

North American wintering mallards infected with highly pathogenic avian influenza show few signs of altered local or migratory movements

Claire S. Teitelbaum^{1,2*}, Nicholas M. Mastro³, Jeffery D. Sullivan⁴, Allison C. Keever³, Rebecca L. Poulson⁵, Deborah L. Carter⁵, Abigail G. Blake-Bradshaw³, Cory J. Highway³, Jamie C. Feddersen⁶, Heath M. Hagy⁷, Richard W. Gerhold⁸, Bradley S. Cohen³, Diann J. Prosser⁴

1. Akima Systems Engineering, Herndon, VA, USA
2. Contractor to U.S. Geological Survey, Eastern Ecological Science Center, Laurel, MD, USA
3. College of Arts and Sciences, Tennessee Technological University, Cookeville, TN, USA
4. U.S. Geological Survey, Eastern Ecological Science Center, Laurel, MD, USA
5. Southeastern Cooperative Wildlife Disease Study, College of Veterinary Medicine, University of Georgia, Athens, GA, USA
6. Tennessee Wildlife Resources Agency, Nashville, TN, USA
7. U.S. Fish and Wildlife Service, National Wildlife Refuge System, Stanton, TN, USA
8. University of Tennessee College of Veterinary Medicine, Knoxville, TN, USA

* Corresponding author: Claire Teitelbaum (claire.teitelbaum@gmail.com)

* Current address: Bay Area Environmental Research Institute and NASA Ames Research Center, Moffett Field, CA, USA

1 Abstract

2 Avian influenza viruses pose a threat to wildlife and livestock health. The emergence of
3 highly pathogenic avian influenza (HPAI) in wild birds and poultry in North America in late
4 2021 was the first such outbreak since 2015 and the largest outbreak in North America to date.
5 Despite its prominence and economic impacts, we know relatively little about how HPAI spreads
6 in wild bird populations. In January 2022, we captured 43 mallards (*Anas platyrhynchos*) in
7 Tennessee, USA, 11 of which were actively infected with HPAI. These were the first confirmed
8 detections of HPAI H5N1 clade 2.3.4.4b in the Mississippi Flyway. We compared movement
9 patterns of infected and uninfected birds and found no clear differences; infected birds moved
10 just as much during winter, migrated slightly earlier, and migrated similar distances as
11 uninfected birds. Infected mallards also contacted and shared space with uninfected birds while
12 on their wintering grounds, suggesting ongoing transmission of the virus. We found no
13 differences in body condition or survival rates between infected and uninfected birds. Together,
14 these results show that HPAI H5N1 clade 2.3.4.4b infection was unrelated to body condition or
15 movement behavior in mallards infected at this location during winter; if these results are
16 confirmed in other seasons and as HPAI H5N1 continues to evolve, they suggest that these birds
17 could contribute to the maintenance and dispersal of HPAI in North America. Further research
18 on more species across larger geographic areas and multiple seasons would help clarify potential
19 impacts of HPAI on waterfowl and how this emerging disease spreads at continental scales,
20 across species, and potentially between wildlife and domestic animals.

21

22 **Introduction**

23 Infectious diseases associated with wildlife have emerged at increasing rates in the last 50
24 years, a trend that is linked to declines in biodiversity and changes in climate and land use ¹⁻⁴.
25 Avian influenza viruses (AIVs) are one such emerging threat to wildlife, domestic animals, and
26 potentially human health. Low pathogenic avian influenza viruses (LPAI) circulate endemically
27 in wild waterfowl populations (ducks, geese, and swans; order Anseriformes) and generally
28 cause little or no clinical disease ⁵. However, since the 2.3.4.4 clade of the
29 A/goose/Guangdong/1/1996 H5N1 lineage of highly pathogenic influenza (HPAI) emerged in
30 2010, it has caused substantial mortality in many sensitive wild bird populations and significant
31 economic impacts to commercial poultry operations ^{6,7}. Outbreaks of HPAI have been
32 concentrated in Eurasia, where these viruses are beginning to be independently maintained in
33 wild birds and cause detrimental effects to many species ^{8,9}. In November 2021, the 2.3.4.4 clade
34 was detected in North America for the first time since 2015 ¹⁰. It has since spread across the
35 contiguous U.S. and Alaska, across 12 Canadian provinces and territories, and into Central and
36 South America ¹¹. Given its pandemic potential in wild birds and poultry ^{9,12}, it is crucial to
37 further understand how HPAI impacts wild bird health and how it spreads within and among
38 wild bird populations.

39 Movement behavior of infected hosts drives the spread of infectious diseases and serves
40 as an important indicator of an infection's pathogenicity. For example, a pathogen that imposes
41 an energetic cost can reduce infected hosts' movement ability, thus reducing contact rates and
42 limiting transmission. Infection with LPAI is sometimes associated with reduced movement in
43 wild waterfowl at both local and migratory scales ¹³, but just as often LPAI infection has no
44 effect on waterfowl behavior ¹⁴. However, HPAI viruses likely have stronger negative effects

45 than LPAI viruses on waterfowl movement behavior. For example, a lesser scaup (*Aythya*
46 *affinis*) infected with HPAI H5N1 in Maryland, USA in January 2022 exhibited reduced local
47 movements and subsequent mortality (cause unknown); despite these reduced movements, this
48 individual still could have contacted multiple uninfected birds while infected with HPAI H5N1
49 ¹⁵. Conversely, a white-faced whistling duck (*Dendrocygna viduata*) infected with a highly
50 pathogenic strain of avian influenza (HPAI H5N2) in West Africa displayed similar movement
51 patterns as uninfected conspecifics ¹⁶. Laboratory studies also show wide variation in responses
52 to HPAI infection across waterfowl species and individuals, including in viral pathogenicity and
53 shedding rates ^{17–20}, which can be modulated by individuals' previous exposure to HPAI and/or
54 LPAI ^{21,22}. Each species' unique relationship between HPAI infection and movement behavior
55 likely influences its role in the dispersal of HPAI at local, continental, and global scales.

56 Among waterfowl, mallards (*Anas platyrhynchos*) and other dabbling ducks are the best-
57 known reservoir species for AIVs ²³. Although mortalities have been reported in wild mallards
58 infected with HPAI, including in the 2021–2022 North American outbreak ²⁴, most mallards
59 experimentally infected with HPAI H5N1 in laboratory settings show few or no clinical signs
60 despite shedding large quantities of virus ^{25–27}. Mallards are also the most abundant waterfowl
61 species globally, are distributed across the Northern Hemisphere, and exhibit complex migratory
62 patterns including within-population variation in migration propensity and distance, making
63 them an important species for both dispersal and local maintenance of AIVs ^{28,29}. Finally,
64 mallards are relatively adaptable to human activities and often occupy urban and agricultural
65 areas ^{30,31}. Their abundance in anthropogenic landscapes makes mallards a potential source of
66 spillover or spillback of AIVs between wild and domestic birds. However, despite their
67 potentially important role for HPAI infection dynamics, we know little about how HPAI

68 infection affects mallard movement behavior, and until now, have had no data on North
69 American mallards' movement responses to newly emerged HPAI H5N1.

70 In January 2022, we detected HPAI H5N1 in 11 wild mallards in Tennessee, USA. These
71 are the first known detections of HPAI in wild waterfowl in the Mississippi Flyway during the
72 2021–2022 North American outbreak²⁴. These mallards, which showed no signs of disease at
73 capture, and 32 uninfected conspecifics were fitted with GPS transmitters that provided hourly
74 locations. We used these data to compare local movement behavior and migration patterns
75 between infected and uninfected individuals, and to identify spatio-temporal interactions
76 between marked birds that could have resulted in HPAI transmission. We expected that HPAI
77 H5N1 infection would have pathogenic effects on mallards, which would be reflected in reduced
78 movement by infected mallards shortly after detection of the virus. We expected that this
79 reduced movement would decrease contact rates and shared space use between infected and
80 uninfected birds. We also hypothesized that energetic costs of infection could carry over to
81 spring migration, which would be reflected in later, slower, and/or shorter-distance migration in
82 mallards infected during winter, compared to those with no known history of HPAI infection.
83 Finally, we compared mortality rates and body condition between infected and uninfected birds
84 to understand whether infection with HPAI H5N1 had apparent energetic or fitness costs.

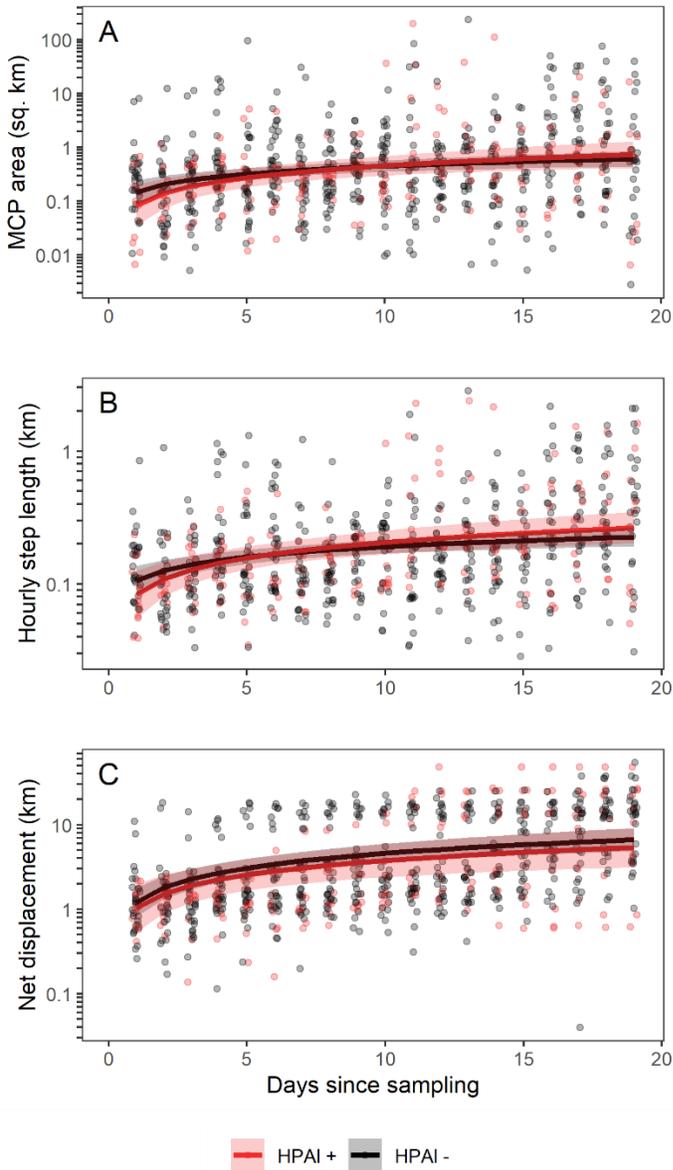
85 **Results**

86 We captured 11 mallards infected with HPAI H5N1 and 32 that were not shedding any
87 AIV in Tennessee, USA in January 2022. HPAI infection prevalence was 0.39 in females ($n =$
88 7/18), 0.16 in males ($n = 4/25$), 0.32 in juveniles ($n = 7/22$), and 0.19 in adults ($n = 4/21$).
89 Prevalence of antibodies to the nucleoprotein of AIV was 0.57 overall ($n = 23/40$; antibody data
90 were unavailable for three individuals) and 0.54 in HPAI-infected birds ($n = 6/11$); detection of

91 antibodies could indicate either prior exposure to influenza (HPAI or LPAI) or seroconversion
92 from a recent infection. No clinical signs of illness were observed at the time of capture.

93 *Local movements*

94 Local movement behaviors in the first 19 days following sampling were unrelated to
95 HPAI infection status (Fig. 1, Table S1–S3); the 19-day period of study was designed to include
96 both active infection and recovery for HPAI-infected birds and ended before any tracked
97 mallards initiated migration. On the first day following sampling, when differences between
98 groups would be expected to be largest, the average area of a HPAI-infected mallard’s daily
99 100% minimum convex polygon (MCP) was 0.085 km² (95% CI: 0.036-0.203), which was
100 indistinguishable from that of the average uninfected mallard (mean: 0.148 km², 95% CI: 0.087-
101 0.250). Regardless of infection status, mallard space use increased following sampling and
102 release, probably indicating temporary effects of capture or transmitter attachment and not
103 infection on movement. We also found no difference in movement behavior by infection status
104 for hourly movement distances or daily net displacement (Fig. 1B–C; Table S2–S3). In a second
105 set of models, we found no evidence that AIV antibody status (which could indicate either
106 seroconversion from the current infection or from a prior infection) moderated the relationship
107 between HPAI active infection and movement behavior (Table S4).



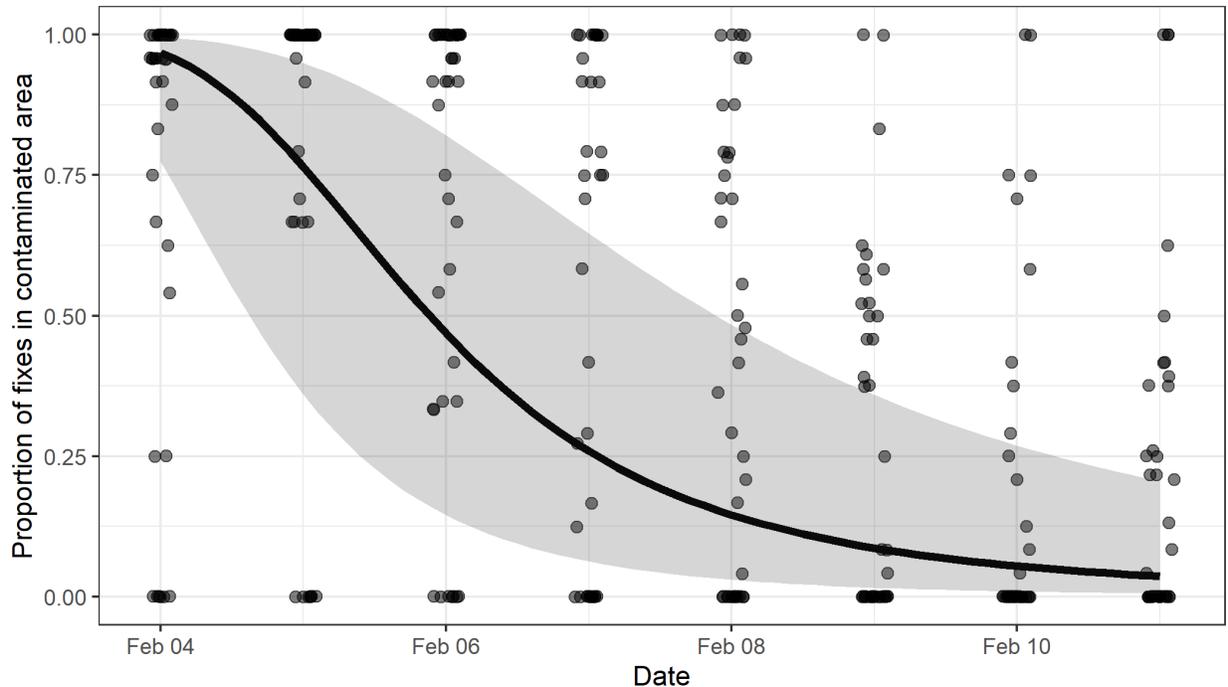
108
 109
 110
 111
 112
 113
 114
 115
 116
 117

Figure 1: Local movement patterns are unrelated to infection with HPAI H5N1 clade 2.3.4.4b in 43 mallards (*Anas platyrhynchos*) sampled in Tennessee, USA during winter 2022. In each plot, points show raw data, lines show estimated means from a linear mixed-effects model, and shaded areas show 95% confidence intervals of the mean. Models also included terms for age, sex, and a temporal autoregressive term for each individual; plots show marginal values averaged across age and sex. For plots that show predictions conditional on random effects, see Fig. S1. (A) Area of a 100% minimum convex polygon (MCP), a measurement of space use. (B) Mean hourly step lengths, a measurement of overall movement. (C) Net displacement, i.e., distance from the first GPS fix, a measurement of dispersal from the capture site.

118 **Contact rates**

119 We observed 375 interactions between pairs of mallards (i.e., a mallard was detected
120 within 25 meters of a known location of another bird within 65 minutes; Fig. S2), of which 80
121 (23%) were potential close or indirect HPAI contacts, i.e., an infected bird followed by an
122 uninfected bird. When we compared this proportion to the expected frequency of contacts in the
123 population, the observed proportion was in the 75th percentile of the randomized data, indicating
124 no significant difference between the observed frequency of contacts and the expected frequency
125 assuming birds were interacting independently of infection status.

126 Infected birds used a cumulative total area of 6.9 km² during the first four days following
127 sampling; birds were likely to be shedding HPAI for at least four days after sampling, so we
128 considered this area potentially HPAI-contaminated (hereafter “contaminated area”). All birds
129 initially spent most of their time in the contaminated area, but use of this area declined as the
130 winter progressed, at similar rates for infected and uninfected birds (Fig. 2, Table S5). Our model
131 estimated that on the first day of measurement (February 4), tracked mallards spent >90% of
132 their time in the contaminated area, but this time decreased to <5% by February 9. There was
133 substantial variation among individuals; two individuals (6%) were never detected in the
134 contaminated area while two others spent all their time in the contaminated area through the end
135 of the study period (individual ID standard deviation = 5.085; AR1 correlation = 0.862). Males
136 spent more time in the contaminated area than females (Table S5).



137
 138 **Figure 2:** Time spent in the potentially HPAI-contaminated area declines prior to initiation of
 139 migration. The contaminated area was defined as the total area of all 95% utilization
 140 distributions of HPAI-infected mallards (*Anas platyrhynchos*) in the first four days following
 141 sampling. The proportion of time was calculated as the proportion of daily fixes for each mallard
 142 within the contaminated area. Points show raw data and are jittered to increase visibility. The
 143 line and shaded area show marginal means from a generalized linear mixed-effects model. The
 144 model also included terms for HPAI infection status, age, and sex; only sex was related to time
 145 spent in the contaminated area (Table S5, Fig. S3).

146 *Migration patterns*

147 We quantified migration patterns for birds with sufficient telemetry data to measure the
 148 beginning of spring migration ($n = 35$) and arrival at summer sites ($n = 29$); some birds lacked
 149 sufficient data due to mortality, lack of transmitter signal, or transmitter failure. The mean spring
 150 migration initiation date was March 15 for infected birds ($n = 9$) and March 20 for uninfected
 151 birds ($n = 26$). Infected birds departed slightly earlier than uninfected birds (13 days, 95% CI: 27
 152 days earlier to 0.2 days later, $R^2 = 0.33$; Fig. 3, Table S6) and males departed earlier than females
 153 (14 days, 95% CI: 1–28 days earlier).

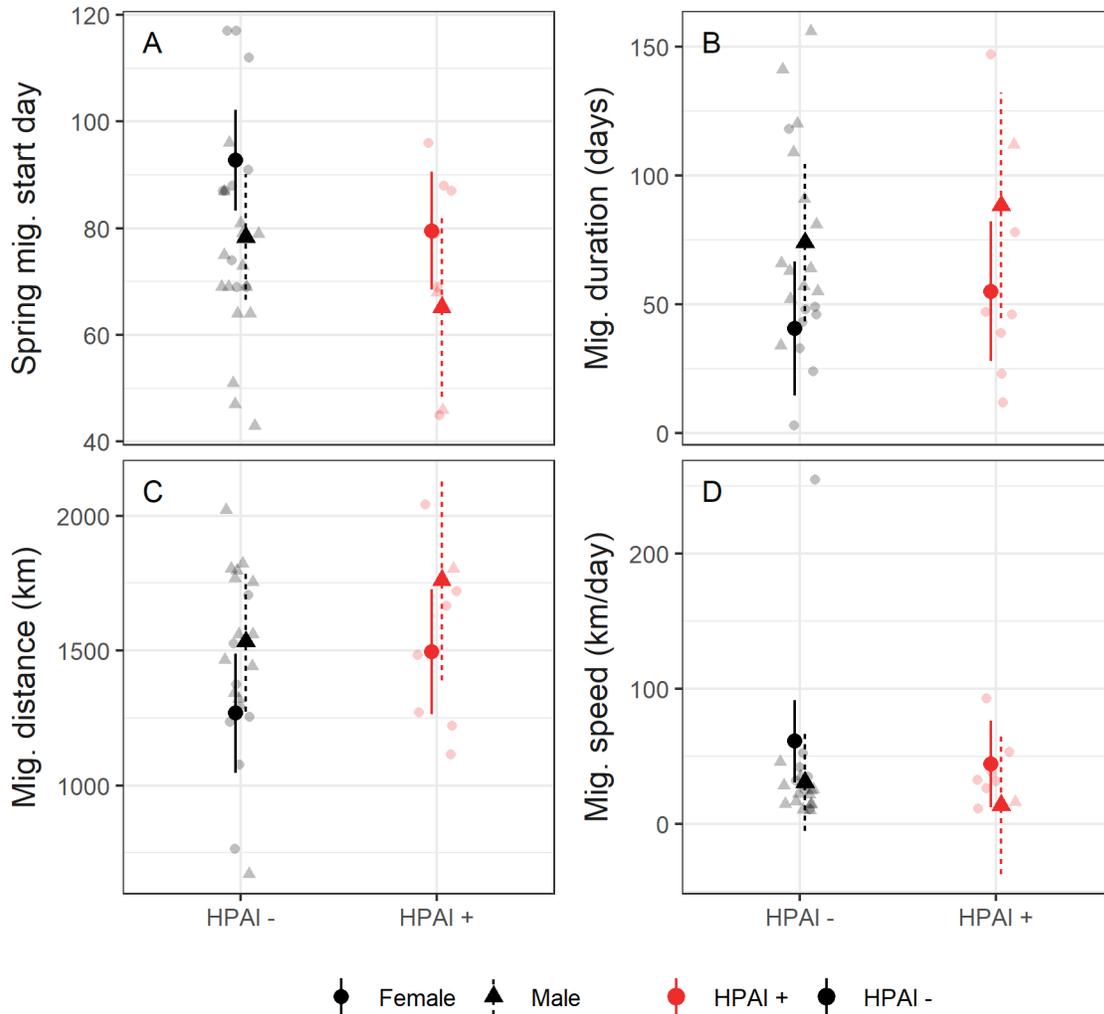
154 The time between winter site departure and summer site arrival (i.e., migration duration)
 155 averaged 63 days for infected birds ($n = 8$) and 69 days for uninfected birds ($n = 21$). Our model

156 indicated no difference in migration duration by infection status or age (Fig. 3, Table S6; $R^2 =$
157 0.24). There was weak evidence that males migrated for longer than females (estimate: 33 days
158 longer, 95% CI: 3 days shorter to 70 days longer).

159 The average migration distance was 1,540 km for infected birds ($n = 8$) and 1,445 km for
160 uninfected birds ($n = 21$). Our model showed no evidence for a difference in migration distances
161 in infected birds (difference: 228 km, 95% CI: 80 km shorter to 536 km farther, $R^2 = 0.17$; Fig.
162 3, Table S6). We found no evidence for a difference in migration distance by age or sex.

163 The average migration speed for infected birds was 38 km/day and for uninfected birds
164 was 36 km/day. We found no relationship between infection status and migration speed
165 (estimate: 16 km/day, 95% CI: 24 km/day slower to 57 km/day faster, $R^2 = 0.14$; Fig. 3, Table
166 S6). We also found no evidence for differences in migration speed by sex or age.

167 We found no evidence that AIV antibody status was related to migration date, duration,
168 distance, or speed (Fig. S4, Table S7).



169
 170 **Figure 3:** Relationships between HPAI infection status, sex, and migration patterns in mallards
 171 (*Anas platyrhynchos*). Each panel shows the estimated mean and 95% confidence interval of the
 172 mean from a linear model. Partially transparent points show raw data. Models also included a
 173 term for age; plots show values for juveniles. (A) HPAI-infected birds departed on spring
 174 migration slightly earlier than uninfected birds and males migrated earlier than females. The y-
 175 axis shows the day of year of spring migration initiation (day 80 = March 21). (B) The duration
 176 of migration was unrelated to infection status. (C) Migration distance was unrelated to infection
 177 status, but males migrated farther than females. (D) Migration speed was unrelated to infection
 178 status or sex.

179 **Body condition and mortality**

180 We found no evidence for differences in body condition at capture between infected and
 181 uninfected birds ($F_{1,47} = 0.073$, $p = 0.787$) or for differences in survival by infection status, age,

182 or sex. Model-estimated mortality rates were 0.38 for infected birds ($n = 7$; 95% CI: 0.01–0.77)
183 and 0.33 in uninfected birds ($n = 14$; 95% CI: 0.05–0.62).

184 **Discussion**

185 We detected infections with HPAI H5N1 clade 2.3.4.4b in 11 of 43 (26%) mallards
186 sampled in Tennessee, USA, during January 2022. These detections represent some of the
187 earliest detections in live birds during the ongoing HPAI outbreak in North America, which has
188 severely impacted wild bird health (e.g., colonially nesting seabirds) and poultry production ¹¹.
189 Collectively, our analyses show that HPAI infection in wild mallards during winter had no
190 detectable effects on movement behavior at local (within 19 days) or migratory scales or on body
191 condition or survival. Importantly, we observed shared space use between infected and
192 uninfected birds on the wintering grounds as well as extensive movement of infected birds on the
193 wintering grounds (up to 50 km from the capture site) and during migration. Together, these
194 results suggest that tolerance of HPAI H5N1 infection could promote transmission within this
195 wintering mallard population and beyond, including to other species and geographic areas.

196 Our finding that HPAI-infected and uninfected birds migrated similarly suggests that
197 mallards had the potential to be effective dispersal agents for this emerging virus during its initial
198 introduction to North America in winter 2021–2022. In general, host-parasite combinations
199 where pathogenicity is low or tolerance is high should be associated most strongly with long-
200 distance pathogen dispersal ³², especially for migratory species ³³. Although the expected
201 duration of infection with this clade of HPAI H5N1 can be up to 14 days (in experimentally-
202 exposed immunologically naïve mallards²⁷) and most migrations began more than 14 days after
203 sampling, the shared space use that we observed suggest potential ongoing transmission during
204 winter. Thus, we strongly suspect that many birds could be actively infected at the time of their

205 migration. In addition, all birds either completed their migrations or made a stopover in less than
206 14 days (Fig. S5), thus providing a potential mechanism for long-distance spread of this
207 pathogen³⁴. However, infection statuses of all marked birds were unknown at the time of
208 migration. It is possible that active and recent HPAI infection affects migration behavior, but that
209 we could not detect these effects because recovery and transmission occurred between the time
210 of sampling and initiation of migration. Nevertheless, our data and analyses show no relationship
211 between infection and movement behavior in the week following sampling, or between infection
212 and body condition, collectively suggesting that HPAI H5N1 infection had minimal negative
213 effects on health or behavior in these wild North American mallards.

214 Laboratory studies show that HPAI infection often has minimal or no effects on duck
215 health or behavior^{27,35,36}, and that in experimental settings, mallards can shed high
216 concentrations of HPAI H5N1 relative to other duck species²⁶. Likewise, we found no
217 differences in body condition or mortality between infected and uninfected birds in this wild
218 population, even though natural settings exhibit higher variability in food availability^{37,38}, body
219 condition^{39,40}, social interactions⁴¹, previous AIV exposure, and influenza viral loads in the
220 environment⁴¹ than laboratory settings, all of which could influence the dynamics and
221 pathogenicity of influenza infection. Still, infections with the same pathogen can differ in their
222 pathogenicity across individuals and across time, depending on body condition, time since
223 infection, behavior, or infection history^{39,42-44}; if the most negatively affected birds are more
224 likely to die or “hunker down,” they would not have been sampled, thus potentially biasing our
225 sample towards individuals that are tolerant of HPAI infection. We also found no evidence that
226 AIV antibodies mediated the effects of HPAI infection on movement behavior. Antibody
227 prevalence is relatively high during winter⁴⁵, which could have limited our ability to detect

228 subtle changes in behavior of infected mallards; this protective benefit of prior exposure could
229 differ at other times of year or in groups of immunologically naïve birds (e.g., juveniles), which
230 could alter infection-movement relationships. A combination of experimental, observational, and
231 theoretical studies across more species and seasons is necessary to fully understand how
232 immunology and the environment interact to determine the impacts of influenza infection on
233 wild bird behavior and health.

234 Mechanistic models are important tools for understanding the maintenance, dispersal,
235 transmission, and reassortment of influenza viruses^{46,47}, but often face uncertainties in parameter
236 values (e.g., HPAI pathogenicity) or model structures (e.g., HPAI transmission routes). This
237 study can inform several important parameters for these models. First, in North American
238 wintering mallards with some prior AIV exposure, infection with HPAI H5N1 is unrelated to
239 movement distances at local or migratory scales based on our fine-scale location data; therefore,
240 modeling movement as homogeneous among infectious groups could be a reasonable assumption
241 in mechanistic models. Second, mallards in our study contacted one another independent of
242 infection status, but shared space use between birds declined over the course of the winter,
243 probably coincident with increases in movement and changes in habitat selection and availability
244 as the hunting season ended and preparation for migration began⁴⁸. This pattern suggests that,
245 while contact rates and contact with virions in the environment might be homogeneous within a
246 population, they might vary within seasons. HPAI was also detected concurrently in
247 heterospecific birds at the same refuge (R. Gerhold, unpubl. data), which could further contribute
248 to environmental contamination. Modeling these spatio-temporal patterns in environmental
249 transmission will require more complex functions than assuming that all birds are equally likely

250 to encounter influenza virus in the environment. Our data and analyses can inform more realistic
251 models that more accurately predict the mechanisms of HPAI transmission and dispersal.

252 The current HPAI H5N1 outbreak in North America has affected over 47 million
253 domestic poultry in the United States and threatens some wild bird species of conservation
254 concern, including seabirds and raptors ^{11,24,49,50}. As this outbreak continues, wildlife managers
255 and farmers must adapt their practices to prevent influenza infection in these sensitive species.
256 Our results suggest that mallard populations – which are important culturally as a game species
257 and for wildlife viewing ⁵¹ – might not be substantially impacted by the ongoing outbreak, at
258 least for wintering mallards with prior exposure to AIV infected with the genotype of HPAI
259 circulating in North America in January 2022. However, reduced wetland availability, as has
260 been observed over the last century ⁵², can promote disease transmission within wild waterfowl
261 populations by increasing local densities, contact rates, and probabilities of environmental
262 transmission ⁵³. Waterfowl densities at these and other state- and federally-owned waterfowl
263 refuges can be high ^{54,55}, meaning that contacts and shared space observed in our study represent
264 only a small fraction of potential direct and environmental transmission among the entire (mostly
265 unmarked) population. As these mallards move locally and northwards on their spring migration,
266 they travel through agricultural areas ⁵⁶ and share stopover sites with other waterfowl species ⁵⁷.
267 We therefore expect that, because of their apparent tolerance to infection and gregarious
268 behavior, wild mallards (and potentially other waterfowl) are important for the epidemiology of
269 HPAI H5N1 in North America. However, because influenza viruses are constantly evolving and
270 some strains exhibit higher pathogenicity than others ^{7,58,59}, it is critical to continue to monitor
271 the effects of HPAI H5N1 across larger samples of multiple wildlife species, especially as the
272 virus continues to reassort with North American-origin LPAI. More broadly, these results

273 highlight that interspecific variation in behavior and responses to an emerging infectious disease
274 can impact how these diseases spread, how long they persist, and their potential impacts on
275 wildlife, domestic animal, and human health.

276 **Methods**

277 *Study area, capture, and sampling*

278 We captured male and female mallards using rocket nets at Lake Isom National Wildlife
279 Refuge (NWR; 36.3049°N, -89.4173°W) on 24, 25, 30, and 31 January 2022 ($n = 20, 8, 5,$ and
280 10 individuals, respectively ⁶⁰. Lake Isom NWR was established in 1938 as Tennessee's first
281 NWR and maintains a diversity of managed wetlands including croplands, forested wetlands, and
282 a large ~150 ha shallow-water lake that is seasonally dried, mechanically manipulated, and
283 flooded during winter to provide moist-soil vegetation and seeds for wintering waterfowl. Lake
284 Isom NWR hosts nearly 40,000 ducks on average in January (January 2022 aerial estimate =
285 36,834 ⁵⁵).

286 We banded all captured mallards with U.S. Geological Survey aluminum tarsal bands and
287 determined sex and age based on cloacal inversion, wing plumage, and bill color ⁶¹. We aged
288 ducks as juvenile (second year) or after adult (after second year). We measured weight (± 0.10 g)
289 and wing cord length (± 1 mm) for all individuals. We collected oropharyngeal and cloacal
290 swabs of all individuals and placed paired swabs into 2 mL viral transport medium (VTM; 64).
291 We also extracted ≤ 3 mL of blood from the brachial artery for each individual and separated the
292 serum fraction. Swabs and sera were stored at -80°C until sent for virologic testing and
293 additional analyses at the University of Georgia (Athens, GA, USA).

294 We attached 20-g solar rechargeable and remotely programmable Global Positioning
295 System-Global System for Mobile (GPS-GSM) transmitters (OrniTrack; Ornitela, UAB

296 Švitrigailos, Vilnius, Lithuania) to birds weighing ≥ 1 kg to ensure deployment package remained
297 below recommended body weight limits (3–5%⁶³). We attached transmitters using dorsally-
298 mounted body harnesses made of automotive moisture-wicking elastic ribbon⁶⁴. Completed
299 harnesses had two body loops knotted and sealed with cyanoacrylic glue above the keel and
300 across the abdomen⁶⁴. Total package of GPS-GSM transmitter and harness weighed ~22 g.
301 Transmitters were remotely programmed to record hourly locations and were not synchronized
302 among individuals. Calibration data on this tag model indicates median location error of <25 m.
303 We used all available telemetry data from AIV-sampled birds from the first capture (24 January
304 2022) until we began analysis (27 October 2022)⁶⁵.

305 All duck capture, handling, and sampling procedures were approved by and carried out in
306 accordance with Tennessee Technological University's Institutional Animal Care and Use
307 Committee (protocol #19-20-0020) and authorized under Federal Banding Permit #05796. This
308 study complies with the relevant portions of the ARRIVE guidelines for observational studies.

309 *Influenza lab methods*

310 We attempted virus isolation on all swab samples by inoculating a total 1mL of VTM
311 into the allantoic cavities of three 9-11 day-old embryonated chicken eggs⁶⁶ and incubating at
312 37°C for 120 hours. Amnioallantoic fluid was collected and tested by hemagglutination assay⁶⁷.
313 RNA was extracted from amnioallantoic egg fluids for all putative virus isolation-positive
314 samples using the QIAamp viral RNA mini kit (Qiagen Inc.; Germantown, MD, USA) following
315 manufacturer recommendations, and screened for the matrix gene of influenza A virus in real-
316 time reverse transcription polymerase chain reaction (rRT-PCR) as previously described⁶⁸.
317 Influenza A-positive samples were further screened for 2.3.4.4 HP H5 via rRT-PCR; suspect
318 positives from this assay were sent to the United States Department of Agriculture National

319 Veterinary Services Laboratory, Ames, Iowa for confirmation. A positive virus isolation result
320 indicates active shedding of influenza at the time of capture.

321 Because no birds displayed visible indications of illness, laboratory testing was
322 completed after capture and release, meaning that infection statuses were unknown at time of
323 release.

324 Serum samples were tested for the presence of AIV antibodies by commercial blocking
325 enzyme-linked immunosorbent assay (bELISA, IDEXX Laboratories, Westbrook, ME) as
326 described by the manufacturer. An initial serum-to-negative control (S:N) absorbance ratio < 0.5
327 represents the cutoff threshold recommended by the manufacturer, so we considered samples
328 with an S:N ratio > 0.5 to be positive. A positive bELISA result represents the presence of
329 antibodies to AIV, which indicates prior infection with any AIV (HPAI or LPAI). Influenza
330 antibodies are estimated to be detectable for 6 months-1.5 years⁶⁹⁻⁷¹ but usually peak within 3
331 weeks of infection^{69,70}

332 *Data analysis*

333 *Local movements*

334 We analyzed daily movement patterns within 19 days of capture to determine whether
335 movement behavior differed between HPAI-infected and uninfected birds, beginning at the time
336 of capture and ending after presumed recovery from infection (≤ 14 days;²⁷). We expected that, if
337 HPAI infection affected local movement behavior, infected and uninfected birds would move
338 differently in the first few days following sampling, but any differences in movement would no
339 longer be observed by the end of the 19-day window. One mallard started migrating 20 days
340 after capture, so we used a 19-day window to include non-migratory movements only.

341 To measure daily movements, we used three related metrics of local movement: the area
342 of a daily 100% minimum convex polygon (MCP), mean hourly step lengths per day, and mean
343 daily net displacement. Daily MCPs draw a convex hull around all daily locations (i.e., GPS
344 fixes); a larger MCP indicates more movement and more exploratory behavior^{72,73}. Mean step
345 length is the average distance between hourly GPS fixes in a day and has been used in prior
346 analyses of influenza in ducks^{14,15}. Finally, mean net displacement measures a bird's daily
347 average distance from its capture location and measures the timing and distance of initial
348 dispersal. We resampled telemetry data to 1-hour intervals with a tolerance of 8 minutes (i.e.,
349 GPS fixes between 52 and 68 minutes apart), then calculated each movement metric per
350 individual per day. We split days at sunrise because ducks usually move between foraging and
351 roosting areas at dawn and dusk^{64,73}, so using sunrise as the beginning of a day helps ensure that
352 movement or resting at a single foraging or roosting site are included as part of the same day. We
353 identified sunrise times using statistical software (*suncalc* package version 0.5.0 in R version
354 4.0.1^{74,75}) and calculated MCPs and step lengths (*amt* package version 0.1.4⁷⁶).

355 For each local movement metric, we fit a linear mixed-effects model with log-
356 transformed area or distance as the response variable (*glmmTMB* package version 1.1.3^{77,78}).
357 Explanatory variables were: active influenza infection status at capture (positive or negative);
358 days since influenza sampling; sex; age; and the pairwise interaction between infection status
359 and days since sampling. This interaction was included to test the prediction that movement
360 would change as birds recovered from infection. We log-transformed days since sampling
361 because we expected that differences in movement between infected and uninfected birds would
362 be largest in the first few days following sampling^{14,79}. We included log-transformed number of
363 GPS fixes as a fixed effect to account for the sensitivity of movement metrics to sample sizes.

364 We also included an AR1 autoregressive random slope for each individual to account for inter-
365 individual variation and temporal autocorrelation in individuals' locations over time ⁸⁰. We
366 evaluated models using standard plots and tests of residuals (*DHARMA* package version 0.4.3 ⁸¹)
367 and calculated post-hoc estimated marginal means and 95% confidence intervals (CIs, *emmeans*
368 package version 1.6.3 ⁸²).

369 Antibodies from a prior infection can protect birds from the most severe effects of
370 infection, and the presence of antibodies can indicate that an individual is relatively late in its
371 current infection; in either case, we hypothesized that effects of HPAI infection on movement
372 behavior might be smaller in individuals with antibodies to influenza. Therefore, we repeated
373 these models using a combination of active infection and antibody status as a predictor variable.
374 This variable had three levels: HPAI+/antibody+, HPAI+/antibody-, and HPAI- with either
375 antibody status. These models were otherwise identical to the models using active infection
376 status only.

377 *Contact rates and environmental transmission*

378 We used observed movement patterns of birds within four days of sampling to identify
379 close and indirect contacts that could have led to transmission. Based on experimental infection
380 data, four days is a conservative estimate of the shedding period for HPAI ²⁷. We defined a pair
381 of locations as a contact if two birds were observed within 25 m of the same location within 65
382 minutes ¹⁵; this 65-min window accounted for different schedules among GPS transmitters,
383 which were not synchronized to provide fixes at the same time as each other, and allowed five
384 minutes for deviations from this hourly schedule. We considered an interaction to be a contact
385 that could lead to transmission if the bird that was present first was infected and the bird that was
386 present second was uninfected.

387 Next, we examined whether contacts that could have led to transmission were more or
388 less common than would be expected if contacts were random with respect to infection status. To
389 do so, we randomized infection statuses among individuals, then calculated the proportion of
390 contacts that were “possible transmission contacts” in the randomized data. We repeated this
391 process 500 times with replacement, then compared the distribution of proportions in the
392 randomized data to the proportion in the observed data.

393 We also assessed the potential for environmental transmission of HPAI from GPS-tagged
394 mallards by estimating shared space use between infected and uninfected birds; note that this
395 analysis does not account for the presence of untagged HPAI-positive birds at the site and
396 therefore represents a conservative estimate of environmental transmission. For each infected
397 bird, we calculated a dynamic Brownian bridge movement model (dBBMM⁸³; *move* package
398 version 4.0.6⁸⁴) for the first four days following sampling (as above, a conservative estimate of
399 the HPAI shedding period). We used a location error of 23.5 m and a raster resolution of 30 m
400 for dBBMMs. We then extracted the 95% utilization distribution (UD) contour for each infected
401 bird, which represents the area where the infected individual spent 95% of its time during the
402 four-day period. We then defined the “HPAI-contaminated area” for the population, which
403 included any location covered by at least one infected bird’s 95% UD (i.e., the union of the 95%
404 UDs across all infected birds).

405 Starting at the end of the four-day period for the latest-captured infected bird (February 4,
406 2022) and continuing until the first date of spring migration (see below; February 11, 2022), we
407 calculated the proportion of time that birds that were uninfected at the time of capture spent
408 inside the HPAI-contaminated area. We started at the end of this period because we had
409 incomplete data on infected birds until the end of this time. For each bird, we calculated the

410 proportion of fixes in the HPAI-contaminated area vs. outside the area for each bird-day. This
411 proportion is a proxy for the daily environmental transmission risk per individual. To understand
412 how this risk varied across individuals, by infection status, over time, and by age or sex, we used
413 a generalized linear mixed-effects model with a logit link to model the proportion of fixes within
414 the contaminated area as a function of days since February 3 (log transformed), HPAI infection
415 status, age, sex, and the interaction between infection status and days since February 3 (using
416 *glmmTMB*^{77,78}). We also included an AR1 autoregressive term for each bird⁸⁰, because each
417 bird's locations on consecutive days are autocorrelated. The model used the number of fixes
418 inside and outside the HPAI-contaminated area as the response variable.

419 *Migration patterns*

420 To measure differences in migration phenology and migration patterns between infected
421 and uninfected mallards (Fig. S5), we first segmented each track into wintering, migration, and
422 summer periods. We used bivariate time-series segmentation on latitude and longitude using the
423 *segclust2d* package⁸⁵. This method uses the mean and/or variance in these two variables across
424 the track to identify discrete segments. We visually inspected each track to identify the number
425 of segments that most accurately separated wintering and summering phases from migration and
426 stopover. Because segmentation accurately identifies break points in segments but includes
427 movement bouts with either the previous or subsequent segment, we further segmented tracks by
428 creating a new segment each time a bird was observed moving 20 km/h or faster; this speed was
429 a clear distinction between dispersive (flight) and non-dispersive (local) movements for most
430 birds⁸⁶.

431 We then classified each segment as winter, migration/stopover, or summer. We defined
432 winter as segments with median locations within 50 km of capture. We defined summer

433 locations as segments lasting at least 30 days and beginning in March-July, with a range of net
434 displacement ≤ 50 km^{87,88}. For birds whose transmitters failed before this 30-day period was
435 over, we assigned the last segment as a summer segment if it was at least 1000 km from the
436 capture location and started in March-July. We verified all classifications manually using plots of
437 net displacement over time and maps of the locations of each segment.

438 From these segmented tracks, we measured four characteristic of each individual's spring
439 migration: (1) the initiation date of spring migration, i.e., the end date of a bird's last wintering
440 segment; (2) the duration of spring migration, i.e., the time elapsed between the last day of
441 wintering and the first day of summering; (3) migration distance, i.e., the median net
442 displacement of all summer locations (i.e., median distance from capture site); and (4) migration
443 speed, i.e., migration distance divided by migration duration. For six individuals, it was possible
444 to calculate migration initiation date but not the other metrics because they did not have
445 sufficient tracking data for the full migration period. For each migration metric, we modeled
446 differences between infected and uninfected birds using linear models. Each model used the
447 migration metric as the response variable and included infection status, sex, and age as
448 predictors.

449 We also developed a separate set of models that measured relationships between these
450 same variables and prior infection (as opposed to active infection status). These models were
451 constructed identically except that infection was measured using bELISA results as well as virus
452 isolation (i.e., active infection) results. We considered an individual as previously infected at the
453 time of migration if it tested positive for antibodies at the time of capture (i.e., a positive bELISA
454 result) or if it was actively infected at the time of capture.

455 *Body condition and mortality*

456 We examined differences in body condition at capture between infected and uninfected
457 birds. We estimated body condition using the residuals from a linear regression between body
458 mass (g) and wing chord length (cm), which represent deviation from the expectation of size-
459 adjusted mass in the population⁸⁹. We found no evidence for differences in this relationship by
460 age or sex, so we did not account for age or sex in our calculation of body condition. We tested
461 for differences in body condition between infected and uninfected birds using a linear model
462 with body condition as the response variable and infection status as the predictor variable.

463 Finally, we evaluated whether survival to the end of the study (October 2022) was related
464 to HPAI infection status at capture. We only included birds confirmed to be dead or alive on
465 October 25, 2022 and omitted birds with unknown fates due to transmitter back-log, lack of
466 cellular connectivity, and/or transmitter failure. We fit a generalized linear model with a logit
467 link that measured mortality as a function of infection status, age, and sex.

468 **Acknowledgements**

469 We appreciate support and funding from Tennessee Wildlife Resources Agency; U.S.
470 Fish and Wildlife Service Refuge System, Southeast Region; and the Center for the
471 Management, Protection, and Utilization of Water Resources (Water Center) at Tennessee
472 Technological University. R. Bealer, L. Bull, and K. Hall assisted in capture of mallards, viral
473 sampling, and deployment of GPS transmitters. At the University of Georgia, Alinde Fojtik
474 assisted in laboratory analysis. USGS scientists are supported in part by the Ecosystems Mission
475 Area. We thank David Stallknecht for valuable comments on an earlier version of the
476 manuscript. Any use of trade, product, or firm names are for descriptive purposes only and do

477 not imply endorsement by the U.S. Government. Views expressed in this article are those of the
478 authors and do not necessarily represent views of the U.S. Fish and Wildlife Service.

479 **Data availability statement**

480 Data are available at USGS ScienceBase ⁶⁵ and code to reproduce all analyses is available at
481 Zenodo ⁹⁰.

482 **Competing interests**

483 All authors declare no competing interests.

484 **Author contributions**

485 NMM, BSC, and ACK designed with study with input from HMH, JCF, and RWG. NMM,
486 AGBB and CJH collected telemetry data. RLP and DLC performed and interpreted influenza
487 analysis. CST analyzed data with input from NMM, DJP, and JDS. CST prepared figures. CST
488 wrote the first draft, with editing from JDS, DJP, NMM, RLP and review by all authors. Project
489 administration and funding acquisition by BSC, JCF, HMH, and DJP.

490 **References**

- 491 1. Keesing, F. *et al.* Impacts of biodiversity on the emergence and transmission of infectious
492 diseases. *Nature* **468**, 647–652 (2010).
- 493 2. Cunningham, A. A., Daszak, P. & Wood, J. L. N. One health, emerging infectious
494 diseases and wildlife: Two decades of progress? *Philos. Trans. R. Soc. B* **372**, (2017).
- 495 3. Tompkins, D. M., Carver, S., Jones, M. E., Krkošek, M. & Skerratt, L. F. Emerging
496 infectious diseases of wildlife: A critical perspective. *Trends Parasitol.* **31**, 149–159
497 (2015).
- 498 4. Jones, K. E. *et al.* Global trends in emerging infectious diseases. *Nature* **451**, 990–993
499 (2008).
- 500 5. Kuiken, T. Is low pathogenic avian influenza virus virulent for wild waterbirds? *Proc. R.*
501 *Soc. B* **280**, 20130990 (2013).
- 502 6. Lycett, S. J., Duchatel, F. & Digard, P. A brief history of bird flu. *Philos. Trans. R. Soc. B*
503 **374**, 20180257 (2019).
- 504 7. Verhagen, J. H., Fouchier, R. A. M. & Lewis, N. Highly Pathogenic Avian Influenza
505 Viruses at the Wild-Domestic Bird Interface in Europe: Future Directions for Research
506 and Surveillance. *Viruses* **13**, 212 (2021).
- 507 8. Pohlmann, A. *et al.* Has Epizootic Become Enzootic? Evidence for a Fundamental Change
508 in the Infection Dynamics of Highly Pathogenic Avian Influenza in Europe, 2021. *MBio*
509 **13**, 1–8 (2022).
- 510 9. Ramey, A. M. *et al.* Highly pathogenic avian influenza is an emerging disease threat to
511 wild birds in North America. *J. Wildl. Manage.* **86**, e22171 (2022).
- 512 10. Bevins, S. N. *et al.* Intercontinental Movement of Highly Pathogenic Avian Influenza
513 A(H5N1) Clade 2.3.4.4 Virus to the United States, 2021. *Emerg. Infect. Dis.* **28**, 1006–
514 1011 (2022).
- 515 11. Harvey, J. A., Mullinax, J. M., Runge, M. C. & Prosser, D. J. The Changing Dynamics of
516 Highly Pathogenic Avian Influenza H5N1: Next Steps for Management & Science in
517 North America. *Biol. Conserv.* **In Press**, 110041 (2023).
- 518 12. Horwood, P. F. Avian influenza H5N1: still a pandemic threat? *Microbiol. Aust.* **42**, 152–
519 155 (2021).
- 520 13. van Gils, J. A. *et al.* Hampered foraging and migratory performance in swans infected
521 with low-pathogenic avian influenza A virus. *PLoS One* **2**, e184 (2007).
- 522 14. Bengtsson, D. *et al.* Does influenza A virus infection affect movement behaviour during
523 stopover in its wild reservoir host? *R. Soc. Open Sci.* **3**, 150633 (2016).
- 524 15. Prosser, D. J. *et al.* A lesser scaup (*Aythya affinis*) naturally infected with Eurasian 2.3.4.4
525 highly pathogenic H5N1 avian influenza virus: Movement ecology and host factors.
526 *Transbound. Emerg. Dis.* **69**, e2653–e2660 (2022).
- 527 16. Gaidet, N. *et al.* Evidence of infection by H5N2 highly pathogenic avian influenza viruses
528 in healthy wild waterfowl. *PLoS Pathog.* **4**, 1–9 (2008).
- 529 17. Hénaux, V. & Samuel, M. D. Avian influenza shedding patterns in waterfowl:
530 Implications for surveillance, environmental transmission, and disease spread. *J. Wildl.*
531 *Dis.* **47**, 566–578 (2011).
- 532 18. Luczo, J. M., Prosser, D. J., Pantin-Jackwood, M. J., Berlin, A. M. & Spackman, E. The
533 pathogenesis of a North American H5N2 clade 2.3.4.4 group A highly pathogenic avian
534 influenza virus in surf scoters (*Melanitta perspicillata*). *BMC Vet. Res.* **16**, 1–10 (2020).

- 535 19. Stephens, C. B., Prosser, D. J., Pantin-Jackwood, M. J., Berlin, A. M. & Spackman, E.
536 The pathogenesis of H7 highly pathogenic avian influenza viruses in lesser scaup (*Aythya*
537 *affinis*). *Avian Dis.* **63**, 230–234 (2019).
- 538 20. Spackman, E., Prosser, D. J., Pantin-Jackwood, M., Stephens, C. B. & Berlin, A. M.
539 Clade 2.3.4.4 H5 North American Highly Pathogenic Avian Influenza Viruses Infect, but
540 Do Not Cause Clinical Signs in, American Black Ducks (*Anas rubripes*). *Avian Dis.* **63**,
541 366–370 (2019).
- 542 21. Costa, T. P., Brown, J. D., Howerth, E. W., Stallknecht, D. E. & Swayne, D. E. Homo-
543 and heterosubtypic low pathogenic avian influenza exposure on H5N1 highly pathogenic
544 avian influenza virus infection in wood ducks (*Aix sponsa*). *PLoS One* **6**, (2011).
- 545 22. Berhane, Y. *et al.* Pre-exposing Canada geese (*Branta canadensis*) to a low-pathogenic
546 H1N1 avian influenza virus protects them against H5N1 HPAI virus challenge. *J. Wildl.*
547 *Dis.* **50**, 84–97 (2014).
- 548 23. Munster, V. J. *et al.* Spatial, temporal, and species variation in prevalence of influenza a
549 viruses in wild migratory birds. *PLoS Pathog.* **3**, e61 (2007).
- 550 24. USDA APHIS. 2022 Detections of Highly Pathogenic Avian Influenza in Wild Birds.
551 [https://www.aphis.usda.gov/aphis/ourfocus/animalhealth/animal-disease-](https://www.aphis.usda.gov/aphis/ourfocus/animalhealth/animal-disease-information/avian/avian-influenza/hpai-2022/2022-hpai-wild-birds)
552 [information/avian/avian-influenza/hpai-2022/2022-hpai-wild-birds](https://www.aphis.usda.gov/aphis/ourfocus/animalhealth/animal-disease-information/avian/avian-influenza/hpai-2022/2022-hpai-wild-birds) (2022).
- 553 25. Brown, J. D., Stallknecht, D. E., Beck, J. R., Suarez, D. L. & Swayne, D. E. Susceptibility
554 of North American ducks and gulls to H5N1 highly pathogenic avian influenza viruses.
555 *Emerg. Infect. Dis.* **12**, 1663–1670 (2006).
- 556 26. Keawcharoen, J., Riel, D. Van, Amerongen, G. Van, Bestebroer, T. & Beyer, W. E. Wild
557 Ducks as Long-Distance Vectors of Highly Pathogenic Avian Influenza Virus (H5N1).
558 *Emerg. Infect. Dis.* **14**, 600–607 (2008).
- 559 27. Spackman, E., Pantin-Jackwood, M. J., Lee, S. A. & Prosser, D. The pathogenesis of a
560 2022 North American highly pathogenic clade 2.3.4.4b H5N1 avian influenza virus in
561 mallards (*Anas platyrhynchos*). *Avian Pathol.* **In Press**, 1–28 (2023).
- 562 28. Hill, N. J. *et al.* Transmission of influenza reflects seasonality of wild birds across the
563 annual cycle. *Ecol. Lett.* **19**, 915–925 (2016).
- 564 29. Hill, N. J. *et al.* Migration strategy affects avian influenza dynamics in mallards (*Anas*
565 *platyrhynchos*). *Mol. Ecol.* **21**, 5986–5999 (2012).
- 566 30. Baldassarre, G. *Ducks, geese, and swans of North America*. (JHU Press, 2014).
- 567 31. Wille, M., Lindqvist, K., Muradrasoli, S., Olsen, B. & Järhult, J. D. Urbanization and the
568 dynamics of RNA viruses in Mallards (*Anas platyrhynchos*). *Infect. Genet. Evol.* **51**, 89–
569 97 (2017).
- 570 32. Altizer, S., Bartel, R. & Han, B. A. Animal migration and infectious disease risk. *Science*
571 **331**, 296–302 (2011).
- 572 33. Fritzsche McKay, A. & Hoyer, B. J. Are migratory animals superspreaders of infection?
573 *Integr. Comp. Biol.* **56**, 260–267 (2016).
- 574 34. Gaidet, N. *et al.* Potential spread of highly pathogenic avian influenza H5N1 by wildfowl:
575 Dispersal ranges and rates determined from large-scale satellite telemetry. *J. Appl. Ecol.*
576 **47**, 1147–1157 (2010).
- 577 35. Kim, J. K., Negovetich, N. J., Forrest, H. L. & Webster, R. G. Ducks: The ‘Trojan Horses’
578 of H5N1 influenza. *Influenza Other Respi. Viruses* **3**, 121–128 (2009).
- 579 36. Van Den Brand, J. M. A. *et al.* Wild ducks excrete highly pathogenic avian influenza virus
580 H5N8 (2014–2015) without clinical or pathological evidence of disease. *Emerg. Microbes*

- 581 *Infect.* **7**, (2018).
- 582 37. Kross, J., Kaminski, R. M., Reinecke, K. J., Penny, E. J. & Pearse, A. T. Moist-Soil Seed
583 Abundance in Managed Wetlands in the Mississippi Alluvial Valley. *J. Wildl. Manage.*
584 **72**, 707–714 (2008).
- 585 38. Hagy, H. M. & Kaminski, R. M. Winter waterbird and food dynamics in autumn-managed
586 moist-soil wetlands in the Mississippi Alluvial Valley. *Wildl. Soc. Bull.* **36**, 512–523
587 (2012).
- 588 39. Arsnoe, D. M., Ip, H. S. & Owen, J. C. Influence of body condition on influenza A virus
589 infection in mallard ducks: Experimental infection data. *PLoS One* **6**, e22633 (2011).
- 590 40. Devries, J. H., Brook, R. W., Howerter, D. W. & Anderson, M. G. Effects of spring body
591 condition and age on reproduction in Mallards (*Anas platyrhynchos*). *Auk* **125**, 618–628
592 (2008).
- 593 41. van Dijk, J. G., Verhagen, J. H., Wille, M. & Waldenström, J. Host and virus ecology as
594 determinants of influenza A virus transmission in wild birds. *Curr. Opin. Virol.* **28**, 26–36
595 (2018).
- 596 42. Sánchez, C. A. *et al.* On the relationship between body condition and parasite infection in
597 wildlife: a review and meta-analysis. *Ecol. Lett.* **21**, 1869–1884 (2018).
- 598 43. Hoyer, B. J., Fouchier, R. A. M. & Klaassen, M. Host behaviour and physiology underpin
599 individual variation in avian influenza virus infection in migratory Bewick’s swans. *Proc.*
600 *R. Soc. B* **279**, 529–534 (2012).
- 601 44. Hill, S. C. *et al.* Antibody responses to avian influenza viruses in wild birds broaden with
602 age. *Proc. R. Soc. B* **283**, 20162159 (2016).
- 603 45. Stallknecht, D. E. *et al.* Naturally Acquired Antibodies to Influenza A Virus in Fall-
604 Migrating North American Mallards. *Vet. Sci.* **9**, (2022).
- 605 46. Rohani, P., Breban, R., Stallknecht, D. E. & Drake, J. M. Environmental transmission of
606 low pathogenicity avian influenza viruses and its implications for pathogen invasion.
607 *Proc. Natl. Acad. Sci.* **106**, 10365–10369 (2009).
- 608 47. Li, X., Xu, B. & Shaman, J. The impact of environmental transmission and
609 epidemiological features on the geographical translocation of highly pathogenic avian
610 influenza virus. *Int. J. Environ. Res. Public Health* **16**, 1–14 (2019).
- 611 48. Casazza, M. L., Coates, P. S., Miller, M. R., Overton, C. T. & Yparraguirre, D. R. Hunting
612 influences the diel patterns in habitat selection by northern pintails *Anas acuta*. *Wildlife*
613 *Biol.* **18**, 1–13 (2012).
- 614 49. Nemeth, N. M. *et al.* Bald eagle mortality and nest failure due to clade 2.3.4.4 highly
615 pathogenic H5N1 influenza A virus. *Sci. Rep.* **13**, 191 (2023).
- 616 50. Puryear, W. *et al.* Outbreak of Highly Pathogenic Avian Influenza H5N1 in New England
617 Seals. 1–10 (2022).
- 618 51. Heusmann, H. W. The history and status of the mallard in the Atlantic flyway. *Wildl. Soc.*
619 *Bull.* **19**, 14–22 (1991).
- 620 52. Dahl, T. E. Wetland Losses in the United States: 1780’s to 1980’s. US Department of the
621 Interior, Fish and Wildlife Service, Washington, DC Jamestown, ND. Northern Prairie
622 Wildlife Research Center (Version 16JUL97). (1990).
- 623 53. Yin, S. *et al.* Habitat loss exacerbates pathogen spread: An Agent-based model of avian
624 influenza infection in migratory waterfowl. *PLoS Comput. Biol.* **18**, e1009577 (2022).
- 625 54. Hagy, H. M. *et al.* *Waterfowl Monitoring Plan for National Wildlife Refuges in the*
626 *Southeast*. (2021).

- 627 55. Hagy, H. M. *et al.* Midwinter Aerial Waterfowl Surveys on National Wildlife Refuges in
628 the Southeast during 2022. <https://ecos.fws.gov/ServCat/Reference/Profile/144230>.
629 (2022).
- 630 56. Kremetz, D. G., Asante, K. & Naylor, L. W. Spring migration of mallards from Arkansas
631 as determined by satellite telemetry. *J. Fish Wildl. Manag.* **2**, 156–168 (2011).
- 632 57. Williams, B. R., Benson, T. J., Yetter, A. P., Lancaster, J. D. & Hagy, H. M. Stopover
633 duration of spring migrating dabbling ducks in the Wabash river valley. *Wildl. Soc. Bull.*
634 **43**, 590–598 (2019).
- 635 58. Webster, R. G., Bean, W. J., Gorman, O. T., Chambers, T. M. & Kawaoka, Y. Evolution
636 and ecology of influenza A viruses. *Microbiol. Rev.* **56**, 152–179 (1992).
- 637 59. Pantin-Jackwood, M. J. & Swayne, D. E. Pathogenesis and pathobiology of avian
638 influenza virus infection in birds. *OIE Sci. Tech. Rev.* **28**, 113–136 (2009).
- 639 60. Sharp, D. E. & Smith, H. I. *Rocket-projected net trap use in wildlife management and*
640 *research, 1979-85.* (1986).
- 641 61. Carney, S. M. *Species, age and sex identification of ducks using wing plumage.* (US
642 Department of the Interior, US Fish and Wildlife Service, 1992).
- 643 62. Hanson, B. A., Stallknecht, D. E., Swayne, D. E., Lewis, L. A. & Senne, D. A. Avian
644 Influenza Viruses in Minnesota Ducks During 1998–2000. *Avian Dis.* **47**, 867–871
645 (2003).
- 646 63. Fair, J. M. & Jones, J. *Guidelines to the use of wild birds in research.* (Ornithological
647 council, 2010).
- 648 64. McDuie, F. *et al.* GPS tracking data reveals daily spatio-temporal movement patterns of
649 waterfowl. *Mov. Ecol.* **7**, 6 (2019).
- 650 65. Teitelbaum, C. S. *et al.* Data showing similar movement ecology between mallards
651 infected and not infected with highly pathogenic avian influenza H5N1. *U.S. Geol. Surv.*
652 *data release* (2023) doi:10.5066/P9AZL1MN.
- 653 66. Stallknecht, D. E., Shane, S. M., Zwank, P. J., Senne, D. A. & Kearney, M. T. Avian
654 influenza viruses from migratory and resident ducks of coastal Louisiana. *Avian Dis.* **34**,
655 398–405 (1990).
- 656 67. Killian, M. L. Hemagglutination Assay for the Avian Influenza Virus. in *Avian Influenza*
657 *Virus* (ed. Spackman, E.) 47–52 (2008).
- 658 68. Spackman, E. *et al.* Development of a Real-Time Reverse Transcriptase PCR Assay for
659 Type A Influenza Virus and the Avian H5 and H7 Hemagglutinin Subtypes. *J. Clin.*
660 *Microbiol.* **40**, 33–44 (2002).
- 661 69. Shriner, S. A. *et al.* Influenza A virus surveillance, infection and antibody persistence in
662 snow geese (*Anser caerulescens*). *Transbound. Emerg. Dis.* **69**, 742–752 (2021).
- 663 70. Fereidouni, S. R. *et al.* Dynamics of specific antibody responses induced in mallards after
664 infection by or immunization with low pathogenicity avian influenza viruses. *Avian Dis.*
665 **54**, 79–85 (2010).
- 666 71. Hoyer, B. J. *et al.* Reconstructing an annual cycle of interaction: Natural infection and
667 antibody dynamics to avian influenza along a migratory flyway. *Oikos* **120**, 748–755
668 (2011).
- 669 72. Spiegel, O., Leu, S. T., Bull, C. M. & Sih, A. What’s your move? Movement as a link
670 between personality and spatial dynamics in animal populations. *Ecol. Lett.* **20**, 3–18
671 (2017).
- 672 73. Bengtsson, D. *et al.* Movements, home-range size and habitat selection of mallards during

- 673 autumn migration. *PLoS One* **9**, e100764 (2014).
- 674 74. Thieurmel, B. & Elmarhraoui, A. suncalc: Compute Sun Position, Sunlight Phases, Moon
675 Position and Lunar Phase. *R Packag. version 0.5.0* (2019).
- 676 75. R Development Core Team. R: A Language and Environment for Statistical Computing.
677 (2020).
- 678 76. Signer, J., Fieberg, J. & Avgar, T. Animal movement tools (amt): R package for managing
679 tracking data and conducting habitat selection analyses. *Ecol. Evol.* **9**, 880–890 (2019).
- 680 77. Magnusson, A. *et al.* Package ‘glmmTMB’. *R Packag. Version 1.1.2* (2017).
- 681 78. Brooks, M. E. *et al.* glmmTMB Balances Speed and Flexibility Among Packages for
682 Zero-inflated Generalized Linear Mixed Modeling. *R J.* **9**, 378–400 (2017).
- 683 79. Teitelbaum, C. S. *et al.* Waterfowl recently infected with low pathogenic avian influenza
684 exhibit reduced local movement and delayed migration. *Ecosphere* **14**, e4432 (2023).
- 685 80. Zuur, A. F. *et al.* *Mixed effects models and extensions in ecology with R*. vol. 574
686 (Springer, 2009).
- 687 81. Hartig, F. DHARMA: residual diagnostics for hierarchical (multi-level/mixed) regression
688 models. *R Packag. version 0.4.3* (2019).
- 689 82. Lenth, R. V. emmeans: Estimated Marginal Means, aka Least-Squares Means. *R Packag.*
690 *version 1.6.3* (2021).
- 691 83. Kranstauber, B., Kays, R., Lapoint, S. D., Wikelski, M. & Safi, K. A dynamic Brownian
692 bridge movement model to estimate utilization distributions for heterogeneous animal
693 movement. *J. Anim. Ecol.* **81**, 738–746 (2012).
- 694 84. Kranstauber, B., Smolla, M. & Scharf, A. K. move: Visualizing and Analyzing Animal
695 Track Data. *R Packag. version 4.0.6* (2020).
- 696 85. Patin, R., Etienne, M. P., Lebarbier, E., Chamaillé-Jammes, S. & Benhamou, S.
697 Identifying stationary phases in multivariate time series for highlighting behavioural
698 modes and home range settlements. *J. Anim. Ecol.* **89**, 44–56 (2020).
- 699 86. McDuie, F. *et al.* Moving at the speed of flight: Dabbling duck-movement rates and the
700 relationship with electronic tracking interval. *Wildl. Res.* **46**, 533–543 (2019).
- 701 87. Hupp, J. W. *et al.* Variation in spring migration routes and breeding distribution of
702 northern pintails *Anas acuta* that winter in Japan. *J. Avian Biol.* **42**, 289–300 (2011).
- 703 88. Sullivan, J. D. *et al.* Waterfowl spring migratory behavior and avian influenza
704 transmission risk in the changing landscape of the East Asian-Australasian Flyway. *Front.*
705 *Ecol. Evol.* **6**, 1–14 (2018).
- 706 89. Green, A. J. Mass/Length Residuals: Measures of Body Condition or Generators of
707 Spurious Results? *Ecology* **82**, 1473–1483 (2001).
- 708 90. Teitelbaum, C. S. Code release for: ‘North American wintering mallards infected with
709 highly pathogenic avian influenza show few signs of altered local or migratory
710 movement’. (2023) doi:10.5281/zenodo.8126569.
- 711