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2 Sub-continental scale carbon stocks of individual trees in African drylands

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28 The distribution of dryland trees and their density, cover, size, mass, and carbon content are not well known 29 at sub-continental to continental scales¹⁻¹⁴. This information is important for ecological protection, carbon 30 accounting, climate mitigation, and restoration efforts of dryland ecosystems¹⁵⁻¹⁸. We assessed over 9.9 31 billion trees derived from over 300,000 satellite images, covering semi-arid Sub-Saharan Africa north of the 32 equator. We attributed wood, foliage, and root carbon to every tree in the 0 mm to 1000 mm year⁻¹ rainfall zone by coupling field-data¹⁹, machine learning²⁰⁻²², satellite data, and high-performance computing. Carbon 33 34 stocks of individual trees ranged from 0.54 Mg C ha⁻¹ and 63 kg C tree⁻¹ in the arid zone to 3.7 Mg C ha⁻¹ 35 and 98 kg tree⁻¹ in the sub-humid zone. Overall, we estimated the total carbon for our study to be 0.84 36 (±19.8%) Pg C. Comparisons with 14 previous TRENDY numerical simulation studies²³ for our area found 37 that the density and carbon stocks of scattered trees have been underestimated by 3 models and over-38 estimated by 11 models, respectively. This benchmarking can help understand the carbon cycle and address concerns regarding land degradation²⁴⁻²⁹. We make available a linked database of wood mass, foliage mass, 39 40 root mass, and carbon stock of each tree for scientists, policymakers, dryland restoration practitioners, and 41 farmers, who can use it to estimate farmland tree carbon stocks from tablets or laptops.

43 Introduction

44 Improved knowledge of dryland trees, defined here as having a green crown area $> 3 \text{ m}^2$ with an associated

45 shadow (Extended Data Fig. 1), is essential to understand their roles in local livelihoods, economies,

46 ecosystems, the global carbon cycle, and the climate system in general. Basic information about the
 47 distribution of dryland trees and their density, cover, size, mass, and carbon content are not well known²⁻⁵.

distribution of dryland trees and their density, cover, size, mass, and carbon content are not well known²⁻⁵.
This knowledge is required for understanding the functional traits of trees in relation to water resources with

49 changes in climate, predicted increase in aridity, and the number and duration of drought events^{30,31}. The

50 sources of information used to estimate carbon stocks in drylands include field surveys at plot scale;

51 ecosystem models²³; and low, moderate and high resolution satellite images⁴⁻¹⁴ which are used to infer bulk

52 properties such as averages of tree cover, dry masses, and carbon density per unit area at a much coarser 53 spatial scale than individual trees.

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55 While most emphasis is put on the development of advanced monitoring techniques for forested ecosystems, 56 none of these sources combine wide/total coverage and representation of each individual tree⁵. Reaching this 57 level of detail is critical for dryland monitoring and management because dryland trees grow isolated and in 58 highly variable size and density. Most current studies producing or using areal averages of tree cover, wood mass, or carbon stocks in drylands are either at the very local level¹², or the information for drylands is 59 derived from global maps¹³, which are rarely trained and validated in drylands, and often apply the same 60 method both on forests and dryland vegetation⁶⁻⁸. While national tree inventories exist for few dryland 61 62 countries, the amount of labor required and their uncertainty are high. As a result, all existing assessments on 63 dryland carbon stocks are highly uncertain, very difficult to validate, and do not provide the means for a detailed characterization at the level of individual trees¹⁴. Furthermore, the contribution of different dry mass 64 65 components - wood, foliage, and root mass - to the overall carbon stock is unknown at large scales.

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At the same time, it remains unknown if ecosystem models quantify the right amount of carbon, and the lack of validation of global models or maps in dry areas fuels narratives of possible under-estimation or overestimation of drylands' carbon stocks and their role in accelerating or mitigating climate change^{12,18}. The missing information on trees at the level of individuals is decisive for improved management of woody resources in drylands: to accurately monitor deforestation spurred by clearing of trees for cropping, mining, infrastructure, and urban development²⁴. In addition, accurate monitoring of the tree resource at the level of individual trees is instrumental for tree planting initiatives, for reporting the correct number of trees and

74 carbon stocks for national reporting schemes, such as the Paris Agreement, or to have a reliable system that

75 allows payments for environmental services to farmers and villages. While deforestation and afforestation

areas can be accurately mapped using current methods and data in forest ecosystems, no monitoring system
 exists for trees outside forests and their carbon pools³².

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79 Currently large amounts of funding are being allocated to dryland restoration activities, and the monitoring 80 of success or failure is based on local inventories lacking large-scale assessments of survival rates of planted 81 trees. The Great Green Wall for the Sahara and the Sahel initiative have recently been subject to renewed interest and increased investments³³⁻³⁵. This initiative was conceived to address the increasing challenges of 82 desertification and drought, food insecurity, and poverty in the wake of climate change. Yet, tracking of 83 84 projects and their successfulness remains a major challenge, as no monitoring system is in place. Equally 85 important, large-scale monitoring of single trees will create a foundation for establishing improved 86 knowledge on the functional traits of dryland trees, such as survival, growth, and mortality, controlled by the 87 complex interplay between biotic and abiotic factors³². Afforestation initiatives should also be rooted in a 88 solid ecological understanding of the local environment to avoid causing water shortages for small-holder 89 farming systems³³.

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91 The combined use of very high-resolution satellite images and artificial intelligence made it possible to 92 identify isolated trees and map their crown area at large scales, covering the western Sahara-Sahel-Sudan 93 areas¹. This approach of mapping individual trees has been extended to a 7.5 times larger area covering the 94 drylands across Africa, from the Atlantic Ocean to the Red Sea from 9.5° to 24° N latitude between the 0 and 95 1000 mm/year isohyets, using 326,523 satellite images at a 50 cm spatial resolution, and coupled with 96 machine learning to map 9.9 billion trees (Fig. 1, Methods). The large-scale mapping of individual tree 97 crowns provides an unprecedented opportunity to apply allometric equations to estimate carbon stocks 98 derived from foliage, wood, and root dry masses at local scales to large regions, here close to 10,000,000 99 km² (Extended Data Fig. 2). We take this step to assess the woody carbon pool by adding up tree-by-tree 100 values, calculated using allometric equations to predict foliage, wood, and root dry masses from crown area 101 multiplied by the average carbon concentration (0.47). These allometric equations were established by 102 destructive sampling of trees from 26, 27, and 5 species, respectively, selected within a rainfall gradient from 103 150-800 mm/year. Comparisons with allometric equations established in wetter tropical areas ensure 104 applicability of these equations to wetter zones, at least up to 1000 mm/year rainfall¹⁹. We estimated the 105 combined uncertainty from the allometric equations and the tree crown detection to be $\pm 19.8\%$.





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- 117 The information of carbon stocks of 9.9 billion trees is compared with a set of state-of-the-art TRENDY
- 118 ecosystem models²³ as well as current satellite observation-based regional carbon stock maps⁶⁻¹¹. We
- 119 introduce a publicly available "viewer", which allows farmers, villagers, policy makers, and all stakeholders
- 120 to retrieve the foliage wood and root masses and the corresponding carbon stock of each tree using a mobile
- 121 device. We expect that this could revolutionize not only the level of information available, but also the

- reporting and monitoring of trees and their carbon stocks at various scales, from the individual field plot to country scales.
- 123 country scales.

125 **Results**

126 Carbon stocks at the tree level

127 We applied a deep learning-based tree mapping on a large number of satellite images and measured 128 9,947,310,221 tree crowns: all woody plants with a shadow and a crown area $>3 \text{ m}^2$ from the hyper-arid (0-129 150 mm/year), arid (150-300 mm/year), semi-arid (300-600 mm/year), and the dry sub-humid (600-1000 130 mm/year) rainfall zones of tropical Africa north of the equator and south of the Sahara (Fig. 1). The average 131 carbon stock of a single tree is 51 kg C in the hyper-arid, 63 kg C in the arid, 72 kg C in the semi-arid, and 132 98 kg C in the sub-humid zone. The individual tree information was projected to the area by calculating the carbon density in Mg C ha⁻¹, which was on average 0.03 Mg C ha⁻¹ in the hyper-arid, 0.54 Mg C ha⁻¹ in the 133 arid, 1.54 Mg C ha⁻¹ in the semi-arid, and 3.73 Mg C ha⁻¹ in the sub-humid zone. While foliage mass has a 134 small overall fraction of the total dry mass (3%), it is an important variable for quantification of browse 135 136 potential and serves as a proxy for other ecosystem processes, such as transpiration, photosynthesis, and

- 137 nutrient cycling. The proportion of root mass is on average 15-20 % of the total mass.
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139 Current carbon map and model comparisons

140 We compared our aboveground carbon density maps (foliage + wood) derived from individual trees with

- 141 current state-of-the-art maps (Fig. 2 and Extended Data Fig. 3) available at moderate spatial resolutions of 30
- 142 to 1000 m. The temporal dynamics were assessed by low-frequency passive microwaves (L-VOD)^{36,37} which
- have emerged as a tool for the assessment of carbon stock dynamics at the 25 x 25 km spatial scale
- 144 (Extended Data Fig. 4). Moreover, we compared carbon density maps and dynamics with dynamic
- 145 ecosystem models from the TRENDY database with a 50 x 50 km grid cell sizes²³. None of these maps were
- 146 designed specifically for drylands; most dynamic ecosystem models and satellite-based models are
- 147 developed and trained for forest ecosystems, and in the case of the TRENDY models, used meteorological
- 148 forcings and prescribed vegetation maps that contain additional uncertainties for comparative purposes.
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Existing carbon density maps compare differently to our assessment based on individual trees and there is little spatial agreement among the maps (Fig. 2a, Fig. 2b). Interestingly, while areas of scattered trees having

- a relatively low carbon density are largely mapped as zero carbon in previous maps except for reference⁹,
- areas of denser tree cover and some areas typically without trees, such as wetlands, irrigated croplands, and
- 154 desert mountains have considerably higher values than our assessment. This leads to an overall higher carbon
- stock of the area compared to our results. Although we do not map herbaceous vegetation in our study, the
- tree cover we map can be used to disaggregate herbaceous vegetation from trees (Extended Data Fig. 5).







159 Fig. 2 | Comparisons between current carbon density maps and our estimations derived from 9.9 **billion trees. a**, Carbon density from state-of-the-art maps using satellite data⁶⁻¹¹. Tree carbon from this study 160 161 is derived from wood + foliage + root mass plotted with ± 1 standard deviation. In the grey zone. **b**, Aboveground carbon stocks aggregated over the 0-1000 mm/year rainfall zone using the legend between Fig. 2 a & 162 163 **b**. Our estimations (grey color) of 0.68 Pg are wood + foliage carbon. The combined uncertainty from neural 164 net area mapping, tree crown omission and commission errors, and allometric conversion of tree crowns into 165 tree wood, foliage, root carbon was \pm 19.8% (see Methods Section). **c**, Vegetation carbon density from the 166 mean of 14 TRENDY dynamic ecosystem models and data from six individual TRENDY models for above and belowground carbon²³ are compared to our tree carbon with aboveground herbaceous carbon added from 167 passive microwaves³⁶. **d**, Aboveground carbon density from the LPJ-GUESS model²³, selected here as it uses 168 169 trees outside of the prescribed forest fraction, and our estimations are compared along the rainfall gradient. L-VOD³⁷ was converted to carbon density using coefficients from a linear correlation with our map 170 (Extended Data Fig. 4). Aboveground herbaceous carbon was derived from³⁶. Our sample size for 0 to 1000 171

172 mm/year was 9,947,310,221 tree crowns >3 m².

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174 At regional scales, dynamic ecosystem model vegetation carbon shows a considerable variability, but the

175 mean follows our estimates of both herbaceous, wood, foliage, and root carbon along the rainfall gradient

176 (Fig. 2c). Notably, while previous studies assumed ecosystem models underestimated dryland carbon stocks,

- 177 our results show overall higher values from the model outputs as compared to the assessment based on
- 178 individual trees, although large variations between models exist. Only considering aboveground carbon, the
- example of LPJ-GUESS shows slightly lower values than our assessment up to about 800 mm/year rainfall(Fig. 2d).
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182 Both ecosystem models and previous satellite-based carbon maps diverge markedly from our results beyond 183 700-800 mm/year rainfall. All other maps assume a continuous increase beyond this rainfall zone, yet our 184 results reach a plateau at 800 mm/year and no further increase in carbon is observed with higher rainfall up 185 to 1000 mm/year. We acknowledge the uncertainty of our results can increase with denser canopy cover, and 186 that we miss all understory vegetation. However, statistical evaluations of the rainfall-tree density 187 relationship from our data confirmed that neither carbon stocks per tree (Fig. 1d) nor tree cover further 188 increased between 800 and 1000 mm/year rainfall (Fig. 3a). Trees with crown area < 50 m² make up 88 % of 189 the total number of trees while trees in the semi-arid and sub-humid zones constitute 90% of the total carbon 190 in our study (Fig. 3).

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193 Fig. 3 | Precipitation, tree carbon, and crown area. a, The tree carbon probability density function 194 computed along the rainfall gradient of the study from the hyper-arid (0-150 mm/year), arid (150-300 195 mm/year), semi-arid (300-600 mm/year), and dry sub-humid (600-1000 mm/year) rainfall zones. The 196 percentage area of each semi-arid zone is shown in blue and the percentage of total carbon in red. The 197 increasing tree carbon probability function shows the importance of precipitation for tree carbon in semi-arid 198 regions. The majority of tree carbon is found in the semi-arid (26%) and dry sub-humid zones (64%) which 199 represent only 30% of the area within our study. The percent carbon density contribution by rainfall zones is 200 linearly related to the tree carbon density (Mg C/ha) reported in Figure 1c by a factor of 2.5. b. A total of 88.4% of our mapped trees had crown areas $<50 \text{ m}^2$. The average tree crown area in 0-150 mm/year zone 201 was 15.1 m²; for the 150-300 mm/year zone it was 18.4 m²; for the 300-600 mm/year zones it was 20.9 m²; 202 203 and for the 600-1000 mm/year zone it 28.1 m². Only 11.6% of our mapped trees had crown areas $>50 \text{ m}^2$ 204 and less than 0.6% had crown areas $>200 \text{ m}^2$.

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206 Application at tree level

207 The comparison with Dynamic Global Vegetation Models and existing biomass maps reveals some similar

208 patterns at coarse scale, yet none of these maps can be used to derive information at the level of individual

trees needed to support policy- and decision-makers. For this reason, we introduce a "viewer" (Fig. 4), which

210 is built on Mapbox and OpenStreetMaps, and can be accessed online by everyone and from anywhere. The

211 viewer includes all 9.9 billion trees as objects, and the wood, foliage, and root mass can be accessed

- 212 individually for each of them. As an example, we show the area of Widou Thiengoly, an area in Senegal
- 213 where tree planting for the Green Wall has been promoted over the past decades (Fig. 4a). While previous
- assessments on the success of tree plantations were based on narratives, visual interpretations, or site visits,
- the "viewer" provides an unbiased tool for evaluating success and failure of initiatives, as well as quantifying
- the carbon stocks gained by each planted tree or lost by each removed/dead tree. The example shown in Fig.
- 217 4 illustrates that high density plantations in this arid region reach carbon density values of about 5 Mg C ha⁻¹
- 218 (Fig. 1c), but the survival rate of planted trees has been a long-lasting concern that needs to be carefully
- 219 monitored to be able to assess the efficacy of Sahelian tree planting programs.



Fig. 4 | Different components of the "viewer". This example shows Widou Thiengoly in semi-arid Senegal 221 222 surrounded by tree plantations, which are partly related to the Great Green Wall³⁴ project aiming to increase 223 tree density and improve livelihoods in the Sahel. **a**, Tree crown segmentations from the neural net mapping. 224 b, Wood, foliage, and root carbon calculated for each tree (see Methods). c, Carbon density per hectare 225 aggregated from carbon stocks of single trees to the hectare scale. d, Our "viewer" includes all information 226 from a-c. This online tool provides information on crown size; foliage, wood, and root carbon of single trees 227 and aggregates carbon to the hectare scale. These data can easily be accessed by policymakers and 228 stakeholders to monitor particular areas of interests. The viewer can be accessed at 229 https://trees.pgc.umn.edu/app.

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Another example shows an agroforestry region in Senegal, north of Khombole that has a relatively high
 density of trees, which has increased the region's carbon stocks considerably. The example area shown in
 Fig. 5 has almost doubled carbon density between 2002 and 2021 (Fig. 5).



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Fig. 5 | Monitoring at the level of single trees from Khombele, Senegal. a, A 50 cm-scale image from 2002; and b, a 50 cm satellite image from 2021 showing an agroforestry area at the same location. Tree cover has increased between 2002 and 2021 and the carbon density of both areas was calculated and increased from 6 to 10 Mg ha⁻¹. A large number of trees grow on farmlands keeping the soils fertile and reducing the need for fallow periods. The greyscale of the background images indicates the carbon density per hectare while the color scale shows the carbon content of individual trees. This is a good example of the tree restoration monitoring potential in our study area.

244 **Discussion**

245 Our assessment is a large-scale estimation of wood, foliage, and root carbon at the level of individual trees. 246 The finding that global ecosystem models and previous carbon density maps estimate higher carbon stocks in 247 African drylands compared to our assessment based on 9.9 billion individual trees seems surprising, as 248 current tree cover maps are not able to correctly account for scattered trees and thus should considerably 249 underestimate the number of trees in these areas¹. The explanation for this apparent paradox--higher tree cover but less carbon--is related to the fact that previous models are rarely developed, trained, and validated 250 251 with plots of very sparse tree cover, thus leaving high uncertainty for drylands with scattered trees. 252 Consequently, areas with scattered trees are often represented by zero-values (Fig. 6), while the carbon 253 density of larger groups of trees may be overestimated in previous assessments, as these areas are wrongly 254 considered as dense forests. In essence, most previous assessments do not accurately map carbon density 255 below 10 Mg C ha⁻¹, if at all, and may overestimate the carbon stocks of dryland "forests". Moreover, if the 256 region is taken as a whole, green crops and herbaceous vegetation impact optical images while steep 257 topography and wetlands/irrigated areas impact the radar backscatter, both predicting higher carbon stocks 258 than our estimations. While we used allometric equations specifically developed from locally sampled field 259 data⁶, 95% of the trees we mapped had a crown area < 78 m². This introduces a small uncertainty in carbon 260 values for the 5% of tree crowns > 78 m² in more humid areas, where trees are taller and/or larger.



Fig. 6 | Comparison of different carbon density maps. We show the dry mass and carbon density for our
study area derived from different sources. Areas beyond 1000 mm/year rainfall are masked out. Data are
from a, Santoro et al.¹¹, b, Baccini et al.⁷, c, Hanan et al.⁹, d, Bouvet et al.⁸, and e, Tucker et al. this paper.
See also Extended Data Fig. 3.

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Nevertheless, the divergence between our results and previous assessments in higher rainfall zones needs to be further investigated and our maps should be used with caution beyond 800 mm/year rainfall. The indirect inclusion of the tree height and the application of the same equation to all tree species are uncertainty factors that will be assessed in future versions of the dataset. Finally, the fact that larger trees shade out smaller trees in areas of dense tree cover, makes the method based on individual tree counting less suited for more humid areas.

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Herbaceous dry mass can contribute considerably to the annual carbon density. However, most herbaceous
plants of the region are annuals that die off each year and do not constitute a residual carbon stock but have a
high inter-annual variability. The herbaceous mass used in our study³⁶ shows the seasonal peak value, which

drops by about 25% within only a few weeks (Extended Data Fig. 1a and Extended Data Fig. 5).

Traditionally, remotely sensed separation of herbaceous vegetation from woody foliage is challenging both
with optical and radar satellite data. We overcome this by measuring individual tree crown areas.

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281 The carbon difference between ecosystem models and our study can be explained by different forest

282 fractions assumed by each model (Extended Data Fig. 6). Most of the dynamic global vegetation models do

283 not simulate trees outside forests and woody carbon is usually a sum of pre-defined forest areas. Differences

- 284 may also result from a simplistic implementation of disturbances, in particular fire, grazing, and the fact that
- 285 we did not include belowground herbaceous carbon in our estimates. Still, the results of the dynamic
- 286 vegetation models are closer to our estimations than originally assumed, and the inclusion of our data may
- 287 improve future modelling results leading to more realistic forecasts of the impact of climate change on
- 288 drylands.289

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290 Dryland trees are not only a carbon stock but provide ecosystem services valuable to the environment and

support local livelihoods, including timber, fuel wood, protection against soil erosion and loss, soil

fertilization, shade, and nutrition for tree crops¹⁵. The benefits of increased tree cover are many and establishing an operational monitoring system for dryland trees is critically needed. The dynamics of growth and mortality of trees outside forests goes undetected by conventional monitoring systems based on satellite imagery with a spatial resolution >10 m. While our current assessment at the level of individual trees does not yet include a temporal dimension (except for the exemplary case provided in Fig. 5), it is a baseline of the number, mass, and carbon stock of trees outside forests at sub-continental scale. The publicly available viewer makes this information accessible for scientists, policymakers, stakeholders, and individual farmers

who can easily quantify woody carbon stocks of a given area, down to the level of a single tree growing in aprivate yard.

A next step will be adding a temporal dimension to the wall-to-wall mapping we describe and we expect it to be possible from this source of data at least with decadal time-steps. This will facilitate addressing the impact of droughts, restoration, and policies at various scales, down to the level of individual trees. High spatial resolution is the key to improved tree inventories in drylands. The ever-increasing availability of satellite images will make continental-scale assessments of carbon pools and dynamics at the individual tree level realistic in near-real time. This will be key to develop robust schemes for dryland management plans needed to achieve the United Nations Sustainability Goals. Our paper is a step in that process.

310 References

510	References	
311	1.	Brandt, M. et al. An unexpectedly large count of trees in the West African Sahara and Sahel.
312		Nature 587, 78-82 (2020). https://doi.org/10.1038/s41586-020-2824-5
313	2.	Axelsson, C. R. & Hanan, N. P. Patterns in woody vegetation structure across African
314		savannas. Biogeosciences 14, 3239-3252 (2017). https://doi.org/10.5194/bg-14-3239-2017
315	3.	Crowther, T. W. et al. Mapping tree density at a global scale. <i>Nature</i> 525 , 201-205 (2015).
316		https://doi:10.1038/nature14967
317	4.	Bastin, JF. et al. The extent of forest in dryland biomes. Science 356 , 635-638 (2017).
318		https://doi:10.1126/science.aam6527
319	5.	Hansen, M. C. et al. High-resolution global maps of 21st-century forest cover change. Science
320		342 , 850-853 (2013). https://doi:10.1126/science.1244693
321	6.	Avitabile, V. et al. An integrated pan-tropical biomass map using multiple reference datasets.
322		Glob. Change Biol. 22, 1406-1420 (2016). https://doi: 10.1111/gcb.13139
323	7.	Baccini, A. et al. Estimated carbon dioxide emissions from tropical deforestation improved by
324		carbon-density maps. Nat. Climate Change 2, 182-185 (2012).
325		https://doi.org/10.1016/j.rse2017.12.12.030

326	8.	Bouvet, A. et al. An above-ground biomass map of African savannahs and woodlands at 25 m
327		resolution derived from ALOS PALSAR. <i>Remote Sens. Environ</i> . 206 , 156-173 (2018).
328		https://doi.org/10.1016/j.rse.2017.12.030
329	9.	Hanan, N.P. Gridded Estimates of Woody Cover and Biomass across Sub-Saharan Africa,
330		2000-2004. ORNL DAAC, Oak Ridge, Tennessee, USA (2020).
331		https://doi.org/10.3334/ORNLDAAC/1777
332	10.	Saatchi, S.S. et al. Benchmark map of forest carbon stocks in tropical regions across three
333		continents. Proc. Natl. Acad. Sci. USA. 108, 9899-9904 (2021).
334		https://doi.org/10.1073/pnas.1019576108
335	11.	Santoro, M. et al. The global forest above-ground biomass pool for 2010 estimated from
336		high-resolution satellite observations. <i>Earth System Sci. Data</i> 13 , 3927-3950 (2021).
337		https://doi.org/10.5194/essd-13-3927-2021
338	12.	Skole, D.L., Samek, J.H., Dieng, M. & Mbow, C. The Contribution of Trees Outside of Forests to
339		Landscape Carbon and Climate Change Mitigation in West Africa. Forests 12, 1652 (2021).
340		https://doi.org/10.3390/f12121652
341	13.	Harris, N.L. et al. Global maps of twenty-first century forest carbon fluxes. <i>Nat. Clim. Change</i>
342		11, 234-240 (2021). https://doi.org/10.1038/s41558-020-00976-6
343	14.	de Foresta, H. et al. Towards the Assessment of Trees Outside Forests (Forest Resources
344		Assessment Working Paper 182 FAO, Rome) 345 p. (2013).
345	15.	Bayala, J., et al. Parklands for buffering climate risk and sustaining agricultural production in
346		the Sahel of West Africa. Current Opinion on Environmental Sustainability 6, 28-34 (2014).
347		https://doi.org/10.1016/j.cosust.2013.10.004
348	16.	Stringer, L. C. et al. Challenges and opportunities in linking carbon sequestration, livelihoods
349		and ecosystem service provision in drylands. Environ. Sci. & Policy 19-20, 121-135 (2012).
350		https://doi.org/10.1016/j.envsci.2012.02.004
351	17.	Brito, J. C. et al. Unravelling biodiversity, evolution and threats to conservation in the Sahara-
352		Sahel. Biol. Rev. Cambridge (UK) Philosophy Soc. 89, 215-231 (2014).
353		https://doi:10.1111/brv.12049.
354	18.	Bastin, JF. et al. The global tree restoration potential. Science 365 , 76-79.
355		https://doi:10.1126/science.aax0848 (2019).
356	19.	Hiernaux, P. et al. Allometric equations to assess Sahel woody plant dry mass and carbon
357		content from high resolution satellite imagery. Forest Ecol. & Management (under review
358		2022).
359	20.	LeCun, Y., Bengio, Y. & Hinton, G. Deep learning. Nature 521, 436-444 (2015).
360		https://doi.org/10.1038/nature14539
361	21.	Reichstein, M. et al. Deep learning and process understanding for data-driven Earth system
362		science. Nature 566, 195-204 (2019). https:// doi:10.1038/nature14539
363	22.	Ronneberger, O., Fischer P. & Brox, T. U-net: convolutional networks for biomedical image
364		segmentation. In Int. Conf. Med. Image Computing & Computer-Assisted Intervention
365		(eds. Navab, N. et al.) 234-241, (Springer 2015). <u>https://doi.org/10.48550/arXiv.1505.04597</u>
366	23.	Sitch S. et al. Recent trends and drivers of regional sources and sinks of carbon dioxide.
367		Biogeosciences 12, 653-679 (2015). https://doi:10.5194/bg-12-653-2015
368		

369	24.	Kirkby, M. Desertification and development: Some broader contexts. J. Arid Environ.
370		193 , 104575 (2021). https://doi.org/10.1016/j.jaridenv.2021.104575
371 372	25.	0.5 million km2 of drylands towards desertification. <i>Nat. Commun.</i> 11 , 3853 (2020).
373		https://doi.org/10.1038/s41467-020-17710-(2020).
374	26.	Darkoh, M. B. K. The nature, causes and consequences of desertification in the drylands of
375		Africa. Land Degradation & Development 9, 1-20 (1998). https://doi.org/10.1002/(SICI)1099-
376		145X(199801/02)9:1<1::AID-LDR263>3.0.CO;2-8
377	27.	Ribot, J. C. A history of fear: imagining deforestation in the West African dryland forests.
378		Glob. Ecol. & Biogeogr. 8, 291-300 (1999). https://doi.org/10.1046/j.1365-
379		2699.1999.00146.x
380	28.	Fairhead, J. & Leach, M. Reframing Deforestation (Routedge, 238 p. 2003).
381	29.	Reed, J. et al. Trees for life: the ecosystem service contribution of trees to food production
382		and livelihoods in the tropics. Forest Policy & Econ. 84, 62-71 (2017).
383		https://doi.org/10.1016/j.forpol.2017.01.012
384	30.	Descroix, L. et al. Evolution of Surface Hydrology in the Sahelo-Sudanian Strip: An Updated
385		Review. Water 2, 10, 748 (2018). https://doi.org/10.3390/w10060748
386	31.	Wendling, V. et al. Drought-induced regime shift and resilience of a Sahelian
387		ecohydrosystem. Environ. Res. Lett. 14, 10 (2019). https://doi: 10.1088/1748-9326/ab3dde
388	32.	Skole, D.L. et al. Trees outside of forests as natural climate solutions. <i>Nat. Clim. Change</i> 11 ,
389		1013-1016 (2021) https://doi.org/10.1038/s41558-021-01230-3
390	33.	O'Connor, D. & Ford, J. Increasing the Effectiveness of the "Great Green Wall" as an
391		Adaptation to the Effects of Climate Change and Desertification in the Sahel. Sustainability 6,
392		7142-7154 (2014). https://doi: 10.3390/su6107142
393	34.	Mirzabaev, A. et al. Economic efficiency and targeting of the African Great Green Wall. <i>Nat.</i>
394		Sustainability 5 , 17-25 (2022). https://doi.org/10.1038/s41893-021-00801-8.
395	35.	Sacande, M. & Berrahmouni, N. Community participation and ecological criteria for selecting
396		species and restoring natural capital with native species in the Sahel. Restoration Ecology 24,
397		479-488 (2016). https://doi10.1111/rec.12337
398	36.	Brandt, M. et al. Changes in rainfall distribution promote woody foliage production in the
399		Sahel. Commun. Biol. 2, 133 (2019). https://doi.org/10.1038/s42003-019-0383-9
400	37.	Wigneron, J-P. et al. SMOS-IC data record of soil moisture and L-VOD: Historical
401		development, applications and perspectives. Remote Sens. Environ. 254, 2021,
402		https://doi.org/10.1016/j.rse.2020.112238 (2021).
403	38.	Dardel, C. et al. Re-greening Sahel: 30 years of remote sensing data and field observations in
404		Mali and Niger. <i>Remote Sens. Environ</i> . 40 , 350-364 (2014).
405		https://doi:10.1016/j.rse.2013.09.011 (2014).
406	39.	Tucker, C. GIMMS Global Agricultural Monitoring System. https://glam1.gsfc.nasa.gov (2022).
407		
408	Methods	
409		
410	Overview	

- 411 This study establishes a framework for mapping carbon stocks at the level of individual trees at a sub-
- 412 continental scale in semi-arid Sub-Saharan Africa north of the equator. We used satellite imagery from the
- early dry season (Extended Data Fig. 1). The deep learning method developed by a previous study¹ allowed 413
- 414 us to map billions of discrete tree crowns at the 50 cm scale from West Africa to the Red Sea. Then we used
- 415 allometry to convert tree crown area into tree wood, foliage and root carbon for the 0 to 1000 mm/year
- 416 precipitation zone where our allometry was collected (Extended Data Fig. 2). We introduce a viewer that
- 417 enables the billions of trees to be viewed at different scales, with information on location; meta data of the
- 418 Maxar satellite image used; tree crown area; and the estimated wood, foliage and root carbon content based
- 419 on our allometry (Fig. 4). We also make available our output data for the 1000 mm/year precipitation zone
- 420 southward to 9.5° N latitude with information on location, precipitation, meta data of the Maxar satellite 421 image used, tree crown area, tree wood carbon, tree root carbon, and tree leaf carbon.
- 422

423 Satellite imagery

- 424 We utilized 326,523 Maxar multi-spectral images from the OuickBird-2, GeoEve-1, WorldView-2, and
- 425 WorldView-3 satellites collected from 2002 to 2020 from November through March from 9.5° to 24° North
- 426 latitude within Universal Transverse Mercator (UTM) Zones 28 to 37 for Africa (Extended Data Table 1a).
- 427 These images were obtained by NASA through the NextView License from the National Geospatial
- 428 Intelligence Agency. Data were assembled over several years with a focus on later years to achieve a relatively recent and complete wall-to-wall coverage.
- 429 430
- 431 When using satellite data from different satellites over a several years, with varying Sun-target-satellite 432 angles, with varying radiometric calibration of satellite spectral bands, and different atmospheric 433 compositions through which the surface is imaged, there are two possibilities for using hundreds of 434 thousands of satellite images together quantitatively. One approach, used extensively in NASA's, 435 NOAA's, and the European Space Agency's Earth-viewing satellite programs, is to quantitatively inter-436 calibrate radiometrically the satellite channels through time; correct these data for time-dependent 437 atmospheric effects such as aerosols, clouds, haze, smoke, dust, and other atmospheric constituent effects, and then normalize the viewing perspective to the same Sun-target-satellite angle⁴⁰. Another approach is to 438 439 use the satellite data as collected; assemble training data of trees viewed from different satellites under 440 different Sun-target-satellite angles, different times, different atmospheric conditions, and employ 441 machine learning with high performance computing to perform the tree mapping at the 50 cm scale. 442 The key to successful machine learning is to account for all the sources of variation within the domain of 443 study in the training data to ensure accurate identification of trees under all circumstances. We included 444 trees viewed substantially off-nadir, trees collected under different aerosol optical thicknesses, trees collected under cirrus cloud conditions, trees viewed in the forward and backward scan directions, trees on 445 446 sandy soils, trees on clay soils, trees on burn scars, trees in laterite areas, and trees in riverine settings. Our 447 training data were collected by one team member and are a carefully selected manual delineation of 448 89,899 individual trees under a range of atmospheric conditions, viewing perspectives, and ecological settings.
- 449
- 450 451 All multi-spectral and panchromatic bands associated with our Maxar images were orthorectified to a
- 452 common mapping basis. We next pan-sharpened all multi-spectral bands to the 0.5 m scale with the
- 453 associated panchromatic band. The absolute locational uncertainty of pixels at the 0.5 m scale from orbit is
- 454 approximately ± 11 m, considering the root mean square location errors among the QuickBird-2, GeoEye-1,
- 455 WorldView-2, and WorldView-3 satellites (Extended Data Table 1). We formed the NDVI⁴² from every
- 456 image in the traditional way from the pan-sharpened red and near-infrared bands. We also associated the
- 457 panchromatic band with the NDVI band and ensured the panchromatic and NDVI bands were highly co-

458 registered. The NDVI was used to distinguish tree crowns from non-vegetated background because the

- 459 images were taken from a period when only woody plants were photosynthetically active in this area³⁶. Our
- 460 training data were labeled on images from the early dry season when only trees have green leaves. Because
- 461 most semi-arid savanna trees continue to photosynthesize in the early dry season after herbaceous vegetation
- senesces, green leaf tree crowns are easily mapped because of their higher NDVI values than their senescent
- 463 herbaceous vegetation surroundings. We substantiate this by analysis of 308 individual trees using NDVI
- time series with 4 m PlanetScope imagery that emphasized the importance of satellite data from the
- 465 November, December, and January early dry-season months (Extended Data Fig. 1).
- 466
- 467 We next formed our data into mosaics by applying a set of decision rules, resulting in a collection of 16 x 16 468 km tiles within each UTM Zone from 9.5° N to 24° N latitude for Africa. The first round of scoring 469 considered percentage cloud cover, sun elevation angle, and sensor off nadir angle: preference was given to 470 imagery that had lower cloud cover, then higher sun elevation angle, and finally view angles closest to nadir. 471 In the second round of scoring, selections were prioritized to favor early dry season months and off-nadir 472 view angles: preference was given to imagery from November, December, and January with off nadir angle 473 $<\pm 15^{\circ}$ degrees; second to imagery from November through January with off nadir angle between $\pm 15^{\circ}$ and 474 30° degrees; third to imagery from February or March with off nadir angle less than $\pm 15^{\circ}$ degrees; and 475 finally to imagery from February or March with off nadir angle between $\pm 15^{\circ}$ and 30° degrees. Image 476 mosaics are necessary to eliminate multiple counting of trees. We formed mosaics using 94,502 images for 477 tree segmentation with 94% of these being from November, December, and January. 90% of our selected 478 imagery was within $\pm 15^{\circ}$ of nadir and 1% of our study area was identified as having insufficient data
- quality. In addition, 87% of our data were between 2010 and 2020 and 94% were from the early dry season
 (Extended Data Fig. 7).
- 481

Possible obscuration of the surface by clouds totaled 4.1 % of our input data area and aerosol optical depth
>0.6 at 470 nm⁴¹ totaled 3.4% of our input data. However, we mapped 691,477,772 trees in our possible
cloud cover- and aerosol-affected areas, indicating cloud and aerosol effects were lower than these numbers.

In addition, 0.9% of our input data did not process. We include a data layer in our viewer for these three
conditions.

488 Mapping tree crowns with deep learning

489 We used convolutional neural network models developed by a previous study¹. The models were trained 490 with manually delineated and annotated 89,899 individual trees along a north-south gradient from 0 to 1000 491 mm/year rainfall¹. Only features that showed a distinct crown area and associated shadow were included, 492 which excluded small bushes, grass tussocks, rocks, and other features that might have green leaves or cast a 493 shadow from our classification. All training data and model training was done in UTM zones 28 and 29. 494 Since tree floristic diversity in the 0 to 1000 mm/year zone of our study is highly similar from the Atlantic Ocean to the Red Sea across Africa⁴³⁻⁴⁵, we added no additional training data as our study moved further 495 496 eastward. We utilized state-of-the-art deep learning to segment trees crowns at the 50 cm scale¹. We used two 497 different models based on a U-Net architecture, one for lower-rainfall desert regions with <150 mm/year 498 precipitation and one for regions with average annual precipitation > 150 mm/year. Details about the 499 network architecture, training process, and hyperparameter choices can be found in reference¹. Previous 500 evaluation showed that early dry season images performed better than late dry season images, which was a 501 limitation of our previous study. We reduced this error by using early dry season images with only 6% of our 502 area being covered by images from February and March. The models were also designed to separate

503 clumped trees by highlighting spaces between different crowns during the learning process, similar to a 504 strategy for separating touching cells in microscopic imagery²².

506 Allometry

505

507 Based on¹⁹ we predicted the wood (w), foliage (f) and root (r) dry mass as functions of the crown area (A) 508 of a single tree as:

509 $\max_{w}(A) = 3.9448 \cdot A^{1.1068} \quad (N_{w} = 698)$ $\max_{f}(A) = 0.2693 \cdot A^{0.9441} \quad (N_{f} = 900)$ $\max_{r}(A) = 0.8339 \cdot A^{1.1730} \quad (N_{r} = 26)$

510 The tree mass components of wood, leaves, and roots were combined to predict the total mass(A) in kg of a 511 tree from its crown area A in m²:

 $mass(A) = mass_w(A) + mass_f(A) + mass_r(A)$

As in reference¹, a crown area of size $A > 200 \text{ m}^2$ was split into [A/100] areas of size 100 m² and one area with the remaining m² if necessary. We converted dry mass to carbon by multiplying with a factor of 0.47^{46} .

515

512

516 Uncertainty analysis

517 We evaluated the uncertainty of our tree crown area mapping and carbon estimation in two ways. First, we 518 quantified our tree crown mapping omission and commission errors by inspecting randomly selected areas 519 from UTM Zones 28 to 37, validating that our neural network generalizes over UTM zones consistently 520 (Extended Data Fig. 8).

521

522 Second, we quantified the relative error of our carbon estimation. We consider the uncertainty Δ_x of a

523 quantity x and the corresponding relative uncertainty δ_x defined by the absolute and relative error,

524 respectively⁵⁰. To assess the relative error in carbon estimation due to errors by the neural network, we

525 considered external validation data from¹ which were not used in the model building process. We considered 526 expert-labeled tree crowns as well as the predicted tree crowns from 78 plots of 256×256 pixels. The hand-

525 experimeted tree crowns as well as the predicted tree crowns from 78 plots of 250 × 250 pixels. The nand 527 labeled set contained 5,925 trees, the system delineated 5,915. The total hand-labeled tree crown area was

528 118,327 m² and the neural network predicted 121,898 m². This gave a relative error in the carbon of $\delta_{area} =$

529 3.3%. We matched expert-labeled and predicted tree crowns and computed the root-mean-square error

530 (RMSE) per tree, taking overlapping areas and missed trees into account, see Extended Data Fig. 8. We

531 estimated the allometric uncertainty ($\delta_{allometric}$) using the data from¹⁹ (see below). The two relative errors δ_{area}

532 and $\delta_{\text{allometric}}$ were combined to an overall uncertainty estimate for the carbon prediction of ±19.8% (see 533 below).

533 534

535 Omission and commission errors. We evaluated our tree crown mapping accuracy by analysis of 1028
536 randomly selected 512 x 256 pixel areas over the 9.5° to 24° north latitude within UTM Zones 28 to 37.

537 Because the drier 60% of our study area only contains 1% of the 9,947,310,221 trees we mapped in the 0 to

538 1000 mm/year rainfall zone, we applied an 80% bias for selecting evaluation areas above the 200 mm/year

539 precipitation line⁴⁷, as >98% of tree identifications were above the 200 mm/year precipitation isoline.

540 Identified tree polygons were further categorized into tree crown area classes from $0 - 15 \text{ m}^2$, 15-50 m², 50-

541 200 m^2 , and >200 m² with a total of 50,570 trees evaluated. While a previous study reported greatest

542 uncertainty in both the smallest and largest area classes¹, our more expansive work found the greatest

543 uncertainty in our smallest tree class. We excluded from evaluation any tiles that had annual precipitation⁴⁷

544 > 1000 mm/year and all areas that were devoid of vegetation, leaving us with 850 areas.

546 Seven members of our team evaluated the accuracy in terms of commission and omission by tree crown area

547 classes for the 850 areas. Input data provided for every area were the NDVI layer, the panchromatic layer,

- and the neural net mapping results in each of the 4 crown area classes. Ancillary data available to evaluators
 included the center coordinates for comparison to Google Earth data, the Funk et al.⁴⁷ rainfall, the acquisition
- 550 date of the area evaluated, and the viewing perspective.
- 551

We identified areas wrongly classified as tree crowns (commission errors), missed trees (omission errors), and crown areas corresponding to clumped trees (see Extended Data Fig. 8). Clumped trees were most common for $>200 \text{ m}^2$ tree crown area. They were rare in the 3 –15 m² and 15 – 50 m² tree classes which comprise 88% of our tree crowns. In the 850 patches, the number of trees ranged from 1 tree to 326 trees, with a total of 50,570 trees evaluated and 3,765 errors identified. Overall, the commission and omission error rates were 4.9% and 2.7%, respectively, a net uncertainty of 2.2%.

558

Allometric uncertainty. The prediction of tree carbon from the crown area for a single tree based on crown area alone is inherently uncertain^{48,49}. As the allometric equations are based on three different datasets, we compute their uncertainties independently, combine them, and put them in relation to the total carbon measured in the three datasets.

563

The allometric equations were established using an optimal least-squares fit of an affine linear model predicting the logarithmic carbon from the logarithmic tree crown area¹⁹. To estimate the uncertainty of the allometric equations, we repeated the fitting using random subsampling. The datasets were randomly split into training data (80%) for fitting the allometric equations and validation data (20%) for assessing the uncertainty. For example, from the root measurements,

569 $(A_1, y_1), \dots, (A_{N_r}, y_{N_r})$ we compute $\mu_r = \frac{1}{N_r} \sum_{i=1}^{N_r} y_i$ and $\hat{\mu}_r = \frac{1}{N_r} \sum_{i=1}^{N_r} \max_r (A_i)$. The corresponding error 570 is

- 572
- 571

573 Because the total carbon for a tree with a certain crown area is the sum of the three carbon components, we 574 add the absolute uncertainties assuming independence⁵⁰

 $\Delta_r = |\mu_r - \hat{\mu}_r|$

- 575 $\Delta_{\text{allometric}} \simeq \sqrt{\Delta_f^2 + \Delta_w^2 + \Delta_r^2}$
- and compute the relative uncertainty as

$$\delta_{\text{allometric}} = \frac{\Delta_{\text{allometric}}}{\mu_{\text{mass}}}$$

578 where the average mass μ_{mass} is given by the sum of the averages for wood (μ_w) , leaves (μ_f) , and root (μ_r) . 579 This process was repeated 10 times, resulting in a mean relative uncertainty of

- $\overline{\delta}_{
 m allometric} = 19.5\%$.
- 580 581

577

Total carbon uncertainty. We combine the uncertainties from the neural net mapping and our allometric equations, which can be viewed as considering $(1+A)^*(1+B)$ with A and B being random variables with standard deviations δ_{area} and $\delta_{\text{allometric}}$. Neglecting higher-order and interaction terms, we combine the two sources of uncertainty to $\delta \simeq \sqrt{\delta_{\text{area}}^2 + \overline{\delta}_{\text{allometric}}^2}$ resulting in an uncertainty in total tree carbon for our study of ±19.8%. See also Extended Data Fig. 9 for the root-mean-square errors (RMSEs) of our predicted crown areas calculated on external validation data from¹, binned based on the 50-quantiles of the hand-labeled crown areas, and converted also into carbon. Supplemental Information Fig. 1 is a flow diagram of ourmethods.

- 591 **Our Viewer.** Visualizing our large tree mapping dataset in an interactive format was essential for quality 592 control purposes, exploration of the data, and hypothesis creation. Creating a web-based viewer serves the 593 purpose of being the initial point of interaction with our dataset for fellow researchers, local stakeholders, or 594 the general public. The visualization of over 10 billion trees in a web browser required maintaining 595 performance, interactivity, and individual metadata for each polygon. Users should be able to zoom in to any 596 area within the dataset to view individual tree polygons and query their statistics, while at the same time 597 accurately depicting the overall trends of the dataset at lower zoom levels. The visualization also needed to 598 clearly denote where data were missing or possibly affected by clouds or aerosols. Finally, the extent and 599 origin of the source imagery, its acquisition date, and a preview of the imagery needed to be available. To 600 accomplish these goals, a vector tile-based approach was taken, with the data visualized in Mapbox GL-JS 601 map within a React web application. In order to create vector tiles covering the entire study area, we 602 developed a data processing pipeline leveraging high performance computing resources to transform the data 603 into compatible formats, as well as to package, optimize, and combine the vector tiles themselves.
- 604

590

605 We used two tracks to store and visualize the results of this study on the web: vector polygon data, and 606 generalized rasters representing tree crown density. At the native spatial resolution of 50 cm, the map 607 displays the full-resolution tree polygon dataset. At lower spatial resolution zoom levels rasterized 608 representations of tree density are shown. Visualizing generalized rasters in place of vector polygons 609 improves performance significantly. As users zoom in to higher spatial resolutions, the raster layer fades 610 away and is replaced by the full resolution polygon layer. Once zoomed far enough to resolve individual 611 polygons, users can click to select a polygon to display a map overlay containing various properties of the 612 tree, as well as the date which the source imagery was acquired on and a link to preview the source imagery.

613

Rainfall data. We used Funks et al. rainfall data to estimate annual rainfall at 5.6 km grids⁴⁷. We averaged
the available data from 1982 to 2017 and extracted the mean annual rainfall for each mapped tree and
bilinearly interpolated it to 100 x 100 m resolution. The rainfall data were also used to classify the study area
into mean annual precipitation zones: hyper-arid from 0-150 mm/year, arid from 150–300 mm/year, semiarid from 300–600 mm/year, and sub-humid 600–1000 mm/year zones.

619

620 Code availability. The tree detection framework based on U-Net is publicly available at
621 https://zenodo.org/record/3978185. Please contact AK, CI, MB, or JM for support and more information.

622 Data availability. The viewer can be accessed via <u>https://trees.pgc.umn.edu/app</u>. The Funk et al. rainfall 623 data⁴⁷ are freely available at (http://chg.geog.ucsb.edu/data/chirps/). Commercial very high-resolution 624 satellite images were acquired through the NASA under the NextView Imagery End User License 625 Agreement. The copyright remains at Maxar Inc. and redistribution is not possible. However, the derived 626 products produced by this study are publicly available at the Oak Ridge National Laboratory's Distributed Active Archive Center: https://doi.org/10.3334/ORNLDAAC/2117. Please contact the CT, MB, or PH for 627 more specific requests. A detailed description of our processed data for the 95,402 selected mosaic images, 628 629 including output data, specific cutlines affected by aerosol optical depth and cloud cover; data distributions 630 for year, month, solar azimuth angle, and off-nadir angle for each UTM Zone segment in our study, can also 631 be found at : https://doi.org/10.3334/ORNLDAAC/2117.

- 632
- 633 References

634	40. Schaaf, C.B. et al. 2002. First operational BRDF, albedo, and nadir reflectance products from		
635	MODIS. Remote Sens. Environ. 83, 135-148 (2002). https://doi:10.1016/S0034-		
636	<u>4257(02)00091-3</u>		
637	41. Lyapustin, A., Wang, Y., Korkin, S. & Huang, D. MODIS collection 6 MAIAC algorithm.		
638	Atmos. Meas. Tech. 11, 5741-5765 (2018). https:// https://doi.org/10.5194/amt-11-5741-2018		
639	42. Tucker, C. J. Red and Photographic Infrared Linear Combinations for Monitoring		
640	Vegetation, <i>Remote Sens. Environ.</i> 8, 127-150 (1979).		
641	43. Hiernaux P. & Le Houerou H-N. Les parcours du Sahel. Secheresse 17, 1-21 (2006).		
642	44. White F. The Vegetation of Africa, (UNESCO Press, Paris 356 p. 1983).		
643	45. Schnell R. Introduction a la phytogeographie des pays tropicaux. La flore et la vegetation de		
644	l'Afrique tropicale. (eds, Gauthier-Villars et al.), Vol. 3 (Bordas, Paris. 460 p. 1976).		
645	46. McGrouddy, M.E., Daufresne, T., & Hedin, L. Scaling of C:N:P stoichimetry in forests		
646	worldwide: implications of terrestrial redfield-type ratios. <i>Ecology</i> 85 (9), 2390-2401 (2004).		
647	47. Funk, C. et al. The climate hazards infrared precipitation with stations-a new environmental		
648	record for monitoring extremes. Sci. Data 2, 150066 (2015).		
649	https://doi.org/10.1038/sdata.2015.66		
650	48. Djomo, A. & Chimi, C. Tree allometric equations for estimation of above, below, and total		
651	carbon in a tropical moist forest: Case study with application to remote sensing. Forest Ecol.		
652	& Management 381, 184-193 (2017). https://doi.org/10.1016/j.foreco.2017.02.022		
653	49. Kuyah, S. et al. Allometric equations for estimating carbon in agricultural landscapes: II.		
654	Belowground carbon. Agriculture, Ecosystems & Environ. 158, 225-234 (2012).		
655	https://doi:10.1016/j.agee.2012.05.010		
656	50. Bevington P. & Robinson, D. K. Data Reduction & Error Analysis. McGraw-Hill, New		
657	York, 3rd edition, 338 p. (2003).		
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670 Author Contributions. CT, MB, PH, and AK contributed equally to the paper and should be considered as 671 the first authors. CT and MB coordinated the study and drafted the manuscript with support by KR, RF, and 672 PH. PH, PS, and B-AI collected, assembled, and developed the allometric equations. Satellite data were 673 prepared by JS, S Sinno, JM, CP, and YF. JP and JM directed the evaluation which was conducted by KM, 674 ER, DM, AM, AK, JS, and CT. The tree detection was processed on the BlueWaters supercomputer by JM, 675 JS, and S Sinno. Neural network implementation was provided by AK and CI. Ecosystem models were run 676 and analyzed by BP and PC. LVOD data were prepared by JPW and MB. Data were analyzed by JP, ER, EG, 677 RF, JM, DM, AM, AK, KM, JS, MB, FR, CT, AK, JP, RK, RM, S Saatchi, and YF. Phenology was provided

by YF, PH, and LK. The viewer was developed by PM and SL.

- 679 **Competing financial interest**. The authors declare no competing financial interests.
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685 Extended Data Fig. 1 | Planet NDVI time series for 308 trees contrasted with the background NDVI 686 from the same area. a, Note the separation between the tree NDVI values and the background NDVI for the dry season months of September to March. b, False color tree crown NDVI image from WorldView-3 at 687 688 30 cm spatial resolution showing green leaf vegetation in red colors. Individual trees are evident with green 689 leaf tree crowns with associated shadows. Our mapping of trees with machine learning is based on trees being defined as having a tree crown of at least 3.0 m² with an associated shadow. Areas of low vegetation 690 do not cast shadows sufficiently to be classified as trees. The areas in **a** and **b** are close to 13° 31' N x 2° 40' 691 692 E.





695 Extended Data Fig. 2 | Allometric equations for converting tree crown areas into dry wood mass, dry 696 root mass, and dry foliage mass. a, Allometric equation based on tree crown area to predict wood mass, 697 established from 698 Sahelian and Sudanian woody plants of 27 species. The power model was fitted using 698 log-log regression. The stored carbon is estimated from dry mass by multiplying with a factor of 0.47. Data were collected within the 0 to 800 mm/year long-term precipitation zone, see reference¹⁹. The plot also 699 shows the cumulative percentage up to 15 m² and 50 m² crown area (50% and 88.4% respectively) of the 700 predicted trees in our study (95% of which had crown areas less than 78 m²). **b**, **c**, Same as **a** but for foliage 701 702 and root mass. Allometric equations established from 900 trees of 26 species, and 26 trees of 5 species, 703 respectively.



Extended Data Fig. 3 | Comparison between current carbon density maps and our estimations derived
 from 9,947,310,221 trees. a, Scatterplots showing the spatial agreement at pixel level with all datasets
 aggregated to 1 x 1 km grids. b, We correlated all 1x1 km pixels for each rainfall zone, defined by 100
 mm/year steps, between our carbon density estimates and current state-of-the-art maps⁶⁻¹¹.

713 Extended Data Fig. 4 | Temporal changes in carbon density. a, Scatterplot between the passive

714 microwave L-VOD³⁶ and our carbon density (wood + foliage) aggregated to 25 x 25 km grids. **b**, The linear 715 relationship seen in (a) was used to convert annual L-VOD to the unit carbon density. L-VOD aboveground woody carbon density as well as TRENDY models²³ values (all vegetation carbon) are averaged over the 716 717 study area for 0-1000 mm/year for the 2011-2019 period. Correlating L-VOD from the dry season, to avoid 718 the complication of herbaceous vegetation, with our carbon density map aggregated to 25 x 25 km resolution 719 shows a moderately high level of agreement (r = 0.72), however, the strong scattering especially in low rainfall areas also shows that the uncertainty is high, impeding the use of L-VOD³⁷ for local applications in 720 721 arid areas. Nevertheless, the linear relationship can be used to convert L-VOD to the unit carbon density to 722 derive temporal dynamics in carbon density, which show stable woody carbon stocks during 2010-2019 723 (~2.0 Mg C year⁻¹) for this region, without clear trend or inter-annual variations, suggesting that neither 724 droughts, deforestation, nor restoration had a measurable impact on carbon stocks over the last decade (Fig. 725 3b). TRENDY models show a variety of responses but the ensemble shows a similar behavior as L-VOD, 726 although when herbaceous vegetation and belowground biomass are included, a variety of different

727 magnitudes result.

729 Extended Data Fig. 5 | Seasonal comparison between MODIS NDVI from a 0.25° x 0.25° area and field 730 data collections of dry herbaceous biomass production in kg/ha near Agoufou, Mali within the MODIS 731 area. a, The MODIS 8-day time-step data show the NDVI maximum and minimum range from 2000 to 732 2021 in the gray colored portion, with MODIS NDVI time series by years for 2004 to 2010, and the average 733 MODIS NDVI from 2001 to 2021. Above-ground herbaceous dry mass in b, are biweekly in-situ measurements for a rangeland field site in Mali³⁸ centered at 15.4625° N by 1.4886° W. These show high 734 735 inter- and intra-annual herbaceous dry mass variability, sharp increases in herbaceous dry mass during the 736 wet season, and rapid decreases in early dry season at the few-meter scale. The MODIS NDVI data show 737 similar temporal trends to the herbaceous dry-mass variations in (b) for 2004 to 2010 and put these into 738 context of the 2000 to 2021 MODIS record for a $0.25^{\circ} \times 0.25^{\circ}$ area. Trees are $\sim 3\%$ of the total vegetation 739 cover in this area and thus herbaceous vegetation dominates. The species composition in (b) is dominated by 740 annual grasses such as Aristida mutabilis, Cenchrus biflorus, Brachiaria xantholeuca, and annual 741 dicotyledons such as Zornia glochidiata, Tribulus terrestris. The data in (b) are from a 1 x 1 km location and 742 were selected to be representative of the area. The MODIS NDVI data from in (a) are available for all users with instructions for use for the MODIS record from 2000 to date³⁹. Our tree crown data provide the means 743 744 to separate primary production into herbaceous and tree fractions for semi-arid areas and will improve 745 carbon residence understanding in areas of mixed tree and herbaceous vegetation. 746

For Extended Data Fig. 6 | Forest fractions of different TRENDY²³ models. The figure shows the percentage of forest areas assumed by each model along the rainfall gradient. Woody biomass in most models mainly comes from pre-defined forest areas and consequently results in a high degree of variation within the semi-arid area of our study. This is a good example of the utility of our tree mapping results for more accurate depiction of semi-arid trees for numerical simulation modeling.

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757 Extended Data Fig. 7 | Candidate and selected satellite images for analysis. a, 326,523 Maxar images

covering 90,097,073 km² from November to March were available for our study and were acquired from 2002 to 2020. **b**, The distribution by month of all satellite data in (**a**); **c**, The selected imagery for processing

760 (see Methods) by year totaled 94,502 images that covered 9,685,324 km² with 87% of the satellite data from

761 2010 to 2020; and **d**, 94% of the selected images were from the early dry season months of November,

December, and January. We had a 9.3:1 ratio of available imagery to selected input data for analysis. Seealso Extended Data Table 1a.

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767 **Extended Data Fig. 8** | Evaluation of the tree crown mapping omission and commission errors. The 768 performance of the tree crown predictions was evaluated for 1028 randomly selected 512 x 256 pixel areas from the UTM zones 28 to 37 with approximately 100 areas for each UTM zone. The patches were extracted 769 with an 80% bias towards precipitation⁴⁷ above 200 mm/year, as the majority of tree identifications are 770 771 above the 200 mm/year isoline. 178 areas were excluded from evaluation because the rainfall was >1000 772 mm/year or the areas were devoid of trees. A total of 50,740 trees were evaluated. **a**, shows the class 773 breakdown of our 3,765 omission and commission errors; **b**, shows the class numbers of all the trees 774 evaluated; and c, Summarizes the number of trees, errors of commission and omission, and number of 775 clumped trees by error classes. The highest percentage of binned-errors classes, from 15% to > 25%, 776 occurred in areas with few trees and resulted from mixed tree and bush confusion for < 8% of all trees 777 evaluated. In contrast, the lowest binned-error classes, from 0% to 15%, had 92% of all trees evaluated. In 778 the 850 patches, the number of trees ranged from 1 tree to 326 trees, with a total of 50,740 trees evaluated 779 and 3,765 errors identified. Overall, the commission and omission error rates were 4.9% and 2.7%, 780 respectively, a net uncertainty of 2.2%. 781

Extended Data Fig. 9 | Tree crown and carbon errors. The root-mean-square errors (RMSEs) of our predicted crown areas calculated on external validation data from¹, binned based on the 50-quantiles of the hand-labeled crown areas. In 78 plots of 256×256 pixels, the hand-labeled set contained 5,925 areas and the system delineated 5,915. These crown areas were matched using inner spatial join. Multiple overlapping hand-labeled or predicted crown areas were merged into multi-polygons before calculating the RMSE. The crown areas of missed tress counted as errors. For calculating the corresponding RMSE of predicted carbon we relied on the allometric equations given in Extended Data Fig. 2 a, b, and c. The abscissa has a logarithmic scale and 95% of our 9.9 billion tree crowns had crown areas $< 78 \text{ m}^2$ (fig. 3b).

796 Extended Data Table 1 | The November to March Maxar satellite images considered and the number

selected for our analysis by UTM Zone. a, We started with 326,523 candidate images and selected 94,502

for processing for the area of 9.5° N to 24° N latitude from the Atlantic Ocean to the Red Sea for the months

of November to March. 87% of data selected for processing were acquired from 2010 to 2020 and 94% of

the selected satellite images were from November to January. Each image had a panchromatic and
 normalized difference vegetation index component that were used to identify trees with canopy area 3 m² or

- 802 greater in the early dry season. **b**, Specific satellite information for the four Maxar satellites used in our
- 803 study. The relative geolocation accuracies of the four Maxar satellites used in our study are expressed in
- 804 circular error probabilities or CE90 units, meaning a given point will be within a specific radius 90% of the
- 805 time, and in terms of root mean square errors, which are one standard deviation of the residuals or distance-
- 806 prediction errors. Our satellite data were resampled to a 50 cm spatial resolution for the panchromatic band
- and that band was used to panchromatically sharpen the NDVI to 50 cm. See also:
- 808 <u>https://gbdxdocs.Maxar.com/docs/geoeye-1;</u> https://gbdxdocs.Maxar.com/docs/quickbird;,
- 809 <u>https://gbdxdocs.Maxar.com/docs/worldview-2</u>; and https://gbdxdocs.Maxar.com/docs/worldview-3

a. UTM Zone	Total Images	Selected for Mosaic	GeoEye-1	QuickBird-2	WorldView-2	WorldView-3
32628	42,388	8,026	1,557	1,501	3,897	1,071
32629	29,244	9,883	2,208	2,198	4,334	1,143
32630	32,671	10,155	2,518	2,130	4,401	1,106
32631	34,611	10,158	2,384	2,049	4,554	1,171
32632	32,210	9,971	2,397	1,876	4,651	1,047
32633	34,918	10,097	2,688	2,082	4,310	1,017
32634	31,555	10,144	2,353	2,288	4,637	866
32635	36,663	9,975	2,269	2,099	4,595	1,012
32626	36,934	10,156	2,003	1,932	5,061	1,160
32637	15,329	5,937	1,079	1,353	2,742	763
ALL	326,523	94,502	21,456	19,508	43,182	10,356

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b. Quickbird-2	GeoEye-1	WorldView-2	WorldView-3
Oct. 2001-Feb. 2015	Nov. 2008	Oct. 2009	Aug. 2014
15 km & 55 cm	15.3 km & 41 cm	16.4 km & 46 cm	13.1 km & 31 cm
	Panchromatic,	Panchromatic,	Panchromatic,
Panchromatic,	Visible, & Near IR	Visible, & Near IR	Visible, Near IR, &
Visible, & Near IR			SWIR
23 m CE90	4.0 m CE90	5.0 m CE90	3.7 m CE90
10.8 m RMSE	2.7 m RMSE	3.0 m RMSE	2.5 m RMSE

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824 Supplemental Information Fig. 1. | Flow chart of the components of our paper and how they are 825 related. We start with the satellite data used, how these were organized for processing with our machine 826 learning software with the requisite training data, and how the resulting segmentation of ten billion tree 827 crown area resulted. We then show the conversion of tree crown area in tree wood, root, and leaf carbon at 828 the tree level for ten billion trees. We then show the use of our viewer to enable use of the data we produced, 829 from tree (1) to tree (10¹⁰).

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832 Supplemental Information Fig. 2a and 2b | UTM Zones 28 and 29 satellite data particulars. a, For

- UTM Zone 28 there were 42,388 candidate Maxar images between 9.5° to 24° N latitude for this UTM Zone
- segment and 8,026 were selected for processing. b, For UTM Zone 29 there were 29,244 candidate Maxar
 images between 9.5° to 24° N latitude for this UTM Zone segment and 9,883 images were selected for
- 836 processing. The distribution of the data with respect to month, solar azimuth, off-nadir angle, and year of
- 837 acquisition are given for each UTM Zone segment.

839 Supplemental Information Fig. 2c and 2d | UTM Zones 30 and 31 satellite data particulars. c, For

840 UTM Zone 30 there were 32,671 candidate Maxar images between 9.5° to 24° N latitude for this UTM Zone

- segment and 10,155 were selected for processing. d, For UTM Zone 31 there were 34,611 candidate Maxar
 images between 9.5° to 24° N latitude for this UTM Zone segment and 10,158 were selected for processing.
- 843 The distribution of the data with respect to month, solar azimuth, off-nadir angle, and year of acquisition are
- 844 given for each UTM Zone segment.

846 Supplemental Information Fig. 2e and 2f | UTM Zones 32 and 33 satellite data particulars. e, For UTM

- Zone 32 there were 32,210 candidate Maxar images between 9.5° to 24° N latitude for this UTM Zone
- segment and 9,971 were selected for processing. **f**, For UTM Zone 33 there were 34,918 candidate Maxar
- 849 images between 9.5° to 24° N latitude for this UTM Zone segment and 10,097 were selected for processing.
- 850 The distribution of the data with respect to month, solar azimuth, off-nadir angle, and year of acquisition are
- 851 given for each UTM Zone segment.

853 Supplemental Information Fig. 2g and 2h | UTM Zones 34 and 35 satellite data particulars. g, For

UTM Zone 34 there were 31,555 candidate Maxar images between 9.5° to 24° N latitude for this UTM Zone

- 855 segment and 10,144 were selected for processing. **h**, For UTM Zone 35 there were 36,663 candidate Maxar
- 856 images between 9.5° to 24° N latitude for this UTM Zone segment and 9,975 were selected for processing.
 857 The distribution of the data with respect to month, solar azimuth, off-nadir angle, and year of acquisition are
- siven for each LITM Zone segment
 - 858 given for each UTM Zone segment.

85903>165>352018-2021860Supplemental Information Fig. 2i and 2j | UTM Zones 36 and 37 satellite data particulars. i, For UTM861Zone 36 there were 36,934 candidate Maxar images between 9.5° to 24° N latitude for this UTM Zone

segment and 10,156 were selected for processing. j, For UTM Zone 37 there were 15,329 candidate Maxar

863 images between 9.5° to 24° N latitude for this UTM Zone segment and 5,937 were selected for processing.

864 The distribution of the data with respect to month, solar azimuth, off-nadir angle, and year of acquisition are

865 given for each UTM Zone segment.