

Graphical Abstract

Estimating groundwater use and demand in arid Kenya through assimilation of satellite data and in-situ sensors with machine learning toward drought early action

Katie Fankhauser, Denis Macharia, Jeremy Coyle, Styvers Kathuni, Amy McNally, Kimberly Slinski, Evan Thomas

DRIP Groundwater Use

Year
2021

County
All

Measurement
Multiple Use Water Service (MUS) [1] (35+ L/pc/day)

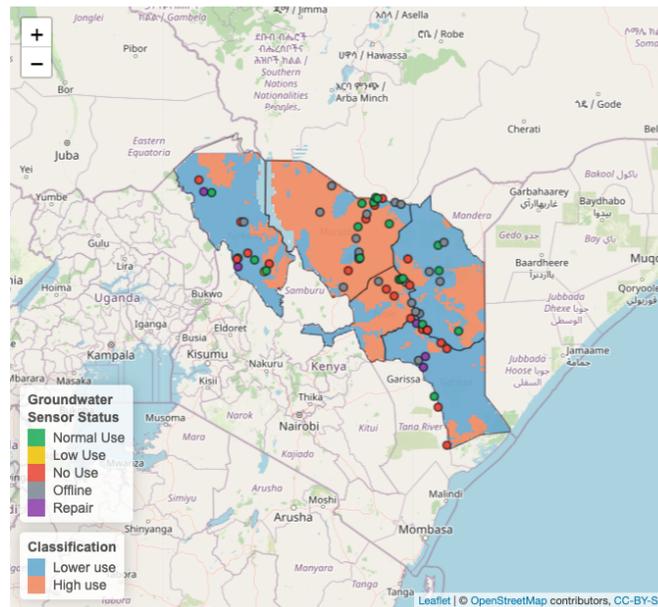
Prediction
Classification of high use

Display
Magnitude

Show current sensor status [2]
 Show projected food security [3]

Month
Jun 2021

[Return to Summary](#)



Expected accuracy: 74%

[1] <https://www.musgroup.net/>
[2] <https://sweetsensors.com/>
[3] <https://fewes.net/>

Highlights

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- Supporting local groundwater management would improve resilience to drought.
- Kenya groundwater use estimated with sensors, satellite data, and machine learning.
- Historical use was modeled with up to 75% cross-validated accuracy.
- Forecasts for the 2021 dry season indicated up to 80% external validation accuracy.

Estimating groundwater use and demand in arid Kenya through assimilation of satellite data and in-situ sensors with machine learning toward drought early action

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Abstract

Groundwater is an important source of water for people, livestock, and agriculture during drought in the Horn of Africa. In this work, areas of high groundwater use and demand in drought-prone Kenya were identified and forecasted prior to the dry season. Estimates of groundwater use were extended from a sentinel network of 69 in-situ sensed mechanical boreholes to the region with satellite data and a machine learning model. The sensors contributed 756 site-month observations from June 2017 to September 2021 for model building and validation at a density of approximately one sensor per $3,700 \text{ km}^2$. An ensemble of 19 parameterized algorithms was informed by features including satellite-derived precipitation, surface water availability, vegetation indices, hydrologic land surface modeling, and site characteristics to dichotomize high groundwater pump utilization. Three operational definitions of high demand on groundwater infrastructure were considered: 1) mechanical runtime of pumps greater than a quarter of a day (6+ hrs) and daily per capita volume extractions indicative of 2) domestic water needs (35+ L), and 3) intermediate needs including livestock (75+ L). Gridded interpolation of localized groundwater use and demand was provided from 2017 to 2020 and forecasted for the 2021 dry season, June - September 2021. Cross-validated skill for contemporary estimates of daily pump runtime and daily volume extraction to meet domestic and intermediate water needs was 68%, 69%, and 75%, respectively. Forecasts were externally validated with an accuracy of at least 56%, 70%, 72% for each groundwater use definition. The groundwater maps are accessible to stakeholders including the Kenya National Drought Management Authority (NDMA) and the Famine Early Warning Systems Network (FEWS NET). These maps represent the first operational spatially-explicit sub-seasonal to seasonal (S2S) estimates of groundwater use and demand in the literature. Knowledge of historical and forecasted groundwater use is anticipated to improve decision-making and resource allocation for a range of early warning early action applications.

Keywords: drought, groundwater, early warning, early action, machine learning, remote sensing, Kenya

1. Introduction

Drought is one of the most persistent, expansive, and damaging of natural disasters. Coupled with weak institutional and civil systems, experience of drought leads to regional food insecurity, water insecurity, disease, and violent conflict for

5 billions of people globally (Wilhite and Glantz, 1985). The Horn of Africa is sus-
6 ceptible to natural exposure and societal vulnerability to drought (Liebmann et al.,
7 2017; NDMA, 2015). As a slow onset disaster, public and private organizations may
8 not effectively mobilize resources in response (Cabot Venton, 2018). An advantage,
9 however, is that the risk of drought can be monitored and forecasted in advance,
10 allowing for the implementation of early warning early action systems (Funk and
11 Shukla, 2020).

12 Drought early warning is an established field with existing indices that rely
13 on climatological and land surface modeling or observing causal factors. For ex-
14 ample, the United States National Integrated Drought Information System (NI-
15 DIS) [drought.gov] produces current condition and 3-month drought outlooks from
16 weather and climate, soil moisture, streamflow, and rainfall and snowpack data in-
17 formed by qualitative expert opinion. In lower resource settings, such as Sub-Saharan
18 Africa, observational data is sparse and satellite-derived and land surface modeled
19 estimates of climate and the hydrologic cycle become the main inputs in drought
20 monitoring. The Famine Early Warning Systems Network (FEWS NET) [few.net]
21 makes the link between drought and food insecurity in low- and middle-income coun-
22 tries explicit.

23 Drought develops from a deficit of precipitation and water storage inadequate
24 to support livelihoods. The four domains of drought – meteorological, hydrological,
25 agricultural, and socioeconomic – compound from an initial lack of rainfall and low-
26 ered water availability to adverse impacts on land productivity and livestock and,
27 ultimately, to water and food insecurity from diminished household returns (Wilhite
28 and Glantz, 1985). The primary rainfall and agricultural season in East Africa has
29 seen a marked decrease in precipitation in recent decades (Nicholson et al., 2018),
30 leading to slower and lower recharge of surface water and diminished flow rates in
31 large rivers. Moreover, the increase in surface temperatures from climate change,
32 generates higher evapotranspiration, aggravating the decline in precipitation and
33 intensifying aridification.

34 The arid and semi-arid lands (ASALs) of Kenya have faced regular drought since
35 at least 2016 with below average rainfall placing 18 million people at risk (UNICEF,
36 2017; KNBS, 2019). During times of low precipitation and decreased surface water
37 capacity, groundwater becomes a critical source of water for domestic, livestock, and
38 agricultural needs (Thomas et al., 2019). Pastoralist communities, in particular, are
39 dependent on predictable access to water, without which they are likely to continue to
40 experience inequitable rates of poverty and poor health (NDMA, 2015). Functional
41 and strategic groundwater distribution, then, may lower exposure and vulnerability
42 to drought for affected populations.

43 For the region, a number of databases are available that address water supply
44 (Hofste et al., 2019; Senay et al., 2013) and demand (Hofste et al., 2019; McNally
45 et al., 2019; Senay et al., 2015). Although these data products often incorporate
46 some information about groundwater, they rely on Gravity Recovery and Climate
47 Experiment (GRACE) satellite data at low resolution (10000 km² - 150000 km²)
48 (Landerer and Swenson, 2012) or mechanistic hydrological modeling without other
49 primary data. AQUASTAT (FAO, 2021) is a rich resource of water data, but it
50 is only aggregated from annual questionnaires in tabulated form by country every 5
51 years, limiting its usability at finer spatial scales and its relevance to real- or near-time
52 response. Therefore, a geospatially explicit dataset that gives estimates of local (<50
53 square kilometers) and acute (sub-seasonal) realized groundwater demand represents
54 an important missing indicator of water resource availability and withdrawal that
55 would improve interpretation of other water and drought indices in the region.

56 The Drought Resilience Impact Platform (DRIP) combines early detection with
57 proactive groundwater management to ensure water availability with various stake-
58 holders and decision support tools (Thomas et al., 2021). A principal feature of
59 the platform is remote monitoring of a sentinel network of mechanical groundwa-
60 ter boreholes in drought-prone regions of East Africa. The sensors have been used
61 to identify non-functionality (Thomas et al., 2021), improve water service delivery
62 and increase water infrastructure uptime (Nagel et al., 2015; Wilson et al., 2017),
63 demonstrate an inverse relationship between rainfall and borehole use (Thomas et al.,
64 2019), and examine a framework of groundwater management for drought resilience
65 (Turman-Bryant et al., 2019). The current study explicitly addresses how monitoring
66 data from the sensors can be used to quantify groundwater and improve delivery of
67 drought early warning services and actions for the entire region.

68 Although groundwater is an essential water source, especially during the dry sea-
69 son, during the 2016-2017 drought, 55% of the pumps needed to access groundwater
70 were non-functional in Kenya (UNICEF, 2017). The ability to direct limited re-
71 sources to repair, maintain, or site the most critically needed boreholes based on pro-
72 jected use prior to drought would enable responsible stewardship of water resources
73 and community resilience to drought. Interpreting use within known limitations of
74 existing infrastructure may also help direct water resources other than groundwater,
75 such as emergency water trucking. Furthermore, observing and predicting trends in
76 groundwater use could be an important indicator of developing drought itself.

77 By leveraging additional remote sensing and in-situ data sources, we developed
78 gridded estimates for historical and forecasted groundwater use in five prominent
79 ASAL counties in Kenya. A statistical machine learning model extrapolates recorded
80 borehole operation to locations that are not instrumented. The final product catego-

81 rizes areas into high vs. low groundwater use based on definitions of pump runtime
82 and per capita volume extraction. This paper describes development of the DRIP
83 groundwater service and presents the first gridded maps of acute groundwater use
84 and demand in this context. It is expected to inform ongoing local management and
85 planning for water and food insecurity during drought in Kenya and demonstrate
86 this for East Africa.

87 **2. Methods**

88 *2.1. Context*

89 This work builds upon efforts of the Kenya Resilient Arid Lands Partnership
90 for Integrated Development (Kenya RAPID) program. In-situ data was collected at
91 all mechanized groundwater pumps identified as critical drought infrastructure and
92 localized groundwater estimates were developed for five ASAL counties: Garissa,
93 Isiolo, Marsabit, Turkana, and Wajir (Figure 1). Collectively the five counties cover
94 approximately 260,000 square kilometers - or nearly half of the total surface area of
95 Kenya - and suffer from water and food insecurity, high poverty rates, and limited
96 access to basic services (NDMA, 2015). Despite several moderate to high produc-
97 tivity sedimentary aquifers underlying the area, only 17% of renewable groundwater
98 resources are being utilized (Mumma et al., 2011). The ASALs are expected to ex-
99 perience more frequent and greater duration dry periods, likely exacerbating existing
100 disadvantage in the absence of sustainable development and dedicated drought risk
101 reduction.

102 The Government of Kenya has committed to ending drought emergencies (EDE)
103 by 2022. The National Drought Management Authority (NDMA), established in
104 2011, was directed to address drought proactively by confronting vulnerability and
105 structural causes (NDMA, 2015). Cooperation between local county water officials
106 and NDMA officials led to the selection of strategic boreholes for drought response
107 that obtain financial and operational support. These boreholes in the five program-
108 matic counties received sensor installation and monitoring. Previously, we observed
109 that the strategic boreholes were more likely to be non-functional and susceptible to
110 demand elasticity (Turman-Bryant et al., 2019). Drought triggers the mobilization
111 of financial and material resources for maintenance of NDMA priority boreholes and
112 the execution of contingency plans. In years where precipitation is higher, national
113 and local resources for assessing drought risk and repair are withheld or reallocated,
114 leading to an increase in the number and duration of malfunctions. The capacity
115 to use strategic boreholes to mitigate future droughts is diminished by a smaller
116 inventory of operating sites and increasing strain on remaining infrastructure. This

117 suggests that resources should consistently be directed to critical boreholes and to
118 areas with restricted surface water availability or where demonstrable demand is
119 already high.

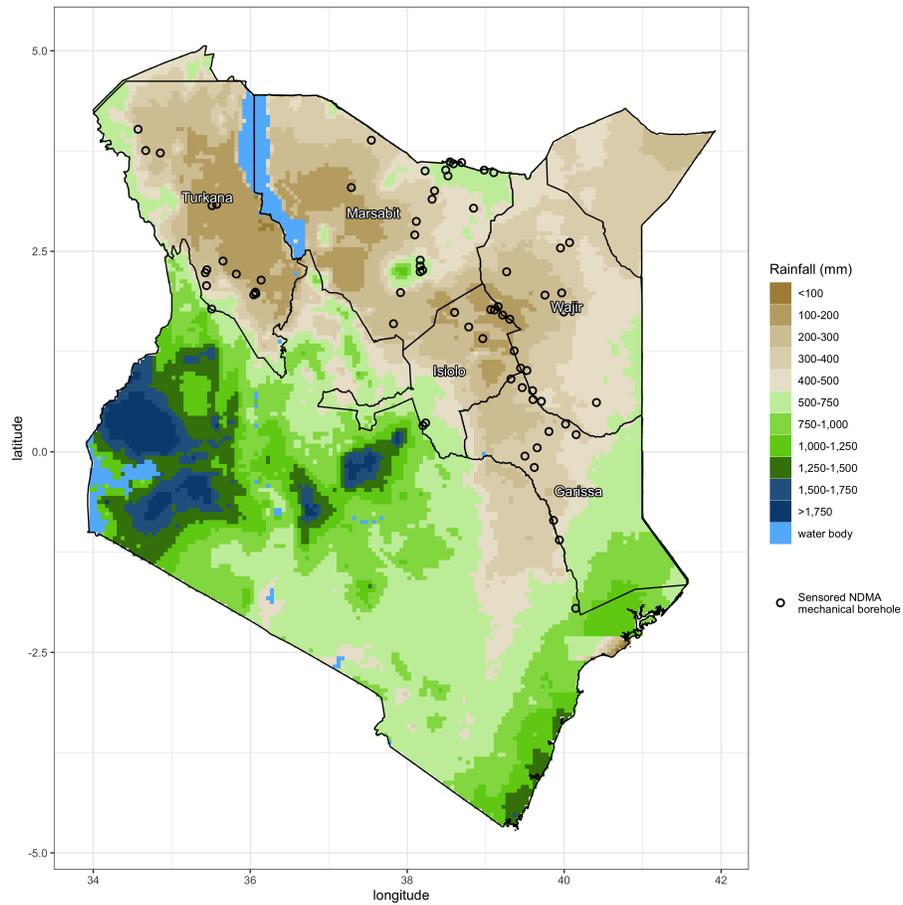


Figure 1: Average annual precipitation, 1981 - 2019. The locations of sensor-monitored NDMA (National Drought Management Authority) strategic drought emergency boreholes are shown in the five programmatic arid and semi-arid counties of Kenya. Data source: Climate Hazards Group InfraRed Precipitation with Station data (CHIRPS) v2.0

120 *2.2. Data*

121 *2.2.1. In-situ data*

122 From 2016 to present, 238 mechanized groundwater pumps in Kenya have been
123 monitored by satellite- and cellular-connected sensors provided by SweetSense Inc.

124 (sweetsense.space). Pressac brand wireless current transformers measure the fre-
125 quency and duration of borehole use by logging the electrical current (amperage)
126 delivered to the mechanical pump over time. Data is transmitted daily to a central
127 database and web application (Thomas et al., 2021). Sub-hourly sensor data was
128 aggregated to mean daily pump runtime and volume by month.

129 NDMA selected the EDE strategic boreholes to prioritize for operations during
130 drought. Further, data was subset to months during Kenya’s main dry season each
131 year – June through September – as the annual period when drought is of biggest con-
132 cern and deliberately surveilled. In-situ measurements from representative NDMA
133 EDE boreholes were included in the sample when the mechanical borehole and sensor
134 were reasonably known to be operational. Borehole and sensor status were assessed
135 from maintenance surveys and the availability of daily data.

136 The final sample included 69 unique mechanical borehole sites (Figure 1) and 565
137 site-month observations from June 2017 to September 2020. The four months of the
138 2021 dry season (June - September 2021) contributed 191 site-month observations
139 from 51 unique sites for external validation.

140 In order to provide comprehensive mapping of groundwater resources for differ-
141 ent stakeholders and applications, three separate, but complementary, binary high
142 groundwater use classifications were made from the sensor data. Each specification
143 informs an aspect of groundwater accessibility, including physical infrastructure and
144 population-based need. They are applied to the project region to be considered
145 singularly or in comparison.

146 The three definitions of categorical high groundwater use by month translated
147 from the sensor data are: 1) mean daily pump runtime over six hours per day (a
148 quarter day) and mean daily per capita volume extracted 2) over 35 liters per person
149 per day (L/pc/d) and 3) over 75 liters per person per day (L/pc/d). From the
150 electrical current delivered to the pump, the number of hours a pump was switched on
151 per day was measured and taken as pump runtime (Thomas et al., 2021). Estimated
152 flow rate at the pump in cubic meters per hour was collected during site visits,
153 and taking the product of runtime and yield provided estimates of volume of water
154 extracted per day.

155 The high use threshold for pump runtime was set at approximately the observed
156 mean while high per capita volume of over 35 L/pc/d was meant to capture basic
157 needs on the multiple-use water services (MUS) ladder. Between 20 and 50 L/pc/d,
158 users are expected to be able to meet most domestic and some livestock and per-
159 sonal agricultural needs (MUS Group, 2013). Additionally, a threshold of over 75
160 L/pc/d was considered to capture the higher water demands of pastoralists with
161 livestock and represent an intermediate service level on the MUS ladder. Thus, lo-

162 cations of groundwater use at or above the high runtime or high volume thresholds
163 indicate, respectively, areas where boreholes are expected to run longer than average
164 or the population is relying on groundwater to meet domestic and livestock needs,
165 presumably because other water resources are unavailable.

166 *2.2.2. Remote sensing data*

167 The variables taken or derived from satellite data were elevation (RCMRD Geo-
168 portal, 2015); precipitation (Funk et al., 2015); surface water availability (Senay
169 et al., 2013); baseflow, runoff, soil moisture, and actual evapotranspiration (Case
170 et al., 2014); Normalized Difference Vegetation Index [NDVI] (USGS FEWS NET
171 Data Portal, 2017); greenness vegetation fraction [GVF] (Vargas et al., 2015); and
172 population density (Stevens et al., 2015). More detail about the actual nature of the
173 features is given below in section 2.4.

174 Data was retrieved from January 2016 to May 2021 to cover the period of sensor-
175 recorded observations of borehole use and accommodate lagging of time-varying co-
176 variates for forecasting. Different temporal resolutions were ultimately averaged or
177 summed by month. Native spatial resolution of the satellite data products ranged
178 from 250 m to 50 km, and values were extracted at point coordinates of the sensors or
179 centroids of a $0.05^\circ \times 0.05^\circ$ ($\sim 30 \text{ km}^2$) grid over the five counties. Higher resolution
180 data were resampled using bilinear interpolation to agree with the reference grid.

181 Gridded population data at 1 km resolution was retrieved from WorldPop [worldpop.
182 org] (Stevens et al., 2015; Lloyd et al., 2019) to scale sensor-derived groundwater vol-
183 ume extraction by local population numbers. Volume per capita has been accepted
184 as a proxy for water demand in previous studies (McNally et al., 2019).

185 *2.3. Supervised machine learning*

186 Supervised machine learning (ML) models are empirical. They optimize model
187 fit to achieve the most similar outcomes to input training data without regard to
188 deterministic or mechanistic explanations. Predictions are derived from the fit ap-
189 plied to new observations. Ensemble ML (also known as Super Learning or model
190 stacking) executes multiple models ("learners") on the same data and selects the
191 optimal combination of them through cross-validation (van der Laan et al., 2007).
192 Our library consisted of 19 candidate learners: nine XGBoost algorithms with dif-
193 ferent hyperparameter selections (Chen and Guestrin, 2016); logistic regression; a
194 Bayesian generalized linear model (Gelman et al., 2008); Ridge regression, LASSO
195 regression, and another elastic net regularization with an alpha parameter equal to
196 0.5 (Friedman et al., 2010); three k-Nearest Neighbors learners (Mouselimis, 2021)
197 considering varying numbers of neighbors; Random Forests (Breiman, 2001); and a
198 null model.

199 Covariates, or features, are the independent variables that have an assumed im-
200 pact on the outcome (see section 2.4). Given the inclusion of algorithms in the
201 ensemble that perform feature selection in addition to a pre-screening procedure,
202 a large number of features, including those that are correlated, can be considered
203 without overfitting the model. Overfitting occurs when a model describes training
204 data well, but cannot generalize to new data. Feature selection and pre-screening
205 reduce the number of features to improve parsimony and cross-validation retains in-
206 dependence between training and testing datasets so that the model can predict new
207 observations of groundwater use with accuracy. Furthermore, ensemble models are
208 proven to perform as well as or better than any one candidate learner in the ensemble
209 and minimizing cross-validated risk controls for overfitting even when a large number
210 of learners are considered (van der Laan et al., 2007; Polley and Laan, 2010).

211 The purpose of modeling in this study was twofold: to extend runtime and volume
212 predictions spatially and forward in time. To provide predictions across the region at
213 a resolution of approximately 30 km², a model trained and validated at instrumented
214 boreholes was fit on independent explanatory features observed at centroid point
215 coordinates of a 0.05° x 0.05° grid over the five programmatic counties. To extend
216 the predictions forward in time, forecasting up to four months was achieved by
217 lagging the features by the same number of months and individual model runs.
218 Since the effect of climatological phenomenon on groundwater supply and demand
219 may be delayed itself, we provided the model with all lagged features available at
220 the respective forecast period and allowed the machine learner to screen and select
221 the best performing lag for each time varying feature separately. For each outcome,
222 the corresponding final model yielded a predicted probability that a certain centroid
223 during a particular month was in a high use category as defined above. To establish
224 binary high vs. low use, a threshold probability was selected based on a trade-off
225 between true positive (high use borehole confirmed by the sensor record) and true
226 negative (low use borehole confirmed by the sensor record) rates (see section 2.5).

227 To replicate operational conditions for the groundwater use forecasts and evaluate
228 performance, we constrained model building between 2017 and 2020 and projected
229 groundwater use forward for the 2021 dry season, June through September. Re-
230 viewing anticipated groundwater use alongside current infrastructure functionality
231 prior to the dry season should inform investments and actions taken by local water
232 officers and national decision-makers to reduce drought risk. Moreover, we evaluated
233 the skill of our forecasts, which serve as a source of external validation for future
234 estimates.

235 Modeling and feature design of contemporary groundwater estimates respected an
236 internal five-fold stratified cross-validation structure. In cross-validation the observed

237 data is sequentially partitioned into independent training and testing subsets. Five
238 equally sized subsamples were generated randomly with all time series observations
239 from one sensor being grouped in the same partition. The model was trained with
240 four of the subsets and tested on the remaining, held-out subsample. This was
241 repeated a total of five times (“folds”) with each of the subsamples being used once
242 as the testing dataset. Cross-validation allows for the calculation of performance
243 statistics and the ability to generalize the model to new data.

244 All data management and analysis was conducted in R statistical computing
245 software (R Core Team, 2020).

246 *2.4. Features*

247 The following variables are conventionally expected to influence drought and wa-
248 ter insecurity and were included in our study: precipitation, land surface tempera-
249 ture, vegetation indices, soil moisture profile, evapotranspiration, other hydrologic
250 system dynamics such as baseflow and runoff, population density, geographic location
251 (longitude, latitude, and elevation), and seasonality (Funk and Shukla, 2020).

252 The Climate Hazards Group InfraRed Precipitation with Station (CHIRPS) v2.0
253 (Funk et al., 2015) is a 30+ year quasi-global rainfall dataset that incorporates satel-
254 lite imagery with available in-situ station data to create $0.05^\circ \times 0.05^\circ$ gridded rainfall
255 time series. However, station data is inconsistent and inadequate over Africa. The
256 Trans-African Hydro-Meteorological Observatory [TAHMO] (tahmo.org) is gradually
257 enhancing weather data collection with the installation of low-cost stations across
258 the continent (Giesen et al., 2014). Currently, they operate over 500 stations in 25
259 countries, including 130 stations in Kenya. To leverage the advantages of both pre-
260 cipitation datasets — the coverage and legacy of CHIRPS with the fine resolution
261 in-situ data of TAHMO — we used inverse distance weighting (IDW) to interpo-
262 late station data and combined the output with the satellite estimates at monthly
263 time-scale through a Simple Bias Adjustment (SBA) merging method. The inter-
264 polation was done using a maximum correlation distance of 250 km, achieving a
265 minimum and maximum number of stations within this distance of two and ten,
266 respectively. If there were no stations in the vicinity to correct the satellite estimate,
267 the underlying CHIRPS grid value was retained, resulting in a spatially consistent
268 bias-corrected gridded product. We adopted this method from the ENACTS merging
269 process (Dinku et al., 2014).

270 Gridded maximum daily temperature from the Global Telecommunication System
271 (GTS) of the World Meteorological Organization (WMO) was retrieved and averaged
272 by month (Physical Sciences Laboratory). NDVI from the Collection 6 Moderate
273 Resolution Imaging Spectroradiometer (MODIS) instrument on the Aqua satellite

274 was taken from processed data distributed by the United States Geological Survey
275 (USGS) Earth Resources Observation and Science (EROS) Center (USGS FEWS
276 NET Data Portal, 2017). GVF from the Visible Infrared Imaging Radiometer Suite
277 (VIIRS) instrument aboard the Suomi National Polar-orbiting Partnership (NPP)
278 and National Oceanic and Atmospheric Administration (NOAA)-20 satellites (Vargas
279 et al., 2015) was based on the differences between the enhanced vegetation index
280 (EVI) for bare soils and dense vegetation.

281 Land surface modeling (LSM) is a method to extend the coverage of hydrologic
282 observations in data sparse contexts, such as remote and developing locations. The
283 Noah-Multiparameterization (Noah-MP) LSM relies on water and energy balances
284 to physically describe land-atmosphere interaction processes under multiple specifi-
285 cations (Niu et al., 2011). Mean daily soil moisture (m^3/m^3) at depths of 0 - 10 cm,
286 10 - 40 cm, 40 - 100 cm, and 100 - 200 cm and total monthly actual evapotranspi-
287 ration (mm), baseflow (mm), and runoff (mm) were derived from local instances of
288 the Noah-MP LSM run per month (Case et al., 2014).

289 Population density was downloaded from WorldPop (worldpop.org) gridded pop-
290 ulation data from combined geospatial and census data (Stevens et al., 2015; Lloyd
291 et al., 2019). Data is provided only through 2020; for 2021, we applied a population
292 growth rate of 2.3% to the 2020 data (The World Bank, 2021). Population den-
293 sity was used as an independent variable when modeling pump runtime, but for per
294 capita volume predictions, it was only used to scale the outcome and not as a model
295 feature.

296 Other variables were included based on a demonstrable or logical relationship to
297 groundwater use and demand. Public water management in Kenya is deregulated to
298 the county level, leading to substantial differences in institutional and operational
299 practices; thus, a county variable was added to describe unmeasured variability in
300 groundwater supply and use due to administrative differences. Generally, deter-
301 minants of the magnitude of groundwater use will be supply, demand, and access.
302 Regionally interpolated relative surface water availability (Senay et al., 2013) was
303 included to express the inverse correlation between rainfall, and by extension surface
304 water supply, and borehole use (Thomas et al., 2019; Thomson et al., 2019). From
305 a data inventory collated by Acacia Water as part of the Kenya RAPID program
306 (kenyarapid.acaciadata.com/map/13) we also considered the presence within 10 km
307 of floodplains, lakes, springs, or river basins. Proximity to installed water infrastruc-
308 ture, such as dams, sand pans, and other boreholes, may limit the use at any one
309 groundwater source, and, therefore, counts of operational infrastructure within 10
310 km were created. Calculated domestic water demand in 2015 provides an estimate
311 of volume per square kilometer per day ($\text{m}^3/\text{km}^2/\text{d}$) by sub-county required for ade-

312 quate human consumption (Tolk et al., 2016). Social and economic networks can be
313 demonstrated by access to main roads, markets, or towns, or, in this case, a count
314 of the number of these amenities within 10 km.

315 Livestock are one of the largest consumers of water in the ASALs. Pastoralists
316 commonly direct their herds to boreholes along migration routes so the presence of
317 a major livestock migration route within 10 km was added as a feature. Estimated
318 water demand for livestock ($\text{m}^3/\text{km}^2/\text{d}$) at the sub-county level was also included
319 (Tolk et al., 2016).

320 We generated several variables from the sensor data to describe typical use behav-
321 ior of boreholes in the region. One, we linearly regressed mean daily pump runtime or
322 volume on static site characteristics, including longitude, latitude, elevation, county,
323 water demand, proximity to natural water sources, other water infrastructure, and
324 amenities, and livestock movement. Two, to leverage the sensor network and char-
325 acterize the behavior of neighboring boreholes, we computed mean use at the five
326 closest, by Euclidean distance, sensed sites as well as an interpolated layer of mean
327 use at the maximally correlated site despite geographic location.

328 An advantage of using an ensemble learner and cross-validation is the ability to
329 consider many features, models, and settings at once with less constraint on assuring
330 a parsimonious model arising from a priori selection of the best set of variables and
331 estimator (Polley and Laan, 2010). Thus, many social, environmental, and economic
332 dimensions of drought risk and vulnerability were included. Ultimately, the ML
333 was provided approximately 30 features from which each algorithm was built after
334 screening.

335 See Appendix A for a detailed summary of all model features.

336 *2.5. Statistical Evaluation*

337 Model performance is evaluated by Receiver Operating Characteristic (ROC)
338 curves and area under the curve (AUC), cross- and externally validated accuracy
339 and rates of misclassification, and reliability diagrams. ROC curves are a perfor-
340 mance measurement for classification problems. The ROC curve is a plot of sensi-
341 tivity (y-axis) – the proportion of correctly identified instances of high use – against
342 one minus specificity (x-axis) – the proportion of wrongly identified cases of high
343 groundwater use among observed low use. The AUC demonstrates how good the
344 model is at discriminating between high and low use. Determining which threshold
345 to set when translating predicted probabilities to categories is always a trade-off be-
346 tween sensitivity and specificity. Youden’s J statistic identifies the closest point on
347 the ROC curve to the uppermost left corner of the plot and assumes both are equally
348 important.

349 Accuracy is the percent number of sites where predicted high use correctly corre-
350 sponds to observed high use after assigning predicted probabilities to above or below
351 a set threshold. The null or no information rate (NIR) is the expected accuracy
352 given random assignment based on the prevalence of the outcome in the data alone.
353 Instances of incorrectly identified high use are evaluated with the false positive rate
354 (FPR) and false negative rate (FNR) rate. The FPR is the proportion of wrongly
355 identified cases of high groundwater use among observed low use. The FNR is relative
356 to the number of misidentified low use among actual instances of high use.

357 Reliability is assessed from a linear relationship between predicted probability
358 and observed frequency of high use relative to the magnitude of the probability. A
359 perfectly reliable model exhibits a 1:1 relationship between the two.

360 **3. Results**

361 *3.1. High resolution maps of groundwater use and demand*

362 The ensemble machine learning models allowed for the creation of high spatial
363 and temporal resolution maps of groundwater use and demand. For each dry month
364 June 2017 - September 2020 contemporary estimates of pump runtime and volume
365 extracted at existing or hypothetical groundwater infrastructure are available at
366 approximately 30 km² resolution. The most recent dry season, June - September
367 2021, was forecasted and validated with ensuing observed data.

368 Figure 2 demonstrates how predicted probability output from the model was
369 translated to high use for domestic needs (> 35 L/pc/d). Red intensity indicates
370 a higher probability of high use and ultimate high use class assignment is outlined.
371 These maps may also suggest the certainty of high use classification. At a probability
372 between 80-100%, we are relatively more confident groundwater use will surpass the
373 high use threshold; if it is closer to 40-60%, we recognize that the model has identified
374 one use category, but it may disagree with observed values on occasion.

375 Site-level accuracy for each 2021 month (72%, 74%, 77%, and 70%) showed that
376 performance was relatively stable over different forecast periods, although it was
377 lowest in September 2021 when the model had a lead of four months. High use
378 was overestimated (i.e., $FPR > FNR$) in June and September and underestimated in
379 July and August (i.e., $FNR > FPR$). When considering all years, specificity was high
380 relative to sensitivity (Figure 4), indicating that generally the rate of false positives
381 was lower than the rate of false negatives despite forecast lead time.

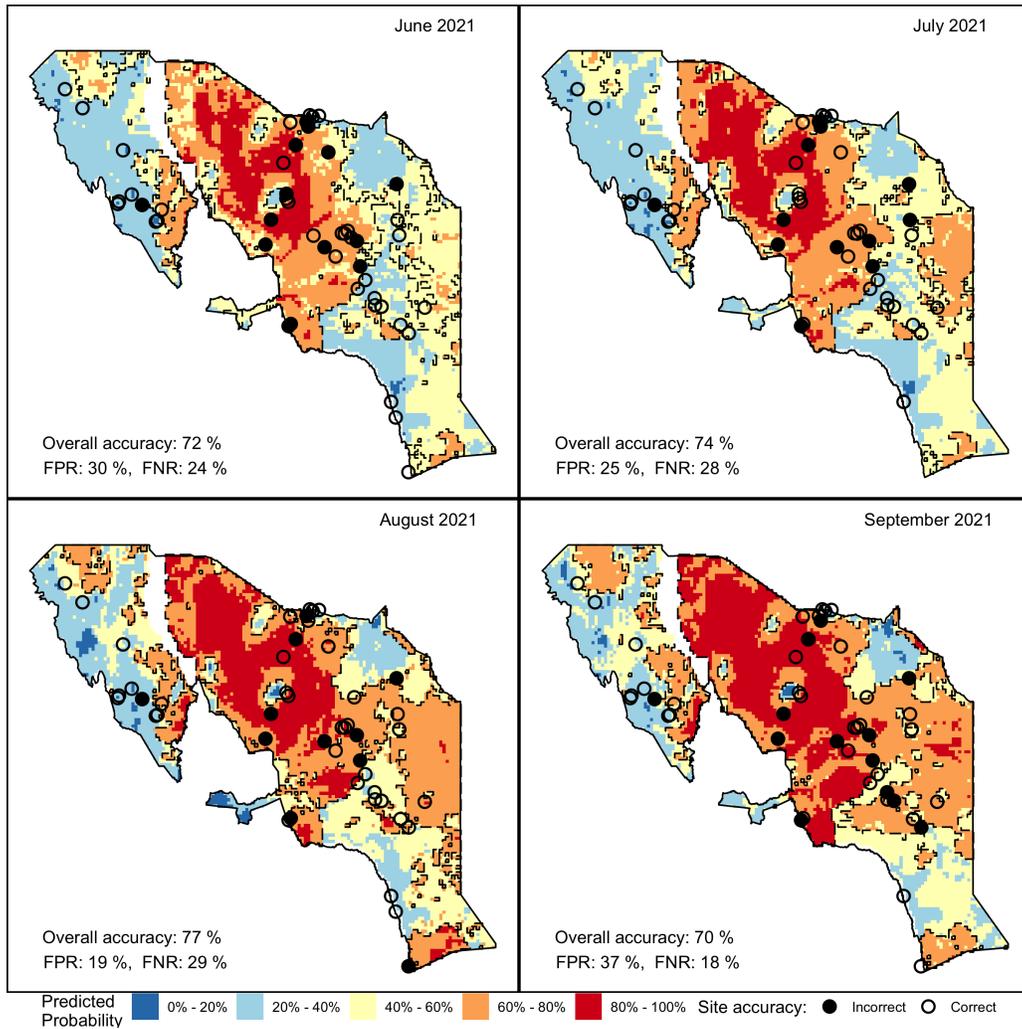


Figure 2: Predicted probability of groundwater demand for daily domestic water needs (>35 L per person per day) forecasted for the 2021 dry season. Categorization of high demand is outlined and accuracy is given for each dry season month. Rates of false positives (FPR) and false negatives (FNR) are based on site-level observations.

382 A similar demonstration was made for intermediate groundwater volume (> 75
 383 L/pc/d) (Figure 3). Areas of northern Turkana County and southwestern Garissa
 384 County showed diminished need for higher volumes indicative of livestock watering.
 385 This was somewhat unexpected as these areas do support livestock. However, these
 386 areas are also places of violent conflict. So while there may be a high need for water,
 387 we anticipate some groundwater points in these region are demonstrating lower user

388 due to insecurity limiting access. There is unmet demand for water and rehabilitating
389 water points or providing alternative water sources could serve to improve resource-
390 based conflict in the area. Site accuracy was high (73-80%) and mis-identification of
391 false negatives was more likely than false positives across the season.

392 Groundwater demand increased as the dry season progressed and was consistently
393 highest in Marsabit and Isiolo counties for both domestic and livestock needs. Here
394 only per capita volume was shown, but categories of pump runtime displayed similar
395 patterns.

396 All maps are accessible via a web-based platform at drip.shinyapps.io/groundwater.
397 We presented only the pre-drought season 2021 volume per capita forecasts, but
398 through the application, users can view historical and projected groundwater use.
399 Other utilities include comparison of trends by month or year, viewing the two-class
400 categorization of high groundwater use informed by the predicted probability, dis-
401 playing the predictions as absolute magnitude or difference from averages, filtering
402 by county, and seeing the status of the sensed boreholes.

403 For the purposes of this paper, we constrained model building prior to 2021 in
404 order to externally validate the forecasts. In practice, the model and application will
405 be updated per month during the dry season with an approximate one-week delay,
406 at which time the concurrent groundwater estimates and forecasts roll forward one
407 month. The model building procedure is repeated with the additional time series
408 so the size and variability of the data will increase and we expect performance to
409 improve over time with a longer data record.

410 *3.2. Performance of gridded contemporary and forecasted estimates*

411 The models achieved high statistical performance (Table 1, Figure 4). Estimation
412 of volume per capita for domestic and intermediate needs proceeded the best.

413 The AUC for determining and forecasting high groundwater use ranged from
414 0.703 to 0.714, from 0.746 to 0.756, and from 0.778 to 0.787 for daily pump runtime,
415 domestic volume per capita, and intermediate volume per capita, respectively (Table
416 1). Regardless if groundwater use is defined by mechanical requirements of the
417 borehole or by water volume, our models contain a 70% or greater probability of
418 accurately discriminating between low and high use. Ultimately, after classifying
419 high use from predicted probabilities, the sensitivity for contemporary estimates, by
420 use definition, was 58%, 57%, and 55% with corresponding specificity of 75%, 82%,
421 and 88%. The rates for the one- to four-month forecasts were similar and can be
422 read from Figure 4A. Since specificity is greater than sensitivity, it is less likely that
423 a site predicted to exhibit high groundwater use would be incorrectly identified and
424 there are less false positives than false negatives in the predictions. Thus, our models

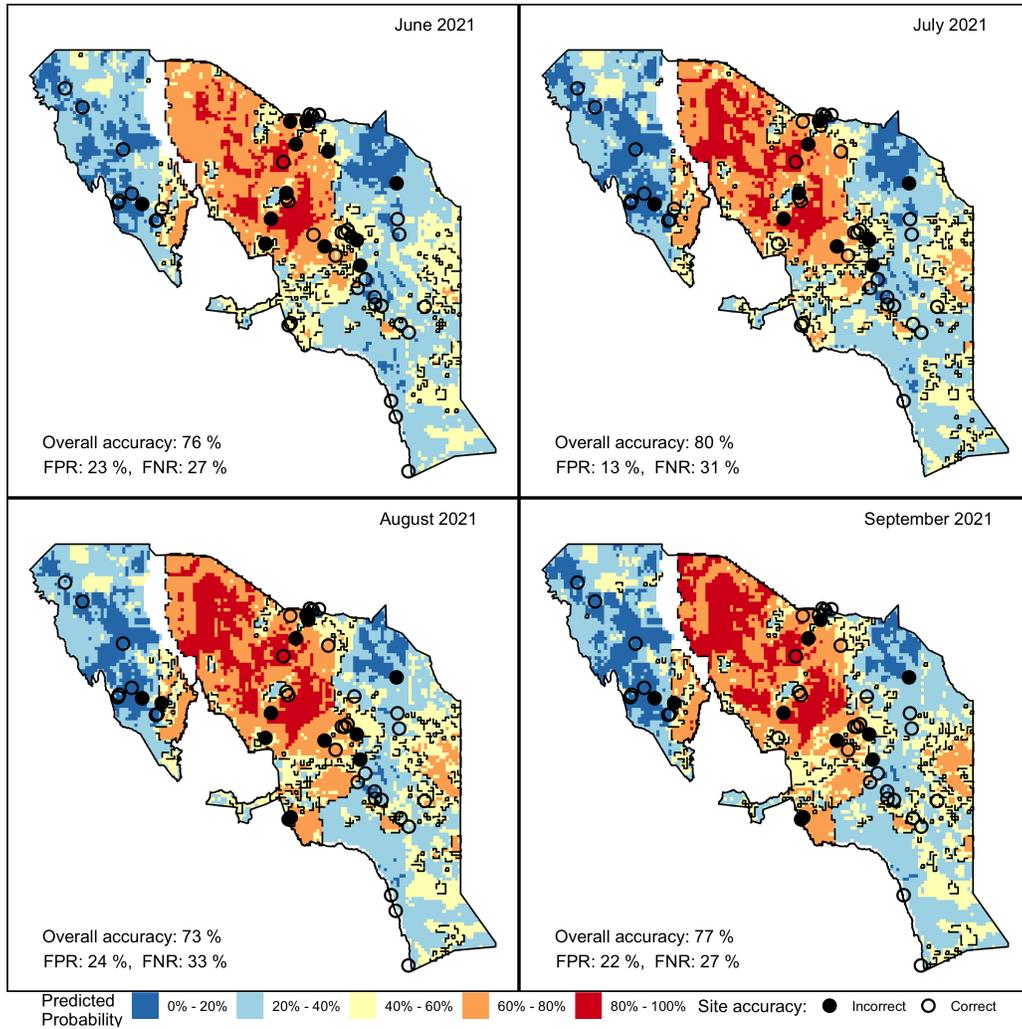


Figure 3: Predicted probability of groundwater demand for daily intermediate water needs (>75 L per person per day) forecasted for the 2021 dry season. Categorization of high demand is outlined and accuracy is given for each dry season month. Rates of false positives (FPR) and false negatives (FNR) are based on site-level observations.

425 are optimized to conserve material resources used to respond to high groundwater
 426 need, such as deploying pump repair teams, at the occasional expense of failing to
 427 capture some regions of high use.

428 Overall model accuracy (Table 1) of the contemporary categorical high runtime
 429 and basic and intermediate volume per capita was 67.6%, 69.4%, and 74.5%, respec-
 430 tively. A similar level of accuracy was observed for the one- to four-month forecasts.
 431 High predictive skill of volume per capita was demonstrated with external validation,
 432 indicating these models were not overfit. Accuracy was at least 70% and as high as
 433 80% for forecasting per capita volume and represented significant improvement over
 434 the NIR. Predictive skill of pump runtime was lower when tested on held-out data.

Table 1: Accuracy statistics for contemporary and one- to four-month forecasts of categorical high pump runtime (6+ hr/d), volume per capita for basic needs (35+ L/pc/d), and volume per capita for intermediate needs (75+ L/pc/d). Sensor data from the 2021 dry season was reserved from model building to provide evidence of operational accuracy for the forecasts from external validation. AUC = Area Under the Curve; NIR = No Information Rate; p-value = significance value of difference between accuracy and NIR

Model	AUC (95% CI)	Accuracy, % (p-value)	Validation Acc, %
Pump Runtime, 6+ hr/d	–	NIR: 57.7	–
Contemporary	0.705 (0.663 - 0.748)	67.6 (<0.001)	–
1 mo forecast - Jun 2021	0.706 (0.663 - 0.748)	66.5 (<0.001)	56.0
2 mo forecast - Jul 2021	0.714 (0.672 - 0.756)	66.9 (<0.001)	58.7
3 mo forecast - Aug 2021	0.703 (0.661 - 0.746)	63.9 (0.002)	56.2
4 mo forecast - Sep 2021	0.708 (0.666 - 0.750)	65.8 (<0.001)	57.4
Volume Per Capita, 35+ L/pc/d	–	NIR: 51.5	–
Contemporary	0.756 (0.717 - 0.796)	69.4 (<0.001)	–
1 mo forecast - Jun 2021	0.756 (0.717 - 0.795)	68.7 (<0.001)	72.0
2 mo forecast - Jul 2021	0.748 (0.708 - 0.787)	66.7 (<0.001)	73.9
3 mo forecast - Aug 2021	0.749 (0.709 - 0.788)	68.3 (<0.001)	77.1
4 mo forecast - Sep 2021	0.746 (0.705 - 0.786)	69.0 (<0.001)	70.2
Volume Per Capita, 75+ L/pc/d	–	NIR: 58.4	–
Contemporary	0.787 (0.749 - 0.825)	74.5 (<0.001)	–
1 mo forecast - Jun 2021	0.785 (0.747 - 0.823)	74.3 (<0.001)	76.0
2 mo forecast - Jul 2021	0.783 (0.745 - 0.821)	74.5 (<0.001)	80.4
3 mo forecast - Aug 2021	0.778 (0.739 - 0.817)	73.8 (<0.001)	72.9
4 mo forecast - Sep 2021	0.781 (0.742 - 0.820)	75.0 (<0.001)	76.6

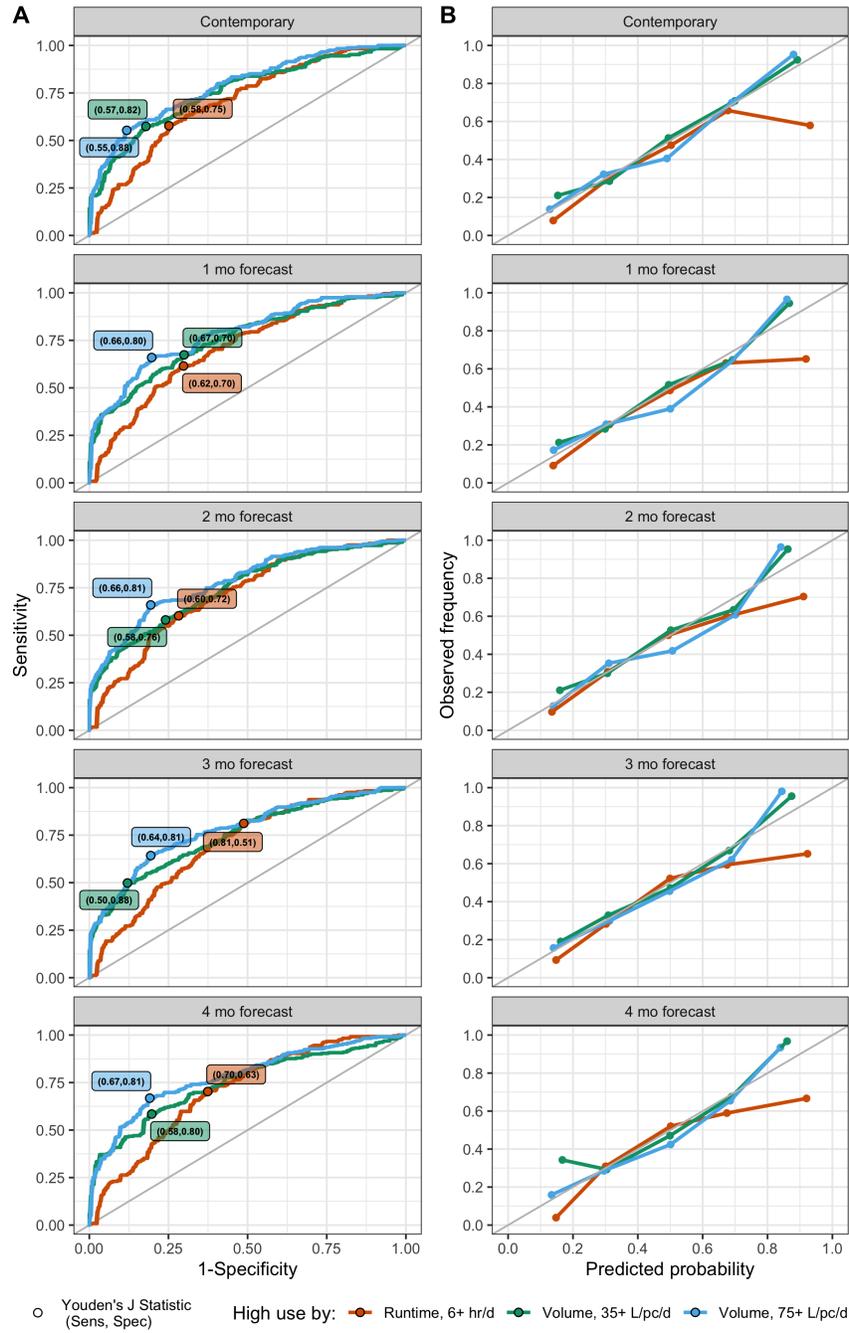


Figure 4: A) Receiver Operating Characteristic (ROC) curves and B) Reliability Diagrams. Performance given for pump runtime and volume per capita and by models used to estimate contemporary and forecasted groundwater use predictions.

435 Volume per capita predictions had the best reliability, exhibiting a consistent
436 linear relationship between predicted probability and observed relative frequency
437 (Figure 4B). Modeling of domestic volume per capita groundwater use assigned a
438 probability of high use comparable to the occurrence of actual high use at each mag-
439 nitude of probability except at the highest frequencies of use, when the probability
440 was underestimated. At a lead time of four months, high use was also underestimated
441 at the lowest frequencies of use. The model for groundwater use for intermediate
442 water needs underestimated use at the highest frequencies and overestimated prob-
443 ability at median frequencies. Prediction of pump runtime tended to overestimate
444 the probability of high use.

445 *3.3. Contribution of model features*

446 Variable importance plots (Figure 5) indicate the relative information gained from
447 the independent explanatory features. Variable importance is measured by the ratio
448 in error, using negative log likelihood loss, between the full ensemble model and a
449 model on modified data where the respective features have been excluded from model
450 building. Features are plotted against the risk ratio, where model error without the
451 features is divided by the error obtained from the full model. See Appendix A for
452 which features are described by each theme. For example, in the one-month lagged
453 models aimed at forecasting groundwater use in June 2021, removing administrative
454 features (county, longitude, latitude, elevation, and proximity to towns, markets,
455 and roads) increased model error by about 8% for pump runtime, between 4% and
456 5% for basic per capita volume, and just over 1% for intermediate per capita vol-
457 ume, indicating that, collectively, these variables were less informative for forecasting
458 groundwater volume per capita needs.

459 The plots demonstrate patterns in what forces the groundwater use predictions
460 for each use definition and lead time. In general, features related to typical, neigh-
461 boring borehole behavior were not instructive of site usage. At a lag of one month,
462 features related to or suggestive of water availability at the surface – such as pre-
463 cipitation, water bodies, evapotranspiration, hydrology, and vegetation greenness –
464 better informed volume per capita than pump runtime. With greater lagged obser-
465 vations, i.e. as the hypothetical dry season progressed, the signal between features
466 and per capita volume forecasts became less clear; although for pump runtime, many
467 of the features shared a similar level of importance throughout. Models with longer
468 lead (three and four months) demonstrated, with a risk ratio less than one, some fea-
469 ture groups were damaging to model performance. This could be suggestive of some
470 overfit or correlated variables within the group; however, since overall performance
471 and accuracy of the forecasts were acceptable we did not investigate this further with

472 a systematic sensitivity analysis.

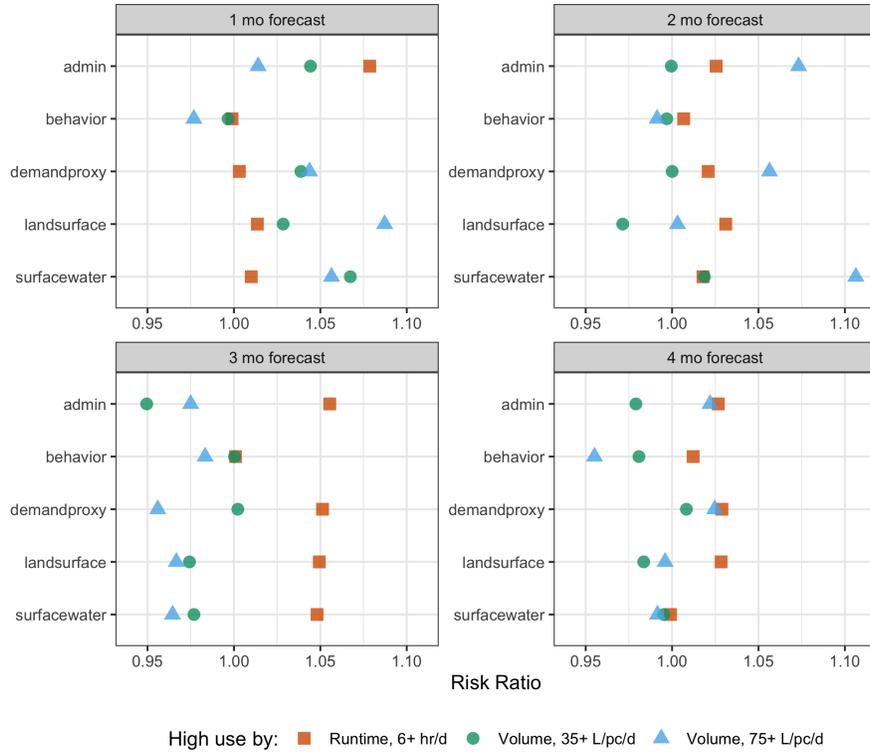


Figure 5: Variable importance plots for models predicting forecasted pump runtime and volume per capita. Covariates are grouped by theme; see Appendix A for which covariates are described by each theme.

473 4. Discussion

474 These maps represent the first operational spatially-explicit sub-seasonal to sea-
475 sonal (S2S) estimates of groundwater use and demand in the literature. The inte-
476 gration of in-situ remote sensors with satellite data and hydrological land surface
477 models through ensemble machine learning directly addresses an identified gap in
478 population-based, near-real time, acute water monitoring (McNally et al., 2019) and
479 supports other services for multi-dimensional drought early warning early action
480 (Funk and Shukla, 2020).

481 The maps show areas of persistent high groundwater use, areas that develop a
482 reliance on groundwater over the dry season, and locations where volume extraction
483 suggests the requirement of groundwater to meet domestic needs for households and

484 watering of livestock for pastoralists. Previous work (Thomas et al., 2019) has iden-
485 tified that pastoral and agropastoral households typically use a mix of groundwater
486 and surface water to meet their needs, but have increased reliance on groundwater
487 during dry seasons or in times of drought. Surface water or rainwater harvesting
488 are preferred sources when available, with rural populations in Kenya reporting 34%
489 less groundwater use during the wet season. The preference for surface water may
490 be even more defined for livestock owners and the most water insecure populations
491 (Thomson et al., 2019). Empirically we also observed an increase in borehole runtime
492 and volume extraction during drought declared years and as the dry season advanced
493 within years (see Section 4.1 below).

494 The ability to discriminate between high and low groundwater use did not atten-
495 uate over the dry season despite lead time up to four months between predictions and
496 predictive features. Groundwater needs become more acute as the dry season pro-
497 gresses, resulting in a stronger measurable signal. Increasingly, rains prior to the dry
498 season provide temporary relief, but are insufficient to provide adequate soil mois-
499 ture for vegetation and recharge of surface water for the duration of the dry season.
500 Thus, characterizing the hydrologic productivity of the rainy season, which occurs
501 one to four months in advance, can explain patterns in subsequent groundwater use.

502 With external validation, forecasts of pump runtime proved to be non-informative
503 during the 2021 dry season. Thus, there is likely additional unexplained variability
504 among sites and activities supported by boreholes. Previous inventories of mechanical
505 groundwater infrastructure in this region indicate that there are substantial differ-
506 ences in withdrawal efficiencies due to size, manufacturer and operator, age, power
507 source, and on-site storage (Thomas et al., 2019). Therefore, despite the introduc-
508 tion of an additional source of error when translating mechanical pump runtime to
509 volume with a yield coefficient, predictive skill of volume per capita was highest,
510 particularly for intermediate (> 75 L/pc/d) use. Instead, we believe that flowrate
511 helped control for differences in pump capacity and implementation. Moreover, a
512 higher volume threshold likely further differentiated between boreholes used for crit-
513 ical drought mitigation during the dry season specifically and other EDE-targeted
514 boreholes that might have been used more indiscriminately.

515 Predictors related to surface water availability – satellite and gauged precipita-
516 tion, relative depth of freshwater water points, and proximity to other natural water
517 bodies – were indicative of volume per capita model performance at one and two
518 month lead times. Many of the other satellite-derived and modeled hydrological
519 features had an inconsistent effect on model error. The change in relative variable
520 importance by forecast lead time highlights that characterizing water resources and
521 drought dynamically and 'on-the-ground' with respect to human behavior is complex.

522 The utility of a statistically driven model is the assimilation of satellite data, land
523 surface models, and field data into accurate estimates that lead to actionable insights
524 in the absence of a clear or complete deterministic understanding. This is especially
525 relevant to Sub-Saharan Africa, a context of sparse monitoring data, limited capac-
526 ity, and developing scientific consensus of the agro-hydrological mechanisms driving
527 drought (Funk and Shukla, 2020).

528 *4.1. Estimated impact of groundwater demand on drought risk and planning*

529 The methodology presented here could not have proceeded without in-situ data,
530 but we have demonstrated that low density sensors can integrate satellite data and
531 mechanistic land surface modeling for local groundwater monitoring. The fine spatial
532 and temporal resolution of our gridded groundwater use and demand maps has not
533 been achievable before. Approximately 70 sensors were used to inform a total area of
534 $260,000 \text{ km}^2$, or one sensor per $3,700 \text{ km}^2$ ($1,500 \text{ mi}^2$). A small network of sentinel
535 sensors provides economical and efficient means to ground-truth and expand the
536 utility of earth observation data.

537 We made a pixel-to-point comparison between the gridded groundwater estimates
538 and observed site data. One-to-one comparisons between gridded satellite data and
539 station observations have been practiced in the literature (Dinku et al., 2014) and
540 while they have been shown to represent area averages less well than pixel-to-pixel
541 comparisons (Dinku et al., 2018), the method is relevant to our context. In the
542 ASALs of northern Kenya residents may walk up to 10 km to collect water for
543 domestic needs and pastoralists walk even greater distances in search of livestock
544 watering. Generating estimates at $0.05^\circ \times 0.05^\circ$, roughly equal to $5 \text{ km} \times 5 \text{ km}$,
545 is appropriate for describing water collection behavior then. Additionally, the use
546 of high vs. low categories smooths differences in groundwater use attributable to
547 unexplained site variability at this scale. We report high cross-validated accuracy at
548 this density and resolution. Thus, the modeled groundwater use/demand pixels did
549 characterize site-level observations well.

550 The distribution of sensors in the five program counties was not uniform, but
551 this was by design with respect to drought emergency. NDMA and county officials
552 identified strategic EDE boreholes for monitoring and water resource managers an-
553 ticipate areas without sensors to be lower priority during drought response because
554 of low population densities, uninterrupted water resources other than groundwater,
555 or other mitigating factors. Despite these differences, similar hydrological, climato-
556 logical, and socioeconomic conditions are captured at sensed sites since the ASALs
557 share many common characteristics (NDMA, 2015). In counties with better distribu-
558 tion of sensors, such as Turkana, predictions were accurate across a range of different

559 conditions. Moreover, in areas of predicted high use without existing EDE prior-
 560 ity, our predictions may prompt re-evaluation of EDE resources and mobilization of
 561 additional in-situ data collection.

562 Gridded groundwater demand contains essential information for decision-makers
 563 in settings experiencing or expecting drought. The percent area and percent of pop-
 564 ulation affected by reliance on groundwater each year can be enumerated from our
 565 maps (Table 2). We delineated trends in groundwater use over time that can be
 566 related to other meteorological and hydrological phenomenon. The national govern-
 567 ment of Kenya declared drought in 2017 after several years of low precipitation. An
 568 above-average long rains season in 2018 helped sustain improvements in water and
 569 food security, but the next year saw a return to drier conditions and higher tempera-
 570 tures and, subsequently, 2019 was another year of drought. However, rainfall toward
 571 the end of the year and during the 2020 long rains promoted recovery. Unfortunately,
 572 the following two rainy seasons were critically below average, and another national
 573 drought disaster was declared in 2021 (OCHA, 2021a). The ASALs, where now over
 574 2.5 million people are facing water and food shortages, have been especially impacted
 575 (OCHA, 2021b). We see relative changes reflected in high groundwater demand dur-
 576 ing drought years and accurately forecasted higher demand prior to the 2021 dry
 577 season (Table 2).

Table 2: Frequency of predicted high groundwater use over northern Kenya by percent of total area and population affected each year at any point during the dry season (June - September) of that year. Population counts were taken from WorldPop unconstrained UN Census adjusted estimates (Lloyd et al., 2019).

Year	Pump Runtime		Volume Per Capita		Volume Per Capita	
	6+ hr/d		35+ L/pc/d		75+ L/pc/d	
	Area (%)	Popl (%)	Area (%)	Popl (%)	Area (%)	Popl (%)
2017	54.2	50.3	64.3	40.4	33.6	10.1
2018	46.5	40.6	59.8	30.1	40.3	12.9
2019	51.5	49.0	57.9	27.4	37.0	10.0
2020	49.6	43.6	57.6	28.1	40.4	11.7
2021	71.4	68.4	63.1	34.7	45.0	14.7

578 A review of the 2021 dry season before it begins (Table 3), provided national and
 579 county officials with the total and percent of their constituents expected to be living
 580 under circumstances contributing to high groundwater use, such as limited surface

581 water and lowered access to other services. Percent change in groundwater use by the
 582 end of the dry season demonstrates how and where need changes. All counties were
 583 expected to experience an increase (positive change) in groundwater reliance as the
 584 dry season progressed with one exception. In Isiolo, the percent of the population
 585 expected to undertake high groundwater use to meet basic water needs decreased
 586 from June to September. Counties that experienced the greatest change, such as
 587 Wajir and Garissa, may engender more attention and resource allocation for early
 588 warning early action services.

Table 3: Impact of 2021 dry season. Total and percent of population predicted to experience high groundwater use at the start of the 2021 dry season (June) by county and the the percent increase in population expected to be relying on groundwater by the end of the dry season (September). Population counts were taken from WorldPop unconstrained UN Census adjusted estimates (Lloyd et al., 2019).

County	Pump Runtime			Volume Per Capita			Volume Per Capita		
	6+ hr/d			35+ L/pc/d			75+ L/pc/d		
	<i># thous.</i>	<i>%</i>	<i>% chg</i>	<i># thous.</i>	<i>%</i>	<i>% chg</i>	<i># thous.</i>	<i>%</i>	<i>% chg</i>
All	1830	35.9	63.4	1050	20.7	61.2	424	8.3	49.5
Garissa	392	24.1	45.5	123	7.6	140.5	34.1	2.1	81.9
Isiolo	72.7	39.5	14.3	105	57.0	-5.2	59.1	32.1	33.1
Marsabit	179	49.2	40.0	202	55.8	10.0	187	51.4	9.1
Turkana	859	72.5	25.8	191	16.1	24.9	65.1	5.5	56.3
Wajir	326	18.8	207.7	432	24.9	94.8	78.5	4.5	138.0

589 Several actions are anticipated from these maps in advance of drought. When
 590 mapped together, sensor data indicating pumps needing repair where high usage
 591 is predicted would be an alert to prioritize maintenance services to these monitored
 592 sites. Where there are no sensors, this kind of assessment would need to proceed from
 593 the institutional knowledge of local water officers, field scientists, and other experts.
 594 Thus, if high usage is predicted in areas without adequate borehole coverage and/or
 595 functionality — known through prior or external sources — then resources should
 596 be devoted to new installations, maintenance, and other water infrastructure.

597 The models had a higher specificity than sensitivity, meaning there were fewer
 598 false positive results and the risk of allocating resources to increase water availabil-
 599 ity unnecessarily is lower. Conversely, this interpretation accepts that that some
 600 instances of need will be missed. We chose a threshold to balance sensitivity and

601 specificity for the highest overall accuracy, but the two parameters can be changed
602 based on programmatic priorities.

603 Operational drought monitoring for Sub-Saharan Africa has been relatively nascent
604 in the last decade, when satellite data and land surface modeling were leveraged to
605 establish historical climatology and forecast parameters related to meteorological,
606 agricultural, and hydrological drought, such as precipitation, soil moisture, vege-
607 tation, and streamflow. Yet, most monitors do not include a unique groundwater
608 component.

609 FEWS NET provides several drought indicators (McNally et al., 2019), and ad-
610 ditionally the water supply and demand product for crops, the Water Requirement
611 Satisfaction Index (WRSI) (Senay et al., 2015), and surface water levels, the Wa-
612 terpoint Viewer (Senay et al., 2013). Acute water stress monitoring related these to
613 population needs (McNally et al., 2019). Although, given that these analyses are typ-
614 ically generated from renewable freshwater resources and that the volume of stored
615 groundwater is estimated to be 100 times that of annual renewable sources and 20
616 times the volume of freshwater lakes in Africa, groundwater as a resource to meet
617 domestic, pastoral, and agricultural needs is missing from water scarcity assessments
618 (MacDonald et al., 2012). These datasets in large part form the evidence base of
619 food security classifications and risk outlooks.

620 Thus, we propose the DRIP groundwater use and demand maps will become
621 another reference dataset for drought indices and expand the knowledge base for
622 decision making in Kenya and other future operational contexts. A challenge will
623 be how to systematically integrate localized groundwater withdrawal estimates as
624 a quantitative feature, but we have identified potential to apply them to drought
625 monitoring through expert interpretation and guidance with several key stakeholders.

626 *4.2. Case study applications*

627 Stakeholder consultations were a critical component of this research. We con-
628 ducted regular consultations with scientists at the eastern and southern Africa SERVIR
629 hub – the Regional Centre for Mapping of Resources for Mapping (RCMRD) – and
630 FEWS NET in order to align the groundwater use and demand products with user
631 needs. We also organized two user engagement workshops. The first workshop was
632 held in February 2021 and was attended by drought management officers from the
633 NDMA. The second meeting was held in June 2021 and focused on the rollout of pro-
634 totype products for Marsabit County in Kenya. Participants in the second meeting
635 were drawn from NDMA, the Kenya Meteorological Department, drought humani-
636 tarian agencies including the Kenya Red Cross Society (KRCS), Mercy Corps, and
637 Food for the Hungry, and the county departments of information technology and

638 water services.

639 We received three important recommendations from these meetings: (i) the need
640 for participatory planning and co-development of community early warning systems
641 that leverage groundwater use and demand information from DRIP, (ii) continuous
642 improvement of the groundwater use and demand products through expert feedback
643 and user-based validations, and (iii) linking the products to existing drought infor-
644 mation dissemination mechanisms through the NDMA national and county drought
645 bulletins. These recommendations are part of future co-development processes. The
646 development of the groundwater products is also synergistic with other rangeland
647 vegetation monitoring services by RCMRD and national drought early warning early
648 actions program led by the KRCS.

649 *4.2.1. The Famine Early Warning System Network (FEWS NET)*

650 FEWS NET scientists and domain experts (co-authors McNally and Slinski),
651 reviewed the DRIP groundwater platform for its potential to inform the FEWS
652 NET analysis of acute food insecurity risk. The risk of acute food insecurity is a
653 function of a particular hazard or shock, the vulnerability of a specific population,
654 and that population's ability to cope, or recover, from the shock. This framework
655 informs the FEWS NET scenario development process that allows for the projection
656 of food insecurity eight months in advance for humanitarian assistance planning.

657 The ASALs of Kenya are subject to shocks that include drought, animal pests
658 and diseases, limited access to dry season grazing, and cattle raiding. Some com-
659 munities additionally experience inadequate access to water for domestic use and
660 watering livestock, especially during the dry season. This results in the occurrence
661 of waterborne disease and poor animal health which are additional shocks to the sys-
662 tem. Much of the population in northern Kenya relies on livestock for food and cash
663 income (FEWS NET, 2011). Thus, given the nature of the hazards and vulnerability
664 of the livestock sector, the ability of this region to cope tends to be low. In late 2021,
665 this region was experiencing Phase III - Crisis food insecurity which is characterized
666 by households that either have above usual acute malnutrition or are marginally able
667 to meet minimum food needs by depleting essential assets or engaging in negative
668 coping strategies.

669 The predictions of groundwater use and demand can be helpful to the FEWS
670 NET scenario development process. As previously mentioned, groundwater demand
671 is indicative of inadequate access to water from other sources, like surface water and
672 shallow water infrastructure due either to non-functionality or drought. In this way,
673 groundwater demand forecasts could alert food security analysis of a shift in behavior
674 toward using available groundwater. In many locations, however, the groundwater

675 maps and overlaid in-situ sensor data show that households have inadequate access
676 to groundwater during crucial periods (e.g., dry season) due to non-functionality of
677 infrastructure at nearby boreholes.

678 This unmet need, which contributes to the risk of food insecurity, could be high-
679 lighted in FEWS NET reporting and in turn addressed by decision-makers (e.g.
680 programmers of humanitarian assistance or national water ministries) in a number
681 of different ways: 1) encouraging water point rehabilitation, including repair and
682 maintenance of pumps, in locations where DRIP indicates that wells are not prop-
683 erly functioning; 2) alternative solutions like water trucking could be employed in
684 locations where wells do not exist or are unable to be repaired; 3) institutionalization
685 of reliable access to groundwater through installation of new boreholes and increased
686 support of pump maintenance in longer term planning documents. Preventive action
687 to ensure sustainable groundwater access may be particularly useful in locations that
688 suffer from unreliable water trucking due to inaccessibility or political factors.

689 Prior to the 2021 dry season and before the current drought in Kenya became
690 apparent, our forecasts would have indicated an increased demand in groundwater
691 relative to 2020 (Table 2). An increased reliance on groundwater in times of drought
692 may, in some cases, involve traveling long distances to Kenya’s EDE strategic bore-
693 holes. These predictions were consistent with the observed situation in the September
694 2021 Key Message Update (FEWS NET, 2021): “an atypically high number of live-
695 stock are migrating to dry season grazing areas driven by the decline in rangeland
696 and water resources. Between July and August, livestock trekking distances to wa-
697 tering points increased by 60-90 percent, likely driving the 13-55 percent decline in
698 milk production compared to the three-year average. [...] Overall, the decline in
699 livestock productivity and body conditions is constraining household access to food
700 and income and maintaining area-level Crisis (IPC Phase 3) outcomes across pastoral
701 areas.”

702 Taken together, knowledge of groundwater demand and FEWS NET risk profiles
703 could have initiated concrete early actions such as borehole rehabilitation and pump
704 repair or alternative reliable access to water through, for example, water trucking to
705 improve water and food security during this year’s drought.

706 *4.2.2. The Kenya National Drought Management Authority (NDMA)*

707 NDMA exercises coordination across drought risk management and establishes
708 mechanisms, either on its own or with stakeholders, that will end drought emergencies
709 in Kenya (NDMA, 2015). NDMA has headquarter offices in Nairobi, Kenya and has
710 established sub-offices in 23 arid and semi-arid counties considered vulnerable to
711 drought. County Steering Groups (CSG) manage the coordination of drought and

712 early warning information at the county level in Kenya (USAID, 2018). The CSG
713 is co-chaired by the county governor and county commissioner while the NDMA is
714 the secretariat. Based on NDMA’s overall mandate and role in guiding the agenda
715 and discussion at CSG meetings, they have been targeted as a critical user of data
716 on groundwater.

717 From predicted groundwater demand, NDMA and other stakeholders can identify
718 mitigation activities against the effects of drought, such as identifying areas
719 with predicted high water demand whose borehole pumps would require preventive
720 maintenance. The Ministry of Water at the county would conduct timely budgeting,
721 procurement of spare parts, and plan visits by borehole technicians to specific sites.
722 Further, by analyzing the predicted water demand in relation to production capacities
723 of boreholes in an area, NDMA and stakeholders could advise communities on
724 migration that is usually triggered by the search for water. Such timely advice has
725 the potential to avert violence between communities over competition of resources
726 that has been common during the drought period. Thus, knowledge of groundwater
727 patterns and forecasts helps public agencies address multi-dimensional impacts of
728 drought, including health, livelihoods, and conflict.

729 *4.3. Limitations and Future Work*

730 In northern Kenya, groundwater supply is not a limiting factor. Recharge and
731 fossil quantities in aquifers are higher than abstraction rates and are capable of
732 providing groundwater (Mumma et al., 2011). This assumption would need to be
733 revisited when applying this framework in new settings. Instead, being able to extract
734 groundwater from functional infrastructure is a limitation that may contribute to
735 unmet demand for water and under or overestimate use in our maps. We know
736 from sensor reports and the motivation behind the DRIP theory of change (Thomas
737 et al., 2020) that mechanical boreholes in this region are, in fact, often non-functional.
738 In this study, our data filters attempted to remove the most persistent periods of
739 non-functionality, and we have begun to identify and correct for functional status
740 (Thomas et al., 2021). Our maps represent areas where dependence on groundwater
741 at strategically located boreholes is high under the assumption that EDE boreholes
742 do exist and are functioning in that area; if neither of these assumptions prove to be
743 valid, that in and of itself may be the justification to initiate a response.

744 The inclusion of forecasted climatology as model features, to supplement or sup-
745 plant the current lagged observations, should be explored. Predicting the dry season
746 from conditions at a one to four month lead was demonstrated here because hydro-
747 logical and agricultural drought, and subsequent reliance on groundwater, should be
748 strongly correlated to conditions during the long rains season, March through May

749 (Shukla et al., 2021). However, incorporation of other forecasted drought indicators,
750 which are often built from decades of historical climatology, could improve our pre-
751 dictive ability during years outside the range of values observed since in-situ borehole
752 monitoring.

753 Further development of the gridded groundwater maps and DRIP service will
754 focus on application and amplifying the wider context of groundwater health and
755 potential for agricultural and rangeland productivity. Installation of groundwater
756 level sensors and modeling of trends in groundwater level will address aquifer health,
757 local overdraft, and sustainability of groundwater use. As groundwater is promoted
758 and leveraged as a drought mitigation strategy, appropriate development and man-
759 agement supported by monitoring will be critical.

760 **5. Conclusion**

761 Groundwater represents an opportunity to increase reliable water supplies in
762 Africa and provide a buffer against drought. Improving accessibility to groundwater
763 through better maintenance of water systems and responsible development of new
764 water schemes are mitigation strategies that reduce drought risk and improve re-
765 siliience. Sustainable, effective, and local management of groundwater resources and
766 infrastructure is a necessary precondition to achieving water security, as is accurate
767 monitoring of changing water needs in a changing climate. The Drought Resilience
768 Impact Platform fine resolution gridded dry season groundwater use and demand
769 maps are a novel data source that can be used directly to allocate resources as well
770 as provide a groundwater component to other water and food insecurity indices as
771 part of multi-sectoral, multi-dimensional drought early warning and early action.

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780 **Appendix A. Characteristics of Data Features**

781 Features, or covariates, provided to groundwater prediction models. Theme relates to grouped features held out to test
 782 variable importance; see Figure 5. Res is the data product’s native resolution. ¹Population density was used as an independent
 783 variable in pump runtime models, but for per capita volume predictions it was only used to scale the outcome and not as a
 784 model feature.

Coded Name	Theme	Description	Units	Res
amenities	admin	presence of major or primary roads, markets, or towns within 10 km	numeric (0 - 4)	–
county	admin	county	discrete (Turkana, Marsabit, Isiolo, Wajir, Garissa)	–
dem	admin	elevation from Digital Elevation Model	m	30 m
x	admin	longitude, east/west location	decimal degrees in WGS84	–
y	admin	latitude, north/south location	decimal degrees in WGS85	–
corrneighbor_*	behavior	average daily pump runtime or volume of interpolated 5 most linearly correlated sensed boreholes	hours or L/per capita	5 km
geoneighbor_*	behavior	average pump runtime or volume of 5 closest by Euclidean distance sensed boreholes	hours or L/per capita	–
geoneighbor_dist	behavior	distance between site or grid centroid and 5 closest sensed boreholes	km	–
proxy_*	behavior	modeled average daily pump runtime or volume based on static site characteristics	hours or L/per capita	–
boreholes_krapid	demandproxy	number of boreholes (inventory identified by Kenya RAPID) within 10 km	numeric	–
dams_or_pans	demandproxy	number of dams and sand pans (identified by Kenya RAPID) within 10 km	numeric	–
livestock_h2o	demandproxy	estimated amount of water required for livestock in 2015	discrete (>0.26, >0.50, >1.0, >1.5, >3.5 m ³ /km ² /day)	subcounty
livestock_route	demandproxy	presence of major livestock migration route within 10 km	binary (0, 1)	–
month_factor	demandproxy	categorical month	discrete (1 - 12)	–

Coded Name	Theme	Description	Units	Res
people_h2o	demandproxy	estimated domestic water demand in 2015	discrete (>1030, >2000, >6000, m ³ /day)	subcounty >4000, >8000
popldens ¹	demandproxy	estimated population density, number of people per square kilometer	numeric	1 km
baseflow	landsurface	total baseflow, streamflow that is sustained between precipitation events	mm	3 km
et	landsurface	total actual evapotranspiration	mm	3 km
gvf	landsurface	greenness vegetation fraction	proportion (0 - 1)	4 km
maxtemp	landsurface	average daily maximum temperature	degrees Celsius	50 km
ndvi	landsurface	average 10-day maximum Normalized Difference Vegetation Index value	proportion (0 - 1)	250 m
runoff	landsurface	total amount of flow of water across the ground surface after it no longer infiltrates the soil	mm	3 km
soilm1	landsurface	average daily soil moisture in top layer at depths of 0 - 10 cm	m ³ /m ³	3 km
soilm2	landsurface	average daily soil moisture at depths of 10 - 40 cm	m ³ /m ³	3 km
soilm3	landsurface	average daily soil moisture at depths of 40 - 100 cm	m ³ /m ³	3 km
soilm4	landsurface	average daily soil moisture at depths of 100 - 200 cm	m ³ /m ³	3 km
chirps	surfacewater	total precipitation from Climate Hazards Group InfraRed Precipitation with Station (CHIRPS) v2.0	mm	5 km
naturalh2o	surfacewater	presence of floodplains, lakes, springs, or river basins with 10 km	numeric (0 - 4)	–
tahmo_chirps	surfacewater	total localized precipitation, CHIRPS v2.0 scaled and bias-corrected with Trans-African Hydro-Meteorological Observatory (TAHMO) weather stations	mm	5 km
waterpoint_depth	surfacewater	average daily interpolated relative surface water depth	percentage (0 - 100)	5 km

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