

1534 Schroder, I., Kuske, T., Zegelin, S., 2014. Eddy Covariance Dataset for Arcturus (2011–2013).
1535 Geoscience Australia, Canberra. <https://doi.org/10.2100/100/14249>.
1536

1537 Senay, G.B., Bohms, S., Singh, R.K., Gowda, P.H., Velpuri, N.M., Alemu, H., Verdin, J.P., 2013.
1538 Operational evapotranspiration mapping using remote sensing and weather datasets: A new
1539 parameterization for the SSEB approach. *Journal of the American Water Resources Association* 49 (3),
1540 577–591. <https://doi.org/10.1111/jawr.12057>.
1541

1542 Senay, G.B., Kagone, S., Velpuri, N.M., 2020. Operational global actual evapotranspiration:
1543 development, evaluation, and dissemination. *Sensors* 20, 1915. <https://doi.org/10.3390/s20071915>.
1544

1545 Sheffield, J., Wood, E.F., Pan, M., Beck, H., Coccia, G., Serrat-Capdevila, A., Verbist, K., 2018. Satellite
1546 remote sensing for water resources management: Potential for supporting sustainable development in
1547 data-poor regions. *Water Resources Research* 54, 9724–9758. <https://doi.org/10.1029/2017WR022437>.
1548

1549 Shuttleworth, W.J., Gurney, R.J., Hsu, A.Y., Ormsby, J.P., 1989. FIFE: the variation in energy
1550 partitioning at surface flux sites. *IAHS Red Book Series no. 186*, 67–74.
1551

1552 Shuttleworth, W.J., 1993. Evaporation. In: Maidment, D.R. (Ed.), *Handbook of Hydrology*. McGraw-
1553 Hill, New York, pp. 4.1–4.53.
1554

1555 Souza, V.D.A., Roberti, D.R., Ruhoff, A.L., Zimmer, T., Adamatti, D.S., Gonçalves, L.G.G.D., Diaz,
1556 M.B., Alves, R.D.C.M., Moraes, O.L.L.D., 2019. Evaluation of MOD16 algorithm over irrigated rice

1557 paddy using flux tower measurements in Southern Brazil. *Water* 11, 1911.
1558 <https://doi.org/10.3390/w11091911>.
1559
1560 Stoy, P.C., Mauder, M., Foken, T., Marcolla, B., Boegh, E., Ibrom, A., Arain, M.A., Arneth, A., Aurela,
1561 M., Bernhofer, C., Cescatti, A., Dellwik, E., Duce, P., Gianelle, D., Van Gorsel, E., Kiely, G., Knohl, A.,
1562 Margolis, H., McCaughey, H., Merbold, L., Montagnani, L., Papale, D., Reichstein, M., Saunders, M.,
1563 Serrano-Ortiz, P., Sottocornola, M., Spano, D., Vaccari, F., Varlagin, A., 2013. A data-driven analysis of
1564 energy balance closure across FLUXNET research sites: the role of landscape scale heterogeneity.
1565 *Agricultural and Forest Meteorology* 171, 137–152. <http://dx.doi.org/10.1016/j.agrformet.2012.11.004>.
1566
1567 Su, Z., 2002. The Surface Energy Balance System (SEBS) for estimation of turbulent heat fluxes.
1568 *Hydrology and Earth System Sciences* 6, 85–100. <https://doi.org/10.5194/hess-6-85-2002>.
1569
1570 Tang, R., Shao, K., Li, Z.-L., Wu, H., Tang, B.-H., Zhou, G., Zhang, L., 2015. Multiscale validation of
1571 the 8-day MOD16 evapotranspiration product using flux data collected in China. *IEEE Journal of selected*
1572 *topics in applied earth observations and remote sensing* 8, 1478–1486.
1573 <https://doi.org/10.1109/JSTARS.2015.2420105>.
1574
1575 Twine, T.E., Kustas, W.P., Norman, J.M., Cook, D.R., Houser, P.R., Meyers, T.P., Prueger, J.H., Starks,
1576 P.J., Wesely, M.L., 2000. Correcting eddy-covariance flux underestimates over a grassland. *Agricultural*
1577 *and Forest Meteorology* 103, 279–300. [https://doi.org/10.1016/S0168-1923\(00\)00123-4](https://doi.org/10.1016/S0168-1923(00)00123-4).
1578
1579 Uuh-Sonda, J.M., Figueroa-Espinoza, B., Gutiérrez-Jurado, H.A., Méndez-Barroso, L.A., 2022.
1580 Ecosystem productivity and evapotranspiration dynamics of a seasonally dry tropical forest of the
1581 Yucatan Peninsula. *Journal of Geophysical Research Biogeosciences* 127, e2019JG005629.
1582 <https://doi.org/10.1029/2019JG005629>.

1583

1584 Vargas, R., Yépez, E.A., Andrade, J.L., Ángeles, G., Arredondo, T., Castellanos, A.E., Delgado-
1585 Balbuena, J., Garatuza-Payán, J., González del Castillo, E., Oechel, W., Rodríguez, J.C., Sánchez-
1586 Azofeifa, A., Velasco, E., Vivoni, E.R., Watts, C., 2013. Progress and opportunities for monitoring
1587 greenhouse gases fluxes in Mexican ecosystems: the MexFlux network. *Atmósfera* 26, 325–336.
1588 [https://doi.org/10.1016/S0187-6236\(13\)71079-8](https://doi.org/10.1016/S0187-6236(13)71079-8).

1589

1590 Velpuri, N.M., Senay, G.B., Singh, R.K., Bohms, S., Verdin, J.P., 2013. A comprehensive evaluation of
1591 two MODIS evapotranspiration products over the conterminous United States: Using point and gridded
1592 FLUXNET and water balance ET. *Remote Sensing of Environment* 139, 35–49.
1593 <http://dx.doi.org/10.1016/j.rse.2013.07.013>.

1594

1595 Verduzco, V.S., Garatuza-Payán, J., Yépez, E.A., Watts, C.J., Rodríguez, J.C., Robles-Morua, A., Vivoni,
1596 E.R., 2015. Variations of net ecosystem production due to seasonal precipitation differences in a tropical
1597 dry forest of northwest Mexico. *Journal of Geophysical Research Biogeosciences* 120, 2081–2094.
1598 <https://doi.org/10.1002/2015JG003119>.

1599

1600 Verduzco, V.S., Vivoni, E.R., Yépez, E.A., Rodríguez, J.C., Watts, C.J., Tarin, T., Garatuza-Payán, J.,
1601 Robles-Morua, A., Ivanov, V.Y., 2018. Climate change impacts on net ecosystem productivity in a
1602 subtropical shrubland of northwestern México. *Journal of Geophysical Research Biogeosciences* 123,
1603 688–711. <https://doi.org/10.1002/2017JG004361>.

1604

1605 Villarreal, S., Vargas, R., 2021. Representativeness of FLUXNET sites across Latin America. *Journal of*
1606 *Geophysical Research Biogeosciences* 126, e2020JG006090. <https://doi.org/10.1029/2020JG006090>.

1607

1608 Vinukollu, R.V., Wood, E.F., Ferguson, C.R., Fisher, J.B., 2011a. Global estimates of evapotranspiration

1609 for climate studies using multi-sensor remote sensing data: Evaluation of three process-based approaches.
1610 Remote Sensing of Environment 115, 801–823. <https://doi.org/10.1016/j.rse.2010.11.006>.
1611
1612 Vinukollu, R.V., Meynadier, R., Sheffield, J., Wood, E.F., 2011b. Multi-model, multi-sensor estimates of
1613 global evapotranspiration: climatology, uncertainties and trends. Hydrological Processes 25, 3993–4010.
1614 <https://doi.org/10.1002/hyp.8393>.
1615
1616 Wan, Z., Zhang, Y., Zhang, Q., Li, Z.-L., 2004. Quality assessment and validation of the MODIS global
1617 land surface temperature. International Journal of Remote Sensing 25 (1), 261–274.
1618 <https://doi.org/10.1080/0143116031000116417>.
1619
1620 Wang, Z.H., Bou-Zeid, E., 2012. A novel approach for the estimation of soil ground heat flux.
1621 Agricultural and Forest Meteorology 154–155, 214–221. <https://doi.org/10.1016/j.agrformet.2011.12.001>.
1622
1623 Wang, K., Dickinson, R.E., 2012. A review of global terrestrial evapotranspiration: Observation,
1624 modeling, climatology, and climatic variability. Reviews of Geophysics 50, RG2005.
1625 <https://doi.org/10.1029/2011RG000373>.
1626
1627 Willmott, C.J., 1982. Some comments on the evaluation of model performance. Bulletin of the American
1628 Meteorological Society 63 (11), 1309–1313. [https://doi.org/10.1175/1520-](https://doi.org/10.1175/1520-0477(1982)063<1309:SCOTEO>2.0.CO;2)
1629 [0477\(1982\)063<1309:SCOTEO>2.0.CO;2](https://doi.org/10.1175/1520-0477(1982)063<1309:SCOTEO>2.0.CO;2).
1630
1631 Wilson, K., Goldstein, A., Falge, E., Aubinet, M., Baldocchi, D., Berbigier, P., Bernhofer, C., Ceulemans,
1632 R., Dolman, H., Field, C., Grelle, A., Ibrom, A., Law, B.E., Kowalski, A., Meyers, T., Moncrieff, J.,
1633 Monson, R., Oechel, W., Tenhunen, J., Valentini, R., Verma, S., 2002. Energy balance closure at
1634 FLUXNET sites. Agricultural and Forest Meteorology 113 (1–4), 223–243.

1635 [https://doi.org/10.1016/S0168-1923\(02\)00109-0](https://doi.org/10.1016/S0168-1923(02)00109-0).

1636

1637 Wolf, S., Eugster, W., Potvin, C., Buchmann, N., 2011. Strong seasonal variations in net ecosystem CO₂

1638 exchange of a tropical pasture and afforestation in Panama. *Agricultural and Forest Meteorology* 151,

1639 1139–1151. <https://doi.org/10.1016/j.agrformet.2011.04.002>.

1640

1641 Yilmaz, M.T., Anderson, M.C., Zaitchik, B., Hain, C.R., Crow, W.T., Ozdogan, M., Chun, J.A., Evans,

1642 J., 2014. Comparison of prognostic and diagnostic surface flux modeling approaches over the Nile River

1643 basin. *Water Resources Research* 50 (1), 386–408. <https://doi.org/10.1002/2013WR014194>.

1644

1645 Yu, G.-R., Wen, X.-F., Sun, X.-M., Tanner, B.D., Lee, X., Chen, J.-Y., 2006. Overview of ChinaFLUX

1646 and evaluation of its eddy covariance measurement. *Agricultural and Forest Meteorology* 137, 125–137.

1647 <https://doi.org/10.1016/j.agrformet.2006.02.011>.

1648

1649 Zhang, K.[Ke], Kimball, J.S., Running, S.W., 2016. A review of remote sensing based actual

1650 evapotranspiration estimation. *WIREs Water* 3, 834–853. <https://doi.org/10.1002/wat2.1168>.

1651

1652 Zhang, K.[Kun], Zhu, G., Ma, J., Yang, Y., Shang, S., Gu, C., 2019. Parameter analysis and estimates for

1653 the MODIS evapotranspiration algorithm and multiscale verification. *Water Resources Research* 55,

1654 2211–2231. <https://doi.org/10.1029/2018WR023485>.

1655

1656 Zhang, Y.[Yongqiang], Peña-Arancibia, J.L., McVicar, T.R., Chiew, F.H.S., Vaze, J., Liu, C., Lu, X.,

1657 Zheng, H., Wang, Y., Liu, Y.Y., Miralles, D.G., Pan, M., 2016. Multi-decadal trends in global terrestrial

1658 evapotranspiration and its components. *Scientific Reports* 6, 19124. <https://doi.org/10.1038/srep19124>.

1659

1660 Zhang, Y.[Yongqiang], Kong, D., Gan, R., Chiew, F.H.S., McVicar, T.R., Zhang, Q., Yang, Y., 2019.

1661 Coupled estimation of 500 m and 8-day resolution global evapotranspiration and gross primary
1662 production in 2002–2017. *Remote Sensing of Environment* 222, 165–182.
1663 <https://doi.org/10.1016/j.rse.2018.12.031>.

1664

1665

CRediT authorship contribution statement

Diego Salazar-Martínez: Formal analysis, Investigation, Visualization, Writing - Original Draft.

Friso Holwerda: Conceptualization, Data curation, Formal analysis, Investigation, Methodology, Supervision, Writing - Original Draft, Writing - Review & Editing.

Thomas R.H. Holmes: Conceptualization, Investigation, Writing - Review & Editing.

Enrico A. Yépez: Investigation, Resources, Writing - Review & Editing.

Christopher R. Hain: Investigation, Writing - Review & Editing.

Susana Alvarado-Barrientos: Investigation, Writing - Review & Editing.

Rest of the authors (alphabetical): Investigation.

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

March 24, 2022, Mexico City, Mexico

Dear Dr Anagnostou,

We are pleased to submit the revision of the manuscript entitled “Evaluation of remote sensing-based evapotranspiration products at low-latitude eddy covariance sites” for publication in Journal of Hydrology.

The manuscript has been submitted and reviewed before by Journal of Hydrology under the manuscript number HYDROL43166_R1. The present submission is the revised version of this manuscript. We also provide point-by-point replies to the reviewers comments.

We look very much forward to your response.

Sincerely,



Dr. Friso Holwerda (corresponding author)

Instituto de Ciencias de la Atmósfera y Cambio Climático
Universidad Nacional Autónoma de México
Circuito de la Investigación s/n
Ciudad Universitaria
Coyoacan, 04510
México D.F.
México
E-mail address: friso.holwerda@gmail.com
Phone: 0052-55-56224088

Figure 1

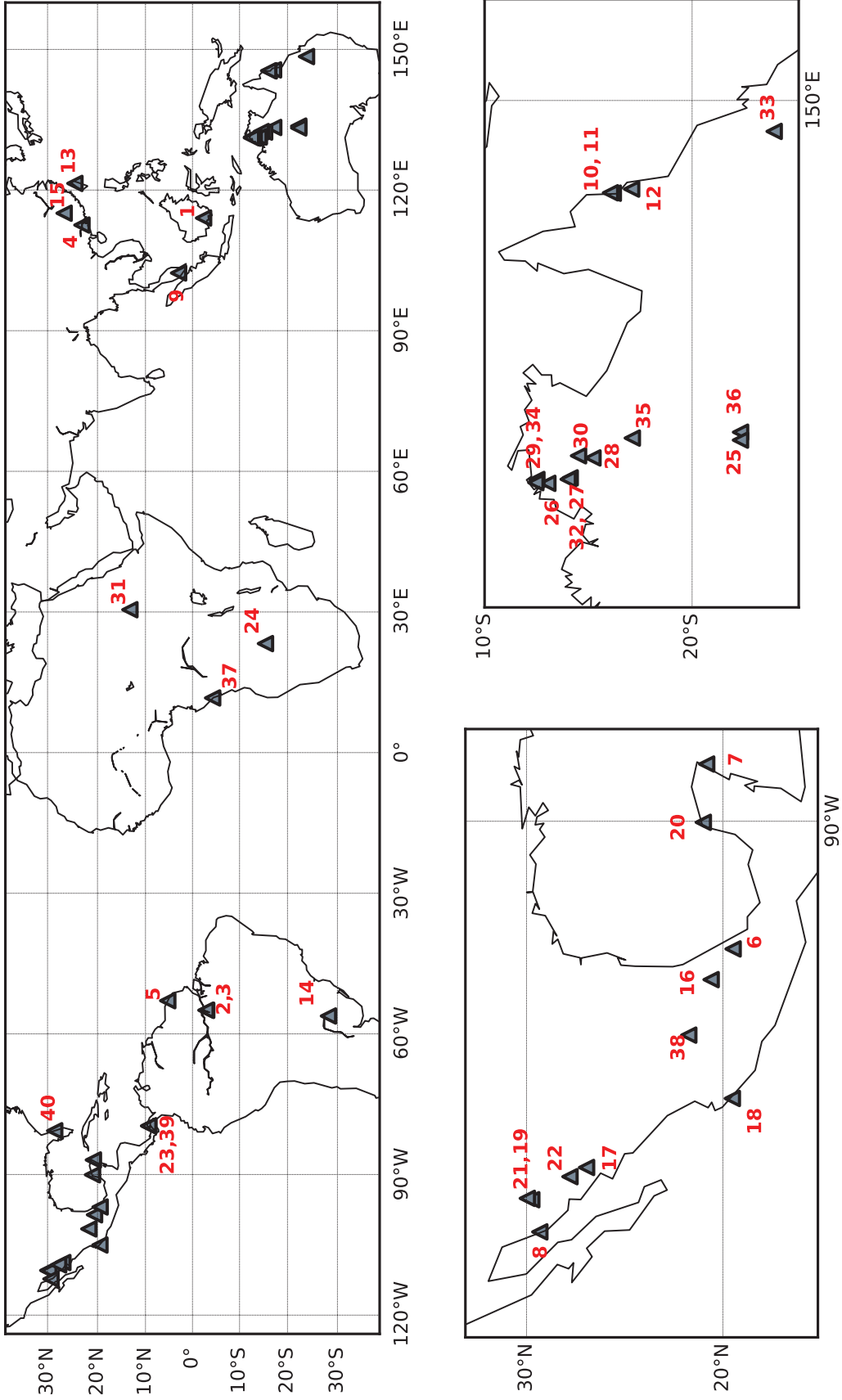


Figure 2

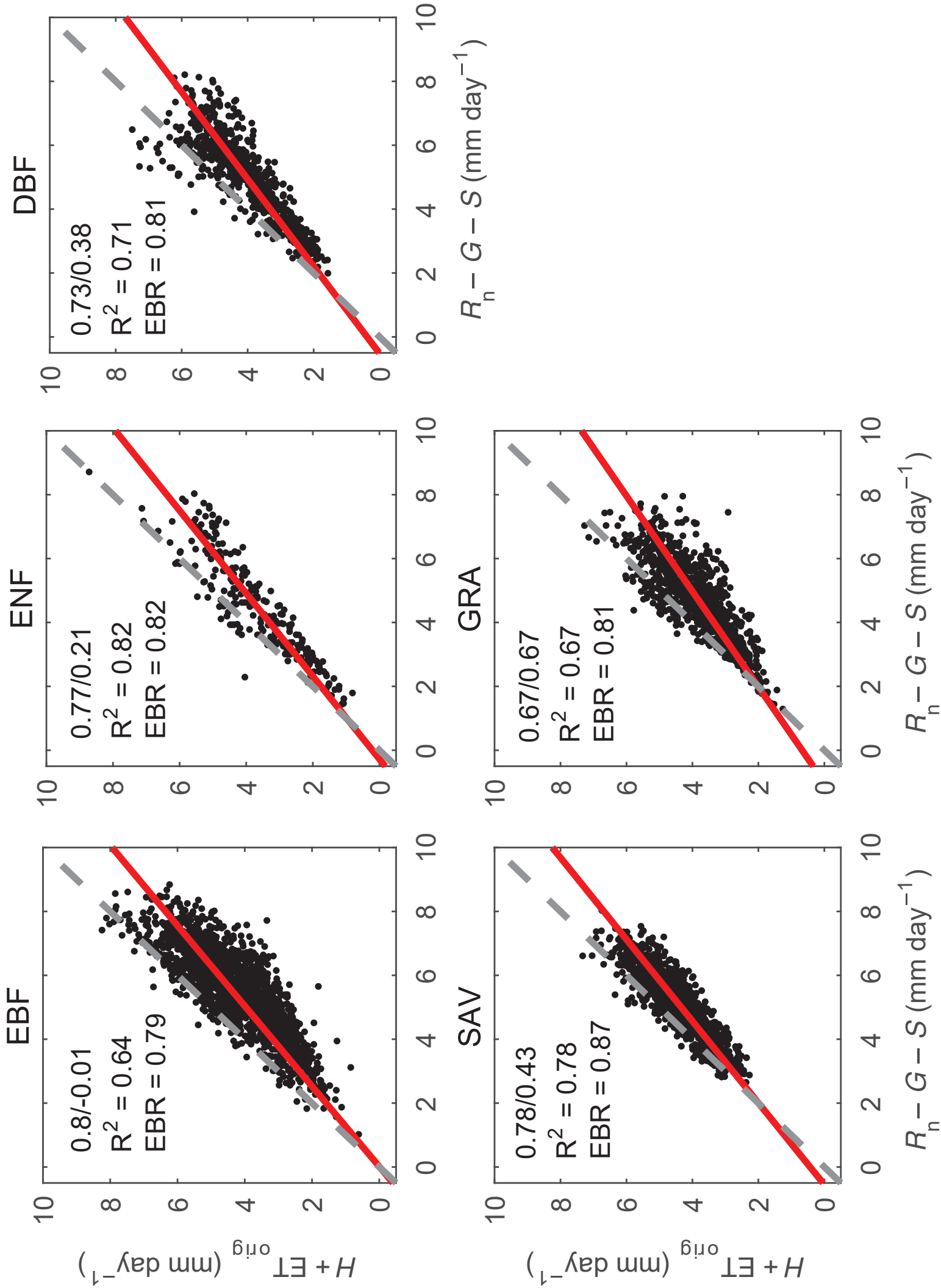


Figure 3

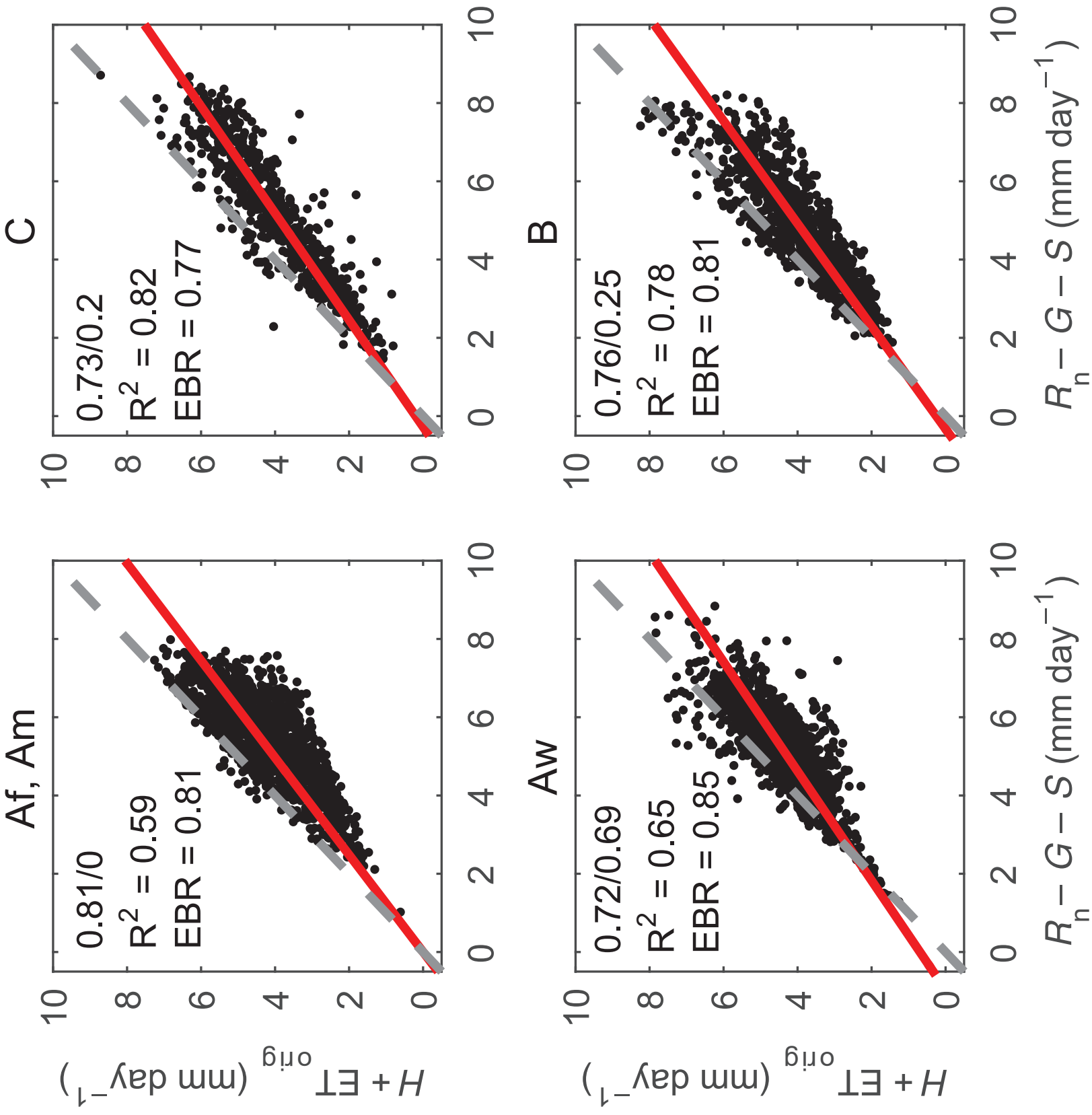


Figure 4

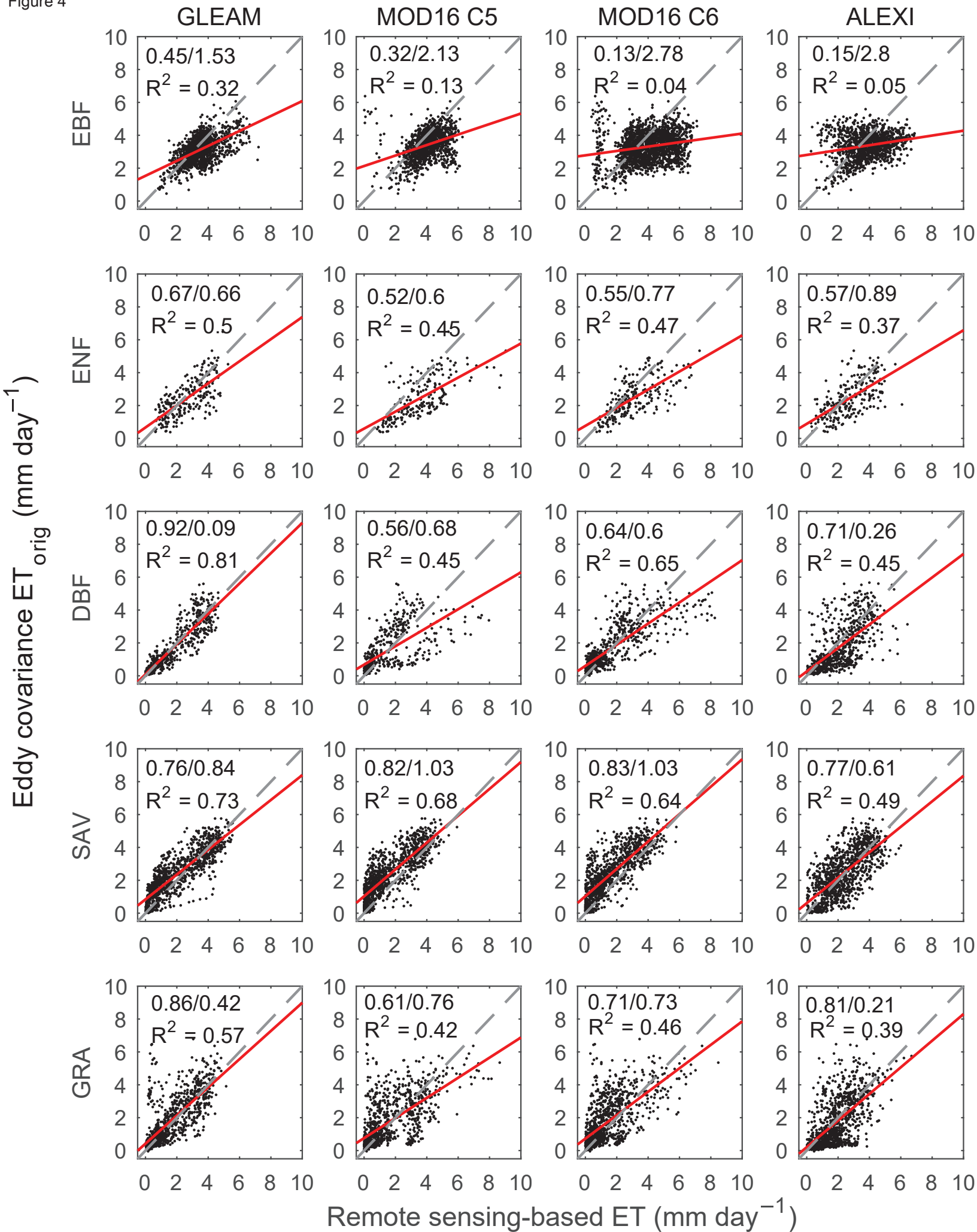


Figure 5

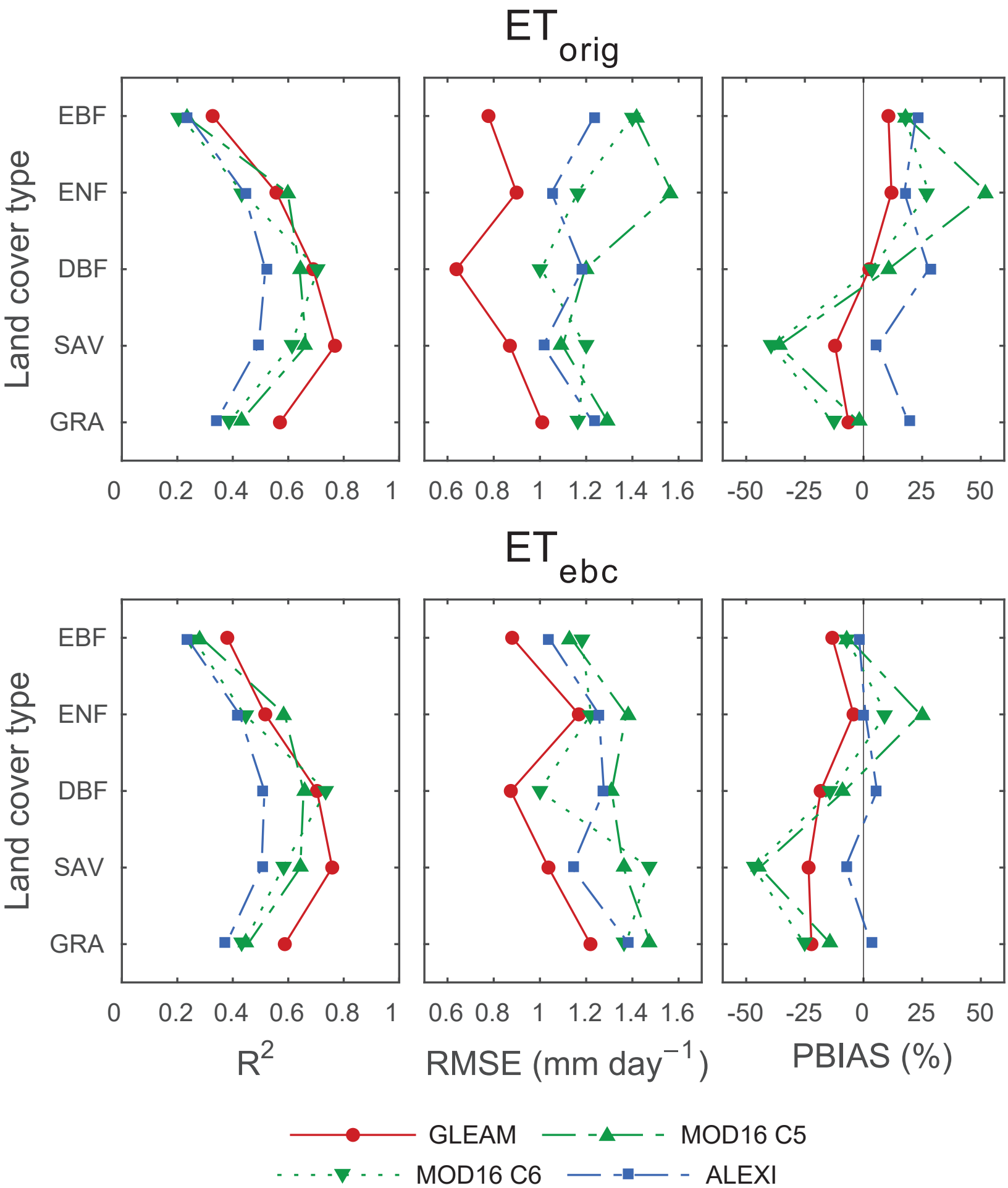
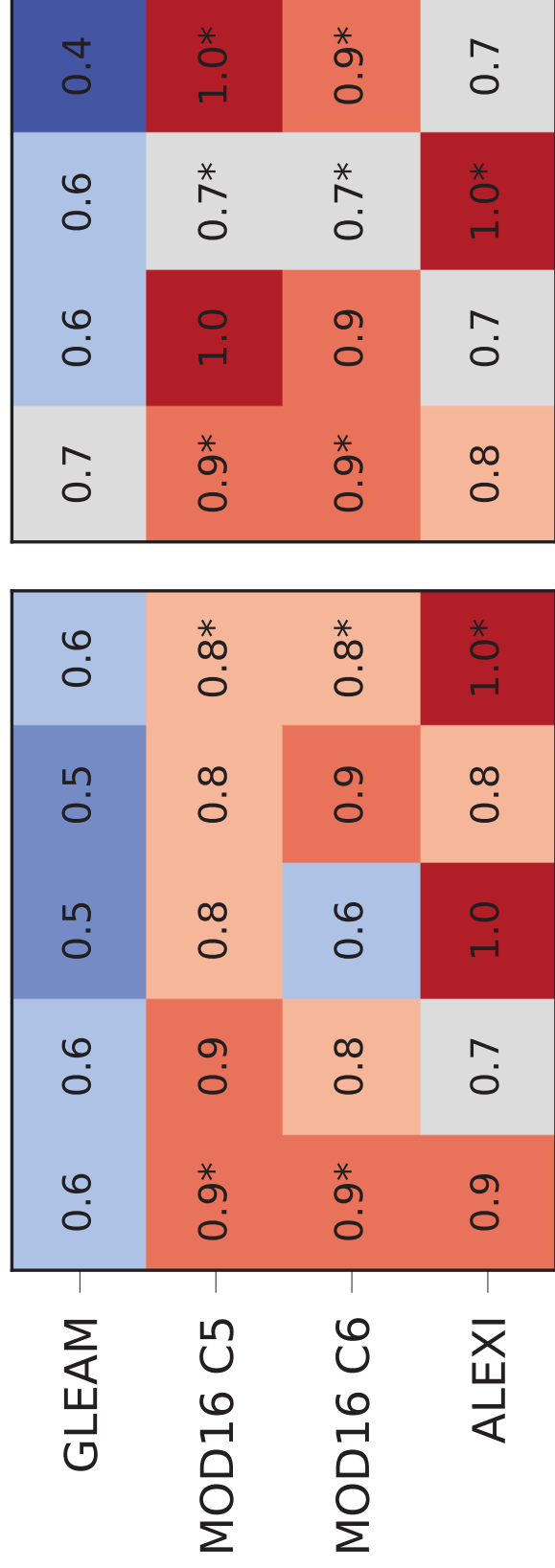


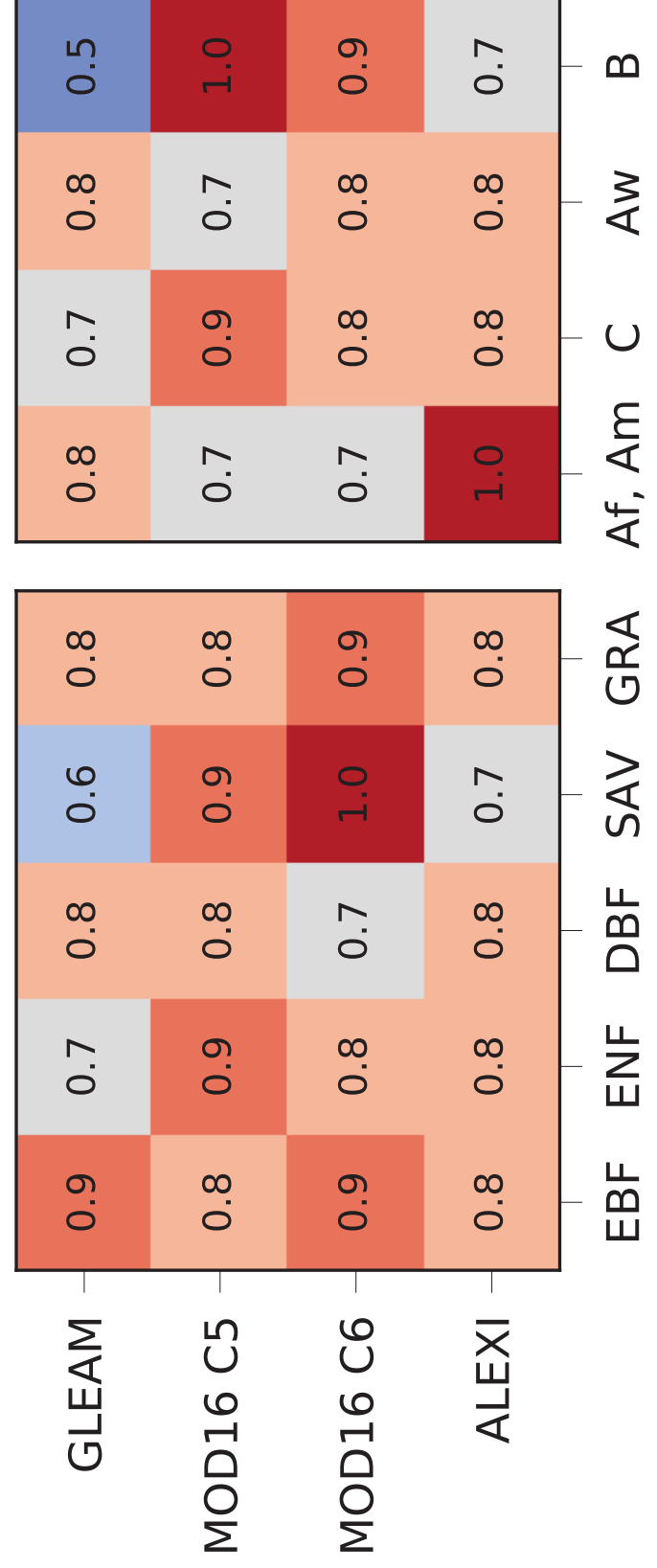
Figure 6

IPE

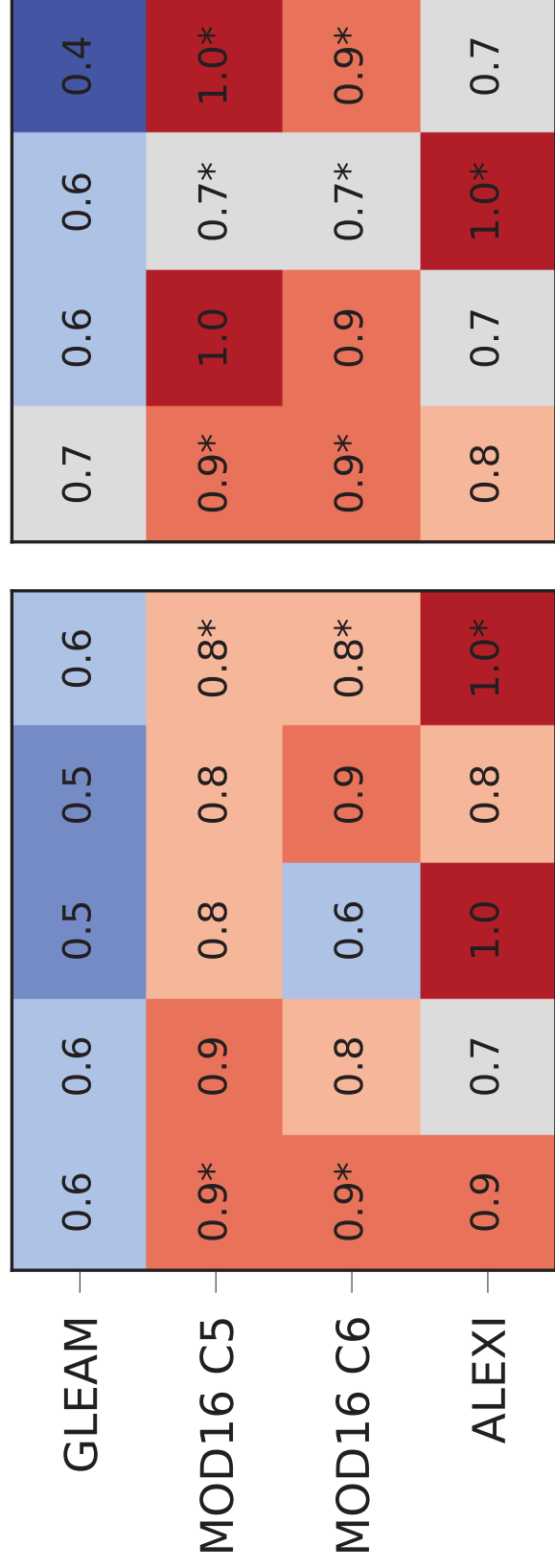
ET_{orig}



ET_{ebc}



ET_{orig}



ET_{ebc}

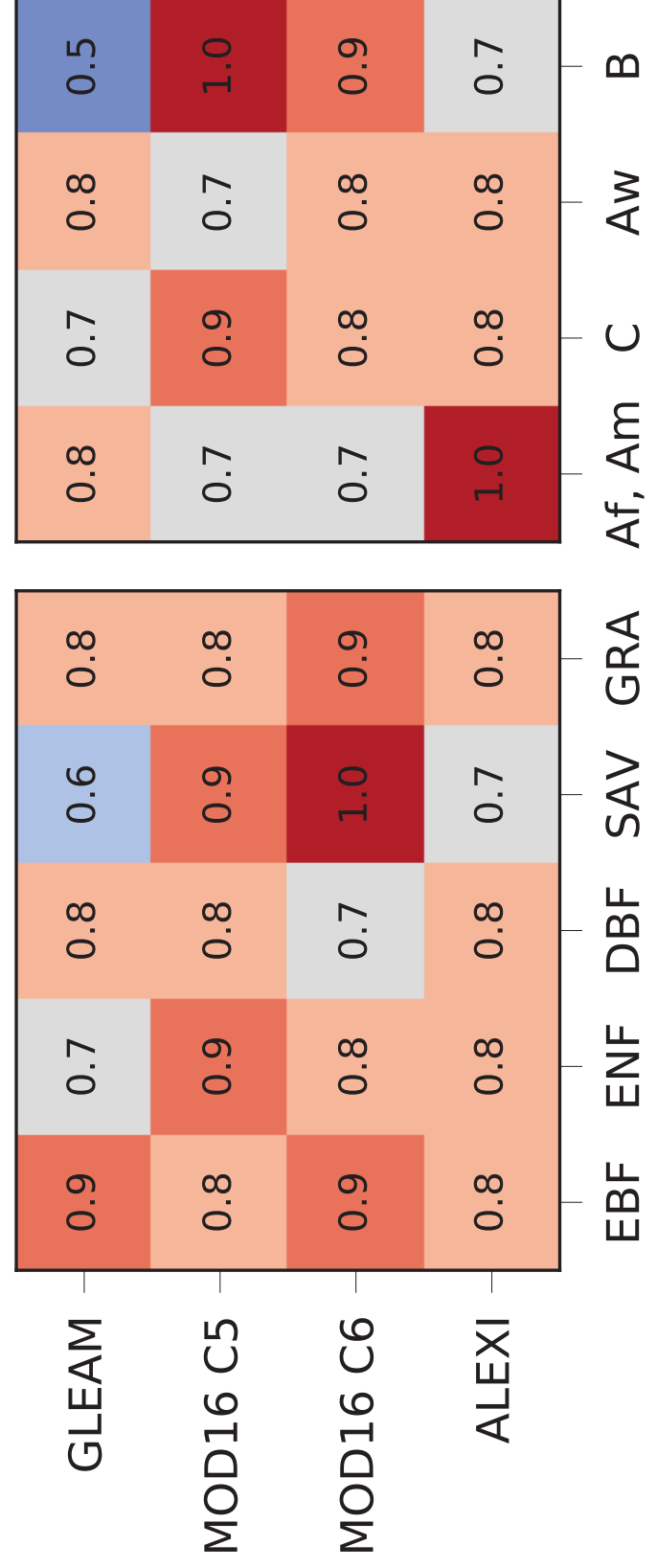


Figure 7

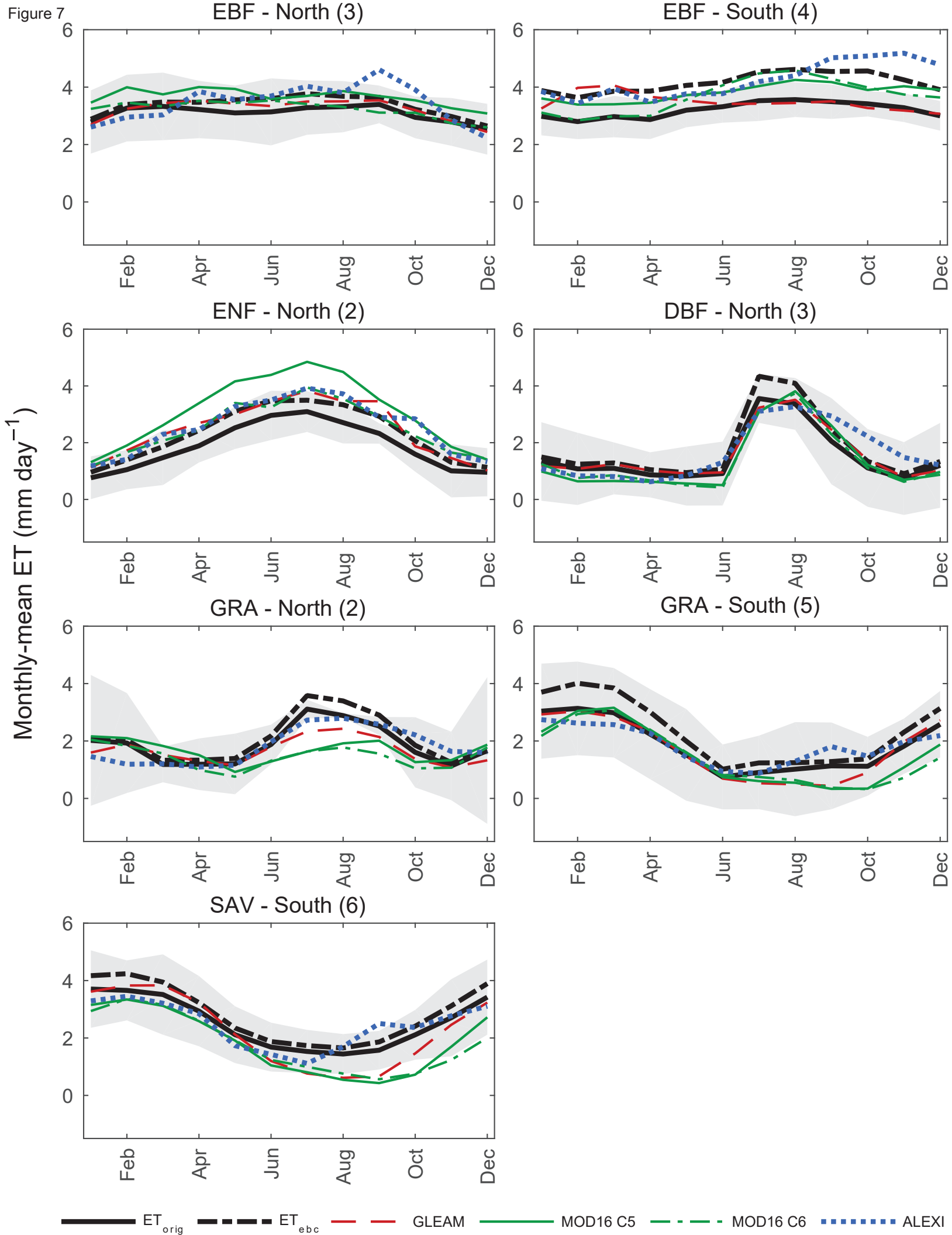


Figure 8

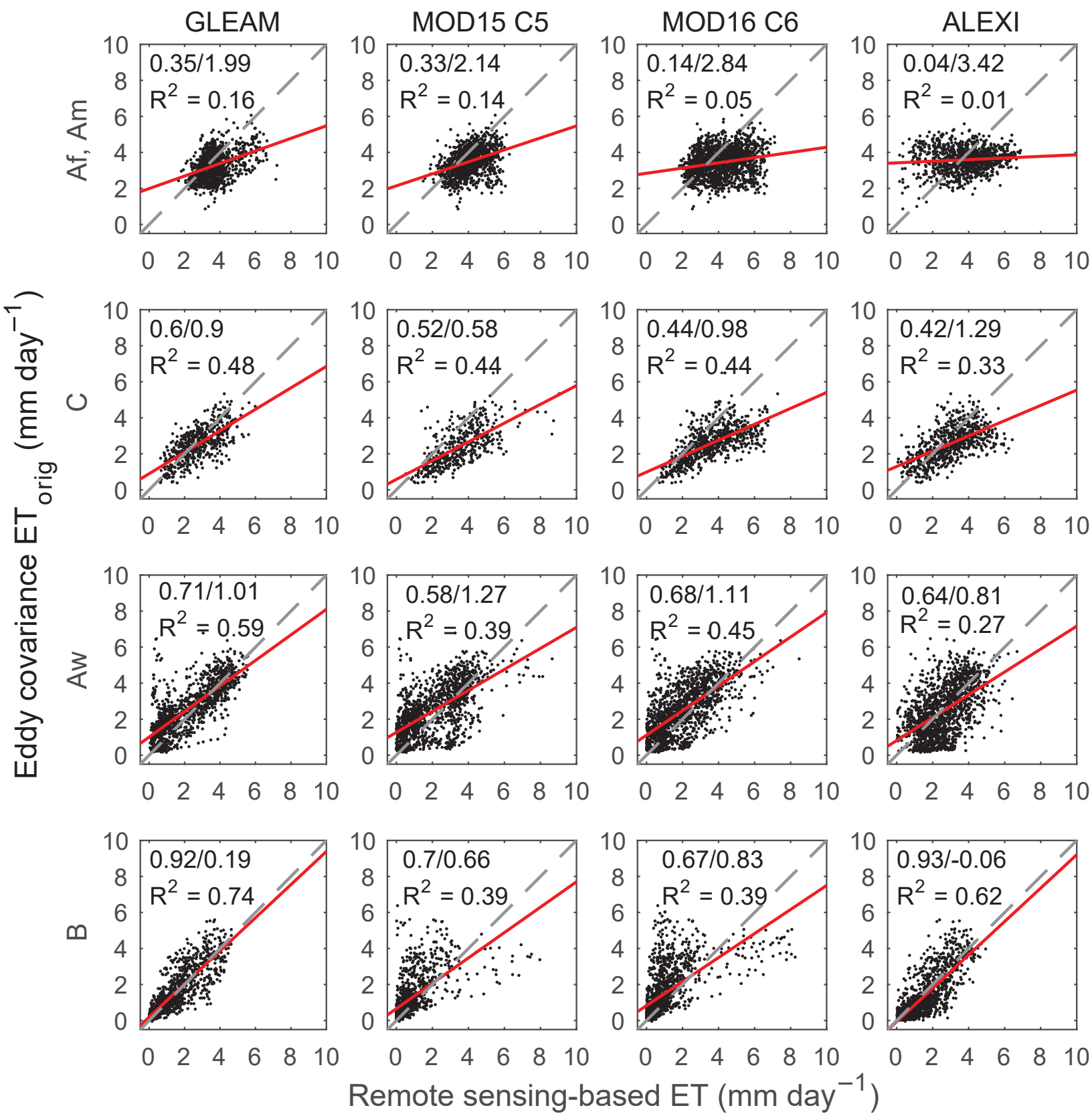
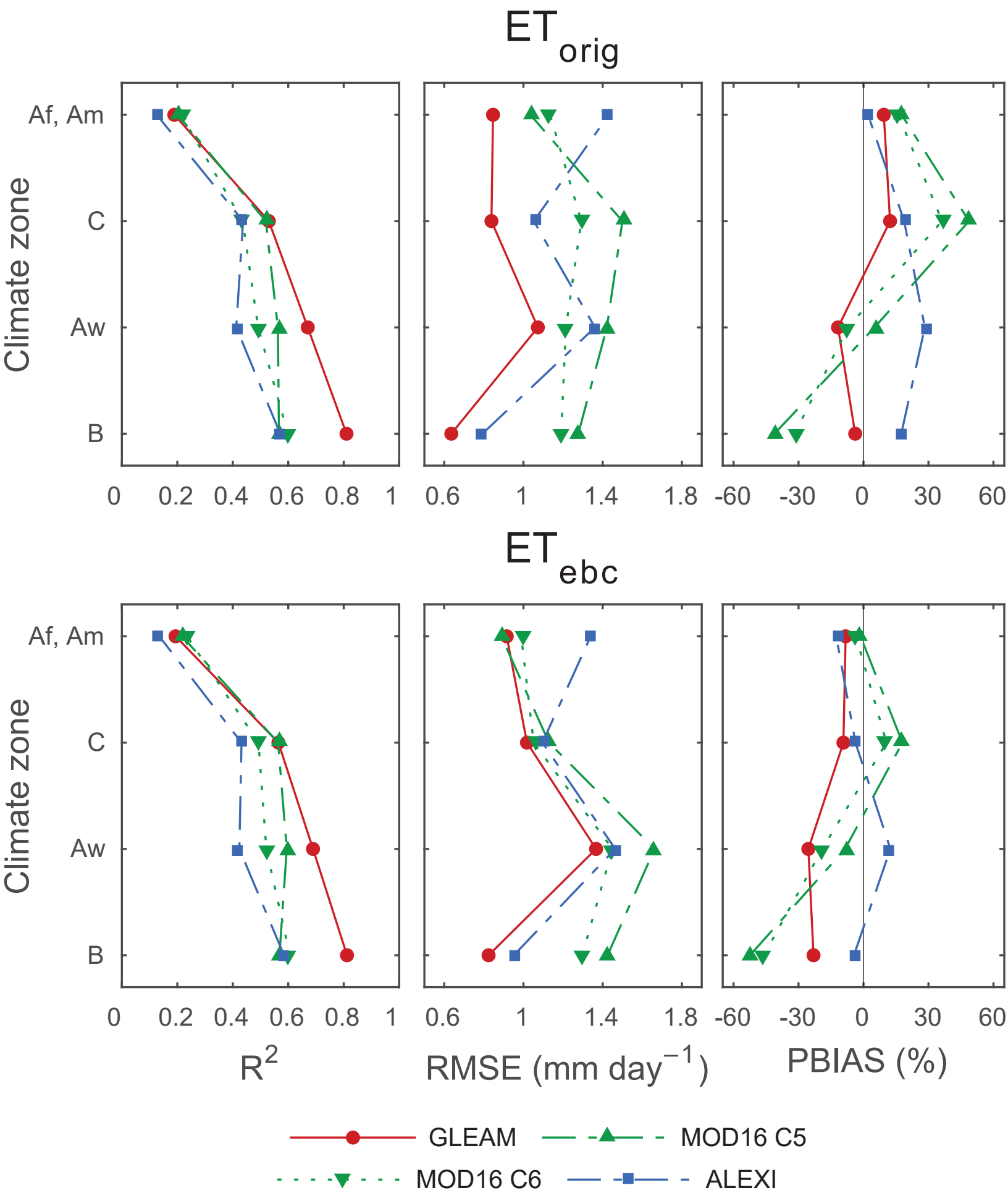
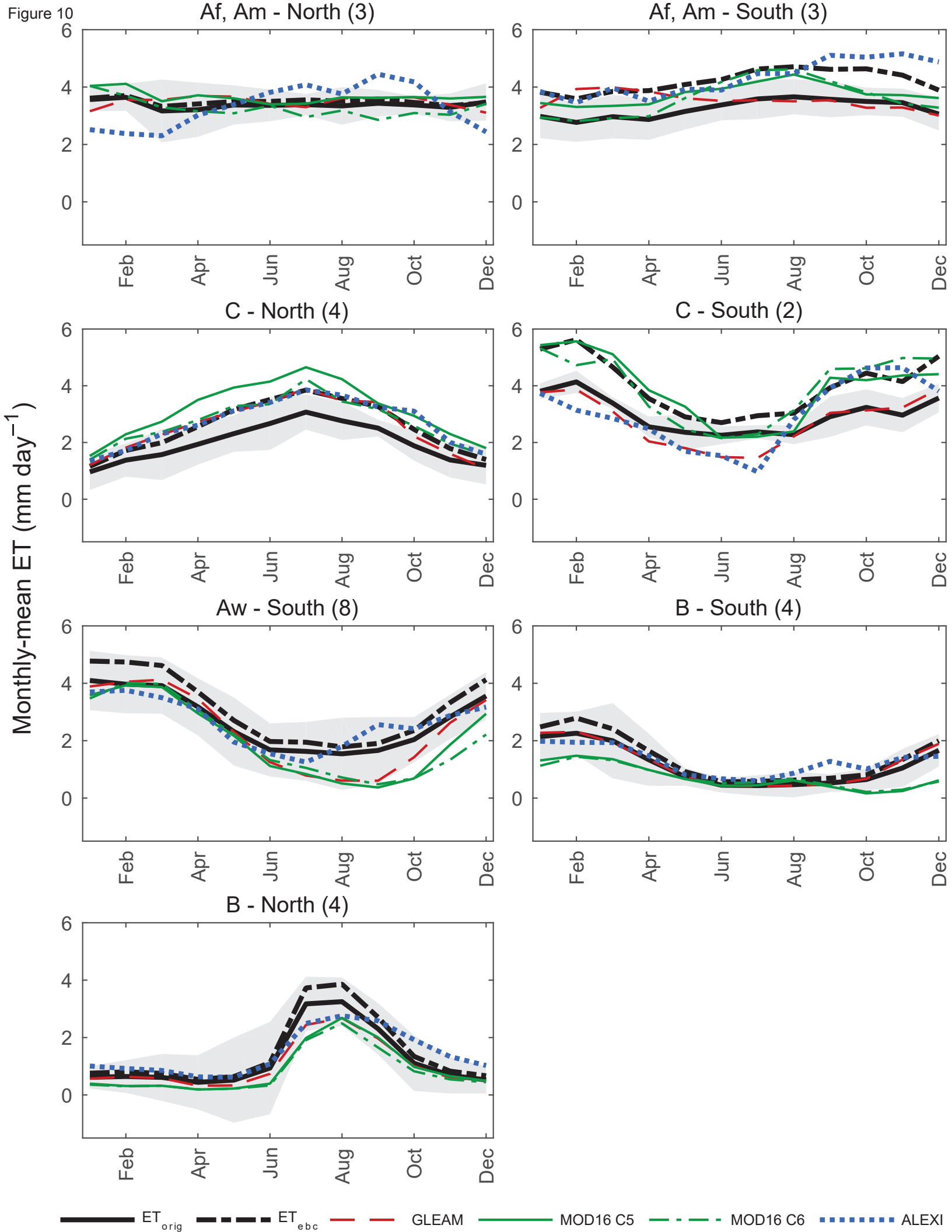


Figure 9





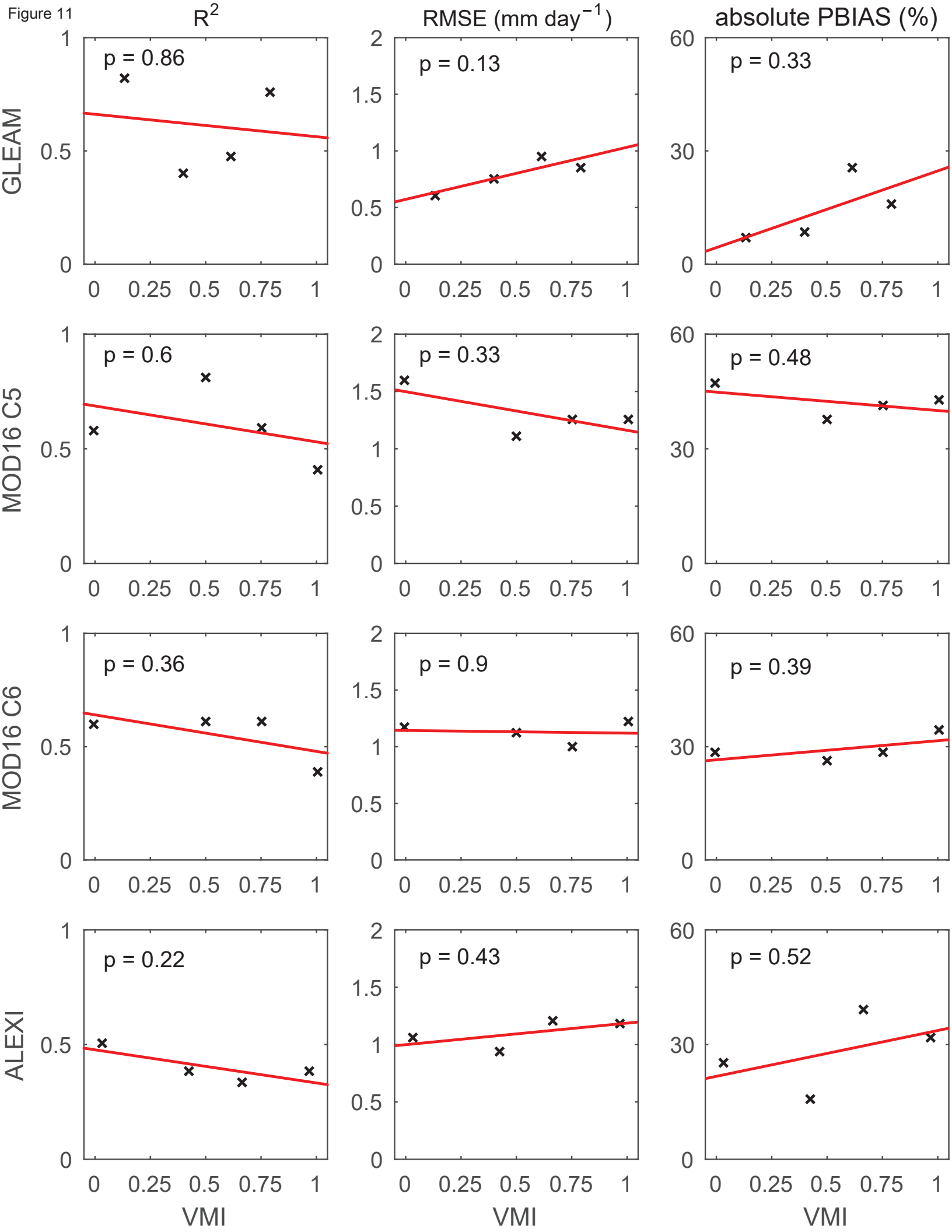
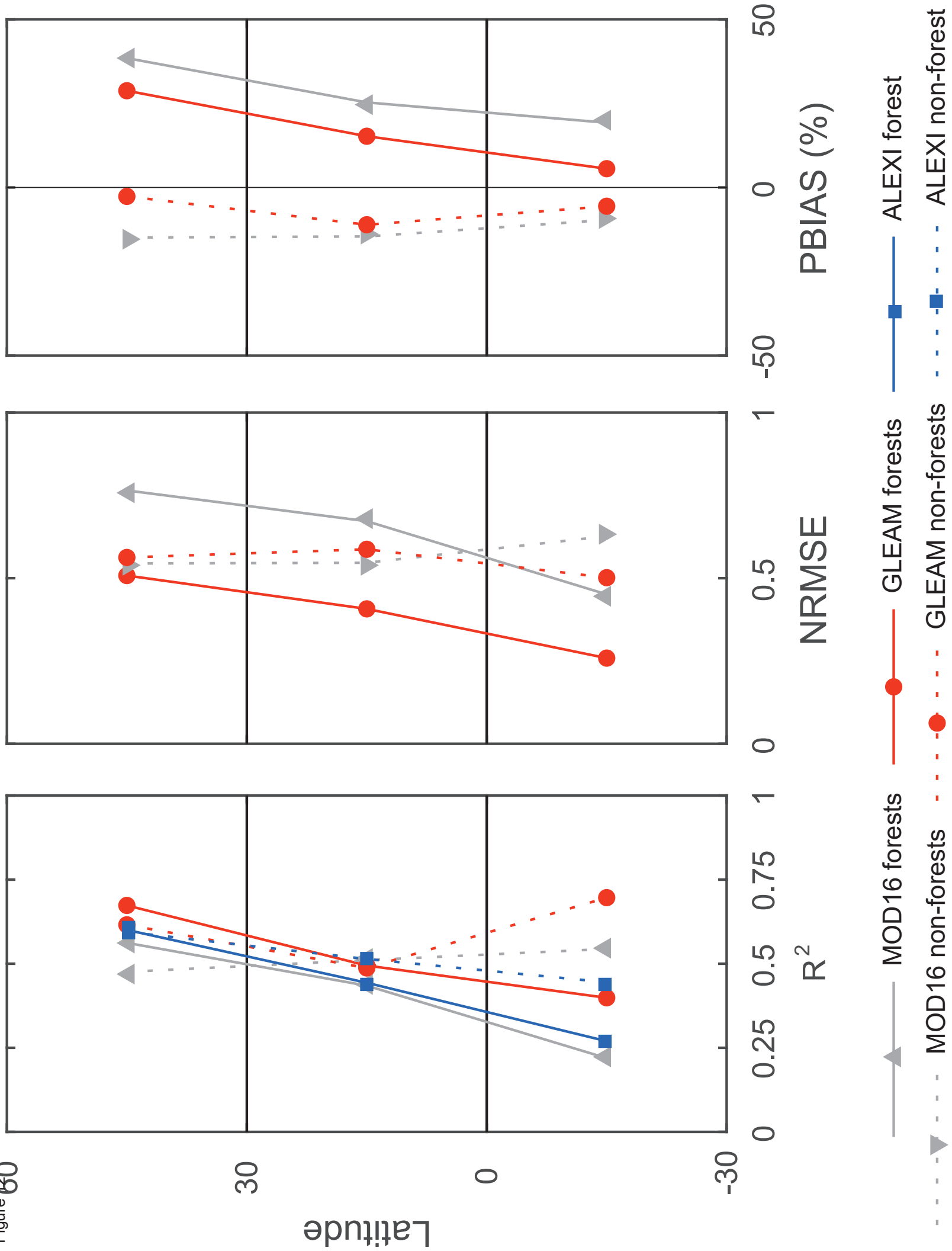


Figure 60



- 1 Fig. 1. Map showing the geographic location of the 40 eddy covariance sites used in the evaluation,
2 zoomed in for Mexico and northern Australia. The numbers identify the sites in Table 2.
3
- 4 Fig. 2. Scatter plots of daytime sums of sensible heat flux (H) and evapotranspiration (ET_{orig}) versus
5 available energy ($R_n - G - S$; all terms in units of millimetres) for different land cover types for the eddy
6 covariance sites used in the evaluation. Shown are the regression slope (value before the slash), the
7 intercept (value after the slash), the coefficient of determination (R^2), the energy balance ratio (EBR),
8 the linear regression line (solid red line), and the 1:1 line (dashed line).
9
- 10 Fig. 3. Scatter plots of daytime sums of sensible heat flux (H) and evapotranspiration (ET_{orig}) versus
11 available energy ($R_n - G - S$; all terms in units of millimetres) for different climate zones for the eddy
12 covariance sites used in the evaluation. Shown are the regression slope (value before the slash), the
13 intercept (value after the slash), the coefficient of determination (R^2), the energy balance ratio (EBR),
14 the linear regression line (solid red line), and the 1:1 line (dashed line).
15
- 16 Fig. 4. Unadjusted eddy covariance ET observations (ET_{orig}) versus remote sensing-based ET for each
17 land cover type for each of the evaluated products. Shown are the regression slope (value before the
18 slash), the intercept (value after the slash), the coefficient of determination (R^2), the linear regression
19 line (solid red line), and the 1:1 line (dashed line).
20
- 21 Fig. 5. Mean performance statistics (R^2 , RMSE, PBIAS) by land cover type for each of the evaluated
22 products for the unadjusted eddy covariance ET observations (ET_{orig}) and those corrected for the lack
23 of energy balance closure (ET_{ebc}).
24

25 Fig. 6. Heat maps of the Ideal Point Error (IPE) for each of the evaluated products for each of the
26 comparisons by land cover type and climate zone for the unadjusted eddy covariance ET observations
27 (ET_{orig}) and those corrected for the lack of energy balance closure (ET_{ebc}). The IPE values are shown on
28 the plot. The lower the IPE, the better the relative performance of the product. Blue/red colors indicate
29 best/worst IPE scores. The asterisks in the heatmap for ET_{orig} indicate where the ranking of a product
30 differed from that based on the IPE scores for the common reference dataset (Figure S3; Section 2.2.4).

31

32 Fig. 7. Average monthly ET for the four ET products together with the average monthly unadjusted ET
33 observations (ET_{orig}) and those corrected for the lack of energy balance closure (ET_{ebc}) for different
34 land cover types in the northern and southern hemispheres. Curves were calculated using the common
35 reference dataset. Only land cover-hemisphere combinations for which data from at least two sites were
36 available are shown. The number of sites in each land cover-hemisphere combination is given between
37 parentheses. The error band represents the standard deviation of the mean monthly ET_{orig} at the
38 different sites.

39

40 Fig. 8. Unadjusted eddy covariance ET observations (ET_{orig}) versus remote sensing-based ET for each
41 climate zone for each of the evaluated products. Shown are the regression slope (value before the
42 slash), the intercept (value after the slash), the coefficient of determination (R^2), the linear regression
43 line (solid red line), and the 1:1 line (dashed line).

44

45 Fig. 9. Mean performance statistics (R^2 , RMSE, PBIAS) by climate zone for each of the evaluated
46 products for the unadjusted eddy covariance ET observations (ET_{orig}) and those corrected for the lack
47 of energy balance closure (ET_{ebc}).

48

49 Fig. 10. Average monthly ET for the four ET products together with the average monthly unadjusted
50 ET observations (ET_{orig}) and those corrected for the lack of energy balance closure (ET_{ebc}) for different
51 climate zones in the northern and southern hemispheres. Curves were calculated using the common
52 reference dataset. Only climate zone-hemisphere combinations for which data from at least two sites
53 were available are shown. The number of sites in each climate zone-hemisphere combination is given
54 between parentheses. The error band represents the standard deviation of the mean monthly ET_{orig} at
55 the different sites.

56

57 Fig. 11. Binned scatter plots between the performance metrics (R^2 , RMSE, PBIAS) and the vegetation
58 match index for each of the evaluated products. Shown are the regression lines and the p -values
59 indicating the statistical significance of the regression slopes.

60

61 Fig. 12. Zonal averages (southern and northern low latitudes and northern middle latitudes) of the
62 performance metrics grouped by forest and non-forest vegetation for MOD16 and GLEAM (R^2 ,
63 NRMSE, PBIAS) and ALEXI (only R^2). Averages were calculated using evaluation results from this
64 study and from the literature. See Section 2.4 for further details.

65

66

1 Table 1. General characteristics of the remote sensing-based ET products evaluated in this study.

| | GLEAM v3.3a | MOD16 C5 | MOD16 C6 | ALEXI |
|---------------------|---|------------------------|---|--|
| Spatial resolution | 0.25° | 1 km | 500 m | 0.05° |
| Temporal resolution | daily | 8-day | 8-day | daily |
| Temporal coverage | 1980–2018 | 2000–2014 | 2000–present | 2002–2019 |
| Principle | Priestley-Taylor | Penman-Monteith | Penman-Monteith | Two-source energy balance |
| Public access | yes | yes | yes | no |
| References | Miralles et al. (2011); Martens et al. (2017) | Mu et al. (2011, 2013) | Mu et al. (2011); Running et al. (2019) | Anderson et al. (1997, 2007, 2011); Hain and Anderson (2007) |

- 2 Table 2. Eddy covariance sites used in the evaluation of the remote sensing-based ET products. Shown for each site are: number to locate site on map
- 3 in Figure 1; site ID or site name used by the flux network; latitude and longitude (decimal degrees); period with data availability; flux network;
- 4 country; IGBP land cover type; Köppen-Geiger climate class; evaporative fraction (EF); energy balance ratio (EBR); site elevation (m); and
- 5 reference(s) to article(s) with additional information or to dataset.

| Site ID or site name | Lat. | Lon. | Data period | Network | Country | Land cover | Climate class | EF | EBR | Elev. (m) | Reference |
|----------------------|--------|---------|-------------|----------|---------------|------------|---------------|------|------|-----------|--|
| 1 PDF | -2.35 | 114.03 | 2002-2005 | AsiaFlux | Indonesia | EBF | Af | 0.77 | 0.82 | 30 | Hirano et al. (2015) |
| 2 BR-Sa1 | -2.86 | -54.96 | 2002-2011 | FLUXNET | Brazil | EBF | Am | 0.76 | 0.71 | 88 | Saleska (2002-2011) |
| 3 BR-Sa3 | -3.02 | -54.97 | 2000-2004 | FLUXNET | Brazil | EBF | Am | 0.77 | 0.82 | 100 | Goulden (2000-2004) |
| 4 CN-Din | 23.17 | 112.54 | 2003-2005 | FLUXNET | China | EBF | Cwa | 0.60 | 0.68 | 300 | Yu et al. (2006) |
| 5 GF-Guy | 5.28 | -52.92 | 2004-2014 | FLUXNET | French Guiana | EBF | Am | 0.75 | 0.85 | 48 | Bonal et al. (2008) |
| 6 La Orduña | 19.47 | -96.93 | 2014-2018 | MexFlux | Mexico | EBF | Cfa | 0.56 | 0.82 | 1210 | Holwerda et al. (2016); Holwerda and Meesters (2019) |
| 7 Puerto Morelos | 20.85 | -86.90 | 2017-2018 | MexFlux | Mexico | EBF | Aw | 0.51 | 0.75 | 10 | Alvarado-Barrientos et al. (2021) |
| 8 El Sargento | 29.34 | -112.28 | 2014-2016 | MexFlux | Mexico | EBF | BWh | 0.59 | 0.99 | 0 | Delgado-Balbuena et al. (2018) |
| 9 MY-PSO | 2.97 | 102.31 | 2003-2009 | FLUXNET | Malaysia | EBF | Af | 0.64 | 0.98 | 112 | MY-PSO (2003-2009) |
| 10 Cape Tribulation | -16.10 | 145.45 | 2012-2018 | OzFlux | Australia | EBF | Am | 0.72 | 0.77 | 40 | Liddell (2013) |
| 11 Cow Bay | -16.24 | 145.43 | 2011-2019 | OzFlux | Australia | EBF | Am | 0.72 | 0.63 | 86 | Liddell (2013) |
| 12 Robson Creek | -17.12 | 145.63 | 2014-2019 | OzFlux | Australia | EBF | Cwa | 0.61 | 0.71 | 710 | Liddell (2013) |
| 13 CLM | 24.59 | 121.42 | 2007-2009 | AsiaFlux | Taiwan | ENF | Cfb | 0.50 | 0.86 | 1638 | Chu et al. (2014) |
| 14 AR-Vir | -28.24 | -56.19 | 2009-2012 | FLUXNET | Argentina | ENF | Cfa | 0.79 | 0.72 | 127 | Posse et al. (2016) |
| 15 CN-Qia | 26.74 | 115.06 | 2003-2004 | FLUXNET | China | ENF | Cfa | 0.70 | 0.76 | 100 | Yu et al. (2006) |
| 16 Atopixco | 20.61 | -98.59 | 2017-2018 | MexFlux | Mexico | ENF | Cwb | 0.43 | 0.99 | 2064 | Hidalgo-Sánchez et al. (2021) |
| 17 Alamos | 27.00 | -108.79 | 2015-2017 | MexFlux | Mexico | DBF | BSh | 0.57 | 0.71 | 368 | Rojas-Robles et al. (2020) |
| 18 Chamela | 19.51 | -105.04 | 2007-2010 | MexFlux | Mexico | DBF | Aw | 0.30 | 0.82 | 73 | González del Castillo et al. (2018) |
| 19 Sierra Los Locos | 29.96 | -110.46 | 2010-2014 | MexFlux | Mexico | DBF | BSh | 0.42 | 0.73 | 1314 | Pérez-Ruiz et al. (2021) |

| | | | | | | | | | | | | |
|----|-----------|--------|---------|-----------|---------|---------------|-----|-----|------|------|------|---|
| 20 | El Palmar | 21.02 | -90.06 | 2016–2018 | MexFlux | Mexico | DBF | Aw | 0.40 | 1.03 | 8 | Figueroa-Espinoza et al. (2021); Uuh-Sonda et al. (2022) |
| 21 | Rayón | 29.74 | -110.53 | 2008–2015 | MexFlux | Mexico | DBF | BSh | 0.23 | 0.81 | 632 | Verduzco et al. (2018); Pérez-Ruiz et al. (2021) |
| 22 | Tesopaco | 27.83 | -109.28 | 2004–2008 | MexFlux | Mexico | DBF | BSh | 0.31 | 0.76 | 426 | Verduzco et al. (2015) |
| 23 | PA-SPn | 9.32 | -79.63 | 2007–2009 | FLUXNET | Panama | DBF | Am | 0.66 | 0.91 | 78 | Wolf et al. (2011) |
| 24 | ZM-Mon | -15.44 | 23.25 | 2007–2009 | FLUXNET | Zambia | DBF | Aw | 0.45 | 0.77 | 1053 | Merbold et al. (2009) |
| 25 | AU-ASM | -22.28 | 133.25 | 2010–2014 | FLUXNET | Australia | SAV | BWh | 0.16 | 0.83 | 600 | Cleverly et al. (2013) |
| 26 | AU-Ade | -13.08 | 131.12 | 2007–2009 | FLUXNET | Australia | SAV | Aw | 0.58 | 0.96 | 100 | Beringer et al. (2011) |
| 27 | AU-DaS | -14.16 | 131.39 | 2008–2014 | FLUXNET | Australia | SAV | Aw | 0.51 | 0.90 | 110 | Hutley et al. (2011) |
| 28 | AU-Dry | -15.26 | 132.37 | 2009–2014 | FLUXNET | Australia | SAV | Aw | 0.46 | 0.80 | 175 | Cernusak et al. (2011) |
| 29 | AU-How | -12.49 | 131.15 | 2003–2014 | FLUXNET | Australia | WSA | Aw | 0.63 | 0.89 | 64 | Beringer et al. (2007) |
| 30 | AU-RDF | -14.56 | 132.48 | 2011–2013 | FLUXNET | Australia | SAV | Aw | 0.40 | 0.87 | 171 | Bristow et al. (2016) |
| 31 | SD-Dem | 13.28 | 30.48 | 2007–2009 | FLUXNET | Sudan | SAV | BWh | 0.40 | 0.83 | 500 | Ardö et al. (2008) |
| 32 | AU-DaP | -14.06 | 131.32 | 2007–2013 | FLUXNET | Australia | GRA | Aw | 0.45 | 0.74 | 67 | Hutley et al. (2011) |
| 33 | AU-Emr | -23.86 | 148.47 | 2011–2013 | FLUXNET | Australia | GRA | BSh | 0.32 | 0.67 | 170 | Schroder et al. (2014) |
| 34 | AU-Fog | -12.55 | 131.31 | 2006–2008 | FLUXNET | Australia | GRA | Aw | 0.72 | 0.80 | 4 | Beringer et al. (2013) |
| 35 | AU-Stp | -17.15 | 133.35 | 2008–2014 | FLUXNET | Australia | GRA | BSh | 0.33 | 0.82 | 225 | Beringer et al. (2011) |
| 36 | AU-TTE | -22.29 | 133.64 | 2012–2014 | FLUXNET | Australia | GRA | BWh | 0.12 | 1.03 | 553 | Cleverly et al. (2016) |
| 37 | CG-Tch | -4.29 | 11.66 | 2006–2009 | FLUXNET | Congo | GRA | Aw | 0.48 | 0.98 | 82 | Merbold et al. (2009) |
| 38 | Ojuelos | 21.78 | -101.61 | 2011–2016 | MexFlux | Mexico | GRA | BSk | 0.41 | 0.83 | 2228 | Delgado-Balbuena et al. (2019) |
| 39 | PA-SPs | 9.31 | -79.63 | 2007–2009 | FLUXNET | Panama | GRA | Am | 0.67 | 0.96 | 68 | Wolf et al. (2011) |
| 40 | US-KS2 | 28.61 | -80.67 | 2003–2006 | FLUXNET | United States | CSH | Cfa | 0.59 | 0.81 | 3 | Drake and Hinkle (2003–2006) |

7 Table 3. Number of eddy covariance sites and site years (between parentheses) available in the
 8 complete dataset for each product by land cover type and climate zone. Note that the number of
 9 site years corresponds to the length of the flux tower records. Actual data availability was lower
 10 due to, for example, missing or erroneous data.

| | | GLEAM | MOD16 C5 | MOD16 C6 | ALEXI |
|-----------------|--------|---------|----------|----------|---------|
| Land cover type | EBF | 9 (49) | 11 (47) | 12 (67) | 8 (47) |
| | ENF | 4 (7) | 3 (6) | 4 (7) | 4 (7) |
| | DBF | 6 (23) | 5 (18) | 8 (27) | 8 (27) |
| | SAV | 7 (33) | 7 (33) | 7 (33) | 7 (33) |
| | GRA | 7 (28) | 8 (29) | 8 (31) | 8 (31) |
| Climate zone | Af, Am | 8 (42) | 9 (46) | 9 (55) | 7 (39) |
| | Aw | 8 (37) | 10 (43) | 12 (46) | 11 (45) |
| | C | 7 (19) | 7 (15) | 8 (24) | 8 (24) |
| | B | 10 (42) | 9 (33) | 11 (45) | 10 (42) |

11
 12

13 Table 4. Number of evaluation results (N_{ER}) from this study and from the literature, broken down
 14 into results for R^2 , NRMSE and PBIAS, and grouped by latitudinal zone, ET product and
 15 vegetation category (see Section 2.4 for further explanation). NA is not Not Available.

| | | MOD16 | | GLEAM | | ALEXI | |
|-------------|-------|--------|------------|--------|------------|--------|------------|
| | | Forest | Non-forest | Forest | Non-forest | Forest | Non-forest |
| 30° N–60° N | R^2 | 19 | 30 | 2 | 9 | 27 | 26 |
| | NRMSE | 11 | 23 | 2 | 9 | NA | NA |
| | PBIAS | 11 | 23 | 2 | 9 | NA | NA |
| 0°–30° N | R^2 | 15 | 9 | 16 | 4 | 14 | 5 |
| | NRMSE | 13 | 9 | 16 | 4 | 14 | 4 |
| | PBIAS | 13 | 9 | 16 | 4 | 14 | 4 |
| 0°–30° S | R^2 | 9 | 16 | 8 | 11 | 6 | 19 |
| | NRMSE | 8 | 16 | 8 | 11 | 6 | 12 |
| | PBIAS | 8 | 16 | 8 | 11 | 6 | 12 |

16
 17