Spatial and temporal dust source variability in northern China identified using advanced remote sensing analysis

A. Taramelli,1* M. Pasqui,2 J. Barbour,3 D. Kirschbaum,3 L. Bottai,2,4 C. Busillo,2,4 C. Calistrini,2,4 F. Guarnieri2,4 and C. Small3

1 ISPRA – Institute for Environmental Research, Rome, Italy
2 Institute of Biometeorology and National Research Council/IBIMET-CNR, Florence, Italy
3 Lamont Doherty Earth Observatory of Columbia University, New York, USA
4 Consortium LaMMa - Laboratory for Meteorology and Environmental Modelling, Sesto Fiorentino, Italy

ABSTRACT: The aim of this research is to provide a detailed characterization of spatial patterns and temporal trends in the regional and local dust source areas within the desert of the Alashan Prefecture (Inner Mongolia, China). This problem was approached through multi-scale remote sensing analysis of vegetation changes. The primary requirements for this regional analysis are high spatial and spectral resolution data, accurate spectral calibration and good temporal resolution with a suitable temporal baseline. Landsat analysis and field validation along with the low spatial resolution classifications from MODIS and AVHRR are combined to provide a reliable characterization of the different potential dust-producing sources. The representation of intra-annual and inter-annual Normalized Difference Vegetation Index (NDVI) trend to assess land cover discrimination for mapping potential dust source using MODIS and AVHRR at larger scale is enhanced by Landsat Spectral Mixing Analysis (SMA). The combined methodology is to determine the extent to which Landsat can distinguish important soils types in order to better understand how soil reflectance behaves at seasonal and inter-annual timescales. As a final result mapping soil surface properties using SMA is representative of responses of different land and soil cover previously identified by NDVI trend. The results could be used in dust emission models even if they are not reflecting aggregate formation, soil stability or particle coatings showing to be critical for accurately represent dust source over different regional and local emitting areas. Copyright © 2012 John Wiley & Sons, Ltd.

KEYWORDS: dust source; spectral mixing analysis; NDVI time series; China
Bryant et al., 2007; Reynolds et al., 2007), comparatively little is known about dust source area characteristics (Okin et al., 2011). Major issues include identification of the most productive areas within an observed dust source area; controlling factors on dust emissions in space and time; mass, size and soil composition of emitted dust; and sensitivity of dust sources to environmental (climate and land cover/use) changes (Reheis and Khl, 1995; Okin et al., 2006). In this context a number of studies have shown that there is a strong link between the seasonal variability of vegetation and the occurrence of dust storms using the Normalized Difference Vegetation Index (NDVI) (Malo and Nicholson, 1990; Nicholson and Farrar, 1994; Schmidt and Karnieli, 2000; Kawabata et al., 2001). Vegetation cover protects the underlying soil and is capable of intercepting dust and precipitation (Engelstaedter et al., 2003; Xu et al., 2003, 2006). The density of the vegetation protects the surface from deflation, while the presence of structured vegetation results in a high surface roughness that reduces surface wind energy and therefore also dust emissions (Engelstaedter, 2003). In contrast, shrublands tend to have less dense vegetation and more bare soil (Okin et al., 2011), relating the emission and transport of dust to the wind strength (Figure 4 in Pasqui et al., 2012). Reduction in perennial vegetation cover has been used to predict the onset of desertification (Xu et al., 2006; Urban et al., 2009) and to map land surface moisture in varying terrains (Vapor and Caselles, 1996; Peters et al., 1997). Decrease in value of the NDVI could therefore highlight areas where different pressures have had a significant impact on the land cover or where large-scale hydrological and atmospheric changes are occurring (e.g. decrease in snowfall). The identification of areas where NDVI values are increasing or decreasing through time could provide evidence of where dust storms may originate in the future (Townshend and Justice, 1986; Schmidt and Karnieli, 2000; Xu et al., 2006; Urban et al., 2009; Okin et al., 2001). Monitoring and tracking changes in waning-signal vegetated areas could lead to future dust sources area identification (Nicholson et al., 1990; Tucker et al., 1991; Maselli et al., 1993; Chladil and Nunez, 1995; William and Thomas, 1996; Weiss et al., 2001) and represent the land cover evolution during seasonal, annual and decadal changes over wide regions (Bach et al., 2007). Reduction in vegetation cover is not the only control on dust emissions; consequently, once vulnerable areas have been identified on the basis of vegetation loss, it is then critical to assess the geomorphology, land use and soil characteristics of the area (Okin et al., 2006; Reynolds et al., 2007; Scheidt et al., 2007; Schepanski et al., 2007; Katra and Lancaster, 2008; Bullard et al., 2011). Of the available remote data sources, Landsat imagery is best suited to this task because it provides a calibrated 30-year baseline and both the spatial and temporal resolution necessary to distinguish among important land cover and soils types and to characterize their temporal variability on seasonal to inter-annual scales (Small and Lu, 2006; Small et al., 2009; Taramelli, 2011). While high spatial resolution (1–10 m) imagery like IKONOS and Quickbird are able to distinguish individual land cover elements (e.g. buildings, roads), the narrow swath width (10–40 km) precludes synoptic regional analysis and the limited spectral resolution (four broad bands in the very near infrared, VNIR) does not provide the short-wave infrared (SWIR) reflectance necessary to discriminate many rock/soil substrates. In principle, other moderate-resolution sensors, like SPOT and Aster, could provide many of the same benefits as Landsat but the cost, non-uniform acquisition and lack of calibration often preclude their use.

As part of a joint Chinese-Italian project (WINDUST; Fratini et al., 2005, 2009; Bach et al., 2007) aimed at estimating the dust emission potential of arid and semiarid areas of northern Asia, the remote sensed imagery from several instruments and missions (MODIS, LANDSAT, AVHRR), in combination with ground data, were integrated to develop an innovative and quantitative method to understand the geomorphological context of dust source areas located in northwest China (Alashan Prefecture, Inner Mongolia; Figure 1). The aim of this paper is to show the remote sensing analysis results to characterize the source areas’ surface properties, and to highlight the potential for emission area detection at the local (hotspot) and regional (source area) scale. Here the effort was a two-step approach:

(a) studying seasonal and inter-annual dynamics at a regional scale using low-resolution remotely sensed data to highlight the response of different plant functional types that is critical to simulate dust emission over vegetated areas;

(b) studying morphological properties within the soil–vegetation change areas using the medium-resolution remotely sensed data in combination with innovative assimilation of data from ground observations to develop a quantitative method to define local hotspots and characterize their surfaces. We addressed the last issue by using (1) spectral mixture analysis to represent land surface reflectance as continuous fields of biophysical land surface properties based on the selection of the most suitable end members (spectral signature for a pure surface cover), (2) field spectroscopy validation to verify the accuracy of the mixture model and (3) decision tree classification to divide the spectral mixing space into discrete dust particle physical properties, (also considering Fratini et al., 2009), in order to develop methods for automated monitoring of dust sources with orbital data.

The overall and final objective of this part of the WINDUST project was to map the spatial extents of spectrally distinct rock and soil substrates that may represent dust sources for use in a model (RAMS – Regional Atmospheric Modeling System) with a continuous land cover grid at 30 m resolutions in order to assess and quantify the impact of environmental changes over China (Pasqui et al., 2012).

Methods and Results

Trend analysis from NDVI

In order to characterize the temporal trends in vegetation losses and potential response of different plant functional types to intra-annual and inter-annual climate variations, a time series of NDVI was analyzed as a measure of the vegetation abundance and to discriminate spatiotemporal changes in vegetation cover that could simulate dust emission (Bullard et al., 2008; Rivera Rivera et al., 2009; Okin et al., 2011). The NDVI product was derived from AVHRR images and processed at the University of Maryland in the Global Inventory Modeling and Mapping Studies (GIMMS). At the time of analysis, the GIMMS NDVI archive provided bimonthly composite images, corrected for several errors, from 1 August 1981 to 31 December 2003 at 8 km spatial resolution (Tucker et al., 2004).

To develop a regional baseline for vegetation trends, we evaluated spatial and temporal patterns in GIMMS 8 km resolution NDVI from the western Alashan Desert to the coast east of Beijing, an area between 37°–43° N and 96°–117° E. To prepare the data for analysis, each global bimonthly GIMMS image was clipped to the study region and joined to create a stack of 540 time-slice images. The dataset was evaluated on a pixel-by-pixel basis (Bach et al., 2007). To obtain a qualitative understanding of how NDVI has changed over the past two decades, we computed April averages for three years at the beginning (1982–1984), middle (1992–1994) and end
Figure 1. Location of the study area. The Alashan Desert is acknowledged as a major source of dust–sand storms that affect Beijing. The large box around the Alashan shows the dust–sand source study region and the small box around Beijing shows the dust–sand target region of interest.

Figure 2. The figures depict relative vegetation change based on the least-squares linear regression fit to the running variance (colour scale equates to the slope of the regression line) calculated from the 5-year running variance where 0.2 is the threshold between negative and neutral, 0.4 is the neutral value and the 0.6 is the threshold between neutral and positive. The top figure used a baseline of 5 years to calculate the regression, while the bottom figure used 8 years of data. These figures provide differing depictions of potential vegetation changes in the southeast portion of the region of interest.
This is a method pioneered by Elvidge et al. (1997a, 1997b) at NOAA's National Geophysical Data Center, who have used time-slice RGB composites to visualize temporal changes in night-time light emissions across the globe (www.ngdc.noaa.gov/dmsp/publications.html). Figure 2 depicts these changes in NDVI over 5- and 8-year time series. The images (Tucker et al., 2005) highlight significant increases in NDVI in the southeastern portion of the image south of Beijing, along the northern portion of the Qilian Shan mountainous region in the southeastern corner of the image, and along the southern margin of the Yellow River between approximately 107°–108° E and 40°–30° N. The images also identify decreases in NDVI from the earlier to later period of this analysis in the Beijing region and generally to the north into southern Mongolia.

We then used the NDVI dataset to identify the spatial distribution and temporal patterns of vegetation using empirical orthogonal function (EOF) analysis. Any multidimensional database can be expressed as a linear sum of elementary basis functions referred to as EOFs. These functions are orthogonal to one another, meaning that when multiplied together their product is equal to zero. The functions are also similar to a Fourier analysis, except that in Fourier analysis the basis functions are sines and cosines of different frequencies and phases. The results of the orthogonal decomposition of the dataset can be used to identify independent spatial and temporal patterns. The first EOF represents the dominant spatial pattern in the data, which contains the largest portion of the dataset variance. Correspondingly, the first principal component (PC) or dominant loading pattern describes the leading pattern in the time series. EOF analyses are computationally intensive and the full-resolution NDVI dataset could not be handled by the available software. To perform the analysis, we resample each time-slice image to 7 arcmin (approximately 10 km) spatial resolution using a nearest-neighbour assignment. The results of the EOF analysis indicate that 64% of the variance in the dataset is described by the first EOF. This image has a strongly seasonal signal that is evident in time series on the top graph of Figure 3. The bottom image in Figure 3 represents the spatial distribution of this dominant sinusoidal trend and explains the seasonality and distribution of vegetation within the test region. Thus the value of the first EOF reflects the relative abundance of vegetation at each pixel.

In order to quantify and compare the strength of the seasonal NDVI signal at each pixel through time and compare them to the multi-year mean NDVI between 1982 and 2000 that were correlated to the frequently storms event according to Xu et al., 2006, we used statistical methodology to determine the difference between low NDVI values between autumn and winter, and high values between spring and summer. The seasonal signal is stronger where vegetation is more dense and weaker where it is less abundant. Thus the strength of the seasonal signal can be represented by the variance of a subset of the data through time. Variance is described as the square of the standard deviation of NDVI over a specified window of data. Variance values can range from 0.03 to 0.08 in highly vegetated areas and are 0.0005 or lower in more arid regions. For this analysis we used variance rather than mean to identify longer temporal trends in NDVI because variance is a better indicator of how vegetation abundance changes seasonally. Thus, if a region that initially has sparse vegetation becomes cultivated due to agricultural expansion, the amplitude of the seasonal NDVI signal would get larger and correspondingly the variance would increase. The same trend can be documented for regions where urbanization causes vegetation to diminish and thus the NDVI variance decreases.

To quantitatively assess changes in vegetation cover through time we computed the variance of the time series at each pixel using a 5-year Hamming weighting function, which is a raised-cosine low-pass filter (Bach et al., 2007). A weighted filter was used in this situation to prevent errors, such as spectral leakage or ringing that results from using a simple boxcar filter, and which essentially decrease the signal variance and introduce more noise into the results. We tested several filter sizes on the data and concluded that a 5-year running variance filter most effectively suppressed seasonality. The filter was able to suppress seasonality or inter-annual signals in the data without masking decadal signals such as El Niño, which generally occur on 7-year cycles. Given the 5-year window used, the resulting variance time series is curtailed at both ends and the total number of bands in the dataset decreases to 421 or 17 years. Although the shape of the Hamming window may introduce some error to the time series, it is clear that there are trends in NDVI variance in several parts of the study area. Figure 4 illustrates examples of changing smoothed variance over time in three separate pixels with very different variance values. Variance values are larger for highly vegetated regions like the Yellow River, moderate for somewhat vegetated regions such as Beijing, and almost zero for desert regions in the Alashan. When the Yellow River and Beijing example pixels are fitted to more appropriate scales, strong increasing and decreasing trends respectively are apparent.
At this stage the smoothed NDVI variance time series appears to give a relatively useful indication of significant temporal changes in degree of seasonality within a region; thus we used the 5-year running variance to estimate possible vegetation changes in the future. We fit each pixel in the study area with a least-squares linear regression line and extrapolated the slope of the trend line to make general estimations of how NDVI variance may change (Figure 7). We used the last 5 and 8 years of the dataset for this regression in order to present two different scenarios to depict whether the variance was positively, negatively or not changing in the future. It appears that the majority of change is occurring in the eastern portion of the image where both vegetation and population are the greatest. In the 2-year data plot there are positive slope values in the southeastern portion of the image and negative slopes to the west between 111° E and 115° E. Contrastingly, the slopes of the 5-year and 10-year data analysis suggest that there is a strong negative trend south of 39° N and 117° E, while the region to the west of this area has a relatively ambiguous pattern. In general, the slopes of the 2-year data figure show more dramatic changes in slopes which most likely correspond to the shorter length of the data used for the analysis, while the slope of the 8-year data shows less dramatic changes, confirmed by the linear model and vegetation development stage (VDS) using the GIMMS NDVI anomalies (de Jong et al., 2011). Both projected NDVI maps for the highlighted six regions in Figure 5 were compared, using a 500 m × 500 m and a 0.005° × 0.005° resolution for the former and the latter, respectively. For the projected grids the total NDVI surfaces were computed considering a cell area equal to 250 000 m², and to the six latitude/longitude grids the ASPHAA algorithm was applied (Santini et al., 2010). When considering the GIMMS NDVI anomalies, the difference in percent seems to increase with the increase of the NDVI index itself (b, f and d regions in Figures 5 and 6) for frequency, following an exponential law and having correlation coefficients equal to 0.85. Given these results it is clear that the length of data used in the linear regression has a considerable impact on the resulting slopes and therefore estimations of future trends in seasonality.

Spectral mixture analysis, classification and field validation of Landsat ETM+ for the Alashan Desert

Data

The analyses are based on Landsat ETM+ imagery. Baseline imagery, ca. 2000, was acquired from the Global Land Cover Facility (GLCF) at the University of Maryland. These data were supplemented by individual scenes acquired in the spring of 2004. The imagery acquired from GLCF was originally processed by the EarthSat Corporation as inputs to the Geocover global Landsat mosaic. Technical details of the imagery are available at:
Details of the orthorectification and at-sensor radiance calculated by restoring linear gains and biases to image Digital Numbers are based on Chand et al., 2009. Exo-atmospheric reflectance is derived from at-sensor radiance by compensating for seasonal variations in Earth–Sun distance and normalizing for solar irradiance as described above.

An unexpected problem arose in the calibration sequence when it was discovered that Geocover scenes acquired prior to 7/1/2000 are not consistent with those acquired after this date. New calibration constants were released for the ETM+ scenes acquired since 7/1/2000 and these constants were used in our calibration. However, we found a serious problem with the scenes acquired before this date. Post-2000 scenes provide generally consistent exo-atmospheric reflectance values when compared to overlapping scenes but pre-2000 scenes produce consistently higher reflectance values than post-2000 scenes.

Two Geocover scenes from the Alashan region were acquired pre-2000. Comparison of sidelap in the pre- and post-2000 Alashan scenes revealed a quasi-linear band-to-band disagreement in the pre-2000 scenes. As a result, it was possible to partially correct the two pre-2000 Alashan scenes. The partial correction was accomplished by fitting a linear trend to the distribution of pre- and post-2000 pixels within the overlap regions in each band and applying a linear correction to offset the estimated trend. This improved the agreement between the pre- and post-2000 scenes but did not completely solve the problem. The magnitude of this effect is apparent in Figure 9.

Fortunately, both affected Alashan scenes were on the southern periphery of the study area and not near the main regions of interest. When notified of the problem with the pre-2000 scenes, a representative at the GLCF suggested that the problem with the pre-2000 scenes may be a result of Level 1 Product Generation Systems (LPGS) versus National Landsat Archive Processing Systems (NLAPS) versus Geocover processing methods.

Figure 6. Time series of 5-year running variance in six different locations identified in Figure 5. Each of the coloured lines is a different pixel within the region and the solid line in each graph represents the average for the pixels in each plot. The left column shows regions where variance appears to be decreasing and the right column indicates areas where variance may be increasing. Best-fit linear regressions were fit to the average in plots (a) and (b) to highlight possible 20-year trend within the data (Bach et al., 2007). This figure is available in colour online at wileyonlinelibrary.com/journal/espl
mosaicing of the 16 ETM+ scenes was by far the most imagery was used for the regional mosaic. Calibration and individual scenes in 'prior to mosaicing. Full-resolution data were used for analysis of was necessary to resample the Landsat scenes to 191 m resolution paths from west - building up the mosaic one path at a time. It was necessary to resample the Landsat scenes to 191 m resolution prior to mosaicing. Full-resolution data were used for analysis of individual scenes in 'hotspot' regions but reduced-resolution imagery was used for the regional mosaic. Calibration and mosaicing of the 16 ETM+ scenes was by far the most another problem arose with the mosaicing of the Alashan scenes. The geographically referenced mosaicing algorithm in ENVI produced numerous projection-related errors when all 16 scenes were mosaiced simultaneously. We believe this is a result of the mosaic spanning more than one UTM zone. The problem was resolved by sequentially mosaicing scenes within individual paths (130–134) and then sequentially mosaicing the paths from west - building up the mosaic one path at a time. It was necessary to resample the Landsat scenes to 191 m resolution prior to mosaicing. Full-resolution data were used for analysis of individual scenes in 'hotspot' regions but reduced-resolution imagery was used for the regional mosaic. Calibration and mosaicing of the 16 ETM+ scenes was by far the most time-consuming phase of the analysis. However, the result produced outstanding sidelap and endlap agreement among adjacent scenes in the Alashan mosaic. While some scene transitions are visible in the mosaic (Figure 9), it is remarkable that the hand-to-hand brightnesses agree so closely. As a result, no histogram matching was necessary. Aside from the two pre-2000 scenes, only one scene (p131r33) contrasts strongly with adjacent scenes. This appears to be a result of a large increase in soil moisture in this scene, as indicated by its lower albedo and the presence of large bodies of standing water within the dune field that dominates the scene. Spectral mixture analysis (SMA) SMA is a methodology whereby an observed radiance is modelled as a linear mixture of spectrally pure endmember radiances. Linear mixture models are based on the observation that, in many situations, radiances reflected from surfaces with different 'endmember' reflectances mix linearly in proportion to the area of each endmember within the instantaneous field of view (IFOV) (Singer and McCord, 1979; Singer, 1981; Johnson et al., 1983). This observation has made possible the development of a systematic methodology for spectral mixture analysis (Adams et al., 1986, 1993; Gillespie et al., 1990; Smith et al., 1990; Sabol et al., 1992) in which land surface reflectance variations are described by a set of endmember fraction images representing spatial variations in the areal abundance of each endmember. Although the physical process represented by the mixture model corresponds to the measurement of a mixed radiance within the sensor IFOV, the model can also be applied to exo-atmospheric reflectances because the conversion equations are linear. If a limited number of spectrally distinct endmembers can be found it is possible to define a mixing space within which mixed pixel spectra can be described as linear mixtures of the endmember spectra. A mixing space is analogous to a spectral feature space but is generally represented with low-dimensional projections of the principal components (PCs) of the image rather than the observed radiance bands. Representing a multispectral feature space with low-order PCs allows the topology of the space to be rendered as a 3D construct. With Landsat imagery, the three primary PCs generally contain more than 95% of the variance in the image (Small, 2004; Taramelli and Melelli, 2009). Given sufficient spectral resolution, a system of linear mixing equations and endmembers can be defined and the best-fitting combination of endmember fractions can be estimated for each of the observed reflectance spectra. The solution to the linear mixing problem can be cast as a linear inverse problem in which the system of mixing equations is inverted to yield estimates of the endmember fractions that best fit the observed mixed reflectances (Boardman, 1993; Settle and Drake, 1993; Boardman and Kruse, 1994). It is important to note that even when the surface within the ground instantaneous field of view (GIFOV) is not a mixture of the unique, spectrally pure endmember materials, it can be represented as such a mixture if it lies within the bounds of the mixing space. Because the methodology provides a general physical representation of mixed reflectances, it has proven successful for a wide variety of quantitative applications with multispectral imagery (e.g. Adams et al., 1986, 1993; Pech et al., 1986; Smith et al., 1990; Roberts et al., 1998; Elmore et al., 2000; White and Bullard, 2009). The Alashan mosaic can be represented as a tetrahedral mixing space bounded by four distinct spectral endmembers (Figure 10). The endmembers correspond to green vegetation, non-reflective dark surface and two rock/soil substrates. The non-reflective dark surfaces represent both transmissive (e.g. clear-water), absorptive (e.g. Fe-rich rocks) and non-reflective (e.g. deep-shadow) targets. The rock/soil substrates represent

Figure 7. Changes in variance extrapolated 5 and 10 years into the future. The values were calculated by fitting a linear trend through the last 5 years of the data and extending the best-fit line into the future.

Figure 8. Percentage differences between areas computed using calculated NDVI and GIMMS NDVI for the different deciles.
SWIR bright sands and a higher-albedo, more spectrally flat reflectance corresponding to both evaporates and mud/silt lithologies. Following the procedure described in detail by Small (2001, 2004), we generated a mixing space for each individual scene in the Alashan mosaic and a composite mixing space for the entire mosaic. As would be expected, the endmembers derived from the composite mixing space bound those found for each individual scene. However, the selection of three or four endmembers per individual scene resulted in 20 endmember suites (16 scenes + double redundancy for four hotspot scenes) that were generally consistent with the endmembers derived from the regional composite – as well as generic endmembers found in the global analysis described by Small (2004). In addition to the four bounding endmembers for the Alashan composite (red), the generic endmembers (black) and bounding binary mixtures (grey) from the global analysis of Small (2004) are shown. RMS misfits between estimated fractions and observed spectra were generally low (<0.05 reflectance units), with no consistent misfit of any major land surface type.

Endmember fraction composites illustrate primary land cover distributions and spatial variations in the areal abundance of the primary biophysical properties of the land surface. Fraction composites are generated by assigning three endmembers to the red, green and blue layers of an image. For Alashan, using one substrate (red), vegetation (green) and dark surface (blue) highlights the agricultural areas and the contrast between high- and low-albedo rock/soil substrates (Figure 11a). Using both mud/silt (red) and sand (green) as well as dark (blue) highlights lithological differences between the low-albedo Fe-rich crystalline rocks, high-albedo mud/silt deposits and evaporates and intermediate albedo sands (Figure 11b).

Field validation
Field observations are necessary to calibrate and validate the spectral mixture model and subsequent thematic soil classification. The purpose of the field observations is twofold. First, the field observations allow us to identify the biophysical land cover types and properties associated with the spectral endmembers. Secondly, field reconnaissance allows us to identify significant intermediary land cover types of particular significance.
In our field campaign we propose to focus the analysis of reflectance variations mainly on vegetation patterns and soil contents. The surfaces in many hotspots are characterized by sparse or no vegetation cover, with perennial shrubs/bushes and annual and ephemeral plants (e.g. grass, herbs) typical of arid areas. Vegetation is an important controlling factor on dust emission by stabilizing the soil through biophysical feedback (Schlesinger et al., 1990; Lavee et al., 1998) and increasing threshold velocity for wind erosion (Wolfe and Nickling, 1993; Marticorena et al., 2006). The field campaign was designed to provide basic reconnaissance of the region and to sample as many specific soil covers as feasible given the time constraints. Soil samples were collected for each spectral endmember as well as for the most commonly observed substrates found in areas of mixed land cover (Figure 9). The results of the field validation campaign are field photographs, laboratory sample photographs and laboratory reflectance spectra. For each waypoint (CH001 to CH094) we collected a GPS location and field photographs of the land cover. At sample points we also collected soil samples and a 1 m scale sample site photograph. Each sample was separated by size fraction and four reflectance spectra were measured with a 2002 ASD FieldSpec Pro. Samples were illuminated with a LowellPro lamp, Ushio halogen bulb (JC14.5 V-50WC) and a Lambda DC power supply. Each lab sample was measured four times from opposing azimuths of illumination. Each 5 MP field site and sample images were shot with a 2004 Pentax Optio 555 exposed for neutral balanced histogram (all macrophotographs and reflectance spectra are available on line at: http://www.ldeo.columbia.edu/~small/Alashan2005/000_SampleSpectra).

Based on the lab sample spectra the presence of three distinct rock/soil substrates and a continuum of Gobi surfaces in the Alashan region were highlighted. Based on that, the unmixing classification has dark endmembers that correspond to Fe-rich crystalline rocks and the other two endmembers, which correspond to sand. Superimposing the reflectance spectra of all the soil samples collected, it is apparent that these soils form a mixing continuum (Figure 11a, b) and that most are intermediate mixtures. To quantify the degree of similarity of the spectra we computed a correlation matrix for all samples (Figure 12a). As shown by the distribution of correlation coefficients, the majority of sample spectra have correlations in excess of 0.8 (Figure 12a).

Working on a mixture model for the lab spectra of the Alashan samples it was possible to discuss diagnostic features. Results are shown in Figure 12b, where a principal component analysis was applied on the lab spectra to generate a sparse mixing space.

Figure 10. Spectral mixing space for Alashan ETM+ mosaic. Colour density shaded (warmer = more pixels) scatter-plots show orthogonal projections of the 3D mixing space spanned by the three low-order principal components (PCs) containing more than 90% of image variance. Dimension 1 corresponds to rock/soil substrate fraction, while dimension 2 corresponds to vegetation fraction and dimension 3 accommodates the divergence of substrate 1 (mud/siltstone) and substrate 2 (sand). Snow and ice in the Qilian Shan mountains form a separate mixing continuum not contained within the primary tetrahedral mixing space bounded by the other four endmembers. Binary mixtures from the edges of the mixing space (light grey) are bounded by the composite mixing space endmembers (dark). This figure is available in colour online at wileyonlinelibrary.com/journal/espl
Figure 11. Fraction composites for Alashan are generated by assigning three endmembers to the red, green and blue layers of an image. (a) Using RGB for S1, V and D, respectively, highlights the agricultural areas and contrast between high- and low-albedo rock/soil substrates. (b) Using RGB for S2, S1 and D, respectively, highlights lithological differences between the low-albedo Fe-rich crystalline rocks, high-albedo mud/silt deposits and evaporites and intermediate albedo sands. The decision tree (c) and spatial distribution of endmember fractions are also shown. D: decision tree classification is shown.

Figure 12. (a) Alashan sample spectra and correlation coefficients. Plot of all soil spectra (top left) shows considerable redundancy within an apparent continuum between the low-albedo moist salt/mud and the high-albedo mud/silt. Nearly flat, very-low-albedo (<0.1) spectra correspond to coal. Correlation matrix (upper right) shows generally high correlations for most of the spectra. Distribution of correlation (lower left) shows at least four modes each corresponding to a spectrum endmember. Cumulative distribution of correlations indicates that 45% of the spectra have correlation > 0.9 and are likely linear mixtures of sand and silt. (b) Principal component analysis on the lab spectra to generate a sparse mixing space.
The central black images show the two primary dimensions of the mixing space (along with some of the bounding spectra and photos of the samples and locations). The primary dimension corresponds to overall albedo, while the second coordinate to spectral slope and depth of the SWIR absorptions. The rotations seem to be strongly influenced by the high-amplitude metamorphic spectra (Small et al., 2009). The main result of that is that the most common surface property shown by the unmixing classification is a mixed Gobi surface consisting of poorly sorted silt, sand, and pebbles overlain by gravel lag deposits. While the reflectance characteristics of the mixed substrate are similar throughout the Alashan, the lithology of the overlying gravel varies with distance from and composition of the crystalline source rocks within the lacustrine system (confirmed by Fratini et al., 2009, measurements). In fact, contrary to expectation, the uniformly high-albedo endmember picked within the Landsat did not correspond to evaporites in many cases but rather to semi-lithified deposits of a fine-grained mud/siltstone (e.g. CH052). On the basis of the spectra (CH054), laboratory analysis the outcrop deposit corresponds to a shallow lacustrine depositional environment as seen in the field (Figure 13). On the contrary, the evaporite deposits we did find in the field were associated with active salt production (CH029) and were located in areas of elevated water table, so they often appear in the Landsat unmixing classification as mixed low albedo, rather than high albedo as expected and confirmed by literature (Wang et al., 2004b, 2007, and references within). This appears to be the result of a large increase in soil moisture evaluable by the p131r33 Landsat scene, as indicated by its lower albedo and the presence of large bodies of standing water within the dune field that dominate the scene itself. This area is just north of the Yellow River arc-shape region identified in the NDVI variance analysis (39–40° N latitude) and is similar within the unmixing classification to the other area that is identified as a long and narrow dust-sand emitting tongue at approximately at 100° E longitude and 40–42° N latitude (Figure 14), where relict palaeochannels and lacustrine depositional environments were recognized in the field (Figure 15). There are other strandline features where lacustrine silt outcrops over the gravel Gobi surfaces, which are not always visible by SMA analysis, but are identified with the analysis of SAR backscatter data that confirmed the landform interpretation of Gilbert-type deltas within the area (Taramelli, 2011). These landforms are complex, large and well established, implying that alluvial areas and lacustrine depositional environments were stable for extended periods, and that hydrological and climatic conditions were very different in the past (see Fratini et al., 2009, and reference within). Thus, in the lab spectra analysis and in the subsequent unmixed classification, soil differences between the low-albedo Fe-rich crystalline rocks and the near-horizontal fine- to coarse-grained size are finally evident. Such results lead to the main conclusion that the Alashan soil is characterized by continuously varying land surface properties dominated by rock and soil substrate gradients that need to be classified as Gilbert delta-related soil classes.

Decision tree classification
In addition to calibrating the Landsat spectral mixture model, laboratory analyses complemented the collected in situ spectra (see previous section) in order to extend and quantify land surface properties for soil classification. This was done because discrete soil classes were required for the next phase of the RAMS implementation (Pasqui et al., 2012). We decided to use the decision tree classification to this end. Decision trees provide a more rational approach to soil cover classification than traditional statistical supervised classification. Decision trees allow the user to specify the exact logical basis of class assignment in the form of a Boolean conditional of arbitrary complexity. In other words, decision trees provide a physical basis for which pixels end up in which classes. In the case of Alashan, the mixing spaces show no evidence for feature space clustering but rather show a single heterogeneous cluster representing mixtures of the spectrally distinct endmembers at the apices of the mixing space. The obvious basis for the final soil classification is then endmember abundance. The decision tree can be used to divide the endmember fraction space on the basis of specific abundances of different endmembers (Figure 11c). Starting from the endmembers and working inward to progressively more mixed land covers, the endmember fraction space is successively divided into discrete classes corresponding to specific abundances of each endmember. While this does impose artificial boundaries on a generally continuous land surface, at least the decision boundaries are based on specific physical quantities (areal abundance of endmembers) rather than the maximum of necessarily low-probability likelihood functions. Once the Landsat imagery has been rendered to a common set of calibrated endmembers, each image was superimposed on the FAO soil classification. Although the FAO soil classification does not accommodate most of the Gobi surfaces found in the Alashan, the general substrate types (mud/silt, sand, Fe-rich crystalline) can be assigned the necessary parameters for the model because they represent distinct biophysical land surface types. Superimposing the different information layers within the decision tree classification identified potential dust-emitting pixels. Textural data were then derived using the FAO global dataset (http://www.fao.org/gtos/tems/index.jsp) with a horizontal resolution of 8 km (Tegen and Fung, 1994) and eddy covariance measurements (Fratini et al., 2009) within the high-litendmember pixels (Figure 16). By projecting the bare soil map on to the simulation domain along with the FAO textural map and the land cover derived from MODIS (Bach et al., 2007), the soil characterization table for emitting pixels was obtained (Kjelgaard et al., 2004; Fratini et al., 2009). For each soil type the representative particle radius (grain size classification according to Shepard’s, 1954, scheme), particle density and dust productivity were also obtained from the FAO dataset and from Nickovic et al. (2001) (Table II).

The results of spectral mixture analysis indicate that the Alashan region is characterized by continuously varying substrate properties. In addition to the relatively homogeneous sand and mud/silt substrates sampled, there is a wide variety of crystalline rocks cropping out throughout the Alashan region and a widespread Gobi surface of strongly mixed composition. While land cover classes can be specified on the basis of both constant and linear thresholds through the fraction space, the resulting boundaries do not represent the discrete transitions
Where natural boundaries are relatively abrupt (e.g. basement/alluvium contact) the classification will be accurate but where they are gradational the classification will introduce a false boundary. This is especially pronounced on the Gobi surfaces. It is not known what effect these boundaries will have on the dust transport/climate model. However, we feel that the endmember fraction maps do capture many of the physical properties of the land surface and could be used to provide physical properties and boundary conditions directly to the RAMS model – without loss of information and introduction of error inherent in the classification (Pasqui et al., 2012). Potential dust emission sources, although in different locations and of diverse extent, appear to have similar surface composition. In particular, where the SMA results show that dark, Fe-rich crystalline rocks and sand correspond to two of the endmembers it gives robustness to the dust particle physical properties, as seen from Tegen and Fung (1994) and Nickovic et al. (2001). At these highlighted potential emission areas, contrary to expectation, the soils did not correspond to evaporites but to semi-lithified deposits of a fine-grained mud/siltstone. The lacustrine depositional environment features for the Alashan area implied by the classification.
represent a potential high-emission area, especially where the alluvium meets the crystalline basement.

**Discussion**

For Alashan, the outstanding questions involved the composition of several mixed reflectance signatures and the presence and abundance of sparse vegetation. A more complete representation of the variety of different substrate types was necessary to isolate potential dust source regions. In order to monitor progressive degradation of the land cover (by grazing and salt formation) a specific effort to quantify and map spatiotemporal changes in sparse vegetation abundance was implemented. This study found that regional vegetation change can be effectively evaluated through the use of coarse-resolution AVHRR NDVI, producing variance measurements of NDVI from 1981 to 2003. Regions that we observe as having a waning vegetation signal could either be due to natural large-scale climatological fluctuations from changes in precipitation or from agricultural impact (see the related phonological–meteorological study in Lee et al., 2002). While at present these regions may not serve as significant sources for dust emission, they could be problematic in the future (Zeng et al., 2005; Baddock et al., 2009). These regions could generate more immediate local affects such as environmental degradation and air-quality deterioration as already seen in the literature (Yu et al., 1999). In addition, the proposed multi-temporal analysis of NDVI changes has highlighted land

![Figure 16](Dust endmembers textural data representation with field sample locations (represented in red). False colour composites (R/G/B = VisIR − high albedo/SWIR − bright/low albedo) are dominated by strong contrast within the selected endmember and field samples. Differences in non-reflective (blue) endmember fraction likely result from differences in endmember calibration related to different soil moisture between August 2002 and April 2004 images. Road tracks shape file shows the accessible sample site during fieldwork.)

<table>
<thead>
<tr>
<th>Type</th>
<th>Endmembers S1/S2/D classification</th>
<th>Particle radius (μm)</th>
<th>Particle density (g cm⁻³)</th>
<th>Dust productivity</th>
</tr>
</thead>
<tbody>
<tr>
<td>Clay</td>
<td>Clay after field validation</td>
<td>0.73</td>
<td>2.50</td>
<td>0.08</td>
</tr>
<tr>
<td>Silt, small</td>
<td>High-albedo mud/silt</td>
<td>6.10</td>
<td>2.65</td>
<td>1.00</td>
</tr>
<tr>
<td>Silt, large</td>
<td>Intermediate-albedo Sands</td>
<td>18.00</td>
<td>2.65</td>
<td>1.00</td>
</tr>
<tr>
<td>Sand</td>
<td>Sand</td>
<td>38.00</td>
<td>2.65</td>
<td>0.12</td>
</tr>
</tbody>
</table>
cover degradation validated by the SMA analysis that could result in new potential source regions. Diagnostic spectral features in the Landsat spectral range made it possible to use SMA to retrieve information related to surface materials. Pixels that are potential dust sources were highlighted by SMA in the absence of field data usually utilized in modelling dust studies (saltating particles, streamwise saltation flux, minimally and fully dispersed particle size distributions – Gillespie, 1992; Crowley and Hook, 1996; Ramsey et al., 1999; Rowan and Mars, 2003; Katra and Lancaster, 2008). Based on the fieldwork done in order to validate the Landsat classification, a more complete representation of the variety of different substrate types is used to isolate potential dust source regions within the lacustrine areas. This classification shows that an increasing sparse vegetation trend is the confirming evidence of specific dust source areas with literature data (Wang et al., 2004a). Loss of vegetation has been identified as a primary focus of their degradation monitoring activities, especially related to the NDVI trend. In addition to its effect on dust generation, vegetation degradation also affects soil erosion and the livelihods of the pastoralists living in the Alashan region. Fortunately, spectral mixture analysis is well suited to monitoring vegetation health and abundance (Small and Lu, 2006), particularly in arid and semiarid environments (Elmore et al., 2000). Mapping at-sensor radiances into endmember fraction abundances converts a physical measure of reflected solar radiance to a physically meaningful estimate of a basic biophysical property of the land surface. As such, endmember fraction maps could provide an ideal input to climate models, vastly superior to the thematic classifications normally used. The advantages of endmember fraction abundances are: (1) physically meaningful quantities (e.g. vegetation fraction, albedo); (2) greater accuracy; (3) no loss of spatial variability information; (4) no introduction of error from class homogenization; and (5) no errors of commission or omission (Small, 2006). The parametrization of the dust source model input parameters has been established through the RAMS procedure (Pasqui et al., 2012). From the RAMS model, the area producing the highest amount of dust is located in the southern part of Alashan Prefecture, between 103° and 106° E longitude. Moreover, a long and narrow dust-sand-emitting tongue at approximately 100° E longitude and 40–42° N latitude was confirmed by the RAMS model using remote sensing inputs (Figure 6 in Pasqui et al., 2012).

The work done so far provides a good foundation for an intensive spatiotemporal monitoring programme for the Alashan region that could be used within the emission–transport modelling system. The combined remote sensing and modelling approach could be used for long-term simulations and it should be considered as a potential monitoring tool in an integrated management to dust storm source area identification and control. The multi-scale remote sensing approach, in fact, represents a practical tool for building possible scenarios both for small- and large-scale areas. This could also provide quantitative evaluation of the effects produced by the different human activities adopted on the potential emission areas, leading to a more systematically integrated monitoring system.

Conclusion

Advanced remote sensing was used, in association with filed sampling, to identify potential dust sources on the basis of vegetation cover and surface sediment characteristics. The methodology is useful for developing an extensive multi-source and multi-temporal analysis and, as a result, helping to identify the baseline environmental conditions necessary to detect changes to potential dust sources at regional and local scales in northern China. A key aspect of the study is characterizing the spatial distribution of surface properties at high resolution in dust emission hotspot regions and identifying the response of these surfaces to changes in climate and land cover/use, since the exposed surface sediments determine the dust composition. Results provide a baseline on which to discriminate land cover for mapping potential dust source regions and to generate inputs to the regional model (RAMS).

The strategy of combining low- and moderate-resolution remote sensing with fieldwork validation highlights different hotspot areas where a combination of low topography, depositional processes, loose/salty soil and sparse/free vegetation act as the main driving factors of dust emission. The approach accounted for previous knowledge of relationships between landforms and dust emissions in northwestern China (Ding et al., 2005). Dust sources in this study were analysed jointly with the areas surrounding them (NDVI analysis), allowing for a better understanding of surface processes. While the major sources of dust occur in extremely arid regions, such as dune fields in the Alashan, which have little to no vegetation throughout the year, areas with a decreasing NDVI signal may also become potential dust sources in the future as highlighted by the SMA. The remote modelling approach shown in this study could be used for long-term dust source identification and it should be considered as a potential monitoring tool in an integrated management approach to dust source control.

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